Importance of isomerization reactions for the OH radical regeneration from the photo-oxidation of isoprene investigated in the atmospheric simulation chamber SAPHIR

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Abstract. Theoretical, laboratory and chamber studies have shown fast regeneration of hydroxyl radical (OH) in the photochemistry of isoprene largely due to previously disregarded unimolecular reactions. Based on early field measurements, nearly complete regeneration was hypothesized for a wide range of tropospheric conditions, including areas such as the rainforest where slow regeneration of OH radicals is expected due to low concentrations of nitric oxide (NO). In this work the OH regeneration in the isoprene oxidation is directly quantified for the first time through experiments covering a wide range of atmospheric conditions (i.e. NO between 0.15 and 2 ppbv and temperature between 25 and 41°C) in the atmospheric simulation chamber SAPHIR. These conditions cover remote areas partially influenced by anthropogenic NO emissions, giving a regeneration efficiency of OH close to one, and areas like the Amazonian rainforest with very low NO, resulting in a surprisingly high regeneration efficiency of 0.5, i.e. a factor of 2 to 3 higher than explainable in the absence of unimolecular reactions. The measured radical concentrations were compared to model calculations and the best agreement was observed when at least 50\% of the total loss of isoprene peroxy radicals conformers (weighted by their abundance) occurs via isomerization reactions for NO lower than 0.2 parts per billion (ppbv). For these levels of NO, up to 50\% of the OH radicals are regenerated from the products of the 1,6 α-hydroxy-hydrogen shift (1,6-H shift) of Z-δ-RO\textsubscript{2} radicals through photolysis of an unsaturated hydroperoxy aldehyde (HPALD) and/or through the fast aldehyde hydrogen shift (rate constant \textasciitilde10 s\textsuperscript{-1} at 300K) in di-hydroperoxy carbonyl peroxy radicals (di-HPCARP-RO\textsubscript{2}), depending on their relative yield. The agreement between all measured and modelled trace gases (hydroxyl, hydroperoxy and organic peroxy radicals, carbon monoxide and the sum of methyl vinyl ketone, methacrolein and hydroxyl hydroperoxides) is nearly independent on the adopted yield of HPALD and di-HPCARP-RO\textsubscript{2} as both degrade relatively fast (<\,1 h), forming OH radical and CO among other products. Taking into consideration this and earlier isoprene studies, considerable uncertainties remain on the oxygenated products distribution, which affect radical levels and organic aerosol downwind of unpolluted isoprene dominated regions.

1 Introduction

The hydroxyl radical (OH) is the main daytime oxidant controlling the removal and transformation of gaseous pollutants in the atmosphere (Levy, 1974). Its high efficiency in the oxidation of trace gases is based on the effective regeneration of OH by radical chain reactions, in which nitric oxide (NO) is oxidized to nitrogen dioxide (NO\textsubscript{2}), linking the OH chemistry to the formation of the tropospheric pollutant ozone (O\textsubscript{3}). Because high levels of OH radicals were observed in field experiments in mainly forested environments with large concentrations of isoprene (Tan et al., 2001; Ren et al., 2008; Hofzumahaus et al.,
In the OH-initiated isoprene oxidation, the first reaction step comprises the formation of six isoprene-RO₂ conformers, which are in equilibrium and can quickly interconvert as first suggested in the Leuven Isoprene Mechanism (LIM) (Peeters et al., 2009) (Fig. 1). The concentration of the different conformers, which is affected by both their losses (unimolecular decomposition and reaction with NO, hydroxyl and RO₂ radicals) and their re-equilibration, can have a large impact on the OH radical concentration. There are three different sets of reaction rate coefficients currently available in the literature for the equilibrium reactions between the isoprene-RO₂ conformers (Table 1). One set is from theoretical calculations in the LIM1 study (Peeters et al., 2014). A second set is currently in use within the master Chemical Mechanisms version 3.3.1 (MCM v3.3.1, Jenkin et al., 2015) where the rate coefficients are as described in LIM1 but all increased by a factor of five. Finally, Wennberg et al. (2018) suggest in their recent review paper on the mechanism of isoprene degradation (Caltech mechanism) to use different rate coefficients based on the experimental study by Teng et al. (2017) (Table 2).

Four of the six isoprene-RO₂ conformers can undergo atmospheric relevant isomerization reactions (Fig. 2). The β-RO₂ radicals directly reform OH radicals, together with the oxygenated organic products methacrolein (MACR), methyl vinyl ketone (MVK) and formaldehyde (HCHO), after a 1,5 hydroxy-hydrogen shift (1,5-H shift) (Da Silva et al., 2010) with a slow reaction rate constant (1.1 \times 10^{-3} \text{s}^{-1} and 0.7 \times 10^{-1} \text{s}^{-1} at 298 K for OH addition on C4 and C1, respectively) (Peeters et al., 2014) making this reaction competitive only in the presence of exceptionally low NO levels (< 10 pptv).

The most relevant isomerization reaction, the 1,6 α-hydroxy-hydrogen shift (1,6-H shift), occurs for the Z-δ-RO₂ radicals with a fast reaction rate coefficient (3.6 s⁻¹ and 0.4 s⁻¹ at 298 K for OH addition on C4 and C1, respectively) (Teng et al., 2017). These measured values are currently in use within the Caltech mechanism and they are in good agreement with the calculated rates in the LIM1 (Peeters et al., 2014) (within 40%). The MCM v 3.3.1 is currently using a factor of ~ 5 slower rate coefficients (Table 2). One of the predicted (Peeters et al., 2009; Peeters et al., 2014) and measured (Crounse et al., 2011; Berndt, 2012; Teng et al., 2017; Berndt et al., 2019) products following the 1,6-H shift is an unsaturated hydroperoxy aldehyde (HPALD), which can photolyze producing OH radicals (Wolfe et al., 2012) (atmospheric lifetime of ~1 h). In addition, formation of a di-hydroperoxy carbonyl peroxy radical (di-HPCARP-RO₂, Fig. 2) was predicted by theoretical calculations (Peeters et al., 2014) and inferred in a recent experimental study (Teng et al., 2017). Its fate is a fast unimolecular decomposition (~ 1 s⁻¹) with formation of OH radical (Wang et al., 2018; Müller et al., 2019) and suggested subsequent elimination of CO and formation of a di-hydroperoxy carbonyl compound (Peeters et al., 2014; Müller et al., 2019). Large uncertainties remain on the yield of these two products. Two experimental studies (Berndt, 2012; Teng et al., 2017) suggest a yield for HPALD in the order of 0.4 to 0.5 as postulated in the LIM1 (Peeters et al., 2014). In contrast, a recent experimental study by Berndt et al. (2019) puts a lower limit of 0.75, where a recent addition (Müller et al., 2019) to the theoretical work within the LIM1 also rationalizes a much higher yield of HPALD (74%) than previously reported in LIM1. Both currently available explicit isoprene oxidation mechanisms, i.e. the Master Chemical Mechanisms (MCM v3.3.1) (Jenkin et al., 2015) and the Caltech mechanisms (Wennberg et al., 2018), use a yield of 0.5 and 0.4, respectively, for HPALD.
To summarize, recent experimental and theoretical studies agree that the most relevant isomerization reaction of isoprene peroxyl radicals is the 1,6-H shift of Z-δ-RO₂. The following kinetic aspects control the impact of the 1,6-H shift of Z-δ-RO₂ on the regeneration of OH radicals and the production of oxygenated products:

- the equilibrium between the isoprene-RO₂ conformers, which determines the fraction of Z-δ-RO₂ radicals that can undergo fast 1,6-H-shift isomerization;
- the temperature dependent rate coefficient for the 1,6-H shift itself;
- the relative yields of HPALD and di-HPCARP-RO₂ formed following the 1,6-H shift;
- the follow-up chemistry of HPALD and di-HPCARP-RO₂.

Despite intensive research as detailed above, there exist significant differences between current chemical mechanisms (Table 2), i.e. different sets of rate coefficients are used for the equilibrium reactions, rate coefficients for the 1,6-H shift differ by up to a factor of five, and the measured yield of HPALD ranges from 0.4 to 0.75.

In this work, new chamber experiments have been performed to test our understanding of the photo-oxidation of isoprene. The experiments are used to test the ability of the explicit mechanisms MCM v 3.3.1 to predict OH radical regeneration from isoprene oxidation over a wide range of NO concentrations (0.15 to 2 ppbv). The chemistry of di-HPCARP-RO₂ has been investigated (Novelli et al., 2018a) with high levels of theory in particular to confirm the role of these radicals in OH radical formation. Model sensitivity studies are applied to identify the isoprene-RO₂ conformers equilibrium constants, 1,6-H shift rate constant and HPALD/diHPCARP branching ratio that provide the best description of the observed radical and trace gases concentrations. The global impact of the optimized isoprene mechanism on the OH radical concentration is shown.

