

Original reviewer comments in bold, author responses beneath.

Referee #2

The manuscript presents a modelling investigation of the impact of recent increase in the emission of one of the major ozone depleting substance banned by the Montreal Protocol, the CFC-11. From the recently published evidence of recent emission increase, authors elaborate several scenarios for future emissions of CFC-11, with consequences on future levels of total inorganic chlorine and total ozone column throughout the 21st century. One interest of the study is that scenario independent metrics are explored to evaluate the relationship between ODS emissions and the timing of total ozone column recovery. The paper is well written and reference to previous work is adequate. It is suitable for publication in ACP provided that following comments and recommendations are considered in the revised version.

We thank the reviewer for their time and comments. Please find our responses to each comment below:

Major comments

1. Description of the model is limited and relates mainly on references. Very little information is provided on main features of tropospheric and stratospheric chemistry schemes. Are VSLs included? Since the model simulation output show a well-defined solar cycle effect, a brief description of the scheme used to integrate solar cycle variation would be valuable. Also, what is the lower boundary of the model?

The model extends from the surface to 84 km. VSLs are included in the chemistry with prescribed, time evolving surface mixing ratios. Further information relating to the model configuration used for this study have been added to the manuscript in section 2. The model description section now reads:

“To explore the impacts of potential future CFC-11 emission scenarios on TCO return dates, a total of 10 transient simulations were performed using version 7.3 of the HadGEM3-A configuration of the Met Office’s Unified Model (Hewitt et al., 2011) coupled with the United Kingdom Chemistry and Aerosol scheme (hereafter referred to as UM-UKCA). This configuration of the model has a horizontal resolution of 2.5° latitude \times 3.75° longitude, and 60 vertical levels following a hybrid sigma-geometric height coordinate, extending from the surface to a model top at 84 km. The chemical scheme is an expansion of the scheme presented in Morgenstern et al. (2009) in which halogen source gases are considered explicitly and the effects of the solar cycle are considered, as described in Bednarz et al. (2016), and includes 45 chemical species, 118 bimolecular reactions, 17 termolecular reactions, 41 photolysis and 5 heterogeneous reactions occurring on the surfaces of polar stratospheric clouds and sulphate aerosols. This chemistry scheme provides a detailed treatment of stratospheric chemistry including the O_x , ClO_x , BrO_x , HO_x and NO_x catalytic cycles, and a simplified tropospheric scheme including the oxidation of a limited number of organic species (CH_4 , CO , CH_3O_2 , CH_3OOH , $HCHO$) alongside detailed HO_x and NO_x chemistry. The chemical tracers O_3 , CH_4 , N_2O , CFC-11, CFC-12, CFC-113, and HCFC-22 are all interactive with the radiation scheme. The halogenated source gases CFC-11, CFC-12, CFC-113, HCFC-22, halon-1211, halon-1301, CH_3Br , CH_3Cl , CCl_4 , CH_2Br_2 , and $CHBr_3$ are considered explicitly, the concentration of each prescribed at the surface as a time evolving lower boundary conditions (LBC).

The version of UM-UKCA used in this study is an atmosphere-only configuration, with each simulation using the same prescribed sea surface temperatures (SSTs) and sea ice fields taken from a parent coupled atmosphere-ocean HadGEM2-ES integration. The configuration of the model used for this study includes the effects of the 11-year solar cycle in both the radiation and photolysis schemes. The top of atmosphere solar flux follows historical observations from 1960 to 2009, after which a repeating solar cycle is imposed which is an amplitude equivalent to the observed cycle 23 (as detailed in Bednarz et al., 2016). Further information on the model configuration used for this study is provided in Keeble et al. (2018). Except for CFC-11 and CFC-12 LBCs, all other chemical forcings in the simulations follow the experimental design of the WCRP/SPARC CCM1 REF-C2 experiment (Eyring et al., 2013), which adopts the RCP6.0 scenario for future GHG and ODS emissions. A baseline experiment performed using CFC-11 and CFC-12 LBCs provided by the WMO (2014) A1 scenario was run from 1960 to 2099. A further 9 simulations were performed, running from 2012 to 2099, using a range of CFC-11 and CFC-12 LBCs, which were designed to cover a large but plausible range of potential future CFC emissions scenarios given the associated uncertainties.”

2. The CFC-11 and CFC-12 emission scenarios are lacking explanation on the particular industrial activities they are based on. For example, what activities lead to emissive use of CFC-11? How realistic are the scenarios regarding industrial use of CFC-11? What is the reference for assuming equal emission of CFC-12 and CFC-11 in SCEN3 scenarios? A table summarizing the various scenarios would also be useful.

The scenarios explored in this study are not based on particular industrial activities, but instead are attempts to bound the potential impacts of recently reported CFC-11 emissions by exploring reasonable assumptions about future CFC-11 emissions and exploring the sensitivities to these assumptions. The direct emissions scenarios are not linked to any industrial activity, but instead explore the scenario in which the observed emissions reflect all the production (arguably the best-case scenario as it means there are no banks produced). This would be the case if the observed CFC-11 was being produced as a by-product in some other chemical manufacturing pathway and simply released. The most widely suggested use of recent CFC-11 production is as a blowing agent for foam insulation, so that the emissions represent some fraction of the total production, with the remainder building up in a newly created bank. In this case, we assume that for every 15 Gg directly emitted, 75 Gg enters the bank, i.e. 16.7% is released immediately, which is consistent with the 15% assumed by Dhomse et al. (2019) based on the study of Ashford et al., 2004. The reference for the ratio of co-production of CFC-12 is the Technology and Economic Assessment Panel. 2018 Assessment Report (http://conf.montreal-protocol.org/meeting/oewg/oewg-41/presession/Background-Documents/TEAP_2018_Assessment_Report.pdf), which states that pure production of CFC-11 is difficult to achieve, with many plants producing CFC-11 and CFC-12 in ratio of 30:70 either way. 50:50 was chosen as the middle of this range. The TEAP assessment has been added as a reference to this section of the manuscript.

3. In Figure 3, the spread of return dates of inorganic chlorine Cly at 40 km is smaller than that of CFC-11 return dates in Figure 2. It would help to show total tropospheric emitted chlorine in order to understand this difference.

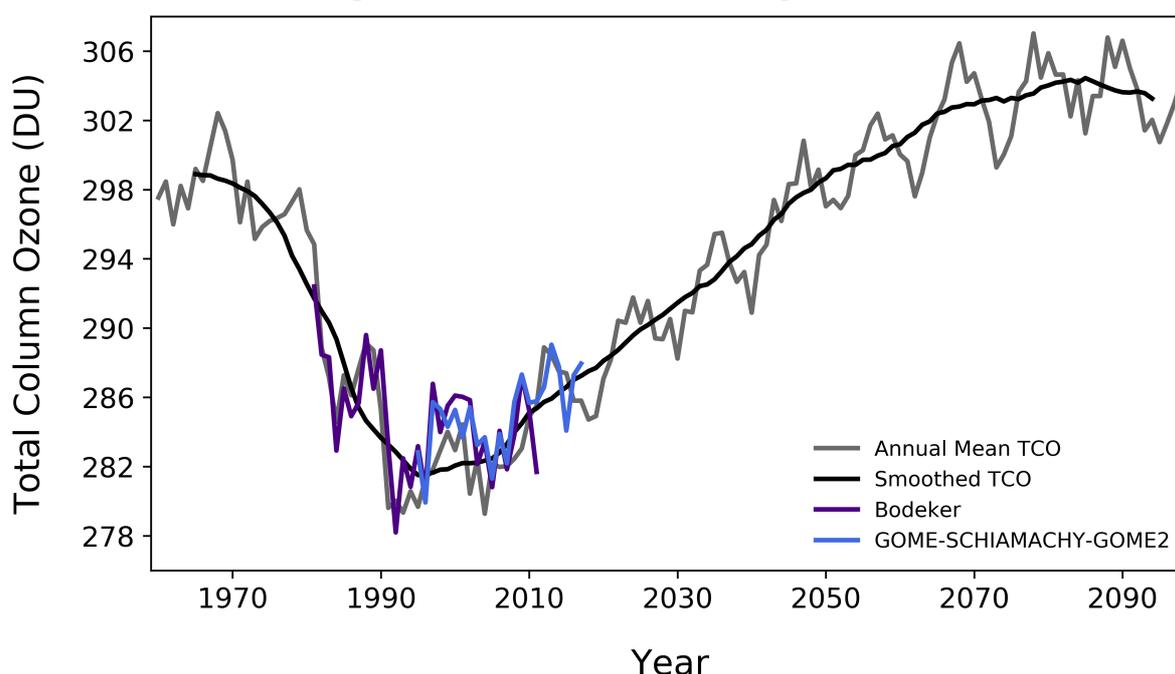
It is perhaps surprising that the spread in Cly at 40 km is smaller than the spread in CFC-11 return dates. We have further explored this, and it is not something peculiar at 40 km but is also true for a range of altitudes and latitudinal averages. It perhaps reflects the fact that CFC-11 is itself only a fraction of the total Cly, and that for many of the scenarios explored here

CFC-11 mixing ratios return to their 1980s values before Cly mixing ratios, suggesting other halogenated source gases are controlling earliest return date of Cly.

We feel that, as the UM-UKCA model uses prescribed lower boundary concentrations, rather than emissions of CFC-11 and other halogenated source gases, showing the total tropospheric emitted chlorine may be misleading.

4. Figure 4: Comparison of baseline scenario with Bodeker dataset: it would be also interesting to compare with one of the datasets used in the last WMO Ozone Assessment (WMO, 2018), e.g. the GOME-SCIAMACHY-GOME-2 (GSG) product from the University of Bremen (Weber et al., 2011). I am not sure that the solar cycle is that pronounced in other data sets.

We have added the GOME-SCIAMACHY-GOME-2 data to the figure. There is some difference between the two products, but overall the model agrees well with both datasets.



5. Figure 5: In order to evaluate the significance of the difference in return dates of the various scenarios, indication of uncertainty as indicated in page 7, 120-25 should be indicated in the figure. The largest spread in return dates is observed for Antarctica, the tropics and in the Arctic. Could the authors elaborate on the processes in the model simulation that explain this difference in spread of return dates? Is the simulated Brewer-Dobson circulation different in the various scenarios? Could that explain the difference?

We are reticent to put the spread of return dates onto Figure 5 as they come from an ensemble of simulations using different halogenated source gas LBCs to those explored in this study. While we include results from this ensemble in Table 1 as a rough guide for uncertainty estimates, we make clear the caveats associated with these ranges in the text, and so including them on the figure may be misleading. As the reviewer states, large uncertainties are seen in the Arctic and tropics. In the Arctic this is the result of large interannual variability in TCO values associated with dynamical variations, while in the tropics the chemical ozone depletion signal is small and TCO variability is dominated by features such as the solar cycle, QBO and ENSO. In all scenarios, the BDC responds to the modelled ozone depletion, further contributing to the variability. Estimating uncertainties in the Antarctic was not possible as the

year of TCO recovery occurs after 2080 for the latitude range 90S-60S, and this is denoted by the '?' in the table.

6. Figure 6: the timing of Cly recovery is chosen from the date at which Cly at 40km averaged in the 10°N-10°S latitude range return to 1980 value. Is there a reason for choosing the 10°N-10°S latitude range?

The 10°S-10°N latitudinal range is chosen as that is within the tropical pipe, in which air is predominantly moved vertically with limited horizontal mixing. This is necessary because Cly varies latitudinally with age of air. This has been added to the discussion of the figure, alongside references to Waugh 1996 and Neu and Plumb, 1999.

7. Figure 7: the added value of additional results from TOMCAT simulations is not clear to me. Also in which time frame the additional TCO depletion is computed in the figure?

The results from the TOMCAT model were included as consistency in the TCO difference vs Cl emission between it and the UM-UKCA model lends confidence to the rest of UKCA results, particularly given the differences in both the model configurations and the choice of CFC-11 emissions scenarios. It also suggests that the dominant driver of the response is the chemistry, and not circulation feedbacks, as the TOMCAT model is run with constant meteorological fields.

