

Insights on Atmospheric Oxidation Processes by Performing Factor Analyses on Sub-ranges of Mass Spectra

Yanjun Zhang¹, Otso Peräkylä¹, Chao Yan¹, Liine Heikkinen¹, Mikko Äijälä¹, Kaspar R. Daellenbach¹, Qiaozhi Zha¹, Matthieu Riva^{1,2}, Olga Garmash¹, Heikki Junninen^{1,3}, Pentti Paatero¹, Douglas Worsnop^{1,4}, and Mikael Ehn¹

¹ Institute for Atmospheric and Earth System Research / Physics, Faculty of Science, University of Helsinki, Helsinki, 00014, Finland

² Univ Lyon, Université Claude Bernard Lyon 1, CNRS, IRCELYON, F-69626, Villeurbanne, France

³ Institute of Physics, University of Tartu, Tartu, 50090, Estonia

⁴ Aerodyne Research, Inc., Billerica, MA 01821, USA

Corresponding author: yanjun.zhang@helsinki.fi

Abstract

Our understanding of atmospheric oxidation chemistry has improved significantly in recent years, greatly facilitated by developments in mass spectrometry. The generated mass spectra typically contain vast amounts of information on atmospheric sources and processes, but the identification and quantification of these is hampered by the wealth of data to analyze. The implementation of factor analysis techniques have greatly facilitated this analysis, yet many atmospheric processes still remain poorly understood. Here, we present new insights on highly oxygenated products from monoterpene oxidation, measured by chemical ionization mass spectrometry, at a boreal forest site in Finland in fall 2016. Our primary focus was on the formation of accretion products, i.e. “dimers”. We identified the formation of daytime dimers, with a diurnal peak at noon time, despite high nitric oxide (NO) concentrations typically expected to inhibit dimer formation. These dimers may play an important role in new particle formation events that are often observed in the forest. In addition, dimers identified as combined products of NO₃ and O₃ oxidation of monoterpenes were also found to be a large source of low-volatile vapors at night. This highlights the complexity of atmospheric oxidation chemistry, and the need for future laboratory studies on multi-oxidant systems. Neither of these two processes could have been separated without the new analysis approach deployed in our study, where we applied binned positive matrix factorization (binPMF) on sub-ranges of the mass spectra, rather than the traditional approach where the entire mass spectrum is included for PMF analysis. In addition to the main findings listed above, several other benefits compared to traditional methods were found.

34 **1 Introduction**

35 Huge amounts of volatile organic compounds (VOC) are emitted to the atmosphere every year
36 (Guenther et al., 1995;Lamarque et al., 2010), which play a significant role in atmospheric chemistry
37 and affect the oxidative ability of the atmosphere. The oxidation products of VOC can contribute to
38 the formation and growth of secondary organic aerosols (Kulmala et al., 2013;Ehn et al., 2014;Kirkby
39 et al., 2016;Troestl et al., 2016), affecting air quality, human health, and climate radiative forcing
40 (Pope III et al., 2009;Stocker et al., 2013;Zhang et al., 2016;Shiraiwa et al., 2017). Thanks to the
41 advancement in mass spectrometric applications, like the aerosol mass spectrometer (AMS)
42 (Canagaratna et al., 2007) and chemical ionization mass spectrometry (CIMS) (Bertram et al.,
43 2011;Jokinen et al., 2012;Lee et al., 2014), our capability to detect these oxidized products, as well
44 as our understanding of the complicated atmospheric oxidation pathways in which they take part,
45 have been greatly enhanced.

46 Monoterpenes ($C_{10}H_{16}$), one major group of VOC emitted in forested areas, have been shown to be a
47 large source of atmospheric secondary organic aerosol (SOA). The oxidation of monoterpenes
48 produces an abundance of different oxidation products (Oxygenated VOC, OVOC), including highly
49 oxygenated organic molecules (HOM) with molar yields in the range of a few percent, depending on
50 the specific monoterpene and oxidant (Ehn et al., 2014;Bianchi et al., 2019). Recent chamber studies
51 have greatly advanced our knowledge of formation pathways for monoterpene HOM products, e.g.
52 “monomers” (typically $C_{9-10}H_{12-16}O_{6-12}$) and “dimers” (typically $C_{19-20}H_{28-32}O_{8-18}$). Dimers, as shown
53 by previous studies, can contribute to new particle formation (NPF) (Kirkby et al., 2016;Troestl et al.,
54 2016;Lehtipalo et al., 2018), and are thus of particular interest.

55 In nearly all atmospheric oxidation chemistry, peroxy radicals (RO_2) are the key intermediates
56 (Orlando and Tyndall, 2012). They form when VOC react with oxidants like ozone, or the hydroxyl
57 (OH) or nitrate (NO_3) radicals, while their termination occurs mainly by bimolecular reactions with
58 nitric oxide (NO), hydroperoxyl (HO_2) and/or other RO_2 . $RO_2+R'O_2$ reactions can form $ROOR'$
59 dimers (Berndt et al., 2018a;Berndt et al., 2018b), and this pathway competes with RO_2+NO reactions,
60 meaning that NO, formed by photolysis of NO_2 , can efficiently suppress dimer formation, as also
61 seen from atmospheric HOM observations (Ehn et al., 2014;Yan et al., 2016). Mohr et al. (2017) also
62 reported daytime dimers in the boreal forest in Finland, coinciding with NPF events. A better
63 understanding of the formation of these daytime dimers would assist elucidating NPF and particle
64 growth mechanisms.

65 At night, nitrogen oxides can also impact the oxidation pathways, when NO_2 and O_3 react to form
66 NO_3 radicals that can oxidize monoterpenes. NO_3 radicals are greatly reduced during daytime due to
67 photolysis and reactions with NO reducing their lifetime to a few seconds (Ng et al., 2017). Yan et

68 al. (2016) reported nighttime HOM initiated by NO₃ in the boreal forest in Finland, but to our
69 knowledge there have been no laboratory studies on HOM formation from NO₃ oxidation of
70 monoterpenes. However, there have been several studies looking into the SOA formation in these
71 systems, finding that certain monoterpenes, like β-pinene, have very high SOA yields, while the most
72 abundant monoterpene, α-pinene, has negligible SOA forming potential. It remains an open question
73 what the role of NO₃ radical oxidation of monoterpenes, and the observed NO₃-derived HOM, in the
74 night-time boreal forest is. Identification of these processes in the ambient environment is
75 fundamental towards better understanding of NPF and SOA.

76 The recent development of CIMS techniques has allowed researchers to observe unprecedented
77 numbers of OVOC, in real-time (Riva et al., 2019). This ability to measure thousands of compounds
78 is a great benefit, but also a large challenge for the data analyst. For this reason, factor analytical
79 techniques have often been applied to reduce the complexity of the data (Huang et al., 1999), e.g.
80 positive matrix factorization, PMF (Paatero and Tapper, 1994; Zhang et al., 2011). The factors have
81 then been attributed to sources (e.g. biomass burning organic aerosol) or processes (e.g. monoterpene
82 ozonolysis) depending on the application and ability to identify spectral signatures (Yan et al.,
83 2016; Zhang et al., 2017).

84 In the vast majority of these PMF applications to mass spectra, the mass range of ions has been
85 maximized in order to provide as much input as possible for the algorithm. This approach was
86 certainly motivated in early application of PMF on e.g. offline filters, with chemical information of
87 metals, water-soluble ions, and organic and elemental carbon (OC and EC), where the number of
88 variables is counted in tens, and the number of samples in tens or hundreds (Zhang et al., 2017).
89 However, with gas-phase CIMS, we often have up to a thousand variables, with hundreds or even
90 thousands of samples, meaning that the amount of data itself is unlikely to be a limitation for PMF
91 calculation. In this work, we aimed to explore potential benefits of dividing the spectra into sub-
92 ranges before applying factorization analysis. This approach was motivated by several issues, which
93 we expected to be resolvable by analyzing several mass range separately. Firstly, the loss rate of
94 OVOC by condensation is strongly coupled to the molecular mass (Peräkylä et al., 2020), likely
95 giving very different behaviors for the high and low mass ranges, even when produced by the same
96 source. Second, dimers are a product of two RO₂, which can have different sources, meaning that
97 they may have temporal profiles unlike anything observable for monomers. Finally, if one mass range
98 contains much less signal than another, it will have very little impact on the final PMF results.

99 In this study, we applied PMF analysis on three different mass ranges of mass spectra of OVOC
100 measured by a chemical ionization atmospheric pressure interface time-of-flight (CI-APi-TOF,
101 Jokinen et al. (2012)) mass spectrometer in the Finnish boreal forest. We utilized our recently

102 proposed new PMF approach, binPMF, to include as much of the high-resolution information in the
103 mass spectra as possible, in a robust way (Zhang et al., 2019). We show the benefits of the sub-range
104 PMF approach to better separate chemical sources, by reducing disturbance from variable loss terms
105 of the OVOC. Much of the analysis focuses on dimer formation pathways, and the role of different
106 nitrogen oxides in these pathways. We find that both daytime dimers and dimers resulting from the
107 combination of different oxidants can be separated with the sub-range approach, but not with the PMF
108 applied to the full mass range. We believe that this study will provide new perspectives for future
109 studies analyzing gas-phase CIMS data.

110

111 **2 Methodology**

112 The focus of this work is on retrieving new information from mass spectra by applying new analytical
113 approaches. Therefore, we chose a dataset that has been presented earlier, though without PMF
114 analysis, by Zha et al. (2018), and was also used in the first study describing the binPMF method
115 (Zhang et al., 2019). The measurements are described in more details below in section 2.1, while the
116 data analysis techniques used in this work are presented in section 2.2.