2 Methods

2.1 Quantum-chemical and theoretical kinetic calculations

The reactants, transition states and products in the studied mechanistic branches of the isoprene chemistry were characterized at the M06-2X and CCSD(T) levels of theory. The conformer space for each of these structures was characterized at the M06-2X/cc-pVDZ level of theory (Dunning, 1989; Zhao and Truhlar, 2008; Alecu et al., 2010; Bao et al., 2017), locating ~24000 distinguishable structures from ~60000 systematically generated starting geometries. The most relevant conformers (~850 structures across all reactions examined) were then fully re-optimized at the M06-2X/aug-cc-pVTZ level of theory (Dunning, 1989) The number of conformers re-optimized at this higher level of theory differs per structure (see Table S1), but enough were included to cover over ~80% of the thermal population at 300K. Intrinsic reaction coordinate (IRC) calculations were performed on the lowest transition states (TS) to verify the nature to the transition state, and to provide the energies used for Eckart tunneling corrections. Finally, single point energy calculations at the CCSD(T)/aug-cc-pVTZ level of theory(Purvis and Bartlett, 1982) were performed on the energetically lowest-lying geometries of each structure, to further refine the energy barrier estimates. The thermal rate coefficients were then obtained using multi-conformer transition state theory (MC-TST) incorporating the energetic and rovibrational characteristics of all conformers (Vereecken and Peeters, 2003).

Temperature-dependent rate coefficients are derived for the temperature range between 200 and 400 K, and both isomer-specific and bulk phenomenological rate coefficients are provided, where the latter incorporate the effect of fast H-scrambling in the hydroperoxide-peroxyl radical isomers. See the supplementary information for a more detailed description of the methodologies involved.

2.2 Atmospheric simulation chamber SAPHIR and experimental procedure

The experiments were conducted in the atmospheric simulation chamber SAPHIR at Forschungszentrum Jülich, Germany. The SAPHIR chamber is designed for the investigation of oxidation processes under atmospheric conditions in a controlled
environment. SAPHIR is made of a double-wall Teflon (FEP) film that is inert, has a high transmittance for solar radiation (Bohn and Zilken, 2005), and is equipped with a shutter system that is opened during photolysis experiments allowing solar radiation to penetrate the chamber. The synthetic air provided to the chamber is mixed from ultra-pure nitrogen and oxygen (Linde, > 99.99990 %). Two fans in the chamber ensure a complete mixing of trace gases within two minutes. The pressure in the chamber is slightly higher than ambient (~ 30 Pa) to avoid external air penetrating the chamber. Due to small leakages and air consumption by instruments, trace gases are diluted at a rate of ~ 6 % h⁻¹ due to the replenishment flow. More details regarding the chamber can be found elsewhere (Rohrer et al., 2005; Poppe et al., 2007; Schlosser et al., 2007).

For the experiments the chamber was cleaned before the experiment by exchanging the chamber air 8 to 10 times with pure synthetic air. Evaporated Milli-Q® water was then introduced into the dark chamber by a carrier flow of synthetic air until a concentration of ~ 5 x 10⁻⁶ cm⁻³ of water vapour was reached. In order to keep the concentration of NO as small as possible after the opening of the shutters, ozone produced by a silent discharge ozonizer (O3onia) was added in the chamber to reach ozone mixing ratios up to 100 ppbv. For experiments at higher concentrations of NO, NO was injected from a gas mixture (Linde, 500 ppm NO in N₂) into the chamber by a mass flow controller. After opening the shutter system of the chamber, nitrous acid (HONO) was photochemically formed on the Teflon surface and released into the chamber (Rohrer et al., 2005), and its subsequent photolysis produced OH radicals and NO. Afterwards, isoprene was injected three times at intervals of about two hours directly from the liquid (99% purity, Sigma Aldrich). The aim was to reach ~ 6 ppbv of isoprene in the chamber after each injection (which corresponds to an OH reactivity between 12 to 15 s⁻¹). Experiments were designed such that chamber-specific sinks (dilution and wall loss of trace gases), and sources of trace gases that are formed in the sunlit chamber except for nitrous acid, did not influence the results.

2.3 Instrumentation

Table S1 summarizes the instruments available during the experiment, giving time resolution, accuracy, and precision for each instrument. The concentrations of OH, HO₂ and RO₂ radicals were measured with the laser induced fluorescence (LIF) instrument permanently in use at the SAPHIR chamber and described previously (Holland et al., 2003; Fuchs et al., 2011). Several studies, showing the presence of an interference in the OH radical detection with the LIF for ambient measurements in some environments, have been recently published (Mao et al., 2012; Novelli et al., 2014; Rickly and Stevens, 2018). The interference depends on the chemical conditions of the sampled air and on the geometry of the different instruments. A laboratory study performed with this LIF instrument (Fuchs et al., 2016) showed only interferences for high ozone concentrations (300 ppbv – 900 ppbv) together with BVOC concentrations up to 450 ppbv, which are far beyond any condition encountered in this study. Therefore, the OH radical concentration measured by the LIF instrument in this study is considered free from interferences. In addition, OH was measured by differential optical absorption spectroscopy (DOAS) (Dorn et al., 1995) for some of the experiments shown within this study. Numerous inter-comparisons between the LIF and the DOAS instrument in the SAPHIR chamber (Schlosser et al., 2007; Schlosser et al., 2009; Fuchs et al., 2012) showed very good agreement between these two instruments, giving high reliability to the OH radical measurements performed in the chamber. For the experiments within this study, a slope of 1.1 for the scatter plot of DOAS OH vs LIF OH was obtained with a correlation coefficient, R², of 0.94.

Several studies have proven that RO₂ radicals can cause an interference signal in the HO₂ radicals measured by conversion to OH after reaction with an excess of NO (Fuchs et al., 2011; Hornbrook et al., 2011; Whalley et al., 2013; Lew et al., 2017). It was shown that a reasonable approach for avoiding the interference is to lower the concentration of NO reacting with the sampled air inside the instrument. During this study, the NO concentration used was low (~ 2.5 x 10¹⁵ cm⁻³) to minimize the possibility of an interference as described in Fuchs et al. (2011). Still, as the RO₂ radicals which originate from the oxidation of isoprene by OH radicals are those able to quickly convert, despite the low NO used, interference from RO₂ radicals was still observed. Tests performed on the LIF instrument used for this study showed that, for the conditions the instrument was...
used in during the experiments, an interference of ~ 30% was observed for isoprene-RO₂. As such, the HO₂ radical measurement was defined as HO₂* to indicate the presence the interference from RO₂ radicals. As the measured RO₂ concentration is obtained from the difference between measured ROₓ and HO₂ radicals, the obtained RO₂ radicals is also underestimated due to the interference observed in the HO₂ measurement and will be marked as RO₂*. The OH reactivity (k-\text{OH}) the inverse lifetime of OH, was measured by a pump and probe technique coupled with a time-resolved detection of OH by LIF (Lou et al., 2010; Fuchs et al., 2017). Isoprene and the sum of MVK and MACR were measured by a proton-transfer-reaction time-of-flight mass spectrometer (PTR-TOF-MS,(Lindinger et al., 1998; Jordan et al., 2009)) and a gas chromatography system (GC,(Wegener et al., 2007)). As the PTR-TOF-MS and the GC were calibrated only for the species listed in table S1, concentrations for other species were not available. Carbon monoxide (CO), carbon dioxide (CO₂), methane (CH₄), and water vapour were measured by an instrument applying cavity ring-down spectroscopy (CRDS, Picarro). NO and nitrogen dioxide (NO₂) were measured by chemiluminescence (CL,(Ridley et al., 1992)) and O₃ by UV absorption (Ansyco). Photolysis frequencies were calculated from measurements of solar actinic radiation by a spectroradiometer (Bohn et al., 2005; Bohn and Zilken, 2005).

2.4 Model calculations

The measured radicals and trace gases were modelled with a zero-dimensional box model using chemical mechanistic information from the Master Chemical Mechanism downloaded via website: http://mcm.leeds.ac.uk/MCM. The MCM v3.3.1 was released in 2015 with newly updated isoprene chemistry described in Jenkin et al. (2015). The most relevant changes for this study are the inclusion of the equilibrium reactions between the OH-isoprene adducts and the isoprene-RO₂ (Fig. 1), and the inclusion of isomerization reactions for isoprene-RO₂ radicals (Table 2).

Several chamber-specific properties were implemented in the model. First, a dilution rate was applied to all trace gases present in the model to account for the dilution from the replenishing flow of the chamber. The background production of HONO and HCHO known to occur in the sunlit chamber (Rohrer et al., 2005; Karl et al., 2006), was parametrized by an empirical function that depends on temperature, relative humidity and solar radiation. For the experiments shown in this study, the background OH reactivity in the chamber was at most 1 s⁻¹ and was parametrized with a co-reactant Y added to the model, which converts OH to HO₂ in the same way as CO does (Fuchs et al., 2012; Fuchs et al., 2014; Kaminski et al., 2017). The concentration of Y was adjusted to match the observed OH reactivity during the zero-air phase of the experiment and was kept constant throughout the experiment. As shown in a previous study (Novelli et al., 2018b), the unknown chemical nature of the background reactivity that dominates the loss of OH radicals for the zero-air phase of the experiment has a negligible impact once the main reactant, in this case isoprene, is added, with total OH reactivity as high as 20 s⁻¹.