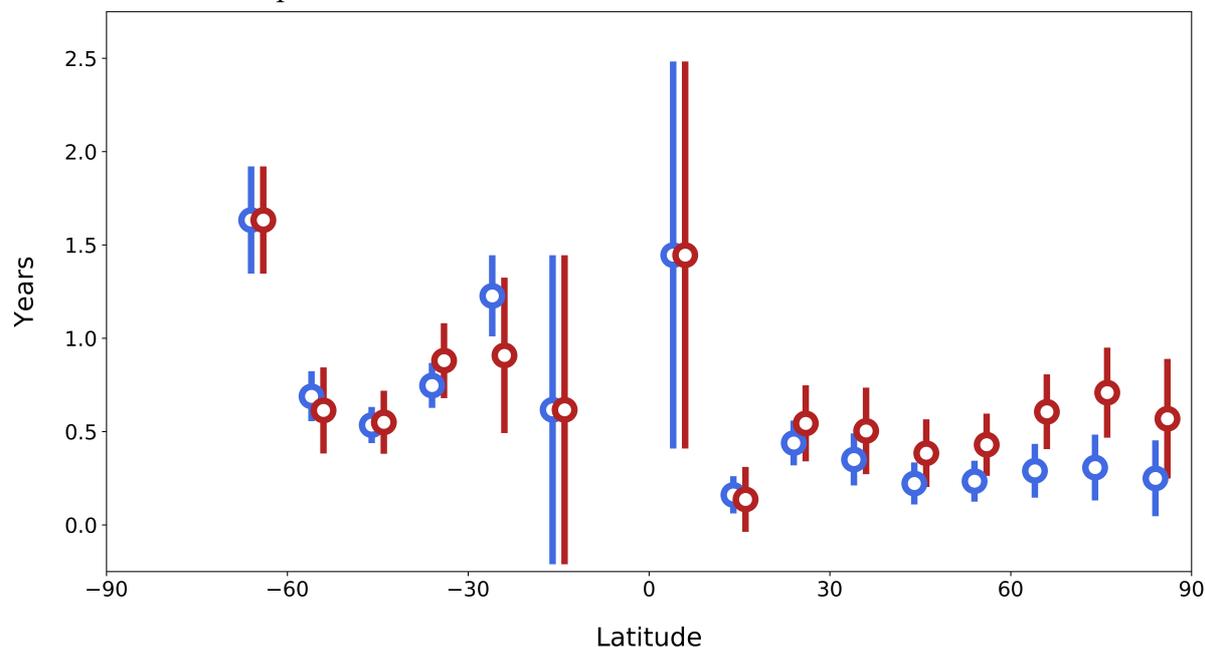
Figure 7 uses data from 2012-2100 for the UM-UKCA simulations, and 2012-2080 for the TOMCAT simulations. Each point on the figure represents one modelled year for a given scenario, and plots the TCO difference in that year between that scenario and base simulation against the cumulative Cl emissions reached in that scenario by that year. This has been clarified in the text.

8. In Figure 8, and other correlation figures, the goodness of the fit is heavily influenced by the SCEN3-30 scenario, which contradicts in some way the statement on the scenario independent metrics. I would assume that for the results ranged in latitude bins, even larger error bars would be expected from the regression in the absence of that extreme scenario. Could the authors comment on that and on the significance of the delay in TCO recovery in some latitude bins in that case?

The goodness of the fit is to some extent dependent on the SCEN3-30 point, but even when it is removed there still exists a strong correlation for the data in both Figure 6 (year of TCO recovery vs Cly recovery) and Figure 8 (year of TCO recovery vs cumulative Cl emissions). For Figure 6, when the SCEN3-30 experiment is ignored, the r^2 changes from 0.92 to 0.85, while the gradient changes from 0.64 to 0.72, while for Figure 8, when the SCEN3-30 experiment is ignored, the r^2 changes from 0.93 to 0.85, while the gradient changes from 0.56 to 0.59. This suggests that the relationships identified in the paper are not heavily influenced by the SCEN3-30 scenario.

This is true for the results in latitude bins as well. Below is a version of Figure 9 from the paper, with the blue points the same as those presented and using all the experiments, while the red points are the same calculations but ignoring the SCEN3-30 scenario. While the uncertainty estimates are changed slightly, and the delay in the Arctic is slightly longer per emission of 200 Gg Cl, there is no significant change, again showing that the SCEN3-30 scenario alone

does not heavily influence the conclusions of the paper. We have left the original Figure 8 the same in the manuscript as it uses all the available data.



Specific comments:

P3 19-12: The text is somewhat confusing. It is not clear whether the increase in CFC-11 emissions is 35 Gg.y-1 or 13 Gg.y-1 greater than expected from the Montreal Protocol.

Observational evidence indicates that CFC-11 emissions increased by ~ 13 Gg yr⁻¹ after around 2012, and that a large fraction of this increase came from East Asia. This increase, in conjunction with the expected decline in global CFC-11 emissions which would have resulted from full compliance with the Montreal Protocol, has resulted in global CFC-11 emissions being ~ 35 Gg yr⁻¹ higher than expected in 2019. This clarification has been added to the text, and the paragraph now reads:

“Against this background, Montzka et al. (2018) showed that the atmospheric abundance of one of the major chlorine-carrying CFCs, CFC-11, is not declining as expected under full compliance with the Montreal Protocol. Using the NOAA network of ground-based observations, they demonstrated clearly that the rate of decline of CFC-11 in the atmosphere between 2015-2017 was about 50% slower than that observed during 2002–2012 and was also much slower than had been projected by WMO (2014). They inferred that emissions of CFC-11 had been approximately constant at ~ 55 Gg yr⁻¹ between 2002 and 2012 and had then risen after 2012 by 13 Gg yr⁻¹ to ~ 68 Gg yr⁻¹. Montzka et al. (2018) argued that this increase could not be explained by increased release from pre-existing banks. Instead, they suggested that production of CFC-11 in east Asia, which is inconsistent with full compliance of the Montreal Protocol, was the likely cause. Using inverse modelling, Rigby et al. (2019) have since shown that the increase in CFC-11 emissions from eastern China between 2008-2012 and 2014-2017 is ~ 7 Gg yr⁻¹, corresponding to approximately 40-60% of the global emission increase identified by Montzka et al. (2018) during that period. The increase in CFC-11 emissions after 2012, in conjunction with the expected decline in global CFC-11 emissions which would have resulted from full compliance with the Montreal Protocol, has resulted in global CFC-11 emissions being ~ 35 Gg yr⁻¹ greater than anticipated by the WMO (2014) A1 scenario in 2019.”

P5 I19-20: explain relation between 1Gg CFC-11 and Gg Cl.

The relationship between the two is $1\text{Gg CFC-11} = 0.77\text{ Gg Cl}$. This is based on the fact that the molar mass of CFC-11 is 137.37 g/mol and contains 3 molecules of Cl (molar mass = 35.453 , multiplied by 3 is 106.359). Therefore, every Gg of CFC-11 contains $106.359/137.37 = 0.77\text{ Gg Cl}$.

P13 I20-22: It is not clear why 200 Gg of CFC-11 emitted in 2020 would not have the same effect on ozone recovery as the same emitted in 2080.

This statement arises from two considerations. The first is related to the non-linear coupling between chlorine emission, background stratospheric composition and climate. The same emission of CFC-11 has a different relative impact on stratospheric ozone for different stratospheric temperatures arising from the different temperature dependences of ozone depleting reactions. Further, changes to CH_4 and N_2O lower boundary conditions also change the relative proportions of the HO_x , NO_x and ClO_x catalytic cycles, and the partitioning of stratospheric chlorine. Studies by Banerjee et al. (2016), and Keeble et al. (2017) have explored this non-linear response of column ozone changes to ODS changes under different future climates in the UKCA model, and these results are consistent with the original findings of Haigh and Pyle (1982) and recent studies from other groups (e.g. Fleming et al., 2011; Portmann et al., 2012; Meul et al., 2015). The second is that 200 Gg of CFC-11 emitted into the atmosphere after the year of ozone recovery when chlorine mixing ratios have already been significantly reduced may not reduce column ozone values below the 1980s/1960s value, and so while this leads to depletion of ozone it does not influence return dates. The text in the manuscript has been modified to reflect these points, and the paragraph now reads:

“It is not expected that an emission of CFC-11 emitted in 2020 would have the same impact on ozone return dates as the same emission of CFC-11 emitted in, for example, 2080. This is due in part to the different background stratospheric temperatures, circulation, and sinks of active chlorine (e.g. the conversion of ClO_x to HCl through reaction with CH_4) at different times throughout the 21st century. Furthermore, any additional chlorine emissions which occur after TCO has returned to its 1960-1980 mean value might not deplete ozone below this value, and so would not affect the return date.

In the text, some exponents are not well written, e.g. r^2 values for the goodness of the fits.

These have been corrected.

Anonymous Referee #3

The paper reports on the impact of recent emissions of CFC-11 on the ozone layer, first documented in Montzka et al., and most recently, in Rigby et al. 2019. The paper explores a number of possible scenarios of CFC-11 emissions, and concludes that if emissions are allowed to continue into the future, significant delays in the recovery of the ozone layer are expected (by a decade or more in the global TCO, and even more in the Antarctic stratosphere). The subject of the paper is of high relevance and interest for the readership at ACP, and for the wider community. The paper is well written and the conclusions of the paper are supported by the data. There are nonetheless a few issues in the paper, which need to be addressed in order to make the paper suitable for publication. One of them is the missing discussion of another paper, which recently came out on the same subject (and in this same journal), namely Dameris et al., ACP 2019 (<https://doi.org/10.5194/acp-19-13759-2019>). I believe this paper is not discussed nor mentioned in the text because it came out quite recently, so the authors did not see it. Dameris et al., similar to this study, investigated the impact of two CFC-11 emission scenarios on ozone projections, and found that a substantial delay of ozone recovery by up to 20 years could occur due to unabated CFC11 emissions. Are the results in this paper consistent with their conclusions (bearing in mind their differences in the CFC11 scenarios)? There are also some issues in the missing discussion & quantification of the chemical mechanism leading to the delayed ozone recovery. Dameris et al. 2019 found that actually, some catalytic cycles are slowed down (e.g. Ox and NOx), leading to ozone increases under increased CFC11 emissions (which compensate the ozone depletion from enhanced ClOx cycles). For example, how important are ClOx cycles, relative to the other depleting cycles? In addition, the authors tend to only cite their own past papers, neglecting the much more vast body of literature on e.g., BDC and its impacts on ozone, and multi-model comparisons. Please cite the first papers that studies these problems, and not only your own ones; see specific comments below. The paper would benefit from addressing these issues.

We thank the reviewer for their time and comments. Please find our detailed responses below. The referee rightly indicates that more comprehensive referencing is required in some areas – we have been happy to do this.

SPECIFIC COMMENTS

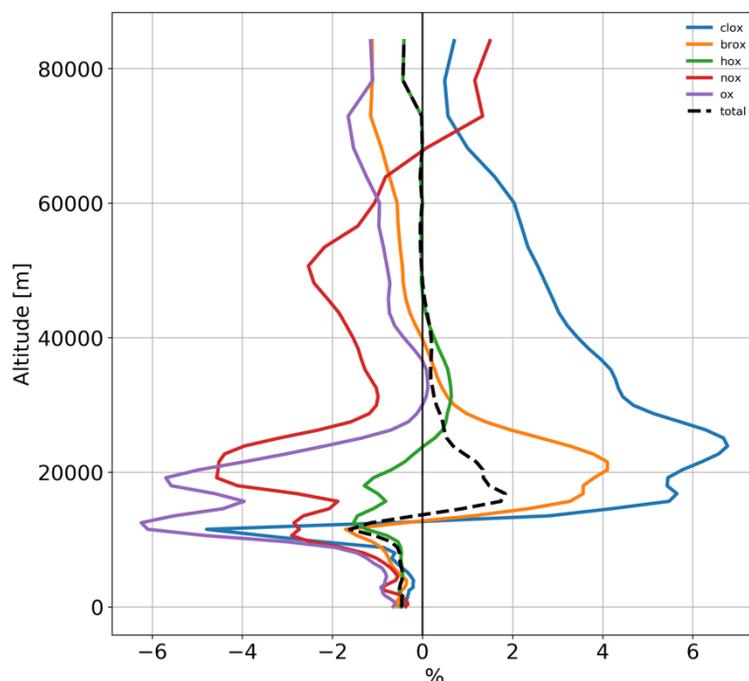
- Can the present results be compared and discussed in the context of the recent paper by Dameris et al., 2019?

The Dameris et al. results have now been discussed in relation to the results presented in this manuscript in the introduction and discussion sections.

- Can the authors please try to pick a more descriptive experiment tag? SCEN_x-y is quite cumbersome and does not convey the basic information of the experiments setup. For example, how about "SCEN35" and "SCEN90" so that the reader immediately knows the total emission in terms of mass of CFC11... or something similar? The authors need to constantly remind the reader what each tag means, when they wish to emphasize some key result. This could be avoided by picking a better experiment tag.

We feel that the scenario names chosen best reflect the emissions scenarios being considered as they include both the scenario type (1 = direct emission, 2 = formation of a bank, 3 = co-production of CFC-12) and also the duration of the assumed CFC-11 production, in years, in as succinct a manner as possible. It is not possible to use only the total emission of CFC-11 to name the scenarios as the emissions change through time for the scenarios which assume a bank is produced, and nor would it reflect the duration of the production.

- Could the authors look at the ClOx and other catalytic cycles, so that the reader can get a glimpse into the key chemical mechanism, and its dependence on height (without just simply assuming that the Molina and Rowland (1974) cycle is enhanced...? While I don't expect the authors to perform a full ozone budget analysis for each of the ensembles, showing the relative change in ozone production rates in a few key experiments would improve the paper.