117

118 **2.1 Measurements**

119 **2.1.1 Ambient site**

120 The ambient measurements were conducted at the Station for Measuring Ecosystem–Atmosphere
121 Relations (SMEAR) II in Finland (Hari and Kulmala, 2005) as part of the Influence of Biosphere–
122 Atmosphere Interactions on the Reactive Nitrogen budget (IBAIRN) campaign (Zha et al, 2018).
123 Located in the boreal environment in Hyytiälä, SMEAR II is surrounded with coniferous forest and
124 has limited anthropogenic emission sources nearby. Diverse measurements of meteorology, aerosol
125 and gas phase properties are continuously conducted at the station. Details about the meteorological
126 conditions and temporal variations of trace gases during IBAIRN campaign are presented by Zha et
127 al. (2018) and Liebmann et al. (2018).

128

129 **2.1.2 Instrument and data**

130 Data were collected with a nitrate (NO_3^-)-based chemical ionization atmospheric pressure interface
131 time-of-flight mass spectrometer (CI-APi-TOF, Jokinen et al., 2012) with about 4000 Th Th^{-1} mass
132 resolving power, at ground level in September, 2016. In our study, the mass spectra were averaged to
133 1 h time resolution from September 6th to 22nd for further analysis. We use the thomson (Th) as the
134 unit for mass/charge, with 1 Th = 1 Da/e, where e is the elementary charge. As all the data discussed
135 in this work are based on negative ion mass spectrometry, we will use the absolute value of the

136 mass/charge, although the charge of each ion will be negative. The masses discussed in this work
137 includes the contribution from the nitrate ion, 62, unless specifically mentioned. Furthermore, as the
138 technique is based on soft ionization with NO_3^- ions, any multiple charging effects are unlikely, and
139 therefore the reported mass/charge values in thomson can be considered equivalent to the mass of the
140 ion in Da.

141 The forest site of Hyytiälä is dominated by monoterpene emissions (Hakola et al., 2006). The main
142 feature of previous CI-API-TOF measurements in Hyytiälä (Ehn et al., 2014; Yan et al., 2016) has
143 been a bimodal distributions of HOM, termed monomers and dimers, as they are formed of either one
144 or two RO_2 radicals, respectively. For the analysis in this study, we chose three mass/charge (m/z)
145 ranges of 50 Th each (Figure 1), corresponding to regions between which we expect differences in
146 formation or loss mechanisms. In addition to regions with HOM monomers and HOM dimers, one
147 range was chosen at lower masses, in a region presumably mainly consisting of molecules that are
148 less likely to condense onto aerosol particles (Peräkylä et al., 2020).

149

150 **2.2 Positive matrix factorization (PMF)**

151 After the model of PMF was developed (Paatero and Tapper, 1994), numerous applications have been
152 conducted with different types of environmental data (Song et al., 2007; Ulbrich et al., 2009; Yan et
153 al., 2016; Zhang et al., 2017). By reducing dimensionality of the measured dataset, PMF model greatly
154 simplifies the data analysis process with no requirement for prior knowledge of sources or pathways
155 as essential input. The main factors can be further interpreted with their unique/dominant markers
156 (elements or masses).

157 The basic assumption for PMF modelling is mass balance, which assumes that ambient concentration
158 of a chemical component is the sum of contributions from several sources or processes, as shown in
159 equation (1).

$$160 \quad \mathbf{X} = \mathbf{TS} \times \mathbf{MS} + \mathbf{R} \quad (1)$$

161 In equation (1), \mathbf{X} stands for the time series of measured concentration of different variables (m/z in
162 our case), \mathbf{TS} represents the temporal variation of factor contributions, \mathbf{MS} stands for factor profiles
163 (mass spectral profiles), and \mathbf{R} is the residual as the difference of the modelled and the observed data.
164 The matrices \mathbf{TS} and \mathbf{MS} are iteratively calculated by a least-squares algorithm utilizing uncertainty
165 estimates, to pursue minimized Q value as shown in equation (2), where S_{ij} is the estimated
166 uncertainty, an essential input in PMF model.

$$167 \quad Q = \sum \sum \left(\frac{R_{ij}}{S_{ij}} \right)^2 \quad (2)$$

168 PMF model was conducted by multi-linear engine (ME-2) (Paatero, 1999) interfaced with Source
169 Finder (SoFi, v6.3) (Canonaco et al., 2013). Signal-to-noise ratio (SNR) was calculated as $SNR_{ij} =$
170 $abs(X_{ij}) / abs(S_{ij})$. When the Signal-to-noise ratio (SNR) is below 1, the signal of X_{ij} will be down-
171 weighted by replacing the corresponding uncertainty S_{ij} by S_{ij}/SNR_{ij} (Visser et al., 2015). Future
172 studies should pay attention to the potential risk when utilizing this method since down-weighting
173 low signals element-wise will create a positive bias to the data. Robust mode was operated in the
174 PMF modelling, where outliers ($\left|\frac{R_{ij}}{S_{ij}}\right| > 4$) were significantly down-weighted (Paatero, 1997).

175

176 **2.3 binPMF**

177 As a newly developed application of PMF for mass spectral data, binPMF has no requirement for
178 chemical composition information, while still taking advantage of the HR mass spectra, saving effort
179 and time (Zhang et al., 2019). To explore the benefits of analyzing separated mass ranges, we applied
180 binPMF to the three separated ranges. The three ranges were also later combined for binPMF analysis
181 as comparison with the previous results. The PMF model requires both data matrix and error matrix
182 as input, and details of the preparation of data and error matrices are described below.

183

184 **2.3.1 Data matrix**

185 Different from normal UMR or HR peak fitting, in binPMF, the mass spectra are divided into small
186 bins after baseline subtraction and mass axis calibration. Linear interpolation was first conducted to
187 the mass spectra with a mass interval of 0.001 Th. Then the interpolated data was averaged into bins
188 of 0.02 Th width. We selected three ranges for further analysis based on earlier studies (Ehn et al.,
189 2014; Yan et al., 2016; Bianchi et al., 2019; Peräkylä et al., 2020).

- 190 - Range 1, m/z 250 – 300 Th, 51 unit masses \times 25 bins per unit mass = 1275 bins/variables,
191 consisting mainly of molecules with five to nine carbon atoms and four to nine oxygen atoms
192 in our dataset.
- 193 - Range 2, m/z 300 – 350 Th, $51 \times 25 = 1275$ bins, mainly corresponding to HOM monomer
194 products, featured with nine to ten C- and seven to ten O-atoms.
- 195 - Range 3, m/z 510 – 560 Th, $51 \times 30 = 1530$ bins, mainly corresponding to HOM dimer products,
196 with carbon numbers of sixteen to twenty and eleven to fifteen O-atoms.

197 To avoid unnecessary computation, only signal regions with meaningful signals in the mass spectra
198 were binned (Zhang et al., 2019). For a nominal mass N , the signal region included in further analyses
199 was between $N-0.2$ Th and $N+0.3$ Th for Range 1 and 2, and between $N-0.2$ Th and $N+0.4$ Th for

200 Range 3. The wider signal regions in Range 3 is due to wider peaks at higher masses. The data were
201 averaged into 1-h time resolution and in total we had 384 time points in the data matrix.

202

203 **2.3.2 Error matrix**

204 The error matrix represents the estimated uncertainty for each element of the data matrix and is crucial
205 for iterative calculation of the Q minimum. Equation (3) is used for error estimation (Polissar et al.,
206 1998),

$$207 \quad S_{ij} = \sigma_{ij} + \sigma_{\text{noise}} \quad (3)$$

208 where S_{ij} represents the uncertainty of $m/z, j$ at time i , σ_{ij} stands for counting statistics uncertainty
209 and is estimated as follows,