Photolysis frequencies for O₃, NO₂, HONO, hydrogen peroxide (H₂O₂), formaldehyde (HCHO), acetone, glyoxal, MVK and MACR were constrained to values calculated from the measured actinic flux. For HPALD and peroxy acid aldehyde (PACALD) photolysis frequencies of MACR scaled up by a factor of 100 and 2, respectively, were used (Fuchs et al., 2013). All the other photolysis frequencies present in the model were first calculated for clear sky conditions according to the MCM and then scaled by the ratio of measured to calculated j(NO₂) to account for clouds and transmission of the chamber film. The model was constrained to measured chamber pressure (= ambient pressure) and temperature, as well as water vapour, NO, NO₂ and O₃ concentrations. Model values were re-initiated at 1 minute intervals. Isoprene injections were implemented in the model by applying an isoprene source only active at the time of the injection, adjusted in strength to reproduce the observed change in OH reactivity at the injection time. The modelled OH reactivity used for comparison against the measurement is the total modelled OH reactivity excluding the reactivity of ISOPOOHs (Fig. 2), as for these compounds the OH radicals are recycled at a time scale much shorter than the OH lifetime observed in the k-\text{OH} instrument, negating their measurable OH reactivity. Measurements of MVK and MACR by PTRMS and GC are affected by
interferences from ISOPOOHs (Rivera-Rios et al., 2014). For this reason the measured data are compared with the sum of MVK and MACR together with ISOPOOHs (all isomers); the same sensitivity for MVK, MACR and ISOPOOHs is assumed. Due to the RO₂ interference in the HO₂ measurement, modelled HO₂ concentrations were increased by a small fraction of modelled RO₂ (10% of methoxy radical and 30% of RO₂ radicals from isoprene). Likewise, the measured RO₂⁺ values are compared against the difference between modelled RO and HO₂⁺, rather than the uncorrected RO₂ concentrations, to account for this interference.

The chemistry of di-HPCARP-RO₂ as investigated within this study was implemented in the model. This model (MCM 331, Table 2) served as the basis of all model calculations done in this work, with variations as defined below. Table 2 summarizes the additional model runs performed:

- M0 was performed by removing all isomerization reactions from the MCM 331 model (see supporting information and table S2).
- M1 was performed by using the rate coefficients for the addition of O₂ to OH-isoprene adducts, and for re-dissociation of isoprene-RO₂ the rate coefficient for the 1,6-H shift of Z-δ-RO₂ radicals and relative yield of HPALD/di-HPCARP-RO₂ as applied in the Caltech mechanism (Table S3).
- M2 is the same as M1 but using the rate coefficients for the addition of O₂ to OH-isoprene adducts, and for re-dissociation of isoprene-RO₂ as available in the MCM v 3.3.1 (Table S4).
- M3 is the same as M2 but using a relative yield of HPALD/di-HPCARP-RO₂ of 0.75:0.25 adapted from Berndt et al. (2019) (Table S5).

2.5 Global model
The ECHAM/MESSy Atmospheric Chemistry (EMAC) model is a numerical chemistry and climate simulation system that includes sub-models describing tropospheric and middle atmosphere processes and their interaction with oceans, land and human influences (Jöckel et al., 2010). It uses the second version of the Modular Earth Submodel System (MESSy2) to link multi-institutional computer codes. The core atmospheric model is the 5th generation European Centre Hamburg general circulation model (ECHAM5) (Roeckner et al., 2006). For the present study we applied EMAC (ECHAM5 version 5.3.02, MESSy version 2.53.0) in the T106L31ECMWF-resolution, i.e. with a spherical truncation of T106 (corresponding to a quadratic Gaussian grid of approximately 1.1 by 1.1 degrees in latitude and longitude) with 31 vertical hybrid pressure levels up to 10 hPa. The applied model setup comprised the submodel MECCA (Module Efficiently Calculating the Chemistry of the Atmosphere) to calculate atmospheric chemistry using the Mainz Organic Mechanism (MOM)(Sander et al., 2011). The mechanism was adapted to the changes proposed in this study. In addition, the submodel MEGAN (Model of Emissions of Gases and Aerosols from Nature) was used to simulate biogenic emissions of tracers, including isoprene (Guenther et al., 2006). Global isoprene emissions were scaled to 595 Tg/year, which is the best estimate by Sindelarova et al. (2014). The model was run for 1.5 years (summer 2011-2012) in which the first half year was used as spin up and 2012 for analysis.

3 Theoretical work on isoprene di-HPCARP-RO₂-I
3.1 Kinetics of the di-HPCARP-RO₂ H-migration reactions
The di-HPCARP-RO₂ chemistry was studied in more detail compared to earlier published work (Wang et al., 2018; Møller et al., 2019), with more reliable kinetic methodologies, to obtain the isomer-specific rate coefficients, the phenomenological (bulk) rate coefficients, and the subsequent chemistry. Table 3 lists the rate coefficients at 300 K, and the Kooij expressions for the key reactions in the di-HPCARP-RO₂-I reaction system for the temperature range 200-400K. As expected from the higher energy barriers and entropic considerations, migration of non-aldehyde H-atoms is not competitive, nor is HO₂.
elimination. The fastest reactions are, for all di-HPCARP-RO₂-I intermediates, the hydroperoxide-H-scrambling across all peroxy sites. This result is anticipated, as fast scrambling has been known since several years. (Miyoshi, 2011; Jorgensen et al., 2016; Knap and Jorgensen, 2017; Møller et al., 2019; Praske et al., 2019) Given that these reactions outrun the next fastest reaction by over an order of magnitude, it can be assumed that the different di-HPCARP-RO₂-I isomers are in steady-state equilibrium, and can be considered as a unified pool of reactants in atmospheric models for the purpose of unimolecular reactivity. The 1,4-aldehyde-H-migration reaction is comparatively slow owing to its higher energy barrier (see Table S1, Fig. 3), and formation of the trihydroperoxide acyl radical tri-HPACYL will thus occur mostly by 1,5- and 1,6-H-migration in the di-HPCARP-RO₂-I reactant pool. Table 3 lists the rate coefficients of the elementary processes, but for atmospheric modelling the more relevant numbers are the bulk k(T) expressions that account for the H-scrambling and the combined flux across the 1,4-, 1,5-, and 1,6-H-migrations, listed at the bottom of the table. The stereo-specific rate coefficients are not all that different, and can be expressed within a factor of 2 to 3 as a single Kooij expression across the temperature range 200-400K as follows (see also Figure S1):

$$k(T) = 6.5 \times 10^{-5}T^{2.5} \exp \left( \frac{17000}{T} \right)$$  

(eq. 1)

The effective rate of acyl radical formation by aldehyde-H-migration, i.e. accounting for the rapid re-equilibration across the di-HPCARP-RO₂-I isomers and the different H-shift channels, is then of the order of 10 s⁻¹ at 300 K. For the (2R,3R) and (2S,3S) stereo-isomers, 63% of the reaction flux at 300K passes through the 1,5-aldehyde-H-migration, with 35% undergoing a 1,6-H-migration. For the (2R,3S) and (2S,3R) isomers, 1,6-aldehyde-H-migration constitute over 99% of the acyl radical formation. The estimated uncertainty on the rate coefficient is about a factor of 5, mostly due to the current use of non-conformer-specific tunneling.

3.2 Elimination of CO from tri-hydroperoxy acyl radicals

It has been assumed, in most models in the literature (Peeters et al., 2009; Peeters et al., 2014; Jenkin et al., 2015; Wennberg et al., 2018), that the tri-hydroperoxy acyl radical formed, tri-HPACYL, will eliminate CO, followed by OH-elimination and forming DHP-MVK (Fig. 3). All these models seem to be based ultimately on an estimate by Peeters et al. (2014) who predicted that the reaction would proceed via a 1,4-H-shift, forming tri-HPACYL intermediates with an internal energy of the order of 22-25 kcal mol⁻¹, rapidly losing CO with a barrier ≤7 kcal mol⁻¹. This work, however, shows that these predicates do not represent the chemistry accurately, and the fate of tri-HPACYL must be examined in more detail.