We agree with the reviewer that it is unlikely that only the ClOx catalysed ozone depleting cycles will be affected by the changes to the CFC-11 and CFC-12 lower boundary conditions. Above is a plot of the relative difference in global mean ozone depletion through the ClOx, BrOx, NOx, HOx and Ox catalytic cycles between the SCEN2_30 simulation and BASE (averaged from 2030-2040). As found in the Dameris study, increases to ozone destruction through the ClOx catalytic cycles are somewhat offset by decreases to ozone loss through other catalytic cycle. However, interpretation of the drivers of these changes are complicated. This is partly through the chemical coupling of the ClOx cycles with the NOx, HOx and BrOx cycles, and partly through changes to the Brewer-Dobson circulation affecting the oxidation of source gases. Diagnosing the changes to these cycles is further complicated by the need to account for the changes to O₃ concentrations. Increases to ClO will deplete stratospheric ozone, and so we may expect, for example, NOx catalysed ozone depletion to slow as 1) there is extra formation of the ClONO₂ reservoir, and 2) there is a lower concentration of O₃. Because of these reasons, we feel that addressing the question of which catalytic cycles are responsible for the TCO changes examined in the paper is beyond the scope of this study. It should be stated that we have not assumed in the text that only the Molina and Rowland cycles

are affected but have added text to the discussion to include the points above and cited the Dameris et al. ozone budget results in the discussion section of the manuscript.

MINOR COMMENTS

L20 page 2: Ball et al., ACP 2018 (<https://doi.org/10.5194/acp-18-1379-2018>) is another key paper that should be cited here.

This citation has been added to the discussion.

L5 page 5: why is the lifetime fixed? Shouldn't it be a strong function of altitude, and depend on the photolysis rate (which in turn is a function of the actinic flux)?

While the local lifetime of CFC-11 is dependent on its destruction through photolysis, the 55 year lifetime is used to convert the emissions fluxes into a mixing ratio at the surface of the model, and so represents its bulk atmospheric lifetime (i.e. the global burden/global loss). Imagining a pulse emission of CFC-11, some is transported into the middle and upper stratosphere where it is rapidly destroyed, but the majority (by mass) remains in the troposphere and lower stratosphere, where its lifetime is much longer. A lifetime of 55 years accurately projects the decline in the mixing ratio observed at the surface, and hence its use in this study. The value of 55 years for the lifetime comes from Chipperfield et al., 2014 (referenced in the manuscript).

L12 page 7: again, Salby et al., 2009, and more recently Ball et al. 2018, also discussed the impact of variability on the detection of a recovery. So, these papers should be cited, too.

We have added the Ball citation to the manuscript. We could not find a Salby et al., 2009 publication, but have cited instead the Salby et al., 2011 paper which discusses the impact of variability on the detection of Antarctic ozone recovery (Rebound of Antarctic ozone, GRL).

L15 page 7: how about the technique used in Eyring et al. 2013 and in all the WMO reports (the TSAM method)? This method seems more customary and it should be at least mentioned what the differences between this paper's method and that one is.

This method has been added to the manuscript, along with the Scinocca et al. (2010) reference. The sentence now reads:

“To mitigate these impacts, the effects of natural processes (such as volcanic eruptions, the QBO, ENSO and solar cycle) can be removed from the data using statistical techniques (such as Multiple Linear Regression, e.g. Staehelin et al., 2001; WMO, 2007 or the Time series Additive Model, Scinocca et al., 2010), or the data can be smoothed by averaging across multiple years (e.g. Dhomse et al., 2018).”

L31 page 8: this is probably subject to a lot of variability and running multiple ensembles may wash out this little increase TCO values. Or perhaps, there may be some compensating effect from NO_x and O_x cycles. It would be good if the authors could look at the production rates (see specific comment above).

As the reviewer states, these differences likely result from variability between the integrations and would likely be removed by the use of more ensemble members. However, it was decided

it would be better to explore a larger spread in possible future CFC-11 production scenarios. Please see our comment above about analysing the ozone budget terms in these simulations.

L6 page 9: the impact of BDC acceleration on TCO has been studied in more studies than just these ones. Most recently, this has been shown in Chiodo et al., J.Clim. 2018 (DOI:10.1175/JCLI-D-17-0492.1). So, at the very least, add this paper to the references here.

Additional references have been added to this discussion. The sentence now reads: “The observed ozone loss in the tropics has been small and, furthermore, future changes in the tropics are driven both by reductions in the stratospheric abundance of halogens, which tend to increase ozone, and the strengthening of the Brewer-Dobson circulation, which tends to decrease column ozone (e.g. Oman et al., 2010; Eyring et al., 2013; Meul et al., 2014; Keeble et al., 2017; Chiodo et al., 2018).”

L13 page 9; this has been first studied in Oman et al., 2009 (<https://doi.org/10.1029/2010JD014362>), so this paper should be cited, too.

This reference has been added to the manuscript.

L4-16 page 10: I don't get what the scenario-independent aspect would be, in this context. Cly clearly depends on the CFC emission scenario so there is no scenario independency here.

By ‘scenario independent’ we mean one could use the linear relationships identified between the amount of chlorine emitted and the timing of TCO return to estimate the delay in TCO return date for any particular chlorine emission. The benefit of this extends to emissions not explored in this study (i.e. not one of the 10 scenarios investigated). As the reviewer states, the Cly projection depends on the assumed CFC emission scenario, but it is not known what possible future emission scenario will be followed. Rather than argue that one of the scenarios explored in this study is more likely than the others, we have looked for ways to say ‘if the emission of CFC-11 was x, then the delay would be y’, which we believe would represent a positive benefit to policy and regulation.

- The paper by Stolarski et al. 2012 (<https://doi.org/10.1029/2012JD017456>) shows that the sensitivity of ozone to temperature decreases as chlorine increases in the stratosphere. Hence, it should increase at a slower rate compared to the reference scenario, as chlorine decreases at a slower rate. This may impact the response of ozone to the upper stratospheric cooling, as a function of much Cly we have in the stratosphere. Could this explain some of the non-linearities found by the authors?

The reviewer is correct to highlight this process as a potential cause for the non-linearities seen here. An additional factor is coupling between the chemistry and dynamics. However, despite these non-linear couplings, and the interannual variability inherent in CCMs, the linear relationships identified in this study between the amount of chlorine emitted and the delay in ozone recovery are robust, with high r² values, suggesting that these factors are minor in the face of the dominant effect of changing stratospheric Cl and depletion of ozone.

L8 page 12: other papers have reported this in the past, a lot more than just Keeble et al., 2017, so they should be cited. At the very least, Butchart et al., 2014 should be cited here.

The Butchart et al. and Meul et al., 2016 references have been added to the discussion.

L20 page 13: where should this non-linearity come from (i.e. the same CFC11 change in 2080 having a different effect than in 2020) ? The Cl + O3 reaction rate changing because of O3 background concentrations? (the Cl change would be the same!)

Please see our response to the same question from referee #2

Modelling the potential impacts of the recent, unexpected increase in CFC-11 emissions on total column ozone recovery

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Abstract. The temporal evolution of [the abundance of](#) long-lived, anthropogenic chlorofluorocarbons [in the atmosphere](#) is a [major factor in determining](#) the timing of total column ozone (TCO) recovery. Recent observations have shown that the atmospheric mixing ratio of CFC-11 is not declining as [rapidly as](#) expected under [full](#) compliance with the Montreal Protocol, and indicate a new source of CFC-11 [emissions](#). In this study, the impact of a number of potential future CFC-11 emissions scenarios on [the timing of](#) TCO [return to the 1960-1980 mean \(an important milestone on the road to recovery\)](#) is investigated using the UM-UKCA model. Key uncertainties related to this new CFC-11 source and their impact on the timing of [the](#) TCO [return date](#) are explored, including: the duration of new CFC-11 [production and emissions](#); the impact of any newly created [CFC-11](#) bank; and the effects of co-production of CFC-12. Scenario-independent relationships are identified between cumulative CFC emissions and the timing of [the TCO return date](#), which can be used to establish the impact of future CFC emissions pathways on ozone recovery in the real world. It is found that, for every 200 Gg Cl ([-258 Gg CFC-11](#)) emitted, the timing of global TCO [return to 1960-1980 averaged values](#) is delayed by ~0.56 years. However, a marked hemispheric asymmetry in the latitudinal impacts of cumulative Cl emissions on the timing of [the TCO return date](#) is identified, with longer delays in the southern hemisphere than the northern hemisphere for the same emission. Together, these results indicate that, if rapid action is taken to curb recently identified CFC-11 production then no significant delay in the timing of TCO [return to the 1960-1980 mean](#) is expected, highlighting the importance of ongoing, long-term measurement efforts to inform the accountability phase of the Montreal Protocol. However, if the emissions are allowed to continue into the future, and are associated with the creation of large banks, then significant delays in the timing of [the TCO return date](#) may occur.

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1 Introduction

The discovery of the ozone hole by Farman et al. (1985) led quickly to the confirmation of the idea put forward by Molina and Rowland (1974) that chlorine radical species, the breakdown products of the chlorofluorocarbons (CFCs), could deplete stratospheric ozone. In the face of the scientific evidence, the Montreal Protocol on substances that deplete the ozone layer was agreed in 1987 and ratified in 1989. The original controls proposed were modest, covering only CFC₁₁ (CFC-11), CFC₁₂ (CFC-12), three further CFCs and three brominated compounds (Halons). However, in line with the developing scientific understanding, the controls were subsequently strengthened in a series of adjustments and amendments to the Protocol. These included stronger regulation on the phase down schedules, the addition of many more CFCs, CCl₄ and transitional hydrochlorofluorocarbon (HCFC) compounds under the London Amendment in 1990 and the inclusion of many more brominated compounds, including methyl bromide under the Copenhagen Amendment of 1992. The initial replacements for the CFCs, the HCFCs, have shorter lifetimes than the CFCs (Chipperfield et al., 2014) and accordingly their impact on stratospheric ozone is less. They, in turn, are being replaced by hydrofluorocarbons (HFCs), compounds which do not directly lead to ozone depletion but some of which are strong greenhouse gases. Regulations to limit the growth of many HFCs were agreed in the Kigali Amendment in 2016.

In consequence, the atmospheric abundances of ozone-depleting chlorine and bromine species are now declining in the atmosphere, following their peaks in the late 1990s (WMO, 2018), leading to the start of ozone recovery. The 2018 Scientific Assessment for the Montreal Protocol (WMO, 2018) reported that the Antarctic ozone hole, while continuing to occur each year, is showing early signs of recovery (e.g. Solomon et al., 2016) and that upper stratospheric ozone has increased by up to 3% since 2000 (e.g. LOTUS, 2019). However, there is as yet no significant recovery trend in global column ozone (e.g. Ball et al., 2018; Weber et al., 2018). Furthermore, because many ODSs are also greenhouse gases, their control has brought significant climate benefits (Velders et al., 2007; Velders et al., 2012; WMO, 2018). The annual reduction in these greenhouse gas emissions is estimated to be about five times larger than the annual emission reduction target for the first commitment period of the Kyoto Protocol (WMO, 2014).

The Montreal Protocol has undoubtedly been successful. Without the Protocol the abundance of ODSs would likely have risen such that, for example, very large ozone depletion could have occurred in the Arctic (Chipperfield et al., 2015). Uncontrolled growth of ODSs would also have severely exacerbated the impact on global warming of the increase in other greenhouse gases, making current climate targets even more challenging to meet. However, there have recently been questions about the completeness of the implementation of the Protocol. The concentrations of some short-lived anthropogenic halocarbons, which are not covered by the Protocol, have increased in the atmosphere (Hossaini et al., 2017; Oram et al., 2017; Fang et al. 2018), with suggestions that they might be by-products in the production of other halocarbons. Furthermore, the concentrations of some of the controlled ODSs have not followed projections based on their phase-out under the Montreal Protocol. For instance, concentrations of carbon tetrachloride, CCl₄, have not fallen as rapidly as expected based on its atmospheric lifetime. A detailed reanalysis of CCl₄ indicates that inadvertent by-product emissions from the production of chloromethanes and

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perchloroethylene, and fugitive emissions from the chlor-alkali process, have contributed to this discrepancy (SPARC, 2016; Sherry et al., 2018; WMO, 2018) and recently Lunt et al. (2018) have shown that emissions of CCl₄ from east China have increased in the last decade. East China as a source of other short-lived halocarbons was also suggested by Ashfold et al. (2015) and Fang et al. (2018).

5 Against this background, Montzka et al. (2018) showed that the atmospheric abundance of one of the major chlorine-carrying CFCs, CFC-11, is not declining as expected under full compliance with the Montreal Protocol. Using the NOAA network of ground-based observations, they demonstrated clearly that the rate of decline of CFC-11 in the atmosphere between 2015-2017 was about 50% slower than that observed during 2002–2012 and was also much slower than had been projected by WMO (2014). They inferred that emissions of CFC-11 had been approximately constant at ~55 Gg yr⁻¹ between 2002 and 2012 and had then risen after 2012 by 13 Gg yr⁻¹ to ~68 Gg yr⁻¹. Montzka et al. (2018) argued that this increase could not be explained by increased release from pre-existing banks. Instead, they suggested that production of CFC-11 in east Asia, which is inconsistent with full compliance of the Montreal Protocol, was the likely cause. Using inverse modelling, Rigby et al. (2019) have since shown that the increase in CFC-11 emissions from eastern China between 2008-2012 and 2014-2017 is ~7 Gg yr⁻¹, corresponding to approximately 40-60% of the global emission increase identified by Montzka et al. (2018) during that 15 period. The increase in CFC-11 emissions after 2012, in conjunction with the expected decline in global CFC-11 emissions which would have resulted from full compliance with the Montreal Protocol, has resulted in global CFC-11 emissions being ~35 Gg yr⁻¹ greater than anticipated by the WMO (2014) A1 scenario in 2019.