$$210 \quad \sigma_{ij} = a \times \frac{\sqrt{I_{ij}}}{\sqrt{t_s}} \quad (4)$$

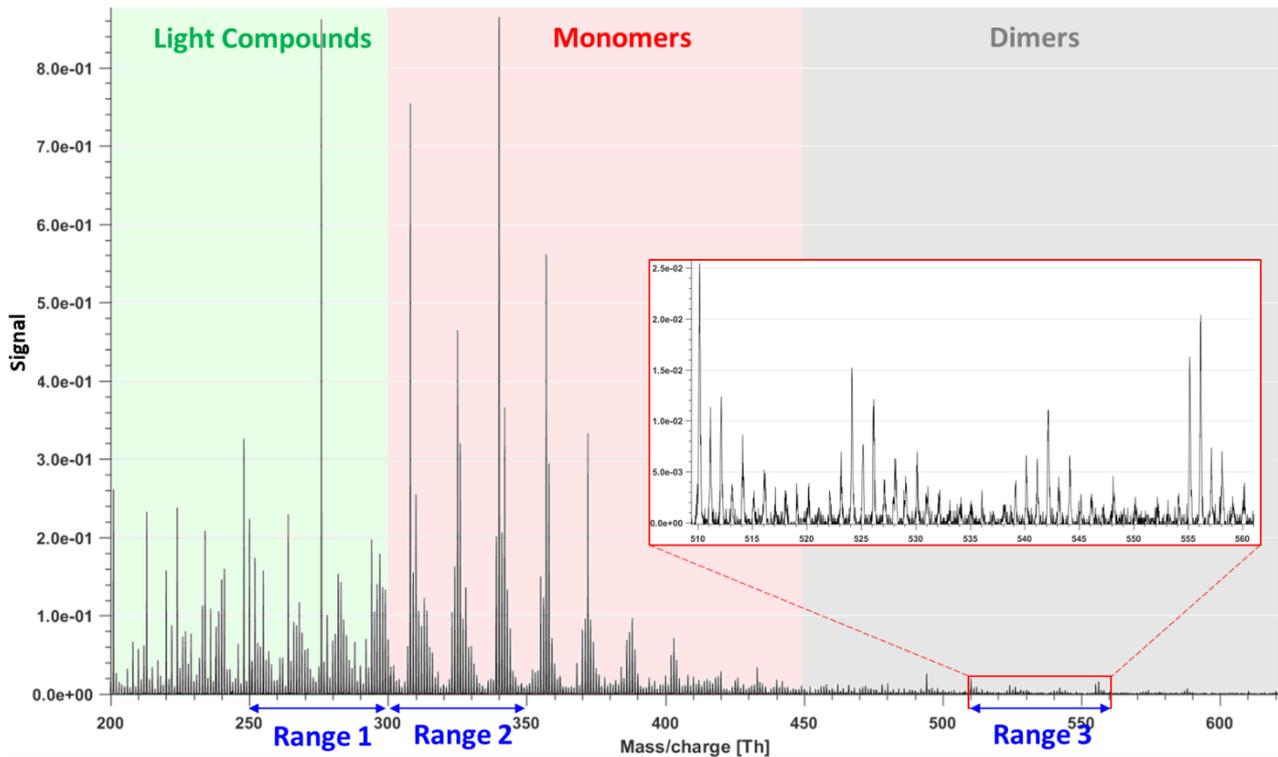
211 where I is the signal intensity term, in unit of counts per second (cps), t_s stands for length of averaging
212 in seconds, while a is an empirical coefficient to compensate for unaccounted uncertainties (Allan et
213 al., 2003; Yan et al., 2016) and is 1.28 in our study as previously estimated from laboratory
214 experiments (Yan et al., 2016). The σ_{noise} term was estimated as the median of the standard
215 deviations from signals in the bins in the region between nominal masses, where no physically
216 meaningful signals are expected.

217

218 **3 Results**

219 **3.1 General overview of the dataset/spectrum**

220 During the campaign, in autumn, 2016, the weather was overall sunny and humid with average
221 temperature of 10.8 °C and relative humidity (RH) of 87% (Zha et al., 2019). The average
222 concentration of NO_x and O_3 was 0.4 ppbv and 21 ppbv, respectively. The average total HOM
223 concentration was $\sim 10^8$ molecules cm^{-3} .



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Figure 1. Example of mass spectrum with 1-h time resolution measured from a boreal forest environment during the IBairn campaign (at 18:00, Finnish local time, UTC+2). The mass spectrum was divided into three parts and three sub-ranges were chosen from different parts for further analysis in our study. The nitrate ion (62 Th) is included in the mass.

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Figure 1 shows the 1 h averaged mass spectrum taken at 18:00 on September 12, as an example of the analyzed dataset. In addition to exploring the benefits of this type of sub-range analysis in relation to different formation or loss pathways, separating into sub-ranges may also aid factor identification for low-signal regions. As shown in Figure 1, there is a difference of 1-2 orders of magnitude in the signal intensity between Range 3 and Ranges 1-2. If all Ranges are run together, we would expect that the higher signals from Ranges 1 and 2 will drive the factorization. While if run separately, separating formation pathways of dimers in Range 3 will likely be easier. As dimers have been shown to be crucial for the formation of new aerosol particles from monoterpene oxidation (Kirkby et al., 2016; Troestl et al., 2016; Lehtipalo et al., 2018), this information may even be the most critical in some cases, despite the low contribution of these peaks to the total measured signal.

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binPMF was separately applied to Range 1, 2, 3, and a 'Range combined' which comprised all the three sub-ranges. All the PMF runs for the four ranges were conducted from two to ten factors and repeated three times for each factor number, to assure the consistency of the results. Factorization results and evolution with increasing factor number are briefly described in the following sections, separately for each Range (sections 3.2 – 3.5). It is worth noting that the factor order in factor

245 evolution does not necessarily correspond to that of the final results. The factor orders displayed in
246 Figures 2-5 have been modified for further comparison between different ranges. More detailed
247 discussion and comparison between the results are presented in Section 4.

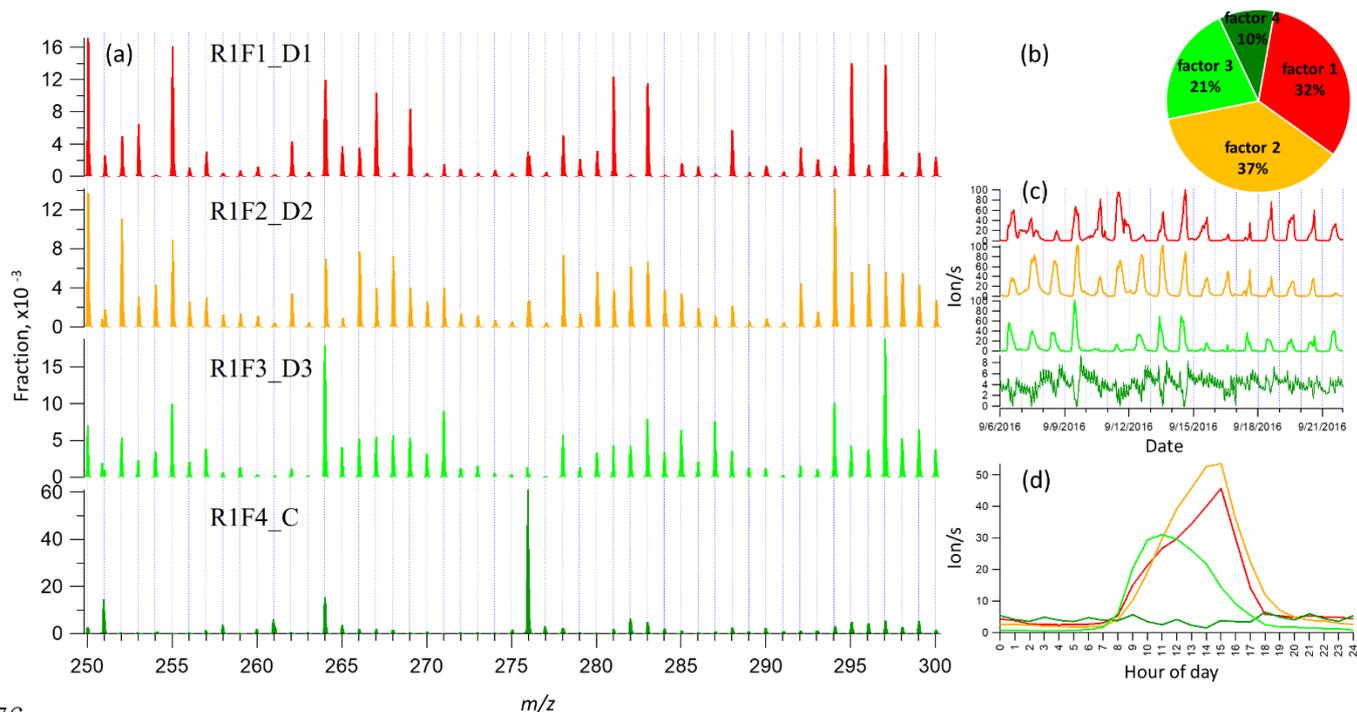
248

249 **3.2 binPMF on Range 1 (250 – 300 Th)**

250 As has become routine (Zhang et al., 2011; Craven et al., 2012), we first examined the mathematical
251 parameters of our solutions. From two to ten factors, Q/Q_{exp} decreased from 2.8 to 0.7 (Fig S1 in
252 supplementary information), and after three factors, the decreasing trend was gradually slowing down
253 and approaching one, which is the ideal value for Q/Q_{exp} as a diagnostic parameter. The unexplained
254 variation showed a decline from 18% to 8% from two to ten factors.

255 In the two-factor results, two daytime factors were separated, with peak time both at 14:00 - 15:00.
256 One factor was characterized by large signals at 250 Th, 255 Th, 264 Th, 281 Th, 283 Th, 295 Th,
257 297 Th. The other factor was characterized by large signals at 294 Th, 250 Th, 252 Th, 264 Th, 266
258 Th, 268 Th, and 297 Th. In Hyytiälä, as reported in previous studies, odd masses observed by the
259 nitrate CI-API-TOF are generally linked to monoterpene-derived organonitrates during the day (Ehn
260 et al., 2014; Yan et al., 2016). When the factor number increased to three, the two earlier daytime
261 factors remained similar with the previous result, while a new factor appeared with a distinct sawtooth
262 shape in the diurnal cycle. The main marker in the spectral profile was 276 Th, with a clear negative
263 mass defect. When one more factor was added, the previous three factors remained similar as in the
264 three-factor solution, and a new morning factor was resolved, with 264 Th and 297 Th dominant in
265 the mass spectral profile, and a diurnal peak at 11:00.

266 As the factor number was increased, more daytime factors were separated, with similar spectral
267 profiles to existing daytime factors and various peak times. No nighttime factors were found in the
268 analysis even when the factor number reached ten. We chose the four-factor result for further
269 discussion, and Figure 2 shows the result of Range 1, with spectral profile, time series, diurnal cycle
270 and averaged factor contribution during the campaign. As shown in Figure 2d, Factors 1-3 are all
271 daytime factors, while Factor 4 has no clear diurnal cycle, but a distinct sawtooth shape. Factor 4
272 comes from a contamination of perfluorinated acids, from the inlet's automated zeroing every three
273 hours during the measurements (Zhang et al., 2019). The zeroing periods have been removed from
274 the dataset before binPMF analysis, but the contamination factor was still resolved. This factor is
275 discussed in more detail in sections 4.1 and 4.4.