The lower energy barriers of the effective aldehyde-H-migration processes implies that the acyl radical tri-HPACYL is formed with a significantly lower energy content, 17-19 kcal mol⁻¹, than estimated by Peeters et al. (2014), reducing the likelihood of chemically activated decomposition. Note that the H-migration reactions are typically found to be in the high-pressure limit (Miyoshi, 2012; Peeters et al., 2014; Xing et al., 2018; Møller et al., 2019), and multi-step chemical activation does not contribute significantly. Furthermore, we found that the CO elimination barrier in tri-HPACYL-I, ~8 kcal mol⁻¹ (see Table S1) is higher than estimated by Peeters et al. (≤7 kcal mol⁻¹), further hampering prompt decomposition. With only ~10 kcal mol⁻¹ excess internal energy, as opposed to 15-18 kcal mol⁻¹ as proposed by Peeters et al. (2014), a significant fraction of the nascent acyl radicals tri-HPACYL could be thermalized, and the resulting slower decomposition process could potentially allow for O₂ addition on the acyl radical site. However, based on MC-TST calculations, incorporating all conformers, we found that thermal decomposition of the tri-HPACYL-I acyl radicals is still sufficiently fast to dominate over O₂ addition, i.e. even when assuming a Boltzmann energy distribution, the predicted rate coefficient of ~2×10⁶ s⁻¹ at 300 K (see Table 3), is significantly higher than the effective O₂-addition rate for acyl radicals, experimentally measured at ≤~3×10⁵ s⁻¹ in atmospheric conditions (Sehested et al., 1998; Blitz et al., 2002; Park et al., 2004; Baulch et al., 2005; Atkinson et al., 2006; Carr et al., 2011). One could counter that the presence of –OOH groups might stabilize the acylperoxy radicals formed in the O₂ addition (e.g. by H-bonding), thus increasing the addition rate coefficient above those reported for the
smaller acyl radicals in the literature. Sample calculations on a smaller proxy substituted acetyl radicals (see Table S1 in the supporting information) found no evidence that a hydroperoxide group interacts with the oxygen atom moiety in a way that reduces the entrance energy barrier (thus increasing the capture rate coefficient), or stabilizes the adduct (thus reducing redissociation). From these results, we conclude that CO elimination will be the dominant fate at 300 K for the tri-HPACYL-I acyl radicals formed in the aldehyde-H-migration, with O$_3$ addition a minor channel. Whether CO elimination occurs promptly or in a thermal reaction is then a moot issue.

Based on a preliminary version of our results (Novelli et al., 2018a), Peeters and coworkers (Müller et al., 2019) now suggest that tri-HPACYL-II acyl radicals (differing from the case I acyl radical by the position of the methyl group) would not eliminate CO, due to a higher CO elimination energy barrier (Méreau et al., 2001) compared to the tri-HPACYL-I acyl radicals we characterized explicitly above. We have as yet been unable to dedicate the significant computational resources for an explicit study of di-HPCARP-RO$_2$-II and tri-HPACYL-II, so this issue cannot be resolved at this time. We can, however, estimate the tri-HPACYL-II thermal CO elimination rate by assuming that case I and case II reactions are entropically similar, and increasing the barrier for tri-HPACYL-II by the difference between a tertiary (case I) and secondary (case II) product radical as calculated by Méreau et al. The obtained rate coefficients at 298 K (~6×10$^8$ s$^{-1}$ for the (R,R) and (S,S) stereoisomers; ~1×10$^7$ s$^{-1}$ for the (R,S) and (S,R) stereoisomers) remain competitive against O$_2$ addition. Though these channels appear no longer truly dominant thermally, CO elimination from tri-HPACYL-II can clearly not be discounted with any degree of confidence based on such estimates, especially as any chemical activation afforded by the preceding H-migration would further shift the subtle competition towards higher CO yields. Quantifying this yield theoretically would be a very committed effort, as it requires explicit calculation of the full conformation space, with chemical activation depending on the energy-specific state density, conformer-specific tunneling, and collisional energy transfer across the thousands of di-HPCARP-RO$_2$-II and tri-HPACYL-II conformers; this is beyond the scope of the current paper.

3.3 Comparison to literature theoretical work

Two earlier studies examined the di-HPCARP-RO$_2$ chemistry theoretically. A first study by Wang et al. (2018) was based on a partial characterization of the conformational space. While some of the provided rate coefficients are comparable to our values, these predictions carry a larger uncertainty due to the limited number of conformers examined, and differences of over an order of magnitude are found compared to our predictions, leading to qualitative changes in the predicted fate of the intermediates. The methodology of the second study, by Möller et al. (2019), is more comparable to that used in our work, with the main difference being the methodology used to screen the conformational space and select the conformers included in the rate calculations. As discussed in more detail in the supporting information, we find that this study only incorporates half or less of the population in the kinetic predictions, relying on cancellation of error to a far greater extent than our more rigorous population description. Though for most reactions the differences in predicted rate coefficient at 298K are small, we find some values that differ by about an order of magnitude. We surmise this is mostly due to the impact of the population truncation in the Möller et al. (2019) study compared to the full population used our work (e.g. 11 conformers versus ~1500 conformers included), making our predictions more robust in this respect. Improved conformer screening methodologies, balancing completeness and accuracy against computational cost, will help to converge the results of the two similar methodologies; this is discussed briefly in the supporting information.

4 Comparison of measured trace gases with model calculations

Figure 4 shows the evolution of trace gas concentrations for an experiment with three separate injections of isoprene and where the NO concentration was below 0.15 ppbv for the entire duration of the experiment. For all three isoprene injections it is possible to observe large discrepancies between measured trace gases and model calculations when using the MCM
sequent degradation products of ISOPOOHs which includes epoxides (St. Clair et al., 2016). The similarity between the two models, despite M1 including a factor of 3 faster 1,6-H shift, is due to the different distribution of the isoprene-RO₂ conformers. Specifically, a much smaller fraction of Z-δ-RO₂ radicals M1, ~0.004, is formed compared with 0.015 for the MCM 331 model, thus reducing the contribution of the 1,6-H shift. Although a larger fraction of Z-δ-RO₂ radicals are formed within the MCM 331 mode, the slow 1,6-H shift used also results in an underestimation of the OH radical concentrations. For both models, for conditions where the NO concentration is lower than 0.2 ppbv, i.e. where isomerization reactions should become more relevant, ~ 30% of the total loss of isoprene-RO₂ conformers (weighted by their abundance) occurs via isomerization reactions. Compared with a simulation without isomerization reactions (M0) there is an improvement in the reproduction of the measured trace gases, but still the importance of the isomerisation reactions is underestimated. The best agreement between measured and modelled trace gases was achieved when the equilibrium reactions between the isoprene-RO₂ conformers providing a larger fraction of Z-δ-RO₂ radicals in combination together with the faster 1,6-H shift are used (M2, Fig. 5). A ratio of measured to modelled data of 0.97 ± 0.10 and 0.98 ± 0.07 for the OH radicals and of the sum of the oxidation products MVK, MACR, and ISOPOOHs, respectively, are found. Also the increase in the carbon monoxide (CO) concentration, of which nearly one quarter is explained by the CO elimination of the di-HPCARP-RO₂, is well captured by the model calculations (ratio of measured to modelled data of 0.98 ± 0.05). In comparison with MCM 331 and M1 model runs, ~ 50% of the total loss of isoprene-RO₂ conforms proceeds via isomerization reactions.

The yields of the chemical compounds formed following the 1,6-H shift of Z-δ-RO₂ radicals carry a large uncertainty. When the isomerisation reactions for isoprene-RO₂ were firstly calculated, a unity yield for HPALD was suggested (Peeters et al., 2009). A subsequent experimental study by (Berndt, 2012) showed a yield of HPALD of less than 0.5. The updated LIM1 adopted a yield for HPALD of 0.5 with formation of the di-HPCARP-RO₂ making up the remainder. Very recently two new experimental studies exploring the 1,6-H shift of Z-δ-RO₂ radicals were conducted with contrasting results. Teng et al. (2017) directly measured HPALD providing a yield of 0.25 with a factor of 2 uncertainty. There was no detection of di-HPCARP-RO₂ compounds; the authors suggest that these might be too reactive and lost on walls before reaching the instrument inlet. The results of this study were incorporated with some modifications in the isoprene Caltech mechanism (Wennberg et al., 2018) with a recommended yield of 0.4 for HPALD and 0.6 for di-HPCARP-RO₂. The very recent experimental study by Berndt et al. (2019), puts a lower limit for the HPALD yield at 0.75 and for the di-HPCARP-RO₂ to 0.01 from the direct measurement of both compounds. These results were subsequently rationalized by Müller et al. (2019), revising the earlier LIM1 HPALD yield of 0.5 based on new theoretical calculations and a mechanistic analysis, deriving a much larger yield of 74% for HPALD, 16% epoxides, and 10% products from di-HPCARP-RO₂ chemistry.

From the above, we find that two HPALD yields are currently described in the literature: one in the range of 0.4 to 0.5, used within the Caltech mechanisms and the MCM v3.3.1, and the other exceeding 0.75, based on the most recent literature. To quantify the impact of the uncertainty on the HPALD yield, a sensitivity model run was performed by changing the yield of HPALD from 0.4 to 0.75, with the yield of di-HPCARP-RO₂ set to 0.25 (M3, Table 2). As it can be seen in figure 5, increasing the yield of HPALD by almost a factor of two does not have a large impact on model reproduction of the measured trace gases. The degradation of HPALD and di-HPCARP-RO₂ are both followed by formation of OH radicals and CO. On the time scale of the experiments in the chamber, it is not possible to distinguish between the relatively slow formation of OH radicals from photolysis of HPALD (~ 1h) from their formation from the fast aldehyde-H shift of the di-HPCARP-RO₂ (~ 0.1 s). This is not entirely true for the CO where the measurement indicates a faster formation rate, better agreeing with the model including a larger yield towards di-HPCARP-RO₂ (M2 model). A better agreement is however...
observed between measured and modelled HO$_2^\ast$ radicals for the HPALD 0.75 model, although this remains within the uncertainty of the measurement, due to the formation of HO$_2$ radicals together with HPALD following the 1,6-H shift.