The exact source of the emissions remains unknown, nor is it known if there is co-production of CFC-12. It is usual that the gases are produced together in an industrial plant (e.g. Siegemund et al., 2000), with the fraction of CFC-11 to CFC-12 20 production varying between 0.3 to 0.7 (UNEP, 2018). There is currently no evidence that CFC-12 concentrations in the atmosphere are also declining at a slower rate than expected but some co-production of CFC-12 along with CFC-11 is always expected.

Compliance with the Montreal Protocol is essential for its continued success in reducing stratospheric Cl_y and ultimately healing the ozone layer. Any non-compliance will inevitably prolong the period when the Antarctic ozone hole will continue 25 to occur and delay the date at which global total column ozone (TCO) returns to its 1980s values, an important milestone on the road to recovery. Recent studies by Dameris et al., (2019) and Dhomse et al., (2019) have explored a range of different future ODS loadings, enhanced above those expected under full compliance with the Montreal Protocol, and find substantial delays in recovery depending on the scenario. It is essential therefore to understand the likely impact of the current non-compliance. Here the UM-UKCA chemistry-climate model is used to assess the possible implications of the change in decline 30 of CFC-11. A number of possible scenarios are explored in a range of sensitivity calculations. These include emissions which cease immediately, or which persist for different periods into the future. These scenarios also consider that some of the non-compliant production may be stored in new CFC-11 banks, for later release, and, that production of CFC-11 may be associated with co-production of CFC-12.

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Until the source of the recent CFC-11 emissions is understood and thoroughly quantified, model calculations can only investigate a range of possible future emissions scenarios. However, models can be used to search for a relationship between the amount of chlorine emitted into the atmosphere and the timing of TCO return dates. Here results from both the UM-UKCA model and the TOMCAT CTM (recently used to study the impact of increased CFC-11 emissions on the behaviour of the Antarctic ozone hole (Dhomse et al., 2019)) are used to investigate the relationship between increased emissions and enhanced ozone depletion. Identification of a robust relationship would allow us to develop a scenario-independent understanding of the impact of uncontrolled CFC emissions on the TCO return date.

In Section 2 the UM-UKCA model and CFC scenarios used in this study are discussed in detail. Section 3 assesses the impact of these CFC emissions scenarios on stratospheric chlorine loading, before the impacts on the TCO return date are examined in Section 4. Section 5 investigates the relationship between cumulative CFC emissions and ozone depletion, using both the CCM and CTM. Finally, further discussion of the results and a summary of our conclusions are provided in Section 6.

2 Model configuration and simulations

To explore the impacts of potential future CFC-11 emission scenarios on the TCO return date, a total of 10 transient simulations were performed using version 7.3 of the HadGEM3-A configuration of the Met Office's Unified Model (Hewitt et al., 2011) coupled with the United Kingdom Chemistry and Aerosol scheme (hereafter referred to as UM-UKCA). This configuration of the model has a horizontal resolution of 2.5° latitude × 3.75° longitude, and 60 vertical levels following a hybrid sigma-geometric height coordinate, extending from the surface to a model top at 84 km. The chemical scheme is an expansion of the scheme presented in Morgenstern et al. (2009) in which halogen source gases are considered explicitly and the effects of the solar cycle are considered, as described in Bednarz et al. (2016), and includes 45 chemical species, 118 bimolecular reactions, 17 termolecular reactions, 41 photolysis and 5 heterogeneous reactions occurring on the surfaces of polar stratospheric clouds and sulphate aerosols. This chemistry scheme provides a detailed treatment of stratospheric chemistry including the O_3 , ClO_x , BrO_x , HO_x and NO_x catalytic cycles, and a simplified tropospheric scheme including the oxidation of a limited number of organic species (CH_4 , CO , CH_3O_2 , CH_3OOH , $HCHO$) alongside detailed HO_x and NO_x chemistry. The chemical tracers O_3 , CH_4 , N_2O , CFC-11, CFC-12, CFC-113, and HCFC-22 are all interactive with the radiation scheme. The halogenated source gases CFC-11, CFC-12, CFC-113, HCFC-22, halon-1211, halon-1301, CH_3Br , CH_3Cl , CCl_4 , CH_2Br_2 and $CHBr_3$ are considered explicitly, the concentration of each prescribed at the surface as a time evolving lower boundary conditions (LBC).

The version of UM-UKCA used in this study is an atmosphere-only configuration, with each simulation using the same prescribed sea surface temperatures (SSTs) and sea ice fields taken from a parent coupled atmosphere-ocean HadGEM2-ES integration. The configuration of the model used for this study includes the effects of the 11-year solar cycle in both the radiation and photolysis schemes. The top of atmosphere solar flux follows historical observations from 1960 to 2009, after which a repeating solar cycle is imposed which is an amplitude equivalent to the observed cycle 23 (as detailed in Bednarz et

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al., 2016). Further information on the model configuration used for this study is provided in Keeble et al. (2018). Except for CFC-11 and CFC-12 LBCs, all other chemical forcings in the simulations follow the experimental design of the WCRP/SPARC CCM1 REF-C2 experiment (Eyring et al., 2013), which adopts the RCP6.0 scenario for future GHG and ODS emissions. A baseline experiment performed using CFC-11 and CFC-12 LBCs provided by the WMO (2014) A1 scenario was run from 1960 to 2099. A further 9 simulations were performed, running from 2012 to 2099, using a range of CFC-11 and CFC-12 LBCs, which were designed to cover a large but plausible range of potential future CFC emissions scenarios given the associated uncertainties.

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2.1 CFC-11 scenarios

There are a number of important details associated with the recently reported CFC-11 emissions from East Asia which are poorly understood. A key factor is whether the identified CFC-11 emissions arise from emissive or non-emissive uses. If they arise from a totally emissive use, then the observed CFC-11 changes represent the total new CFC-11 production, with this new source of CFC-11 being released into the atmosphere during either production or use. Conversely, if they arise from a non-emissive use (e.g. foam insulation), then the observed changes to CFC-11 represent only a fraction of the total production, with a large component entering a new bank. In order to address this, two scenario types were created, which reflect the new CFC-11 production which is in addition to that implied in the WMO (2014) A1 scenario. In SCEN1, which represents the emissive use scenario, constant emissions of 35 Gg CFC-11 yr⁻¹ (~27 Gg Cl yr⁻¹) were assumed while in SCEN2 total production was assumed to be 90 Gg CFC-11 yr⁻¹ (~70 Gg Cl yr⁻¹), with 15 Gg CFC-11 yr⁻¹ of this total directly emitted into the atmosphere, while the remaining 75 Gg CFC-11 yr⁻¹ entered a bank with an assumed release fraction of 3.5 % yr⁻¹. The SCEN2 scenario is designed to give an additional emission increment of ~35 Gg CFC-11 yr⁻¹ in 2019, similar to the emissions value used in SCEN1 and consistent with the CFC-11 emissions increment, in addition to that expected assuming only release from known banks, identified by Montzka et al. (2018.) We emphasize that the 35 Gg CFC-11 yr⁻¹ and 90 Gg CFC-11 yr⁻¹ values represent the extra emissions/production increment assumed in addition to those of the WMO (2014) A1 scenario.

A second key factor is the duration of the illegal production. In order to address this question, three sets of simulation were performed for both SCEN1 and SCEN2, in which uncontrolled production either stops this year (2019), or continues into the future until 2027 or 2042, giving total emissions periods of 7, 15 or 30 years, respectively. All simulations assume that uncontrolled production and emission began in 2012. Simulations are named such that the scenario name is followed by the emission period, separated by an underscore (i.e. SCEN1_15 uses SCEN1 emissions from 2012 for 15 years).

A third consideration is the potential co-production of CFC-12. While there is no evidence currently that CFC-12 concentrations in the atmosphere are declining at a slower rate than expected (Montzka et al., 2018), CFC-12 is commonly co-produced alongside CFC-11 in ratio of 30:70 either way (TEAP, 2019). Accordingly, an additional scenario (SCEN3) was developed in which 90 Gg yr⁻¹ of both CFC-11 and CFC-12 is produced, i.e. a ratio of 50:50, towards the middle of the expected range. The same assumptions are made about the relative fraction entering the bank and the subsequent bank release rate as

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for SCEN2. SCEN3 was performed for the three different emissions periods used by SCEN1 and SCEN2, and simulations follow the same naming convention. As we consider both CFC-11 and CFC-12, all future emission/production values will be given in Gg Cl, with 1 Gg CFC-11 equal to ~ 0.77 Gg Cl, and 1 Gg CFC-12 equal to ~ 0.59 Gg Cl.

The emissions for these various scenarios are shown (in Gg Cl) in Figure 1a, while Figure 1b shows the cumulative emissions and Figure 1c shows the size of the newly created bank. The SCEN3 scenarios include both CFC-11 and CFC-12, both scaled to Gg Cl and summed.

Figure 1a highlights that the SCEN1 scenario emissions stop at the end of the assumed production, while for the SCEN2 and SCEN3 scenarios emissions continue throughout the 21st century long after the cessation of production due to the newly created bank. For SCEN2 and SCEN3 the shapes of the emissions curves are controlled by the combination of direct emission and newly created bank. While production continues, the yearly direct emissions remain constant, but the bank grows, leading to larger total emissions per year. At the moment production stops the direct emissions also stop, resulting in a marked step down in the emissions. After that point, all emissions emanate from the newly created bank. The maximum size of emission is dictated by the duration of production. For comparison, Daniel et al. (2007) estimate that peak CFC-11 production, before the adoption of the Montreal Protocol, reached ~ 300 Gg Cl yr⁻¹ in 1986-87. For SCEN1 scenarios, cumulative emissions increase only during the period of assumed production; there is no newly created bank. For the SCEN2 and SCEN3 scenarios cumulative emissions grow most rapidly during the period when direct emissions occur but they continue to increase throughout the 21st century due to emissions from the bank. The size of the newly created bank is dependent on the duration of production and the release rate from the bank (Figure 1c), which was assumed to be 3.5% yr⁻¹.

As discussed above, the UM-UKCA model uses prescribed lower boundary condition (LBC) mixing ratios of CFCs. As a result, each of the emissions scenarios described above was converted into LBC mixing ratios using a simplified box model which uses only the emissions flux and a CFC-11 lifetime of 55 years. This model reproduces to within good agreement the observed 1994-2017 CFC-11 surface mixing ratio when initialized with the 1994 values and using the emission estimates of Montzka et al. (2018). The time variation of CFC-11 at the surface is shown in Figure 2 for the different scenarios.

3 Stratospheric Chlorine

Increases in stratospheric chlorine will lead to ozone depletion, and so uncontrolled production of CFCs could obviously pose a serious threat to the continued success of the Montreal Protocol. Modelled 40 km stratospheric inorganic chlorine (Cl_y) mixing ratios, averaged from 10°S-10°N, are shown in Figure 3, and stratospheric the Cl_y return date (the date at which Cl_y mixing ratios, averaged from 10°S-10°N at 40 km, return to the BASE 1980 value) are given in table 1. We chose the 10°S-10°N latitude range for calculating Cl_y return dates as this is within the tropical pipe in which air is predominantly moved vertically with limited horizontal mixing (e.g. Waugh, 1996; Neu and Plumb, 1999), a necessary consideration as Cl_y varies latitudinally with age of air. In the BASE simulation, stratospheric Cl_y mixing ratio is projected to return to its 1980 value by

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2058. Only small differences in the stratospheric Cl_y return date are modelled for the SCEN1 simulations, with a maximum delay of 3 years occurring in the SCEN1_30 simulation, which assumes the longest duration of additional CFC-11 emissions. However, large delays in the stratospheric chlorine return date are modelled in the SCEN2 simulations, which assume a large bank is also being produced alongside the direct atmospheric emissions (see figure 1). In the SCEN2_7 scenario, which assumes CFC-11 production stops in 2019, the stratospheric Cl_y return date is delayed by 2 years, and for SCEN2_30, which assumes CFC-11 production continues until 2042, the stratospheric Cl_y return date is delayed by 8 years. The delays highlight the potential importance for ozone depletion of any bank produced and its subsequent emission into the atmosphere. The delay in the stratospheric Cl_y return date is larger still if co-production of CFC-12 is considered, with the stratospheric Cl_y return date being delayed by 14 years in the SCEN3_30 scenario considered here.