276

277 Figure 2 Four-factor result for Range 1, for (a) factor spectral profiles, (b) averaged factor contribution during
 278 the campaign, (c) time series and (d) diurnal trend. Details on the factors' naming schemes are shown in Table
 279 1.

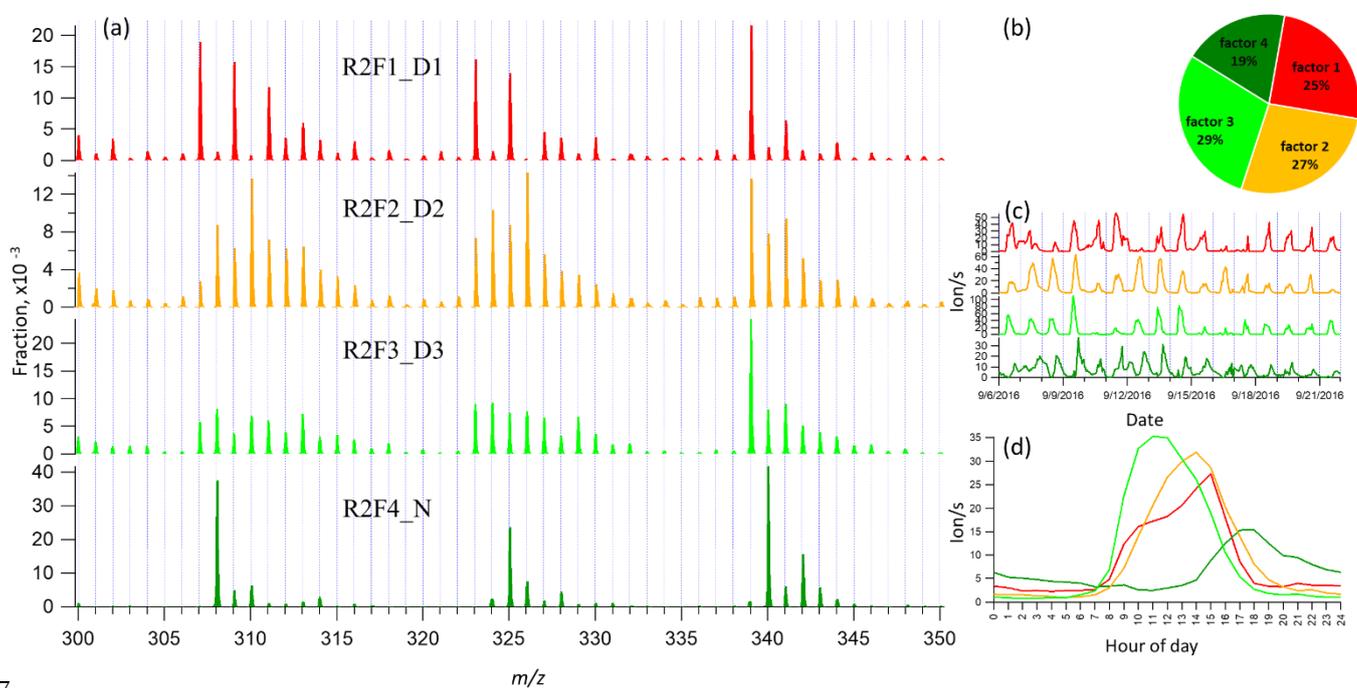
280

281 3.3 binPMF on Range 2 (300-350 Th)

282 This range covers the monoterpene HOM monomer range, and binPMF results have already been
 283 discussed by Zhang et al. (2019) as a first example of the application of binPMF on ambient data.
 284 Our input data here is slightly different. In the previous study, the 10 min automatic zeroing every
 285 three hours was not removed before averaging to 1 hour time resolution while here, we have removed
 286 this data. Overall, the results are similar as in our earlier study, and therefore the result are just briefly
 287 summarized below for further comparison and discussion in Section 4. Similar to Range 1, both the
 288 Q/Q_{exp} (2.2 to 0.6) and unexplained variation (16% to 8%) declined with the increased factor number
 289 from two to ten.

290 When the factor number was two, one daytime and one nighttime factor were separated, with diurnal
 291 peak times at 14:00 and 17:00, respectively. The nighttime factor was characterized by masses at 340
 292 Th, 308 Th and 325 Th (monoterpene ozonolysis HOM monomers (Ehn et al., 2014)) and remained
 293 stable throughout the factor evolution from two to ten factors. With the addition of more factors, no
 294 more nighttime factors got separated while the daytime factor was further separated and more daytime
 295 factors appeared, peaking at various times in the morning (10:00 am), at noon or in the early afternoon
 296 (around 14:00 pm and 15:00 pm). High contribution of 339 Th can be found in all the daytime factor

297 profiles. As the factor number reached six, a contamination factor appeared, characterized by large
 298 signals at 339 Th and 324 Th, showing negative mass defects (Figure S2 in the Supplement). The
 299 factor profile is nearly identical to the contamination factor determined in Zhang et al. (2019), where
 300 the zeroing periods were not removed, causing larger signals for the contaminants. In our dataset,
 301 where the zeroing periods were removed, no sawtooth pattern was discernible in the diurnal trend,
 302 yet it could still be separated even though it only contributed 3% to Range 2. More about the
 303 contamination factors from different ranges will be discussed in Section 4.4. We chose to show the
 304 four-factor result below, to simplify the later discussion and comparison. Figure 3 shows four-factor
 305 result of Range 2, with spectral profile, time series, diurnal cycle and averaged factor contribution
 306 during the campaign.



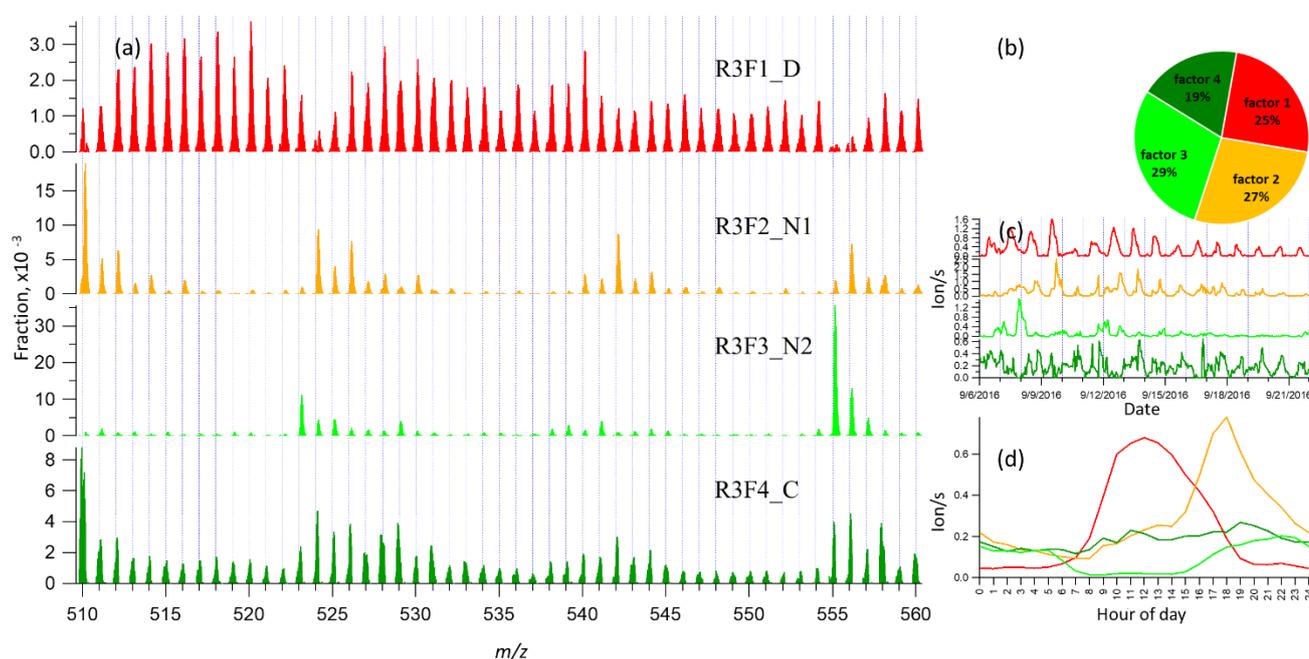
307
 308 Figure 3 Four-factor result for Range 2, for (a) factor spectral profiles, (b) averaged factor contribution during
 309 the campaign, (c) time series and (d) diurnal trend. Details on the factors' naming schemes are shown in Table
 310 1.

311

312 3.4 binPMF on Range 3 (510-560 Th)

313 Range 3 represents mainly the monoterpene HOM dimers (Ehn et al., 2014). Similar to Range 1 and
 314 2, both the Q/Q_{exp} (1.5 to 0.6) and unexplained variation (18% to 15%) showed decreasing trend with
 315 the increased factor number (2-10). As can be seen from Figure 1, data in Range 3 had much lower
 316 signals, compared to that of the Range 1 and 2, explaining the higher unexplained variation for Range
 317 3.

318 In the two-factor result for Range 3, one daytime and one nighttime factor appeared, with diurnal
 319 peak times at noon and 18:00, respectively. The nighttime factor was characterized by ions at 510 Th,
 320 524 Th, 526 Th, 542 Th, and 555 Th, 556 Th, while the daytime factor showed no dominant marker
 321 masses, yet with relatively high signals at 516 Th, 518 Th and 520 Th. As the number of factors
 322 increased to three, one factor with almost flat diurnal trend was separated, with dominant masses of
 323 510 Th, 529 Th, 558 Th. Most peaks in this factor had negative mass defects, and this factor was
 324 again linked to a contamination factor. The four-factor result resolved another nighttime factor with
 325 a dominant peak at 555 Th, and effectively zero contribution during daytime. As the factor number
 326 was further increased, the new factors seemed like splits from previous factors with similar spectral
 327 profiles. We therefore chose four-factor result also for Range 3 (results shown in Fig. 4) for further
 328 discussion.