For experiments transitioning from high (1.5 ppbv) to low (0.2 ppbv) NO (Fig. 6) all the models are able to reproduce (within 10%) the measured trace gases and the OH reactivity for the first injection of isoprene when the NO concentration is above 0.5 ppbv. At this NO level, the OH production is mainly controlled by the reaction between HO$_2$ radicals and NO therefore the impact of the isomerization reactions to the OH radical production is marginal. As soon as the concentration of NO decreases below this threshold, larger discrepancies between the model calculations and the measured trace gases can be observed.

One additional model run (not shown) was performed by re-implementing the original LIM1 within the MCM 331 model (Table 2), which includes a factor 5 slower equilibrium reactions between the isoprene-RO$_2$ conformers and a factor of 5 faster 1,6-H shift. Despite the large reduction in the equilibrium reactions between the isoprene-RO$_2$ conformers the LIM1 model run can reproduce the measured data as well as M2 and M3 for all concentration of NO investigated in this study as a change of only 5% in the fraction of Z-δ- RO$_2$ radicals formed is observed.

5 Modelled contributions to the measured OH radical regeneration efficiency

The extensive range of NO concentrations in the experimental studies, reaching up to 2.0 ppbv, allowed exploring the ability of the models to reproduce the measured data and to quantify the efficiency of regeneration of OH radicals across a wide range of atmospheric conditions (Fig. 7) by drastically changing the competition between the isomerization reactions of RO$_2$ and the RO$_2$ + NO reactions.

The efficiency of the OH regeneration, noted RE henceforward, is defined as the number of OH radicals that are produced after one OH radical has reacted with isoprene. It is calculated as the ratio of the OH regeneration rate $R$ and the OH loss rate $L$. The modelled $R$ quantifies the OH production via radical chain reactions (HO$_2$+NO, HO$_2$+O, and isomerization of isoprene-RO$_2$) and the photolysis of HPALD as produced in the isomerization reactions. It also includes the OH regenerated from the direct products of HPALD photolysis and aldehyde-H shift of di-HPCARP-RO$_2$ (Table S6). $L$ represents the OH loss by reaction with isoprene and its products. As such the model values for the OH regeneration efficiency represent a lower limit. The measured RE is obtained from the difference between the total OH loss rate and the primary OH production rate (ozone and nitrous acid photolysis) divided by the total OH loss rate.

$$RE = \frac{R}{L} = \frac{[\text{HO}_2] - ([\text{OH}] + [\text{O}])}{[\text{HO}_2] + [\text{H}_2\text{O}] + y \times \text{HONO} + \text{HONO}^\ast}$$

Eq. 2

Here most values used ([OH], [O], $[\text{O}_3]$), $[\text{H}_2\text{O}]$, jO($^1\text{D}$), jHONO) are experimentally measured quantities. Only the HONO concentration was not measured, but taken from the model. $y$ is the fraction of O($^1\text{D}$) reacting with water vapour, multiplied by the OH yield of the O($^1\text{D}$) + H$_2$O reaction.

Good agreement is found between measured and modelled OH regeneration efficiency at all values of NO within the uncertainty of the measurements when using either M2 or M3 (Fig. 7) suggesting that all relevant OH production pathways are included.

In environments influenced by anthropogenic emissions, with NO values higher than 0.2 ppbv, 75% of OH radicals are regenerated by the reaction of HO$_2$ radicals with NO. In contrast, at the lowest NO values, representative of rainforest conditions, only 10% of OH radicals reacting with isoprene are reformed by HO$_2$ + NO reaction. This decrease in the OH RE from radical reactions with NO is partly compensated by an increased contribution of regeneration from RO$_2$ isomerization reactions, such that the total OH RE is still approximately 0.5 at the lowest NO concentration investigated. Though contributions of isomerization reactions to the OH RE diminish with increasing NO concentrations, their reaction rate coefficients are high enough to still constitute a competitive loss path for Z-δ-RO$_2$ radicals even at 2 ppbv of NO, with 10% of the OH radicals consumed still regenerated by RO$_2$ isomerization reactions. The differences in contribution of the
isomerization reactions to the OH RE found for [NO] above 0.3 ppbv are mainly due to differences in ambient temperature impacting the isomerisation rate coefficient. Although no experiments are available for levels of NO lower than 0.15 ppbv, a model simulation for the OH RE for up to 0.005 ppbv of NO indicates that the value of OH RE remains constant at around 0.5.

Among the isomerisation reactions, the 1,5-H shift contributes less than 5% to the OH radical regeneration with the large majority of the OH radical regenerated by the products following the 1,6-H shift. This is not surprising as the 1,5-H shift rate coefficient is ~2 order of magnitude slower than the 1,6-H shift. A more detailed analysis of the contribution of HPALD vs. di-HPCARP-RO2 and their products to the OH regeneration is hindered by the uncertainty on their relative yield. Anyway, it is interesting to see that when a larger yield for di-HPCARP-RO2 (0.6, M2) is applied, the largest fraction of OH radical is regenerated via the aldehyde-H shift of di-HPCARP-RO2 and its direct di-hydroperoxy carbonyl products (~70%, Fig. 7a). When using a yield of HPALD of 0.75 (Fig. 7b) photolysis of HPALD and following direct products increase their contribution from 12% to 60%. Still, despite a much smaller yield (0.25) compared to HPALD, di-HPCARP-RO2 and direct products contribute up to 30% to the OH radical regeneration due to the fast aldehyde-H shift.

The magnitude of the OH RE observed in this study, however, remains much lower than anticipated from OH concentration measurements in field campaigns under similar conditions, requiring an OH RE of nearly one to reproduce the observations (Rothr et al., 2014). The reason for the discrepancy is still not fully understood. On the one hand, the experiments in the chamber refer only to isoprene chemistry while the field studies, although performed in areas with large emissions of isoprene, include several different organic compounds which could contribute to the OH concentration. On the other hand, it could be that part of the measured OH radical concentrations in field campaigns was due to an interference. However, the same design LIF instrument as used in this study was deployed in several field campaigns in China with the addition of a chemical titration device to separate ambient OH radicals from interferences, showing at max an interference of 10% during daytime (Tan et al., 2017; Tan et al., 2018; Tan et al., 2019).

6 Global impact

Results from the simulation chamber experiments are used to investigate the impact on the global distribution of the OH regeneration efficiency due to either radical reactions with NO or isomerization reactions by implementing the detailed isoprene chemistry as derived in this study within M2 model (Table S7). The global atmospheric chemistry model EMAC (Jöckel et al., 2010) was applied including a modified version of the Mainz Organic Mechanism (MOM) (Sander et al., 2019) representing an advanced isoprene oxidation mechanism with a complexity comparable to the MCM.

In regions with low NO concentrations, the OH regeneration by HO2 + NO is suppressed but compensated by the OH regenerated from RO2 radical isomerization reactions. These reactions globally have the largest impact in the tropics due to high isoprene concentrations and high temperatures (Peeters et al., 2014). The inclusion of OH regeneration routes gives an OH regeneration efficiency that is at least 60% globally over all land masses covered with vegetation (Fig. 8). As a consequence, in areas where isoprene is the most important reactant for the OH radicals, their concentration at the surface is enhanced by more than a factor of 3 compared to model predictions neglecting RO2 isomerization reactions (Fig. S3). Several studies showing the global impact of isomerisation reactions were performed (Bates and Jacob, 2019; Möller et al., 2019; Müller et al., 2019) all showing, similarly to the results within this study, an enhanced concentration of OH radical in particular at the tropics where high concentrations of isoprene and high temperatures can be observed. A detailed comparison between the different models is not straightforward as they all contain different sets of reactions and are performed with different model parameters (e.g. NOx and isoprene emissions). The three published studies are based on the isoprene Caltech mechanism (Table 2) which underestimate the measured OH radical concentration from this study due to a low yield of formation of the Z-δ-RO2 radical (Fig. 4). In addition, the model used in the study by Müller et al. (2019) includes a large
yields towards HPALD (0.75) as suggested from the study by Berndt et al. (2019) which is different from the other models. Thus one additional global model test run within this work was performed with a yield of HPALD of 0.75 and of di-
HPCARP-RO₂ of 0.25 to verify if the alternative branching ratio would result in significant differences. However, no change in the expected OH radical concentration was observed, in agreement with the model runs for the chamber experiments. However, for small oxygenated VOC (OVCOs) like formaldehyde, formic acid, methanol, glyoxal, methyl glyoxal, hydroxyacetone and peroxy acetyl nitrate, a change of up to 30% was found at ground level in isoprene dominated regions. In the same regions, the change of CO concentration was less than 5%.