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10 4 Modelled total column ozone response

Figure 4 shows the annual mean total column ozone (TCO) data from the BASE simulation (grey line) from 1960-2100, averaged over 60°S-60°N. Consistent with previous studies, TCO values decrease sharply from 1980 to the late 1990s as a result of increasing stratospheric chlorine loadings, before gradually increasing throughout the 21st century. Superimposed on these long-term trends is an 11-year oscillation resulting from the solar cycle. Observed annual mean TCO values from version 2.8 of the Bodeker Scientific total column ozone dataset (Bodeker et al., 2005) are shown in purple. There is generally good agreement between modelled TCO values and the Bodeker dataset; decadal total column ozone changes, the response of column ozone to the solar cycle and the magnitude of interannual variability are all well captured by the model throughout the time period during which the observations and model data overlap.

As discussed by Keeble et al. (2018; following WMO, 2007; Chipperfield et al., 2017), three stages of ozone recovery can be identified: (i) a slowed rate of decline and the date of minimum column ozone, (ii) the identification of significant positive trends and (iii) a return to historic values. Here, we focus on the impact of uncontrolled CFC-11 emissions on the last of these metrics: return to historic values, with the historic baseline value defined as the total column ozone average from 1960-1980.

This definition of the baseline period was chosen to avoid any sensitivity of the TCO return date to the choice of any individual year (which may be anomalously high or low). In actuality, global (60°S-60°N) TCO averaged from 1960-1980 in the BASE simulation is 298.1 DU, and the values for 1960 and 1980 are 297.5 DU and 298.0 DU respectively, and so there is little difference between the timing of global TCO return to these three values. However, differences do occur in regions with high interannual variability (e.g. the Arctic).

As discussed in previous studies (e.g. Salby et al., 2011; Ball et al., 2018), the identification of total column ozone recovery is complicated by the effects of interannual variability. To mitigate these impacts, the effects of natural processes (such as volcanic eruptions, the QBO, ENSO and solar cycle) can be removed from the data using statistical techniques (such as Multiple Linear Regression, e.g. Staehelin et al., 2001; WMO, 2007 or the Time series Additive Model, Scinocca et al., 2010).

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or the data can be smoothed by averaging across multiple years (e.g. Dhomse et al., 2018). Following Dhomse et al. (2018), annual mean UM-UKCA data is smoothed using an 11-point boxcar smoothing to reduce both the effects of short-term variability and the signature of the 11-year solar cycle. The smoothed data are shown on figure 4 as the black line. For the analysis in subsequent sections, all data from the BASE and SCEN simulations are smoothed using this method.

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5 Ideally, to provide uncertainty estimates for our various integrations, a multi-member ensemble would be run for the BASE and each SCEN calculation. However, in order to explore the largest possible range of future CFC-11 emissions scenarios, only one integration was performed for each scenario. Instead, data from a separate 5-member ensemble of 1980-2080 transient simulations (see Bednarz et al, 2016) is used to provide a rough estimate of the uncertainty range of return dates. These simulations used the same model configuration as used for our BASE and SCEN calculations but were forced with the older WMO (2011) CFC-11 and CFC-12 LBCs. Uncertainty estimates were estimated simply by identifying the ensemble member with the largest difference in return date from the ensemble mean, and defining this difference as the uncertainty range (e.g. for 5 ensemble members where the return dates might be 2061, 2062, 2060, 2061 and 2066, the mean return date would be 2062 and the largest deviation is 2066, so the uncertainty estimate would be ± 4 years). This provides an approximate indication of uncertainty for all latitude ranges considered here, except for the latitude range 90°S-60°S, where TCO return to the 1960-1980 mean occurred after 2080, and so lies outside the range of the ensemble.

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For the BASE integration under the WMO (2014) scenario, a global (60°S-60°N) TCO return date of 2054 \pm 2 years was calculated, with the return dates for other latitudes shown in Table 1. In the next section, the impact of CFC-11 and CFC-12 emissions from the different SCEN scenarios on these return dates is assessed.

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4.1 Global total column ozone

20 Figure 5 shows smoothed TCO values for the BASE and SCEN simulations, averaged from 60°S-60°N. All simulations show return to the baseline period of the 1960-1980 average between 2054 and 2064 (return dates provided in Table 1). The BASE simulation, which adopts the WMO (2014) LBC for ODSs and as such assumes the lowest anthropogenic Cl_y emissions, returns to the 1960-1980 mean, the earliest, in 2054, as discussed above. The impact of the additional CFC-11 and CFC-12 production scenarios investigated here is to delay the return date. For SCEN1_7 and SCEN1_15, the delay is small and within the range of return dates calculated from the 5-member ensemble. Only SCEN1_30 of the SCEN1 scenarios shows a significant delay in return date of 3 years. In contrast, for the SCEN2 simulations, which assume the creation of a new bank and subsequent emissions of CFC-11 from that bank, substantial delay in the return date of global TCO is modelled in both the SCEN2_15 and SCEN2_30 simulations (4 and 7 years, respectively). Only in the case that the assumed production of 90 Gg/yr stops this year (2019) is no significant delay in the return of TCO values to the baseline period modelled. The SCEN3 scenarios, which 30 assume the co-production of CFC-12 alongside CFC-11, all show larger delays in the return date, being 10 years for SCEN3_30.

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4.2 Regional total column ozone

Regional TCO projections for the BASE and SCEN simulations are also shown in figure 5, from the Antarctic to the Arctic, and TCO return dates for each region is given in Table 1. Annual mean TCO values over Antarctica (90°S-60°S) return to the 1960-1980 average by 2084 in the BASE simulation, 30 years after the global (60°S-60°N) TCO average is expected to return.

5 Substantial further delays in the date at which Antarctic TCO returns to the 1960-1980 mean are modelled for a number of the SCEN simulations. Again, the impact of the SCEN1 scenarios, which assumes no newly created bank, is modest with no delay modelled, except for SCEN1_30, which returns to the 1960-1980 average in 2088. In contrast, large delays are modelled for both the SCEN2_15 and SCEN2_30 simulations, which have projected return dates of 2093 and 2095 respectively. If co-production of CFC-12 is considered, SCEN3_15 and SCEN3_30 suggest that 11-year averaged TCO values will not return to

10 the 1960-1980 baseline period by the end of the 21st century. In the SH midlatitudes (60°S-30°S), delays in the return date are modelled for a number of SCEN simulations. If production stops in 2019 there is essentially no delay, while scenarios with higher emissions or longer duration lead to delays between 6 (SCEN2) and 9 years (SCEN3_30, which includes co-production of CFC-12).

15 While most SCEN simulations project a delay in the return of Antarctic and SH midlatitude annual mean TCO to the 1960-1980 mean, SCEN1_7, SCEN1_15 and SCEN3_7 all have earlier return dates than the BASE simulation. In the case of SCEN1_7 these changes in the SH midlatitudes are outside the model range calculated from the Bednarz et al (2016) 5-member ensemble, occurring 4 years earlier than in the BASE simulation. This may be because the uncertainty estimates calculated here from the 5-member ensemble do not fully capture the true system uncertainty, or that atmospheric chemistry-climate feedbacks may result in increased TCO values in some locations despite increased stratospheric ODS values. For example,

20 Keeble et al. (2014) show modelled wintertime TCO increases in the northern midlatitudes resulting from increased polar ozone depletion and associated changes in the lower branch of the BDC.

The observed ozone loss in the tropics has been small and, furthermore, future changes in the tropics are driven both by reductions in the stratospheric abundance of halogens, which tend to increase ozone, and the strengthening of the Brewer-Dobson circulation, which tends to decrease column ozone (e.g. Oman et al., 2010; Eyring et al., 2013; Meul et al., 2014;

25 Keeble et al., 2017; Chiodo et al., 2018). Here, tropical (30°S-30°N) TCO values are projected to return to the 1960-1980 mean by 2057 in the BASE simulation, and all SCEN simulations show significant delays to this return date except for SCEN1_7 and SCEN2_7 (i.e. those simulations which assume that uncontrolled production of CFC-11 stops in 2019, and there is no co-production of CFC-12). While TCO values are projected to return to the 1960-1980 average for the broad definition of the tropics used here, if a narrower definition is used (e.g. 5°S-5°N), TCO values do not return to the 1960-1980 average at any

30 point in the 21st century. This is consistent with the impacts on lower stratospheric ozone of the increased speed of the BDC resulting from anthropogenic GHG changes (e.g. Oman et al., 2010; Eyring et al., 2013; Meul et al., 2016; Keeble et al., 2017). In the NH midlatitudes (30°N-60°N), TCO under the BASE projection is modelled to return in 2047, considerably earlier than the SH midlatitudes. As at other latitudes, significant delay in the return date is modelled in the majority of SCEN simulations.

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In the Arctic (60°N-90°N), annual mean TCO values are projected to **return to the 1960-1980 mean** in the BASE simulation in 2048, again substantially earlier than the Antarctic **return date**. While significant delays for Arctic ozone are modelled in the majority of SCEN simulations, unlike at other latitude ranges, the latest **return dates** are not associated with the SCEN3 simulations, which assume co-production of CFC-12. Instead, the latest **return date** of 2057 occurs in the SCEN2 30 simulations. We ascribe this to the very high dynamical variability of the Arctic polar vortex, its subsequent impact on total column ozone, and the large uncertainties in defining **a return date** in this region. Bednarz et al. (2016), also using the UM-UKCA model, showed that, although springtime Arctic ozone was projected to return to 1980 values by the late 2030s, episodes of dynamically-driven very low ozone could be found well into the second half of this century, consistent with our annual mean results presented here.

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5 Identifying scenario-independent relationships between CFC emissions and TCO **return to the 1960-1980 mean**,

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While the SCEN simulations used in this study were designed to cover a broad range of potential future CFC-11 production scenarios, it is unlikely that future CFC-11 emissions will follow any of the scenarios described here. Therefore, we aim here to identify scenario-independent relationships between future CFC emissions pathways and the impact on TCO. For example, Dhomse et al. (2018) found relationships between **the chlorine return date**, and a number of indicators of ozone recovery for a range of models (see, e.g., their Figure 8). In this study this relationship is further explored by linking TCO differences to emissions. Three emerging relationships are explored in the following sections: i) the timing of **the global TCO return date** as a function of Cl_y return date; ii) the magnitude of annual mean TCO depletion in a year as a function of the cumulative CFC emissions up to that year; and iii) the **TCO return date** as a function of the cumulative additional CFC emissions by the end of the simulation.

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5.1 Cl_y return date vs. TCO return date,

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The future evolution of stratospheric ozone mixing ratios follows closely the evolution of stratospheric Cl_y (e.g. WMO, 2018). Dhomse et al. (2018) found correlations between modelled return dates of stratospheric chlorine and ozone recovery dates for Antarctic and Arctic spring across a range of CCMI models. A similar relationship between the timing of Cl_y **return to 1980 values** and the timing of global TCO **return to the 1960-1980 mean** is identified in the SCEN simulations performed as part of this study, shown in Figure 6. As discussed above, the Cl_y **return date** is defined as the date at which Cl_y mixing ratios at 40 km return, averaged from 10°S-10°N, **return** to their 1980s value. The relationship between the global TCO **return date** and **the Cl_y return date** is robust, with an r² of 0.92, and indicates that for every year Cl_y **return** is delayed, the TCO **return date** is delayed by 0.64 years (given by the gradient of the linear fit through the points). The Cl_y **return date** itself is strongly linked to the assumed emissions. A robust linear relationship, with an r² value of 0.96, was identified between the total cumulative additional Cl emissions and the delay in Cl_y return date. This relationship indicates that, for every additional 200 Gg of Cl (258 Gg CFC-11 equivalent) emitted by 2099 above those implied in the WMO (2014) scenario, **the Cl_y return date is** delayed

by 0.86 years. It should be noted that the Cl_y return date occurs ~5 years later on average than the global TCO return date in the BASE and SCEN simulations (see Table 1), indicating that even after the time TCO values have returned to the 1960-1980 average, stratospheric chlorine mixing ratios remain elevated above the 1980 value.

5.2 TCO depletion vs cumulative emissions

5 In order to explore the magnitude of annual mean TCO depletion in any year as a function of the cumulative Cl emissions up to that year, results from the UM-UKCA SCEN simulations are supplemented by simulations performed with the TOMCAT chemistry transport model (CTM; Chipperfield et al., 2017). Both models have full stratospheric chemistry schemes but are independent of one another. The control CTM simulation (CTM_C) was performed until 2080 with repeating year 2000 meteorology and time-dependent future source gas surface mixing ratios. Two further simulations (described in detail in
10 Dhomse et al. (2019)) were performed with additional future CFC-11 emissions (i) at constant 67 Gg yr⁻¹ (CTM_Fix) and (ii) including simulation of a bank and production decreasing to zero over 10 years (CTM_Bank). Note that simulation CTM_Bank gives larger emissions than CFM_Fix until about the year 2040.