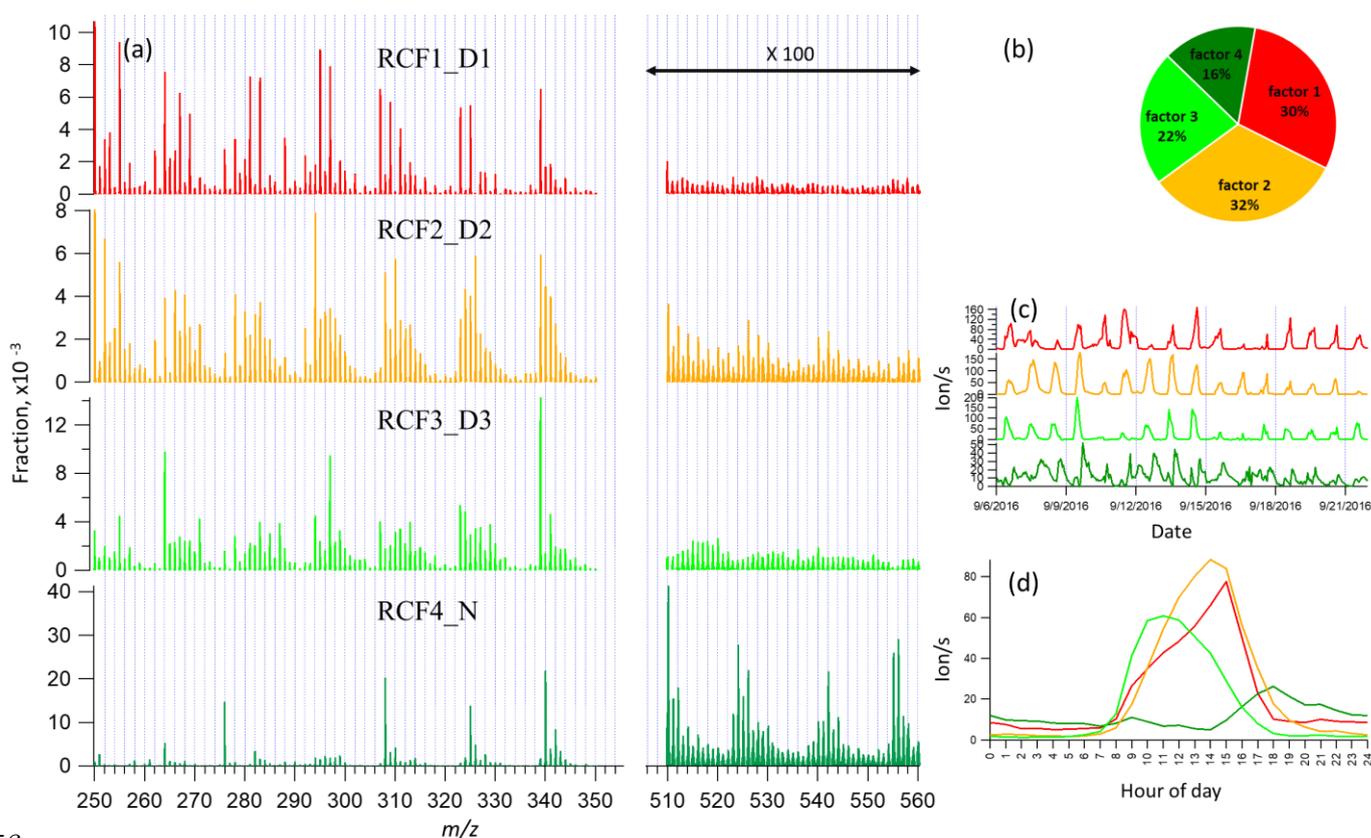


329
 330 Figure 4 Four-factor result for Range 3, for (a) factor spectral profiles, (b) averaged factor contribution during
 331 the campaign, (c) time series and (d) diurnal trend. Details on the factors' naming schemes are shown in Table
 332 1.

334 3.5 binPMF on Range Combined (250-350 Th & 510-560 Th)

335 As comparison to the previous three ranges, we conducted the binPMF analysis on Range Combined,
 336 which is the combination of the three ranges. The results of this range are fairly similar to those of
 337 Ranges 1 and 2, as could be expected since the signal intensities in these ranges were much higher
 338 than in Range 3. As the number of factors increased (2-10), both the Q/Q_{exp} (1.3 to 0.6) and
 339 unexplained variation (16% to 8%) showed a decreasing trend.

340 In the two-factor result, one daytime factor and one nighttime factor were separated. In the nighttime
 341 factor, most masses were found at even masses, and the fraction of masses in Range 3 was much
 342 higher than that in daytime factor. In contrast, in the daytime factor, most masses were observed at
 343 odd masses and the fraction of signal in Range 3 was much lower. During the day, photochemical
 344 reactions as well as potential emissions increase the concentration of NO, which serves as peroxy
 345 radical (RO₂) terminator and often outcompetes RO₂ cross reactions in which dimers can be formed
 346 (Ehn et al., 2014). Thus, the production of dimers is suppressed during the day, yielding instead a
 347 larger fraction of organic nitrates, as has been shown also previously (Yan et al., 2016).
 348 With the increase of the number of factors, more daytime factors were resolved with different peak
 349 times. When the factor number reached seven, a clear sawtooth-shape diurnal cycle occurred, i.e. the
 350 contamination factor, caused by the zeroing. As more factors were added, no further nighttime factors
 351 were separated, and only more daytime factors appeared. To simplify the discussion and inter-range
 352 comparison, we also here chose the four-factor result for further analysis. Figure 5 shows the four-
 353 factor result of Range Combined, with spectral profile, time series, diurnal cycle and averaged factor
 354 contribution during the campaign. The signals in range of 510-560 Th were enlarged 100-fold to be
 355 visible.



356

357 Figure 5 Four-factor result for Range Combined, for (a) factor spectral profiles, (b) averaged factor
 358 contribution during the campaign, (c) time series and (d) diurnal trend. Details on the factors' naming
 359 schemes are shown in Table 1.

360

361 4 Discussion

362 In Section 3, results by binPMF analysis were shown for Ranges 1, 2, 3 and Combined. In this section,
 363 we discuss and compare the results from the different ranges. To simplify the inter-range comparison,
 364 we chose four-factor results for all the four ranges, with the abbreviations shown in Table 1. From
 365 Range 1, three daytime factors and a contaminations factor were separated. In Range 2, three daytime
 366 factors and one nighttime factor (abbreviated as R2F4_N) were resolved. The R2F4_N factor was
 367 characterized by signals at 308 Th ($C_{10}H_{14}O_7 \cdot NO_3^-$), 325 Th ($C_{10}H_{15}O_8 \cdot NO_3^-$), and 340 Th
 368 ($C_{10}H_{14}O_9 \cdot NO_3^-$), and can be confirmed as monoterpene ozonolysis products (Ehn et al., 2014; Yan et
 369 al., 2016). With the increase of factor number to six, the contamination factor got separated also in
 370 this mass range. In Range 3, one daytime factor, two nighttime factors and a contamination factor
 371 were separated. The first nighttime factor (R3F2_N1) had large peaks at 510 Th ($C_{20}H_{32}O_{11} \cdot NO_3^-$)
 372 and 556 Th ($C_{20}H_{30}O_{14} \cdot NO_3^-$), dimer products that have been identified during chamber studies of
 373 monoterpene ozonolysis (Ehn et al., 2014). The molecule observed at 510 Th has 32 H-atoms,
 374 suggesting that one of the RO_2 involved would have been initiated by OH, which is formed during
 375 the ozonolysis of alkenes such as monoterpenes at nighttime (Atkinson et al., 1992; Paulson and
 376 Orlando, 1996). The other nighttime factor (R3F3_N2) was dominated by ions at 523 Th
 377 ($C_{20}H_{31}O_8NO_3 \cdot NO_3^-$) and 555 Th ($C_{20}H_{31}O_{10}NO_3 \cdot NO_3^-$), representing nighttime monoterpene
 378 oxidation involving NO_3 . As these dimers contain only one N-atom, and 31 H-atoms, we can assume
 379 that they are formed from reactions between an RO_2 formed from NO_3 oxidation and another RO_2
 380 formed by ozone oxidation. These results match well with the profiles in a previous study by Yan et
 381 al. (2016). The results of Range Combined are very similar to Range 2, with one nighttime factor and
 382 three daytime factors. The contamination factor was separated with increase of factor number to seven.