7. Remaining uncertainties

7.1 Yield of di-HPCARP-RO₂ versus HPALD

The LIM1 mechanism by Peeters et al. (2014) proposed a low ratio of HPALD to di-HPCARP-RO₂ formation, in agreement with the Teng et al. (2017) experiments. Since then, the experimental study of Berndt et al. (2019) found a much higher HPALD yield, with small contributions only of di-HPCARP-RO₂. The HPALD to di-HPCARP-RO₂ ratio is governed by the chemistry of the HOO-hydroxyallyl radicals formed after the dominant 1,6-H-shift in the Z-δ-1,4-ISOPOO radicals (Fig. S4). Müller et al. (2019) rationalize the high HPALD yield by Berndt et al. (2019) based mostly on the stereoselectivity of the Z-δ-1,4-ISOPOO H-migration and the subsequent O₂ addition, where new theoretical work (Müller et al., 2019) showed a comparatively high barrier for internal rotation of the HOCH₂-C=O moiety in the HOO-hydroxy-allyl radicals, due to the partial double bond. The nascent stereo-specificity (ZE versus ZZ, see figure S4 box A) then remains essentially unchanged throughout the subsequent chemistry of the HOO-hydroxy-allyl radicals, this contrary to earlier assumptions in Peeters et al. (2014). HPALD can be formed rapidly from both stereoisomers by O₂ addition on the HOCH₂ radical carbon (figure S4 box D), followed by fast HO₂ elimination. Di-HPCARP-RO₂ is formed only from the ZZ’-HOO-hydroxy-allyl radicals (figure S4 box E), as the fast enol H-migration requires the geometric proximity of the –OH and –OO groups.

Müller et al. (2019) suggest, based on the stereo-specific chemistry, that the E'-enolperoxy radicals formed from O₂ addition on C₃ of the Z,E'-HOOC-allyl radicals (figure S4 box B) will solely undergo redissociation, and subsequent re-addition of O₂ on either of the radical sites of the Z,E'-HOOC-allyl radical, until "indirect" (Müller et al., 2019) HPALD is formed. Thus, the high yield of HPALD is explained based on its formation from all Z,E'-HOOC-hydroxy-allyl radicals, and part of the Z,Z'-HOOC-hydroxy-allyl radicals, whereas diHPCARP-RO₂ is only formed from part of the Z,Z' radicals.

While this mechanism can indeed lead to numerical agreement with the Berndt et al. yields, the argumentation is not based on actual quantitative theoretical work on each reaction step, and may thus be unable to discriminate between alternative mechanisms or yields in this subtle, complex chemistry; in this paragraph, we briefly examine a few aspects of the mechanism that warrant further investigation. Different yields would be obtained if the stereo-specific yields (determined based on minimum-energy pathways) are affected by non-statistical dynamics induces by chemical activation or post-barrier energy release. Likewise, the site-specificity of O₂ addition, based on radical spin densities as a first-order approximation rather than on characterizations of the addition TS, carries a large uncertainty. The rate of redissociation of the HOOC-enolperoxy adducts are suggested by Müller et al. (2019) to be very similar to the initial Z- and E-δ-OH-peroxy radicals from isoprene. However, H-scrambling between the –OOH and –OO groups (see figure S4 box C) is expected to be orders of magnitude faster (Miyoshi, 2011), leading to an equilibrated population where the terminal peroxy radical has no access to a rapid redissociation channel, leading to a lower overall phenomenological rate coefficients for redissociation. Without characterization of the impact of the terminal peroxy radical in the bulk rate coefficient it is difficult to assess whether sufficient redissociation/re-addition events can occur on the experimental timescale to ensure complete reconversion to HPALD. H-scrambling also allows access to enol reaction pathways ignored by Müller et al. (2019). We examined ring
closure and H-migration reactions for the \( \text{HOCH}=\text{C}(\text{CH}_3)\text{CH},\text{CH}_2\text{OO}^- \) proxy molecule, predicting a ring closure rate coefficient of \( 1\times10^3 \) s\(^{-1} \) (see figure S2) that dominates redissociation. The impact of the additional \( -\text{OOH} \) group in \( \text{HOCH}=\text{C}(\text{CH}_3)\text{CH}(\text{OOH})\text{CH},\text{CH}_2\text{OO}^- \) on the ring closure rate is hard to estimate without computational work, and these proxy results may thus not be applicable to the isoprene chemistry. At the very least, one expects significant differences between the case I and II rate coefficients. However, even a slower rate of ring closure would disrupt the reaction flow as laid out by Müller et al. (2019), and affect the predicted HPALD yield.

At this moment, there are contradicting experimental data on the HPALD vrs. diHPCARP-RO\(_2\) yield (Berndt, 2012; Teng et al., 2017; Berndt et al., 2019) where the authors acknowledge (Berndt et al., 2019) that the reasons behind the disagreement are not clear. The earlier mechanism by Peeters et al. (2014) was used to rationalize the low HPALD observations of Teng et al. (2017) but did not appropriately account for stereospecificity. The Müller et al. mechanism is compatible with the high HPALD yield by Berndt et al. (2019), but is based partially on mechanistic argumentations that may not be compatible with new and existing quantitative theoretical work, despite the excellent apparent numerical agreement with experiment. We must stress that none of the aforementioned considerations listed here on the Müller et al. (2019) reaction scheme have the strength by themselves to invalidate the proposed mechanism, and are mostly indications that caveats apply when implementing this scheme in chemical models. In this complex chemistry, with subtle competition between many channels, we feel there is as yet not enough quantitative theoretical work to claim theory-based support for either experimental yield.

It is therefore not possible from this study to unequivocally determine the correct yield of HPALD versus di-HPCARP-RO\(_2\), highlighting the need for further studies measuring their degradation products to pinpoint their yield of formation.

7.2 Fate of HPALD and di-hydroperoxy carbonyl compounds

HPALD is assumed to photolyse at a relative fast rate (~ 2 x 10\(^{-4}\) s\(^{-1}\) at 30° solar zenith angle) producing CO and OH radicals among other trace gases (Wolfe et al., 2012). In addition to photolysis it will react with OH radicals, where MCM v3.3.1 includes 3 site-specific reactions with a total rate of 5.2 x 10\(^{-11}\) cm\(^3\) s\(^{-1}\) (Wolfe et al., 2012). This reaction can compete with photolysis when OH radical concentrations reach 4 x 10\(^{4}\) cm\(^{-3}\) ( ~ 2 x 10\(^{-4}\) s\(^{-1}\)), but the products distribution has not been yet measured. The fate of the di-hydroperoxy carbonyl compounds (DHPMEK and DHPMPAL in the MCM v3.3.1 and MVK3OOH4OOH and MACR2OOH3OOH in the Caltech mechanisms), formed from the aldehyde shift of the di-HPCARP-RO\(_2\), is more uncertain and within the M2 model they are predicted in relatively large concentrations (~ 10\(^{10}\) cm\(^{-3}\)).

In the MCM v3.3.1 model, the two di-hydroperoxy carbonyl compound degrade either by reacting with OH radicals, regenerating the OH radical, or by photolysis. Only two of the possible five site-specific reactions (Jenkin et al., 2018) are included, with a total rate coefficient in the order of ~ 3.3 x 10\(^{-11}\) cm\(^3\) s\(^{-1}\). Within the MCM v3.3.1 this is the main loss path for these species as photolysis is slower than 1.3 x 10\(^{-4}\) s\(^{-1}\) and does not compete with the pseudo first order rate coefficient for ambient concentration of OH radicals. A recent theoretical study by Liu et al. (2018) suggests, based on calculation for the proxy molecule 2-hydroperoxympropanal, a much faster rate for the photolysis (~ 1 to 5 x 10\(^{-4}\) s\(^{-1}\) at 30° SZA) with a yield of ~ 20% OH radicals. At ambient concentration of OH radicals lower than 4 x 10\(^{6}\) cm\(^{-3}\), photolysis would then become the dominant path. The Caltech mechanism, based on the Liu et al. study, then has photolysis as the only degradation path for the di-hydroperoxy carbonyl compounds. As long as the yield of HPALD remains uncertain, it is difficult to assess the importance of these reactions. This underlines the need for additional studies on the degradation of di-hydroperoxy carbonyl compounds and HPALD as they can have, locally, a large impact on the type of oxygenated products obtained.

8 Summary and Conclusion

Photooxidation experiments on isoprene, the globally dominant biogenic volatile organic compound emitted, were performed in the atmospheric simulation chamber SAPHIR for a range of NO mixing ratios to explore the importance
isomerisation reaction have on OH radical regeneration. Measurements of OH reactivity, OH, HO₂⁺, and RO₂⁺ radicals concentrations, and other important trace gases, were compared to results from different model calculations all based on a state of the art chemical mechanistic model (MCM v3.3.1) (Jenkin et al., 2015).

It was found that the MCM v3.3.1 mechanism for the isoprene degradation initiated by OH radicals is not able to reproduce the measured trace gas concentrations in the experiments despite the inclusion of isomerisation reaction for isoprene-RO₂ following the LIM1 mechanism for NO mixing ratio < 0.2 ppbv. Large discrepancies are observed in particular for OH radicals, with a ratio of modelled to measured OH of 0.7 ± 0.07, and for the sum of MVK, MACR and ISOPOOHs (all isomers) with a model to measurement ratio of 0.6 ± 0.08.