Figure 7 shows, for both the UM-UKCA and TOMCAT models, the cumulative additional Cl emissions plotted against the additional TCO depletion driven by the increased emissions compared with a base integration, averaged over 60°S-60°N. UM-
15 UKCA values are calculated as the difference between each SCEN simulation with respect to the BASE simulation, while TOMCAT values are calculated as CTM_Fix - CTM_C and CTM_Bank - CTM_C. For both models, all available years from each scenario are plotted, with each point representing a single year and showing the TCO difference between the scenario and base simulation for that year plotted against the cumulative Cl emissions reached by that year.

The estimated TCO depletion from both UM-UKCA and CTM simulations follows a reasonably compact, linear relationship
20 with cumulative emissions. The CTM simulations use identical meteorology, without any chemistry-climate feedbacks, and so a comparison of two simulations gives a direct measure of the chemical ozone impact of the additional emissions. In contrast, UM-UKCA generates consistent time-varying meteorology for each of the SCEN simulations and so includes both a forced response and internally generated natural variability. Despite this difference, the modelled relationship between cumulative emissions and global ozone depletion is remarkably similar for the two models: a total emission of 3500 Gg (Cl) gives a near-
25 global mean ozone decrease of between 2.5-3 DU. The agreement between the two models provides confidence in the diagnosed chemical impact of the CFC-11 emissions and shows a means by which results from different model simulations, with different CFC-11 emissions scenarios, can be compared. Further, it allows an estimate to be made of the impact of any given CFC-11 emission scenario on TCO depletion, which can in turn be applied to the real-world future emissions of CFC-11, which remain uncertain.

30 In CTM run CFC_Bank the maximum CFC-11 emissions by 2080 are around 2700 Gg CFC-11. By this stage a significant part of the large initial pulse of emissions (see Dhomse et al 2019) would have been removed from the atmosphere (consistent with its ~55 year lifetime) so that the CTM line deviates from the simple linear fit (black line Figure 7). This is also true for

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the SCEN simulations performed with UM-UKCA, although the effects are small compared to the dynamically driven variability. The SCEN1 simulations are not shown in Figure 7 as their cumulative emissions stop increasing when the direct emissions stop (see Figure 1) and so for a large period of the simulations the differences are dominated by variability.

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5.3 TCO return date vs total cumulative emissions

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- 5 Figure 8 shows the relationship between the total cumulative additional Cl emissions, in Gg, and projected global TCO return date in the BASE and SCEN simulations. These values represent the additional Cl emissions assumed for each scenario in addition to the WMO (2014) A1 scenario. A robust ($r^2=0.93$) linear relationship is found, indicating that for every extra 200 Gg of Cl emitted, global TCO return to the 1960-1980 mean is delayed by 0.56 years. Repeating this analysis for 10-degree latitude bins gives the latitudinal dependence of the impacts of cumulative Cl emissions on TCO return date (Figure 9).

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- 10 Uncertainty estimates are calculated for the regression using the standard error of the estimate, given as $\sigma_{est} = \sqrt{\frac{\sum(y-y')^2}{N}}$

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- 15 where y is the TCO return date, y' is the predicted TCO return date from the regression mode, and N is number of simulations. For every 200 Gg of Cl emitted, the date of TCO return to the 1960-1980 mean is delayed by between 0.18-1.60 years, with a marked hemispheric asymmetry evident in the response. Large delays (~1.6 years per 200 Gg Cl) are modelled in the Antarctic, where heterogeneous processing of chlorine reservoirs on polar stratospheric cloud (PSC) surfaces allows for large ozone depletion even for relatively small chlorine concentrations. On annual mean timescales these low values mix into the southern hemisphere midlatitudes, resulting in larger delays to the return date there (0.5-0.8 years per 200 Gg Cl) than in the northern hemisphere midlatitudes and Arctic (0.1-0.5 years per 200 Gg Cl), where the effects of PSC processing are less pronounced due to the higher temperatures and high dynamical variability of the Arctic polar vortex. No values are given south of 80°S, as TCO values do not return to the baseline period by the end of the simulations.

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- 20 The delay in tropical TCO return date is also long (0.5-1.3 years per 200 Gg Cl) and associated with large uncertainties. As discussed above, the observed depletion of ozone in the tropics has been very small (WMO, 2018) and quantifying TCO recovery is complex, depending on not just the ODS loading of the stratosphere but on other factors including the future levels of greenhouse gases, changes to tropospheric ozone and the projected acceleration of the Brewer-Dobson circulation (e.g. Butchart, 2014; Meul et al., 2016).

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25 6 Discussion and conclusions

The Montreal Protocol has been successful in reducing emissions of ODSs into the atmosphere, which in turn has led to the onset of ozone recovery. However, recent observational evidence indicates that atmospheric mixing ratios of CFC-11 are declining at a slower rate than that expected under full compliance with the Montreal Protocol. It seems likely that emissions resulting from new production, in contravention of the Montreal Protocol, are the likely cause of the change in decline rate,

with an important source in east Asia (Montzka et al., 2018; Rigby et al., 2019). However, there remain large uncertainties associated with these emissions: their source remains unidentified, changes to the release rate from the historical bank are unknown, the size of any newly created bank is uncertain and undetected co-production of other chlorinated ODSs is possible.

Given these uncertainties, the impact of a range of plausible future CFC-11 emissions scenarios on the timing of TCO return to the 1960-1980 mean, a key milestone on the road to ozone recovery, was explored using the UM-UKCA model. Making a range of assumptions the scenarios are intended to cover a breadth of future emission pathways. We consider the size of emissions and their duration; we compare emissive versus non-emissive use (where in the latter the bank is enhanced) and we consider possible co-production of CFC-12. While none of the scenarios developed here is expected to accurately predict the future CFC-11 emissions pathway of the real-world, they provide a sensitivity range to guide possible future trajectories of the ozone layer.

If the recently identified CFC-11 emissions result from an emissive use (i.e. there is no new bank being created and estimated emissions are equal to the total production) then, provided the source of new CFC-11 production stops within the next decade, results from the SCEN1 scenarios indicate that there will be no significant delay in the return of global total column ozone to the 1960-1980 baseline. Only in the case of prolonged emissions would significant delay in the return date of global column ozone be expected.

However, if the recently identified CFC-11 changes result from non-emissive use, results from the SCEN2 scenarios indicate that, unless stopped immediately, the production has the potential to delay global total column ozone return to the 1960-1980 mean by up to 7 years, depending on the duration of the production and the size of the annual increment to the bank. Further, results from the SCEN3 scenarios suggest that if CFC-12 has been co-produced with CFC-11, then global total column ozone return could have already been delayed by 4 years, and that if, under the assumptions made here, production continues for up to 30 years from 2012, the total column ozone return date may be delayed by a decade.

Our results are, of course, dependent on the assumptions made in each of the SCEN scenarios. Therefore, it is important to identify scenario-independent metrics which can be used to establish the impact of future CFC emissions pathways on TCO return date. Three such relationships were identified: i) global TCO return date as a function of Cl_y return date; ii) the magnitude of annual mean TCO depletion in a year as a function of the cumulative CFC emissions up to that year; and iii) global TCO return date as a function of the total cumulative additional CFC emissions by the end of the simulation. The second of these relationships was further verified by comparing with results from the TOMCAT CTM, and despite differences between the assumed emissions scenarios used by both models, and the fundamental differences in the treatment of meteorology, good agreement was found between the two models, with 2.5-3 DU TCO depletion occurring for an additional 3500 Gg Cl. The robust linear relationship found between the total cumulative additional Cl emissions and the global TCO return date indicates that for every 200 Gg of Cl (~258 Gg CFC-11) emitted, global TCO return to the 1960-1980 mean is delayed by 0.56 years. However, a marked hemispheric asymmetry in the impacts of cumulative Cl emissions on the TCO return date at particular latitudes was identified, with longer delays in the southern hemisphere than the northern hemisphere for the same emission.

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While these scenario-independent relationships are useful, they come with some caveats. All the scenarios developed in this study assume that new CFC-11 production began in 2012, and that despite new CFC-11 production, atmospheric CFC-11 mixing ratios continue to decline, consistent with the observations presented by Montzka et al. (2018). ~~It is not expected that an emission of CFC-11 emitted in 2020 would have the same impact on the ozone return date as the same emission of CFC-11 emitted in, for example, 2080. This is due in part to the different background stratospheric temperatures, circulation, and sinks of active chlorine (e.g. the conversion of ClO_x to HCl through reaction with CH₄) at different times throughout the 21st century. Furthermore, any additional chlorine emissions which occur after TCO has returned to its 1960-1980 mean value might not deplete ozone below this value, and so would not affect the return date. Additionally, while the simulations analyzed in this study highlight the role of CFC-11 emissions on stratospheric ozone recovery, coupling between the ClO_x, BrO_x, NO_x, HO_x and O_x catalytic cycles means that not all the changes to the timing of the ozone return date modelled here are solely due to stratospheric chlorine changes. Dameris et al. (2019) highlight that increases in surface CFC-11 concentrations lead to increased ozone depletion through reactions of ClO_x and BrO_x, but relatively decreased depletion through O_x, NO_x and HO_x reactions. Due to these factors, in addition to the cumulative total, the temporal evolution of CFC-11 emissions is likely an important control on the relationships identified in this study. In this study we assume steady changes in emissions, consistent with a continuous anthropogenic source of additional CFC-11, rather than changes which might be large but sporadic. As such, while the relationships identified here likely give a good indication of the TCO response to the recently identified source of CFC-11, they may not prove robust for any unexpected CFC-11 emissions later in the century.~~

The detection of the change in the rate at which CFC-11 concentrations are declining in the atmosphere, and the inferred change in emissions, are important contributions to the Montreal Protocol during its accountability phase, during which the impact of the Protocol on the atmosphere is being assessed. It is clear that long-term monitoring of ODSs, as well as ozone, is an absolutely critical component of the atmospheric science response to the Protocol and its input to policy negotiations. Continued modelling of the impact of these emissions on the projected timing of the TCO return date is also required.

Results presented here highlight the need for rapid action in tackling any uncontrolled production of CFC-11. Unless emissions are stopped rapidly, we anticipate potentially significant delays in recovery. ~~The date at which global TCO returns to its 1960-1980 mean could be delayed by about a decade, on the basis of our assumed emissions, and Antarctic ozone might not recover at all this century. New knowledge concerning the nature of the ODS emissions is required, which, in concert with increased atmospheric measurements of the ODSs, can inform the on-going discussions of the Montreal Protocol and ensure its future success.~~

Data availability

Data from all simulations are available on the UK Research Data Facility (<http://www.archer.ac.uk/documentation/rdf-guide>). v2.8 of the Bodeker dataset can be found at <http://www.bodekerscientific.com/data/total-column-ozone>, while v3 of the

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Competing interests

The authors declare that they have no conflict of interest.

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	Cl _v return date	Date of total column ozone (TCO) return to 1960-1980 mean					
	10°S-10°	60°S-60°	90°S-60°S	60°S-30°S	30°S-30°N	30°N-60°N	60°N-90°N
Base	2058	2054±2	2084±?	2065±1	2057±3	2047±2	2048±3
SCEN1_7	2059	2055	2078	2061	2060	2050	2050
SCEN1_15	2057	2055	2082	2064	2062	2050	2052
SCEN1_30	2061	2057	2088	2068	2063	2049	2048
SCEN2_7	2060	2055	2084	2065	2060	2047	2049
SCEN2_15	2062	2058	2093	2071	2065	2051	2052
SCEN2_30	2066	2061	2095	2071	2069	2052	2057
SCEN3_7	2061	2058	2081	2064	2063	2052	2054
SCEN3_15	2064	2058	No return	2068	2066	2053	2053
SCEN3_30	2073	2064	No return	2074	2081	2052	2052

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5 **Table 1** Cl_v return date (defined as the date at which Cl_v mixing ratios, averaged from 10°S-10°N at 40 km, return to the BASE simulation 1980 value) and the date TCO returns to the 1960-1980 mean in the BASE and SCEN simulations. Uncertainty estimates for the TCO return date in the BASE simulation are provided by a separate 5-member ensemble, as described in the text. TCO return to the 1960-1980 mean occurs after 2080 for the latitude range 90°S-60°S, and so no uncertainty estimate can be provided for this latitude band, denoted by the '±?' in the table.