383

384

Table 1. Summary of PMF results for the different mass ranges

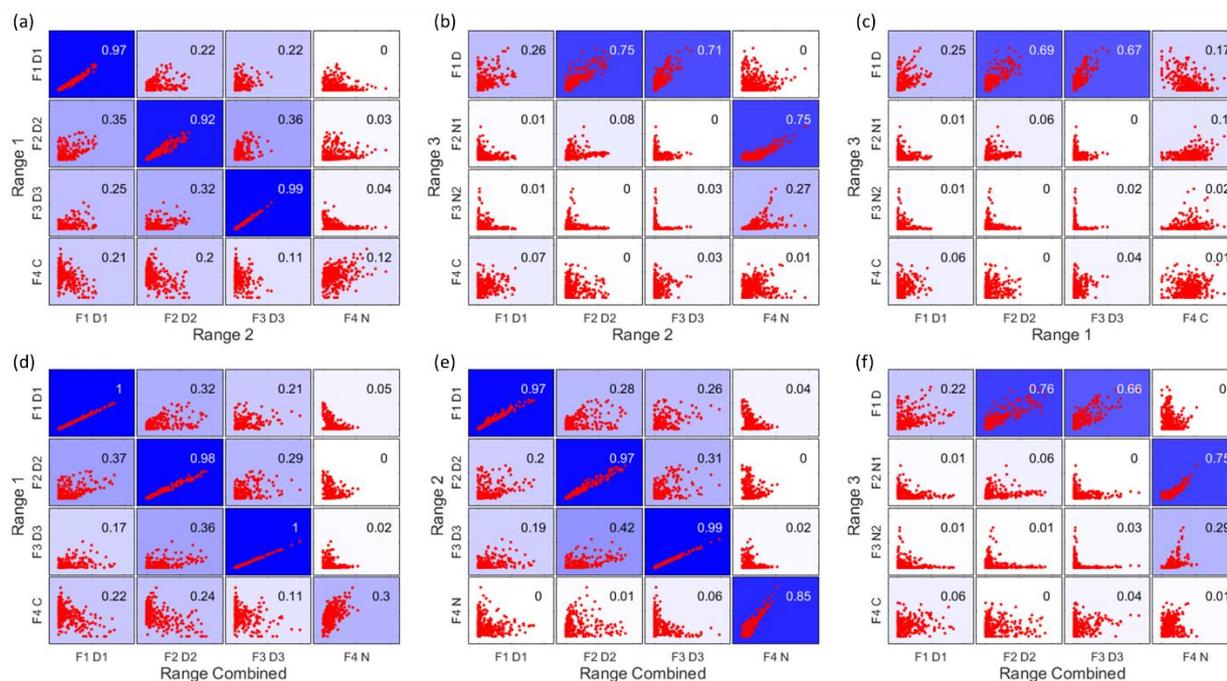
Range	Factor number	Factor name ^a	Dominant peaks	Peak time
1 (250 - 300 Th)	1	R1F1_D1	250, 255, 295, 297	15:00
	2	R1F2_D2	250, 252, 294	15:00
	3	R1F3_D3	264, 297	11:00
	4	R1F4_C	276	^b
2 (300 - 350 Th)	1	R2F1_D1	307, 309, 323, 325, 339,	15:00
	2	R2F2_D2	310, 326, 339,	14:00
	3	R2F3_D3	339	11:00
	4	R2F4_N	308, 325, 340	18:00

3 (510 – 560 Th)	1	R3F1_D	516, 518, 520, 528, 540	12:00
	2	R3F2_N1	510, 524, 542, 556	18:00
	3	R3F3_N2	523, 555	22:00
	4	R3F4_C	510, 558	- ^b
Combined (1, 2, 3)	1	RCF1_D1	250, 255, 295, 339	15:00
	2	RCF2_D2	250, 252, 294, 339	14:00
	3	RCF3_D3	264, 297, 339	11:00
	4	RCF4_N	308, 340, 510, 524, 555, 556	18:00

385 ^a Factor name is defined with range name, factor number and name. For example, RxFy represents Factor y in Range x.
 386 RC stands for Range Combined. For the factor name, D is short for daytime, N for Nighttime, C for contamination.
 387 ^b The contamination factor in Range 1 shows sawtooth pattern; while in Range 3 shows no diurnal pattern.
 388

389 4.1 Time series correlation

390 In Figure 6, the upper panels show the time series correlations among the first three ranges. As
 391 expected based on the results above, generally the daytime factors, and the two nighttime
 392 monoterpene ozonolysis factors (R2F4_N and R3F2_N1) correlated well, respectively. However, the
 393 contamination factors did not show strong correlation between different ranges, even though
 394 undoubtedly from the same source. More about the contamination factors will be discussed in Section
 395 4.4. The lower panels in Figure 6 displays the correlations between the first three ranges and the
 396 Range Combined, and clearly demonstrates that the results of Range Combined is mainly controlled
 397 by high signals from Range 1 and 2. More detailed aspects of the comparison between factors in
 398 different ranges is given in the following sections. The good agreements between factors from
 399 different subranges also help to verify the robustness of the solutions.



400
 401 Figure 6 Time series correlations among Range 1, 2, 3 (upper panels a-c), and between the first three ranges
 402 and the Range Combined (lower panels d-f). The abbreviations for different factors are the same in Table 1,
 403 with F for factor, D for daytime, N for nighttime and C for contamination, e.g. F1D1 for Factor 1 daytime 1.

404 The coefficient of determination, R^2 , is marked in each subplot by a number shown in the right upper corners
405 and by the blue colors, with stronger blue indicating higher R^2 .

406

407 **4.2 Daytime processes**

408 **4.2.1 Factor comparison**

409 As mentioned above, with increasing number of factors, usually more daytime factors will be resolved,
410 reflecting the complicated daytime photochemistry. The three daytime factors between Range 1 and
411 2 agreed with each other quite well (Figure 6a). However, R1F1_D1 and R2F1_D1 did not show
412 strong correlation with the only daytime factor in Range 3 (R3F1_D), while the other two daytime
413 factors in both Range 1 and 2, i.e. R1F2_D2, R1F3_D3, R2F2_D2, R2F3_D3, correlated well with
414 R3F1_D from Range 3.

415 The 1st daytime factors from Range 1 and 2, R1F1_D1 and R2F1_D1, were mainly characterized by
416 odd masses 255 Th, 281 Th, 283 Th, 295 Th, 297 Th, 307 Th, 309 Th, 311 Th, 323 Th, 325 Th, 339
417 Th. The factors are dominated by organonitrates. Organic nitrate formation during daytime is
418 generally associated with the termination of RO_2 radicals by NO. This termination step is mutually
419 exclusive with the termination of RO_2 with other RO_2 , which can lead to dimer formation. If the NO
420 concentration is the limiting factor for the formation of these factors, the low correlations between
421 the NO-terminated monomer factors, and the dimer factors, is to be expected. In contrast, if the other
422 daytime factors mainly depend on oxidant and monoterpene concentrations, some correlation
423 between those, and the daytime dimer factor, is to be expected, as shown in Figure 6b, c.

424 All the spectral profiles resolved from Range Combined binPMF analysis inevitably contained mass
425 contribution from 510 – 560 Th, even the daytime factor from Range Combined (RCF1_D1) which
426 did not show clear correlation with R3F1_D from Range 3 (Figure 6e).

427 The 2nd and 3rd daytime factors in Range 1 and 2, R1F2_D2, R1F3_D3, R2F2_D2, R2F3_D3, had
428 high correlations with R3F1_D in Range 3. Daytime factors in Range Combined (RCF2_D2 and
429 RCF3_D3) also showed good correlation with R3F1_D in Range 3. However, if we compare R3F1_D
430 and the mass range of 510 – 560 Th of the daytime factors in Range Combined, just with a quick look,
431 we can readily see the difference. The daytime factor separated in Range 3 (R3F1_D) has no obvious
432 markers in the profile. With the increase of factor number (up to ten factors), no clearly new factors
433 were separated in Range 3, but instead the previously separated factors were seen to split into several
434 factors. However, the spectral pattern in R3F1_D is different from that in the mass range of 510 –
435 560 Th in RCF2_D2. The factorization of Range Combined was mainly controlled by low masses
436 due to their high signals. The signals at high masses were forced to be distributed according to the

437 time series determined by small masses. Ultimately, this will lead to failure in factor separation for
 438 this low-signal range.

439 4.2.2 Daytime dimer formation

440 Dimers are primarily produced during nighttime, due to NO suppressing $RO_2 + RO_2$ reactions in
 441 daytime (Ehn et al., 2014; Yan et al., 2016). However, in this study, we found one clear daytime factor
 442 in Range 3 (R3F1_D, peak at local time 12:00, UTC+2) by sub-range analysis. With high loadings
 443 from even masses including 516, 518, 520, 528, 540 Th, this only daytime factor in dimer range
 444 correlated very well with two daytime factors in Ranges 1 and 2 (R1F2_D2, R1F3_D3, R2F2_D2,
 445 R2F3_D3) (Figure 6b and c). Table 2 include the correlation matrix of all PMF and factors and
 446 selected meteorological parameters. Strong correlation between R3F1_D with solar radiation was
 447 found, with $R = 0.79$ (Table 2). This may indicate involvement of OH oxidation in producing this
 448 factor.

449 Table 2 Correlation between factors and meteorological parameters and gases

	R1F1 _D1	R1F1 _D2	R1F1 _D3	R1F1 _C	R2F1 _D1	R2F2 _D2	R2F3 _D3	R2F4 _N	R3F1 _D	R3F2 _N1	R3F3 _N2	R3F4 _C	RCF1 _D1	RCF2 _D2	RCF3 _D3	RCF4 _N
O ₃	0.51	0.59	0.35	-0.18	0.47	0.57	0.36	0.43	0.55	0.33	0.27	0.22	0.49	0.57	0.33	0.34
NO	0.13	-0.01	0.24	-0.03	0.18	-0.02	0.24	-0.22	0.13	-0.19	-0.17	0.03	0.13	0.00	0.26	-0.18
NO _x	-0.05	-0.22	-0.10	0.09	-0.01	-0.23	-0.11	-0.13	-0.16	-0.21	-0.04	0.04	-0.04	-0.22	-0.09	-0.11
RH	-0.46	-0.80	-0.63	0.30	-0.43	-0.82	-0.64	-0.27	-0.78	-0.39	-0.07	-0.07	-0.43	-0.82	-0.60	-0.21
T	0.66	0.72	0.40	-0.24	0.65	0.66	0.41	0.39	0.65	0.30	0.14	0.19	0.66	0.68	0.38	0.24
UVB	0.52	0.63	0.82	-0.40	0.52	0.68	0.84	-0.30	0.79	-0.08	-0.27	0.08	0.49	0.68	0.83	-0.29

450
 451 As previous studies have shown, dimers greatly facilitate new particle formation (NPF) (Kirkby et
 452 al., 2016; Troestl et al., 2016; Lehtipalo et al., 2018), and this daytime dimer factor may represent a
 453 source of dimers that would impact the initial stages of NPF in Hyytiälä. Mohr et al. (2017) reported
 454 a clear diel pattern of dimers (sum of about 60 dimeric compounds of $C_{16-20}H_{13-33}O_{6-9}$) during NPF
 455 events in 2013 in Hyytiälä, with minimum at night and maximum after noon, and estimated these
 456 dimers can contribute ~5% of the mass of sub-60 nm particles. The link between the dimers presented
 457 in that paper and those reported here will require further studies, as will the proper quantification of
 458 the dimer factor identified here.