Summarizing the theoretical analysis, we find that the main fate of di-HPCARP-RO₂-I is migration of the aldehyde H-atom followed by rapid CO loss, leading to an unstable αOOH alkyl radical that will eliminate an OH radical.(Vereecken et al., 2004) forming DHP-MVK. The rate limiting reaction is the aldehyde H-migration, with an effective rate coefficient k(300K) = ~10⁻¹⁰ s⁻¹. Alternative reaction channels are found to be uncompetitive. The mechanism leading to these results is significantly more complex than originally proposed (Peeters et al., 2014), with a rate coefficient significantly higher than the original 0.1 s⁻¹ estimate and possibly contribution of thermalized reactions of the tri-HPACYL intermediate. Though, the nett product formation remains identical as that incorporated in the mechanisms by Peeters et al. (2014) and Wennberg et al. (2018), and implemented in e.g. the Master Chemical Mechanism v3.3.1.(Jenkin et al., 2015). Extrapolating these results to the di-HPCARP-RO₂-II radicals by accounting for the expected barrier difference due to the different position of the methyl group, we find there is competition between CO elimination versus O₂ addition in the tri-HPACYL-II radicals formed, especially when assuming absence of chemical activation.

The kinetic aspects controlling the impact of the 1,6-H shift of Z-δ-RO₂ on the regeneration of OH radicals and the production of oxygenated products were carefully checked based on what available in the literature (Table 2). It was found that the best agreement between measured and modelled trace gases is observed when up to 50% of the isoprene-RO₂ conformers (weighted by their abundance) isomerized. This is achieved when including within the MCM v 3.3.1 a faster rate coefficient (3.6 s⁻¹ and 0.4 s⁻¹ at 298 K for OH addition on C4 and C1, respectively) for the 1,6-H shift of the Z-δ-RO₂ radical based on a recent experimental study (Teng et al., 2017). Large uncertainties remain regarding the relative yield of HPALD and di-HPCARP-RO₂ following the 1,6-H shift of Z-δ-RO₂ radicals. Within this study, no meaningful differences between the results of different model calculations could be observed when the yield of HPALD was varied from 0.4 to 0.75. Both HPALD and di-HPCARP-RO₂ produce OH radicals and CO in a relative short time scale (less than 1h). Therefore as long as the 1,6-H shift is fast and there are sufficient Z-δ-RO₂ radicals undergoing isomerisation, the measurements from this study are not sensitive to the HPALD to di-HPCARP-RO₂ ratio.

A detailed study of the path contribution to the OH radical regeneration highlights how, for NO mixing ration < 0.2 ppbv, ~50% of the OH radical is regenerated from the products following the 1,6-H shift: photolysis of HPALD and aldehyde shift of the di-HPCARP-RO₂. These processes help maintaining the OH radical regeneration efficiency up to 0.5 in environment with low NO mixing ratios where the regeneration via reaction of HO₂ with NO becomes less important. For environment where the NO concentration is higher, the regeneration via HO₂ plus NO dominates (> 75%) and even models not including isomerisation reactions are able to reproduce the measured trace gases. The observed OH radical regeneration efficiency in this chamber study at low NO mixing ratios study is, however, lower than what is observed/needed in the field to explain the measured OH radical concentrations in isoprene-dominated environments (Rohrer et al., 2014).

A semi-explicit global model which includes the chemistry highlighted in this study shows how isomerisation helps maintaining an OH regeneration efficiency up to 0.6 globally. In the amazon at ground level, the inclusion of isomerisation reactions increases the OH radical concentrations up to a factor of three although this has no relevant impact on the global budget of methane or CO.
Code and data availability
The data of the experiments in the SAPHIR chamber used in this work are available on the EUROCHAMP data home page (https://data.eurochamp.org/, last access: 1 October 2019, Eurochamp, 2019).

Author contributions
HF and AH designed the experiments. AN analysed the data, performed the box model simulations and wrote the paper together with HF. LV did the theoretical calculations. SR and DT performed the global model simulations. All other coauthors participated in data collection, and experiment operations, and all coauthors participated to paper discussion.

Competing interests
The authors declare that they have no conflict of interest.

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References


Table 1. Rate coefficients for the addition of O$_2$ to OH-isoprene adducts, and for re-dissociation of isoprene-RO$_2$ (Fig. 1). The rate coefficients for the oxygen additions (kf) are in cm$^3$ s$^{-1}$ and are typically temperature independent or provided at 298.15 K. The rate coefficient for the re-dissociations (kr) are in s$^{-1}$ and provided at 298.15 K. The temperature dependent rate coefficients are given in table S8.

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<th>Caltech$^c$</th>
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<td>kf2</td>
<td>0.6×10$^{-12}$</td>
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<td>0.8×10$^{-12}$</td>
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<td>kf3</td>
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<td>0.1×10$^{-12}$</td>
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<td>0.7×10$^{-12}$</td>
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<td>kf7</td>
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<td>0.7×10$^{-12}$</td>
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<td>kf8</td>
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<td>0.5×10$^{-12}$</td>
<td>0.5×10$^{-12}$</td>
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<tr>
<td>kr1</td>
<td>4.0</td>
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<td>18</td>
</tr>
<tr>
<td>kr2</td>
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<td>2.0</td>
<td>1.8</td>
</tr>
<tr>
<td>kr3</td>
<td>0.05</td>
<td>0.3</td>
<td>0.3</td>
</tr>
<tr>
<td>kr4</td>
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<td>11</td>
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<td>0.2</td>
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<tr>
<td>kr7</td>
<td>0.03</td>
<td>0.2</td>
<td>0.3</td>
</tr>
<tr>
<td>kr8</td>
<td>0.1</td>
<td>0.6</td>
<td>4.3</td>
</tr>
</tbody>
</table>

a. Peeters et al. (2014)  
c. Wennberg et al. (2018)
Table 2: Summary of the relevant differences for assessing the 1,6-H shift impact between available chemical models. All models versions within this work are based on the MCM v 3.3.1 (accounting for chamber properties) with only specific rates and yields included in the table adapted from different studies. They also include the follow-up chemistry of di-HPCARP-RO2 as theoretically investigated within this study.

<table>
<thead>
<tr>
<th>Mechanism</th>
<th>k 1,6-H shift (298K)(^a)</th>
<th>Yield</th>
<th>R + O(_2) ⇔ RO(_2)</th>
</tr>
</thead>
<tbody>
<tr>
<td>LIM1</td>
<td>0.5 (C1), 5.8 (C4)</td>
<td>0.5 : 0.5</td>
<td>LIM1</td>
</tr>
<tr>
<td>MCM 331</td>
<td>0.1 (C1), 1.2 (C4)(^b)</td>
<td>0.5 : 0.5</td>
<td>LIM1 × 5</td>
</tr>
<tr>
<td>Caltech</td>
<td>0.4 (C1), 3.6 (C4)</td>
<td>0.4 : 0.6</td>
<td>Values from Teng et al. (2017)</td>
</tr>
</tbody>
</table>

This work:
- M0: No Isomerization
- M1: Caltech → Caltech → Caltech
- M2: Caltech → Caltech → MCM 331
- M3: Caltech → 0.75 : 0.25\(^c\) → MCM 331

\(^a\), \(^b\), \(^c\): same rate coefficient as in LIM1 diminished by a factor of 5. 
\(^d\): adapted from Berndt et al. (2019)
Table 3: Stereo-specific rate coefficients (s⁻¹) at 300K for the relevant reactions of di-HPCARP-RO₂-I. The temperature dependence is given as a Kooij expression: \( k(T) = A \cdot T^n \cdot \exp(-E_a/T) \) for the temperature range 200-400 K. The effective bulk rates of reactions are also given, accounting for hydroperoxide H-atom scrambling and aldehyde H-migration across all channels. The bottom expression averages the stereo-specific rate coefficients, for use in simplified models.