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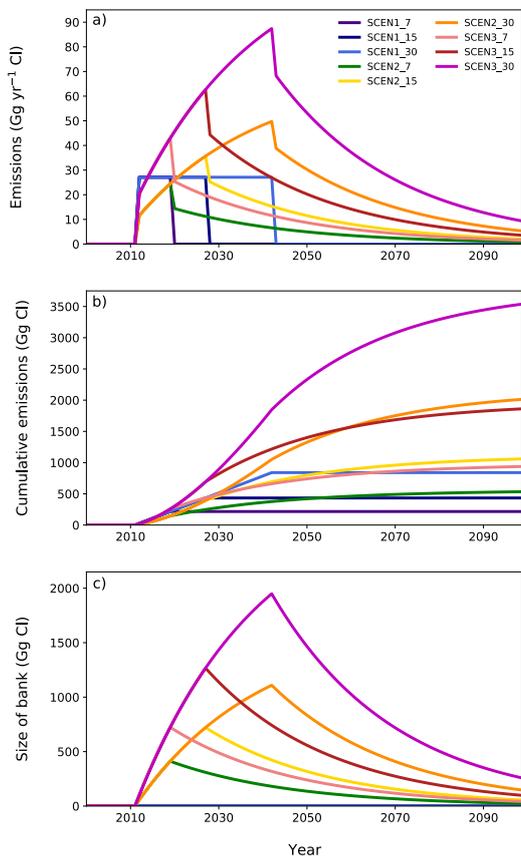
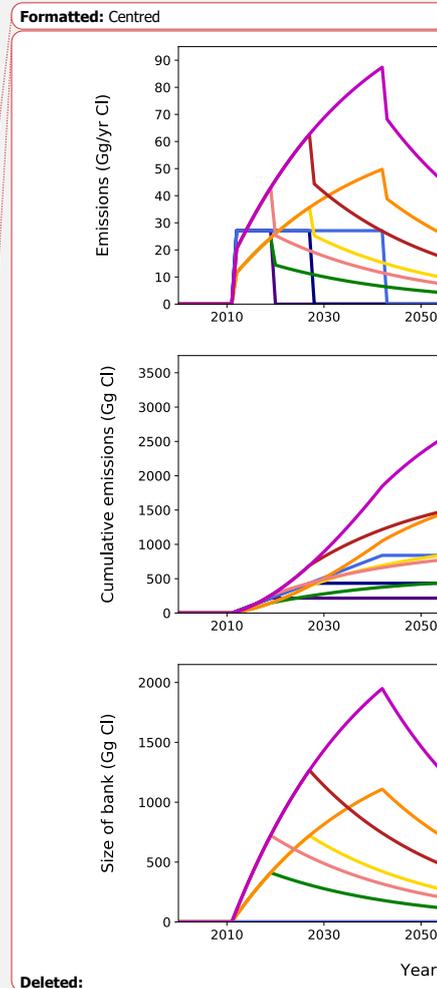


Figure 1 (a) Additional emissions (in Gg Cl yr⁻¹) assumed for each of the SCEN simulations on top of those already assumed by the WMO (2014) A1 scenario. For all scenarios, the additional, uncontrolled source of emissions is assumed to start in 2012. (b) Cumulative additional emissions for each of the SCEN simulations. (c) The size of the newly created bank for the SCEN2 and SCEN3 scenarios.



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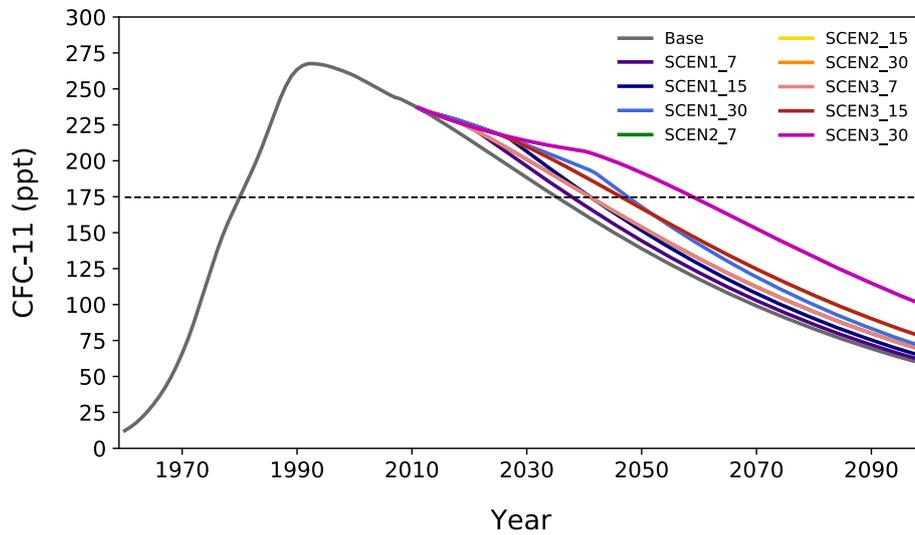
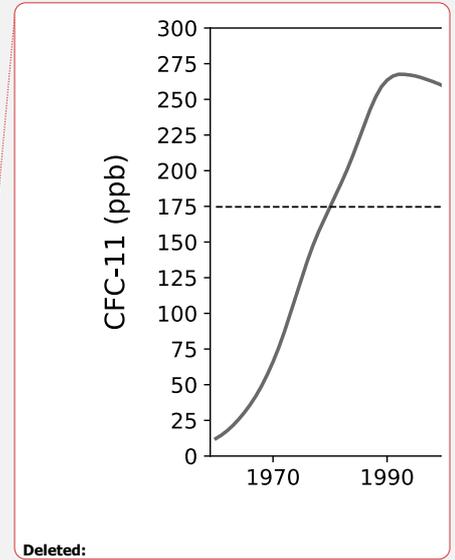
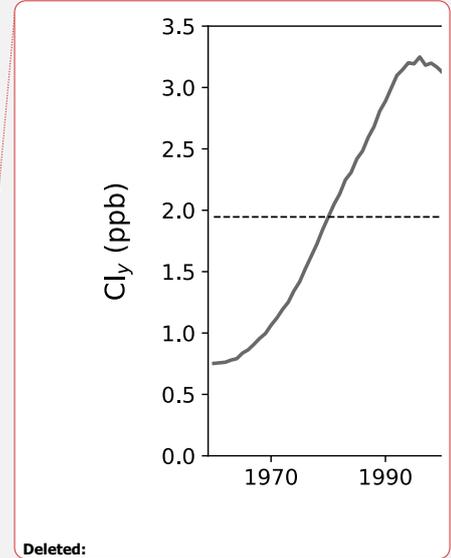
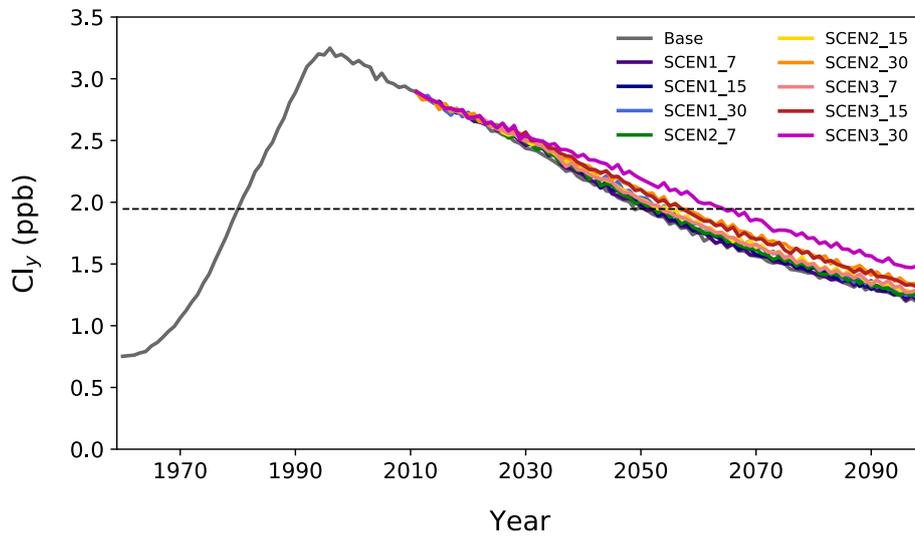


Figure 2 Prescribed global mean, annual mean surface lower boundary mixing ratio of CFC-11 used for the BASE and SCEN simulations. The dashed line represents the 1980 CFC-11 surface mixing ratio in the BASE simulation. Note that the SCEN2 and SCEN3 simulations have the same CFC-11 lower boundary conditions.



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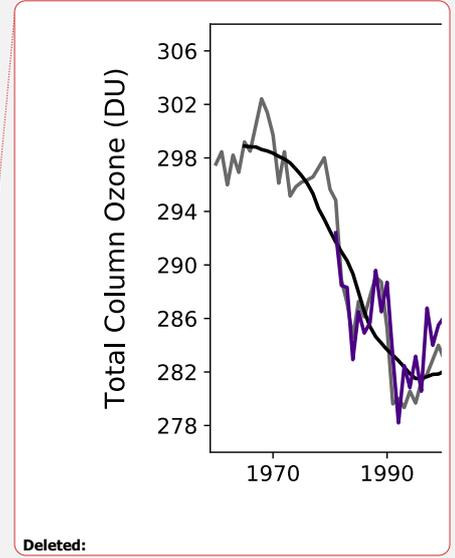
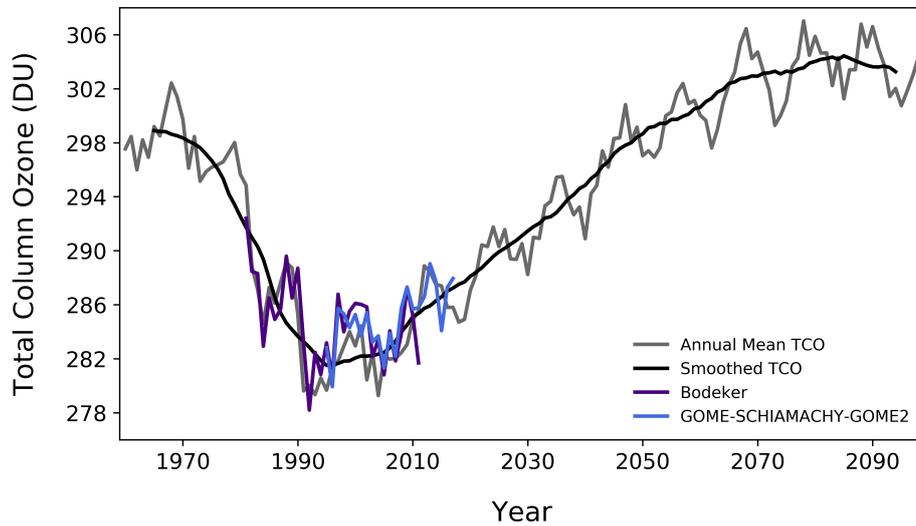


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Figure 3 Modelled annual mean 40 km inorganic chlorine (Cl_y) mixing ratios, averaged from 10°S-10°N, for the BASE and SCEN simulations. The dashed line represents the 1980 Cl_y mixing ratio at 40 km in the BASE simulation, used as the value

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Figure 4 Modelled annual mean TCO values (in DU), averaged from 60°S-60°N, for the BASE (grey line) and smoothed data (black line) using an 11-point boxcar smoothing to reduce both the effects of variability (following Dhomse et al., 2018). Also shown are TCO values, averaged over 60°S-60°N, from v2.8 of the Bodeker dataset (purple line; Bodeker et al., 2005) and v3
 5 of the GOME/SCIAMACHY/GOME2 Merged Total Ozone dataset (blue line; Weber et al., 2011).