459 4.3 Nighttime processes

460 4.3.1 Factor comparison

461 Since high-mass dimers are more likely to form at night due to photochemical production of NO in
 462 daytime, which inhibits $RO_2 + RO_2$ reactions, Range 3 had the highest fraction of nighttime signals
 463

464 of all the sub-ranges. While Range 3 produced two nighttime factors, Ranges 2 and Combined showed
465 one, and Range 1 had no nighttime factor. The difference between the two results also indicates the
466 advantage of analyzing monomers and dimers separately.

467 The two nighttime factors in Range 3 can be clearly identified as arising from ozonolysis (R3F2_N1)
468 and a mix of ozonolysis and NO₃ oxidation (R3F2_N2) based on the mass spectral profiles, as
469 described above. The organonitrate at 555 Th, C₂₀H₃₁O₁₀NO₃·NO₃⁻, is a typical marker for NO₃
470 radical initiated monoterpene chemistry (Yan et al., 2016). However, several interesting features
471 become evident when comparing to the results of Range 2 and Combined. Firstly, only one nighttime
472 factor (R2F4_N, RCF4_N) was separated in each of these ranges, and that shows clear resemblance
473 with ozonolysis of monoterpenes as measured in numerous studies, e.g. Ehn et al. (2012);(2014).
474 Secondly, the high correlation found in Figure 6b between the ozonolysis factors (i.e., R2F4_N,
475 R3F2_N1, RCF4_N), further supports the assignment. However, factor R2F4_N is the only nighttime
476 factor in the monomer range, suggesting that NO₃ radical chemistry of monoterpenes in Hyytiälä does
477 not form substantial amounts of HOM monomers. The only way for the CI-APi-TOF to detect
478 products of monoterpene-NO₃ radical chemistry may thus be through the dimers, where one highly
479 oxygenated RO₂ radical from ozonolysis reacts with a less oxygenated RO₂ radical from NO₃
480 oxidation.

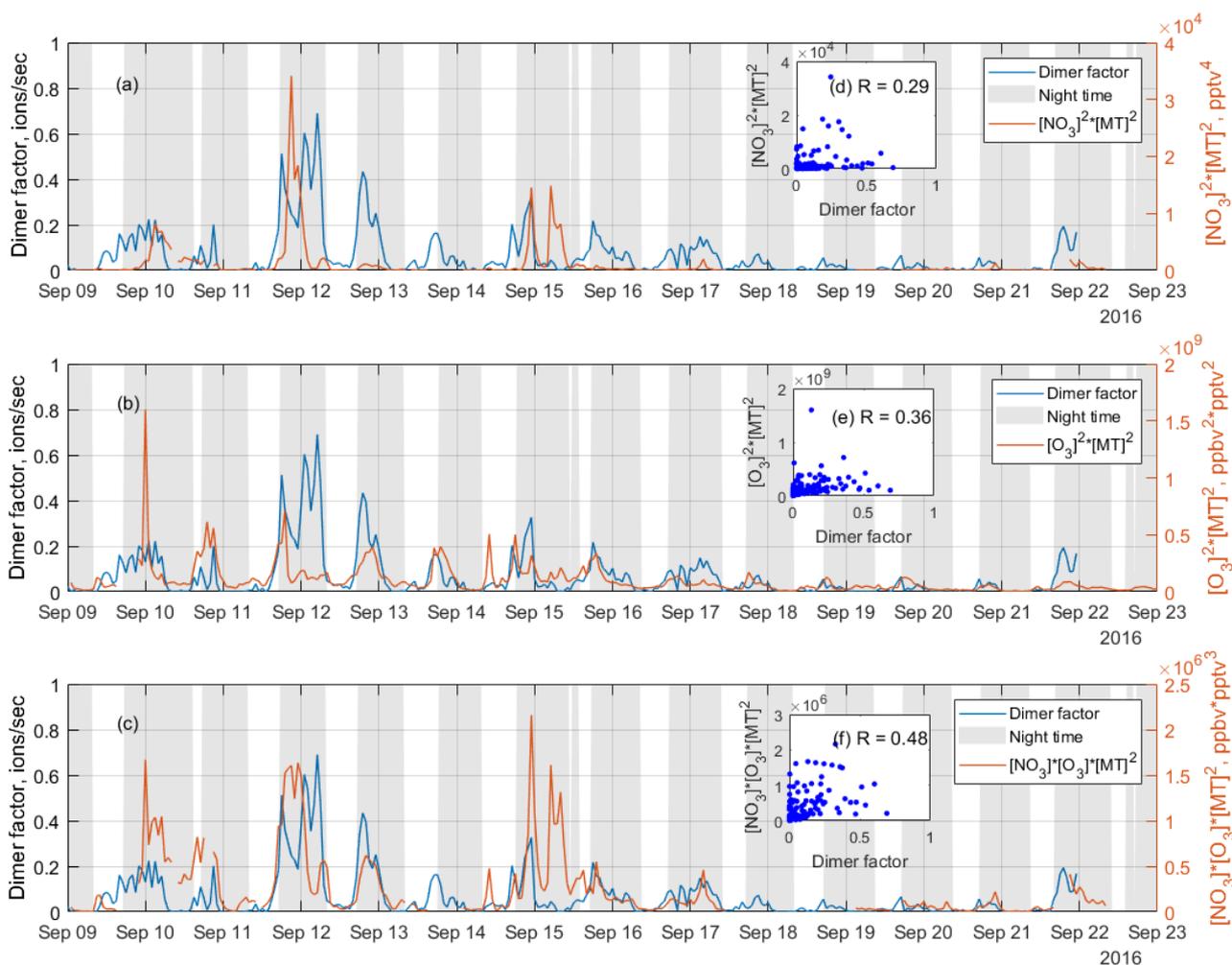
481 In the results by Yan et al. (2016) the combined UMR-PMF of monomers and dimers did yield a
482 considerable amount of compounds in the monomer range also for the NO₃ radical chemistry factor.
483 There may be several reasons for this discrepancy. One major cause for differences between the spring
484 dataset of Yan et al. (2016) and the autumn dataset presented here, is that nighttime concentrations
485 of HOM was greatly reduced during our autumn campaign. The cause may have been fairly frequent
486 fog formation during nights, and also the concentration of e.g. ozone decreased nearly to zero during
487 several nights (Zha et al., 2018). It is also possible that the NO₃ radical-related factor by Yan et al.
488 (2016) is probably a mixture of NO₃ and O₃ radical chemistry, while the monomer may thus be
489 attributed to the O₃ part. Alternatively, the different conditions during the two measurement periods,
490 as well as seasonal difference in monoterpene mixtures (Hakola et al., 2012), caused variations in the
491 oxidation pathways.

492

493 **4.3.2 Dimers initiated by NO₃ radicals**

494 Previous studies show that NO₃ oxidation of α-pinene, the most abundant monoterpene in Hyytiälä
495 (Hakola et al., 2012), produces fairly little SOA mass (yields <4 %), while β-pinene shows yields of
496 up to 53 % (Bonn and Moorgat, 2002;Nah et al., 2016). The NO₃+β-pinene reaction results in low
497 volatile organic nitrate compounds with carboxylic acid, alcohol, and peroxide functional groups (Fry

498 et al., 2014;Boyd et al., 2015), while $\text{NO}_3 + \alpha\text{-pinene}$ reaction will typically lose the nitrate functional
 499 group and form oxidation products with high vapor pressures (Spittler et al., 2006;Perraud et al.,
 500 2010). Most monoterpene-derived HOM, including monomers, are low-volatile (Peräkylä et al.,
 501 2020) and thus a low SOA yield indicates a low HOM yield. Thus, while there are to our knowledge
 502 no laboratory studies on HOM formation from NO_3 oxidation of $\alpha\text{-pinene}$, a low yield can be expected
 503 based on SOA studies.



504

505 Figure 7 Time series of the NO_3 oxidation dimer factor (blue line), and the product of (a) $[\text{NO}_3]^2 \times$
 506 $[\text{monoterpene}]^2$, (b) $[\text{O}_3]^2 \times [\text{monoterpene}]^2$, and (c) $[\text{NO}_3] \times [\text{O}_3] \times [\text{monoterpene}]^2$, where $[\]$ represents
 507 concentration in unit of pptv for NO_3 radicals and monoterpene, ppbv for O_3 , while the scatter plots are shown
 508 as inserts, (d), (e), (f), respectively. The scatter plots and correlation coefficients R are only calculated from
 509 nighttime data, which is selected based on solar radiation, to eliminate the influence from daytime oxidation
 510 processes.