<table>
<thead>
<tr>
<th>Reaction</th>
<th>( k(300K) )</th>
<th>( A (s^{-1}) )</th>
<th>( n )</th>
<th>( E_a (K) )</th>
</tr>
</thead>
<tbody>
<tr>
<td>(2R,3R)-2-Me-3,4-diOOH-butanal-2-peroxyl (A)</td>
<td></td>
<td></td>
<td></td>
<td></td>
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<tr>
<td>1,4-aldehyde-H-migration to D</td>
<td>1.15×10⁹</td>
<td>1.21×10⁻⁸³</td>
<td>30.69</td>
<td>-4811</td>
</tr>
<tr>
<td>1,4-α-OOOH-H-migration</td>
<td>1.18×10⁸</td>
<td>4.59×10⁻⁵²</td>
<td>28.94</td>
<td>-3276</td>
</tr>
<tr>
<td>1,5-α-OOOH-H-migration</td>
<td>4.88×10⁴</td>
<td>2.87×10⁻¹⁸</td>
<td>15.11</td>
<td>3586</td>
</tr>
<tr>
<td>1,6-OOOH-migration to B</td>
<td>1.84×10⁴</td>
<td>7.26×10⁻⁸⁰</td>
<td>15.39</td>
<td>-3656</td>
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<tr>
<td>1,7-OOOH-migration to C</td>
<td>6.29×10⁴</td>
<td>2.75×10⁻⁵¹</td>
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<tr>
<td>HO₂ elimination</td>
<td>≤ 1×10⁻⁷</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>(2R,3R)-2-Me-2,4-diOOH-butanal-3-peroxyl (B)</td>
<td></td>
<td></td>
<td></td>
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<tr>
<td>1,5-aldehyde-H-migration to D</td>
<td>3.34×10⁹</td>
<td>1.12×10⁻⁶⁷</td>
<td>25.15</td>
<td>-4273</td>
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<tr>
<td>1,6-OOOH-migration to A</td>
<td>2.97×10⁴</td>
<td>1.36×10⁻³⁸</td>
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<tr>
<td>1,6-OOOH-migration to C</td>
<td>1.88×10⁴</td>
<td>1.17×10⁻³⁴</td>
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<td>(2R,3R)-2-Me-2,3-diOOH-butanal-4-peroxyl (C)</td>
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<tr>
<td>1,6-aldehyde-H-migration to D</td>
<td>2.28×10⁹</td>
<td>3.57×10⁻⁶¹</td>
<td>16.93</td>
<td>788</td>
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<tr>
<td>1,7-OOOH-migration to A</td>
<td>2.35×10⁴</td>
<td>3.06×10⁻⁴³</td>
<td>13.03</td>
<td>-2486</td>
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<tr>
<td>1,6-OOOH-migration to C</td>
<td>1.27×10⁴</td>
<td>1.35×10⁻²²</td>
<td>9.96</td>
<td>-906</td>
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<td>(2R,3S)-2-Me-3,4-diOOH-butanal-2-peroxyl (A')</td>
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<tr>
<td>1,4-aldehyde-H-migration to D'</td>
<td>3.71×10⁹</td>
<td>7.07×10⁻⁶⁵</td>
<td>27.88</td>
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<tr>
<td>1,4-α-OOOH-H-migration</td>
<td>8.44×10⁴</td>
<td>2.52×10⁻⁶⁴</td>
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<td>-18.17</td>
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<tr>
<td>1,5-α-OOOH-H-migration</td>
<td>5.02×10⁵</td>
<td>8.69×10⁻⁶⁷</td>
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<td>1,6-OOOH-migration to B'</td>
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<td>1,7-OOOH-migration to C'</td>
<td>6.20×10⁴</td>
<td>4.02×10⁻⁶⁰</td>
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<tr>
<td>HO₂ elimination</td>
<td>≤ 1×10⁻⁷</td>
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<tr>
<td>(2R,3S)-2-Me-2,4-diOOH-butanal-3-peroxyl (B')</td>
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<td></td>
</tr>
<tr>
<td>1,5-aldehyde-H-migration to D'</td>
<td>3.05×10⁹</td>
<td>4.28×10⁻⁶⁵</td>
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<td>-25.89</td>
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<tr>
<td>1,6-OOOH-migration to A'</td>
<td>2.07×10⁴</td>
<td>7.58×10⁻⁴⁹</td>
<td>7.91</td>
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<tr>
<td>1,6-OOOH-migration to C'</td>
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<td>(2R,3S)-2-Me-2,3-diOOH-butanal-4-peroxyl (C')</td>
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<td>1,6-aldehyde-H-migration to D'</td>
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<td>6.99×10⁻⁶⁷</td>
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<td>1,7-OOOH-migration to A'</td>
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<td>1,6-OOOH-migration to B'</td>
<td>1.52×10⁵</td>
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<td>(2R,3R)-2-Me-2,3,4-diOOH-1-oxo-1-butyl (D)</td>
<td>CO elimination</td>
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<tr>
<td>Effective aldehyde H-migration to D</td>
<td>5.03×10⁹</td>
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<td>(2R,3S)-2-Me-diOOH-butanalperoxyl (A+B+C)</td>
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23
Effective aldehyde H-migration to D’

<table>
<thead>
<tr>
<th></th>
<th>2.76×10^8</th>
<th>5.04×10^{-35}</th>
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<td>2-Me-diOOH-butanalperoxyl aldehyde H-migration *</td>
<td>11.8</td>
<td>6.52×10^{-35}</td>
<td>20.52</td>
<td>-1669</td>
</tr>
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</table>

*average of the data for (2R.3R)- and (2R.3S) conformers.
Figure 1. Schematic of the equilibrium reactions between OH-isoprene adducts and isoprene-RO$_2$ conformers, and their formation reactions. The names used for the different molecules are as in the MCM v3.3.1 (regular text), and in the LIM1 study (bold). The model-specific rate coefficients for each reaction are summarized in table 1.
Figure 2. Reaction schematic following the OH addition to the isoprene on the carbon C1. For simplicity, only the most relevant reaction paths for the OH radical formation are shown; the full extended mechanism including all six isoprene-RO\textsubscript{2} isomers was implemented in this work. The schematic illustrates the RO\textsubscript{2} unimolecular H-shift reactions (1,5-, 1,6- and aldehyde-H shifts), but all RO\textsubscript{2} species undergo competing reactions with NO, HO\textsubscript{2} and RO\textsubscript{2} radicals (not shown). Rate coefficients and yields shown are obtained from the theoretical work from this study and from recent experimental and theoretical studies. The same yields for the product of the 1,6-H shift, rate coefficient of the aldehyde-H shift, and photolysis frequencies for HPALD, are applied to the chemistry following the OH addition on C4. All rate coefficients are shown for 298 K and 30° solar zenith angle.
Figure 3. Potential energy surface for the aldehyde-H shift reaction showing the multiple competing reactions. A set of fast H-migration reactions ultimately lead to the formation of a tri-hydroperoxy acyl radical, tri-HPACYL. The main fate of this radical is shown by explicit theoretical calculations to be CO elimination, forming DHP-MVK; O₂ addition to a tri-hydroperoxy acylperoxy radical has only a minor contribution.
Figure 4. Comparison of modelled and measured trace gases for an experiment with NO < 0.2 ppbv. Measured time series of radicals and OH reactivity (LIF), isoprene and MVK+MACR+ISOPOOHs (GC) and CO (Picarro) are compared to model calculations. Vertical dashed lines indicate the times when isoprene was injected. No good agreement is observed when using the MCM v3.3.1 or a modified version (M1, Table 2) including isoprene-RO\textsubscript{2} conformers equilibrium reactions as included in the Caltech mechanism. Error bars represent 1 σ standard deviation.
Figure 5. Comparison of modelled and measured trace gases for an experiment with NO < 0.2 ppbv. Measured time series of radicals and OH reactivity (LIF), isoprene and MVK+MACR+ISOPOOHs (GC) and CO (Picarro) are compared to model calculations. Vertical dashed lines indicate the times when isoprene was injected. Good agreement is observed when using M2 or M3 (Table 2) which uses a different yield for HPALD, 0.40 and 0.75, respectively. Within both models, ~ 50% of isoprene-RO$_2$ radicals (weighted by their abundance) are lost via isomerisation reactions. Error bars represent 1 σ standard deviation.
Figure 6. Comparison of modelled and measured trace gases for an experiment with variable NO concentrations, \(1.5 > \text{NO} > 0.2\) ppbv. Measured time series of radicals and OH reactivity (LIF), isoprene and MVK+MACR+ISOPOOHs (GC) and CO (Picarro) are compared to model calculations. Vertical dashed lines indicate the times when isoprene was injected. All models are able to reproduce the observed trace gases within their uncertainties for NO > 0.2 ppbv.
Figure 7. OH regeneration efficiency at different NO concentrations. The experimental OH regeneration efficiency (RE) is compared with the modelled one (M2 panel a, M3 panel b) for different NO values. The modelled OH RE is color-coded by its main contributors (see table S6 for more details). The RE is sustained at low levels of NO by the contribution from isomerization reactions, in particular by the aldehyde-H shift and its products. For [NO] higher than 0.3 ppbv, most of the OH recycling originates from the reaction of HO$_2$ with NO but, although in a small amount, isomerization reactions still contribute up to 2 ppbv NO as the 1,6-H shift is still a competitive loss path for Z,δ-RO$_2$ radicals. The relative large contribution to RE by the reaction between HO$_2$ radicals and O$_3$ at low NO is due to the large concentration of O$_3$ (90 ppbv) needed in the simulation chamber to maintain low values of NO, and is not representative of ambient conditions. Error bars (1σ) for the measured OH regeneration efficiency include the accuracy of the measurements.
Figure 8. Global model of the OH regeneration efficiency at the surface (M2). The top panel shows the OH regeneration efficiency when considering only the reaction between HO$_2$ radicals with NO and with O$_3$. The bottom panel includes the remaining contributions by isomerization reactions and photolysis of HPALD. Isomerization reactions are very efficient in recycling the OH radicals and thus maintaining the oxidation capacity of the atmosphere in environments characterized by high isoprene and low NO.