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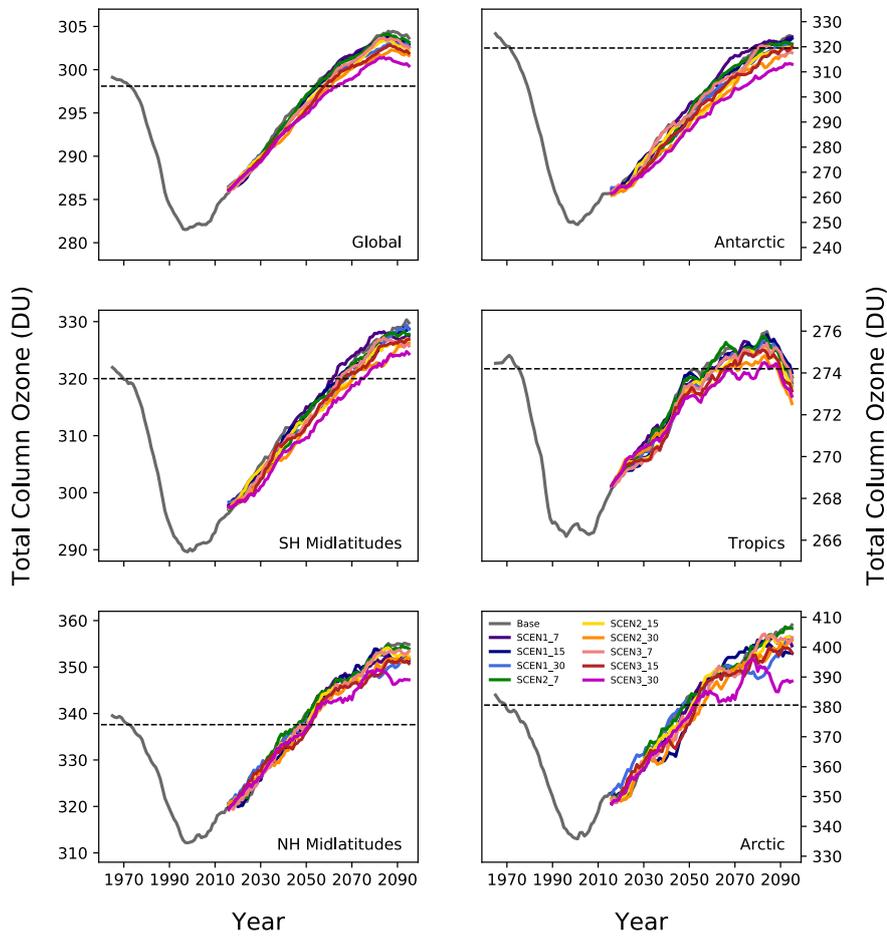
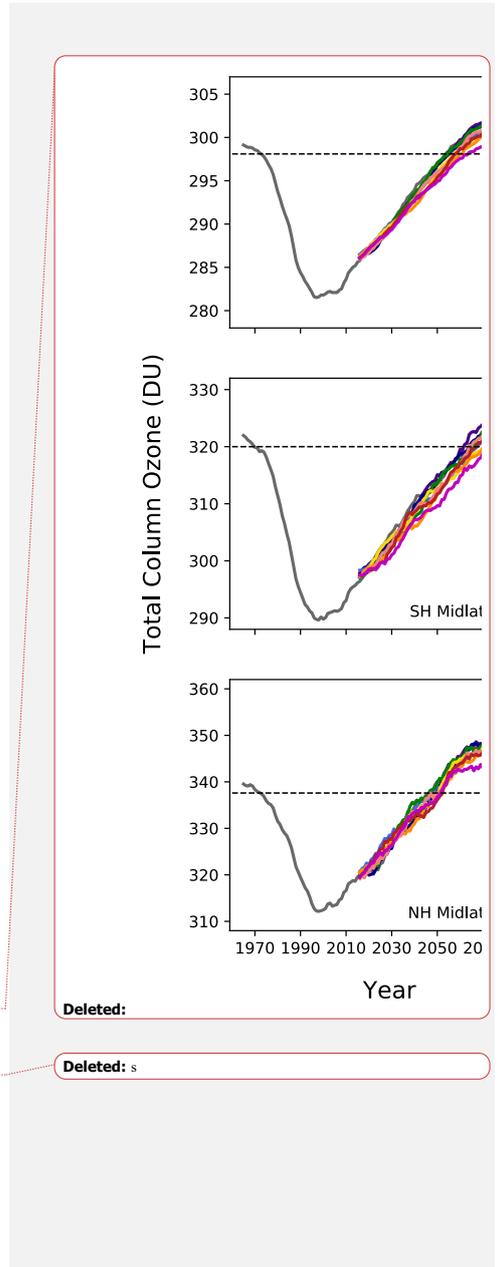


Figure 5 Smoothed TCO (in DU) for the BASE and SCEN simulations, for the global (60°S-60°N) and regional column values. Dashed line denotes the 1960-1980 baseline period, used as the value for calculating the TCO return date.



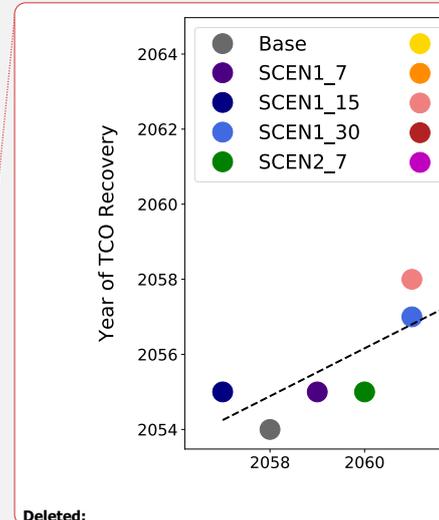
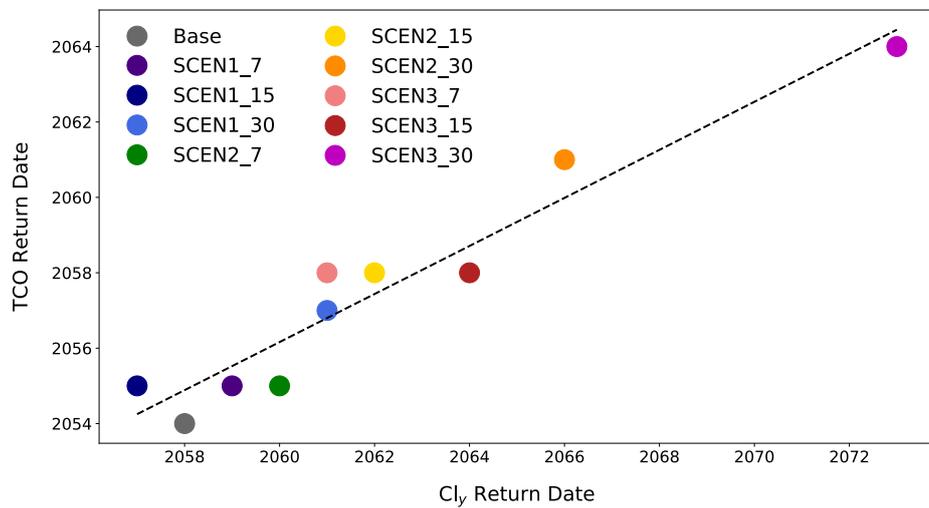
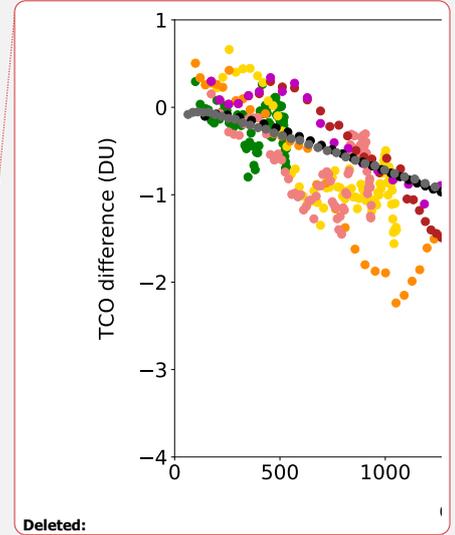
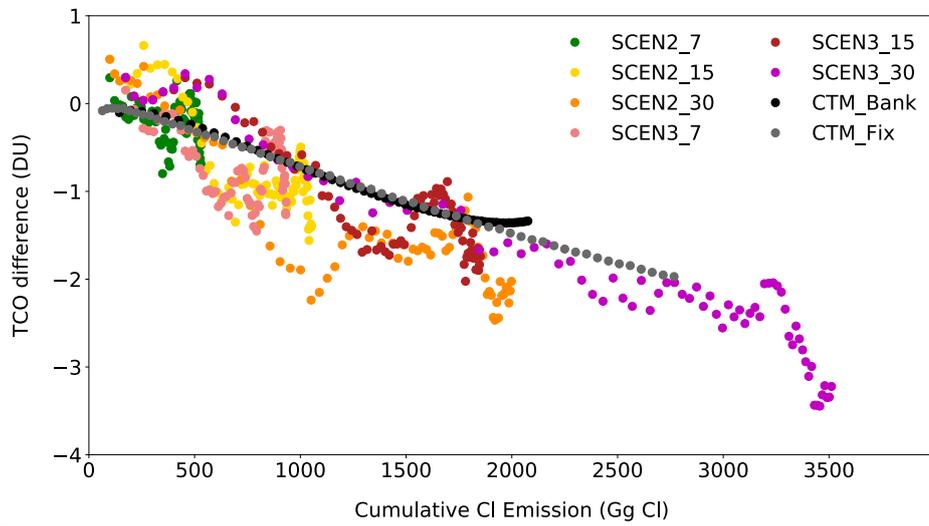


Figure 6 TCO return date (defined as the date at which 60°S-60°N averaged TCO values return to the 1960-1980 average) for the BASE and SCEN simulations versus Cl₂ return date (defined as the date at which Cl₂ mixing ratios at 40 km, averaged from 10°S-10°N, return to the 1980 BASE value). The dashed line gives the linear fit through the points.

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Figure 7 Annual mean TCO difference (in DU) for the UM-UKCA SCEN2 and SCEN3 simulations with respect to the BASE simulation, averaged over 60°S-60°N vs cumulative emissions (Gg Cl) from 2012 to that year. Grey and black points show the same, but are for the CTM_Fix and CTM_Bank simulations performed with the TOMCAT CTM. For these simulations, TCO differences are calculated with respect to the CTM_C baseline simulation. TOMCAT simulations described in Dhomse et al. (2019).

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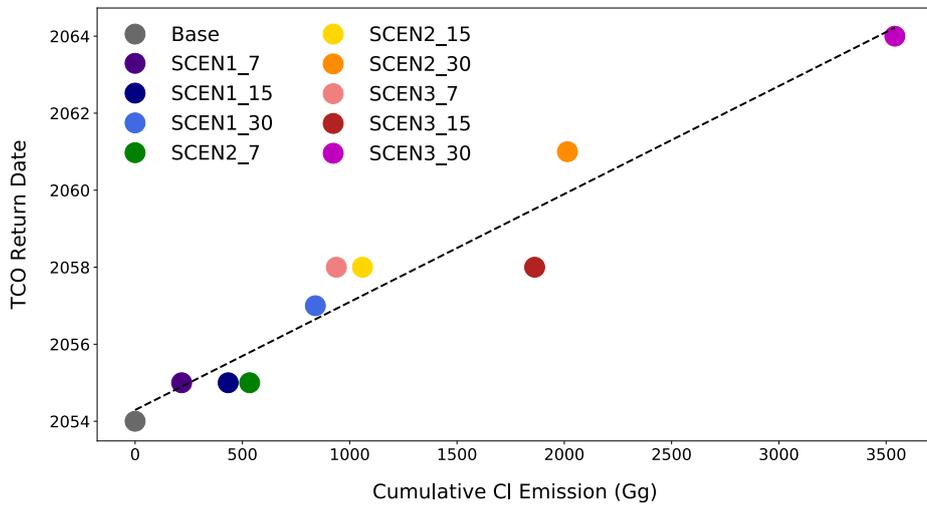
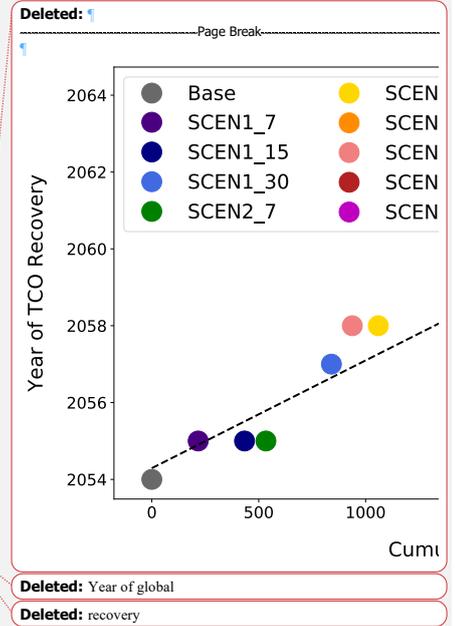


Figure 8 TCO return date (defined as the date at which TCO values return to the BASE 1960-1980 average) for the BASE and SCEN simulations vs cumulative additional emissions (Gg CI) from 2012 to the end of the simulation. The dashed line gives the linear fit through the points.



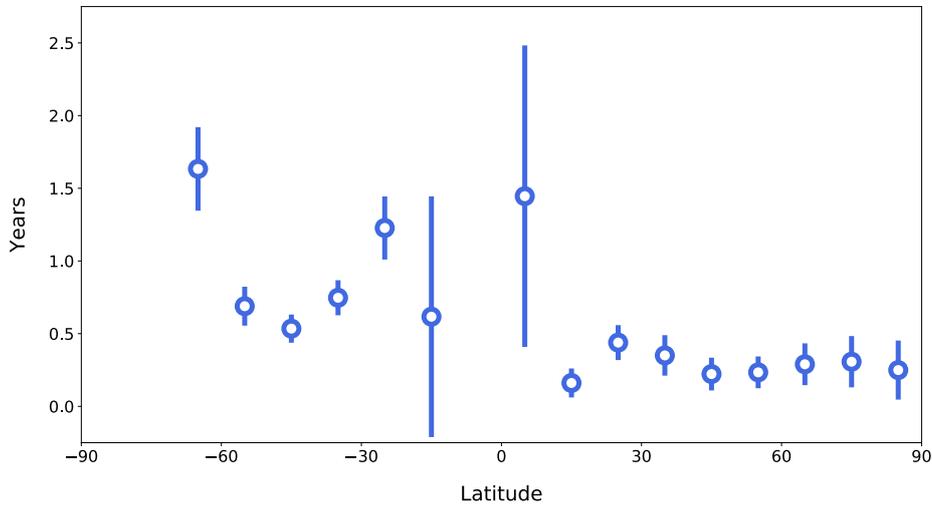


Figure 9 Delay in TCO return date (in years) per 200 Gg Cl emissions for 10-degree latitude bins. Uncertainty bars represent the standard error of the estimate (calculation method provided in text). No values are given for 90°S-80°S, 80°S-70°S and 10°S-0°N as TCO in these latitude bands does not return to the 1960-1980 mean by the end of the model simulations.

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