511
 512 As discussed above, a dimer factor (R3F2_N2) was identified as being a crossover between NO_3
 513 radical initiated and O_3 initiated RO_2 radicals. Figure 7 shows the time series of this factor, as well as
 514 the product of $[\text{NO}_3]^2 \times [\text{monoterpene}]^2$, $[\text{O}_3]^2 \times [\text{monoterpene}]^2$, and $[\text{NO}_3] \times [\text{O}_3] \times [\text{monoterpene}]^2$.
 515 These products are used to mimic the formation rates of the RO_2 radicals reacting to form the dimers,

516 either from pure NO₃ oxidation (Fig. 7a), pure O₃ oxidation (7b), or the mixed reaction between RO₂
517 from the two oxidants (7c). The NO₃ concentration was estimated in Liebmann et al. (2018) for the
518 same campaign. Monoterpenes were measured using a proton transfer reaction time of flight mass
519 spectrometer (PTR-TOF-MS). More details on measurement of NO₃ proxy and monoterpene can be
520 found in in Liebmann et al. (2018).

521 As shown in Figure 7, the time series of the dimer factor tracks those of [NO₃] × [monoterpene] and
522 [O₃] × [monoterpene] reasonably well, but shows the highest correlation with the product of [NO₃] ×
523 [O₃] × [monoterpene]². This further supports this dimer formation as a mixed processes of ozonolysis
524 and NO₃ oxidation. The heterogeneity of the monoterpene emissions in the forest, and the fact that
525 no dimer loss process is included, partly explain the relatively low correlation coefficients. The
526 sampling inlets for PTR-TOF were about 170 m away from the NO₃ reactivity measurement
527 (Liebmann et al., 2018), which in turn was some tens of meters away from the HOM measurements.
528 Thus, this analysis should be considered qualitative only.

529 The nitrate dimer factor (R3F2_N2) was dominated by the organonitrate at 555 Th,
530 C₂₀H₃₁O₁₀NO₃·NO₃⁻. However, unlike the pure ozonolysis dimer factor which had a corresponding
531 monomer factor (R = 0.86 between factor R2F4_N and R3 F2_N1), this NO₃-related dimer factor did
532 not have an equivalent monomer factor. This suggests that the NO₃ oxidation of the monoterpene
533 mixture in Hyytiälä does not by itself form much HOM, but in the presence of RO₂ from ozonolysis,
534 the RO₂ from NO₃ oxidation can take part in HOM dimer formation. This further implies that,
535 different from previous knowledge based on single-oxidant experiments in chambers, NO₃ oxidation
536 may have a larger impact on SOA formation in the atmosphere where different oxidants exist
537 concurrently. This highlights the need for future laboratory studies to consider systems with multiple
538 oxidants during monoterpene oxidation experiments, to truly understand the role and contribution of
539 different oxidants, and NO₃ in particular.

540 **4.4 Fluorinated compounds**

541 During the campaign, an automated instrument zeroing every three hours was conducted. While the
542 zeroing successfully removed the low-volatile HOM and H₂SO₄, the process also introduced
543 contaminants into the inlet lines, e.g. perfluorinated organic acids from Teflon tubing. Each zeroing
544 process lasted for 10 min. In the data analysis, we removed all the 10-min zeroing periods, and
545 averaged the data to 1-h time resolution, but contaminants were still identified in all ranges by
546 binPMF. However, the correlation between contamination factors from different ranges is low (Figure
547 6c).

548 To future investigate the low factor correlations of the same source, three fluorinated compounds with
549 different volatilities, $(\text{CF}_2)_3\text{CO}_2\text{HF}\cdot\text{NO}_3^-$ (275.9748 Th), $(\text{CF}_2)_5\text{C}_2\text{O}_4\text{H}$ (338.9721 Th), and
550 $(\text{CF}_2)_6\text{CO}_2\text{HF}\cdot\text{NO}_3^-$ (425.9653 Th), were examined in fine time resolution, i.e. 1 min. The time series
551 and 3-h cycle of the three fluorinated compounds were shown in Figure S3 and S4 in Supplement.
552 The correlation coefficients dropped greatly before and after the zero period was removed, from 0.9
553 to 0.3 for R^2 between 276 Th and 339 Th, and 0.8 to 0.1 between 276 Th and 426 Th (Fig. S5a, b).
554 Similar effect is also found with the 1 h averaged data (Fig. S5c, d). It is evident that the three
555 fluorinated compounds were from the same source (zeroing process), but due to their different
556 volatilities, they were lost at different rates. This, in turn, means that the spectral signature of this
557 source will change as a function of time, at odds with one of the basic assumptions of PMF.
558 The analysis of the fluorinated compounds in our system was here merely used as an example to show
559 that volatility can impact source profiles over time. In Figure S5, it can be clearly seen that the profile
560 of Range Combined is noisier than that of Range 3, probably due to the varied fractional contributions
561 of contamination compounds to the profile. In ambient data, products from different sources can have
562 undergone atmospheric processing, altering the product distribution. This analysis highlighted the
563 importance of differences in the sink terms due to different volatilities of the products. This may be
564 an important issue for gas phase mass spectrometry analysis, potentially underestimated by many
565 PMF users, as it is likely only a minor issue for aerosol data, for which PMF has been applied much
566 more routinely. If failing to achieve physically meaningful factors using PMF on gas phase mass
567 spectra, our recommendation is to try applying PMF to sub-ranges of the spectrum, where IVOC,
568 SVOC and (E)LVOC could be analyzed separately.

569

570 **4.5 Atmospheric insights**

571 Based on the new data analysis technique binPMF applied in sub-ranges of mass spectra, we were
572 able to separate two particularly intriguing atmospheric processes, the formation of daytime dimers
573 as well as dimer formation involving NO_3 radicals, which otherwise could not have been identified
574 in our study.

575 With a diurnal peak around noon time, the daytime dimers identified in this study correlate very well
576 with daytime factors in monomer range. Strong correlation between this factor and solar radiation
577 indicate the potential role of OH oxidation in the formation of daytime dimers. By now, very few
578 studies have reported the observations of daytime dimers. As dimers are shown to be able to take part
579 in new particle formation (NPF) (Kirkby et al., 2016), this daytime dimer may contribute to the early
580 stages of NPF in the boreal forest.

581 The second process identified in our study is the formation of dimers that are a crossover between
582 NO₃ and O₃ oxidation. Such dimers have been identified before (Yan et al., 2016). However, we were
583 not able to identify corresponding HOM monomer compounds. This finding indicates that while NO₃
584 oxidation of the monoterpenes in Hyytiälä may not undergo autoxidation to form HOM by themselves,
585 they can contribute to HOM dimers when the NO₃-derived RO₂ react with highly oxygenated RO₂
586 from other oxidants. Multi-oxidant systems should be taken into consideration in future experimental
587 studies on monoterpene oxidation processes.

588

589 **5 Conclusions**

590 The recent developments in the field of mass spectrometry, combined with factor analysis techniques
591 such as PMF, have greatly improved our understanding of complicated atmospheric processes and
592 sources. In this study, we applied the new binPMF approach (Zhang et al., 2019), to separate sub-
593 ranges of mass spectra measured using a chemical ionization mass spectrometer in the Finnish boreal
594 forest. By using this method, we were able to identify a daytime dimer factor, presumably initiated
595 by OH/O₃ oxidation of monoterpenes, forming from RO₂+RO₂ reactions despite competition from
596 daytime NO. This compound group, showing a diurnal peak around noon, may contribute to new
597 particle formation at the site. In addition, we successfully separated NO₃-related dimers which would
598 not have been identified from this dataset without utilizing the different sub-ranges. The NO₃-related
599 factor was consistent with earlier observations (Yan et al., 2016), with the exception that we did not
600 observe any corresponding monomer factor. This may be explained by the observed nitrate-
601 containing dimers being formed from two RO₂, where one is initiated by oxidation by O₃, and the
602 other by NO₃. If the NO₃-derived RO₂ are not able to form HOM by themselves, there will not be any
603 related monomers observed. To validate this hypothesis, future laboratory experiments that target
604 more complex oxidation systems will be useful in order to understand the role of NO₃ oxidation in
605 SOA formation under different atmospheric conditions.

606 Apart from these two major findings, we also found several other benefits of applying PMF on
607 separate sub-ranges of the mass spectra. First, different compounds from the same source can have
608 variable loss rates due to differences in volatilities. This leads to increased difficulty for PMF to
609 separate this source, but if the PMF analysis is run separately on lighter masses (with higher volatility)
610 and heavier masses (with lower volatility), the source may become easier to distinguish. Secondly,
611 chemistry or sources contributing only to one particular mass range, e.g. dimers, can be better
612 separated. Thirdly, mass ranges with small, but informative, signals can be more accurately assigned
613 as their contribution becomes larger than if the entire mass range was analyzed at once. Finally,
614 running PMF on separate mass ranges also allows comparing the factors between the different ranges,

615 helping to verify the results. In summary, while we do not suggest that this type of sub-range analysis
616 should always be utilized, we recommend other analysts of gas-phase mass spectrometer data to test
617 this approach in order to see whether additional useful information can be obtained. In our dataset,
618 this method was crucial for identifying different types of dimers and dimer formation pathways,
619 which are of great importance for the formation of both new particles and SOA.

620

626 **Data availability.** The data used in this study are available from the first author upon request: please
627 contact Yanjun Zhang (yanjun.zhang@helsinki.fi).

635 **Author contribution.** ME and YZ designed the study. QZ and MR collected the data; data analysis
636 and manuscript writing were done by YZ. All coauthors discussed the results and commented the
637 manuscript.

638 **Competing interests.** The authors declare that they have no conflict of interest

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646

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