

In the following, editor comments (EC) are reproduced (black) along with our replies AC (blue) and changes made to the text (red) in the revised manuscript.

Editor comments:

However, the presentation of the work falls below normal publication standards. The overall quality and clarity of the presentation greatly hinders the communication of the scientific results.

On the basis of the reviewer comments and my review, I am requesting that the authors make major revisions to their manuscript to improve the communication, clarity, logic, (remove) unnecessary repetition ("As mentioned above" type text), (remove) meaningless general statements, English, and grammar. In addition, the presentation of the scientific results for the molecules under study could be made more concise and earlier to follow for the reader. The tables and figures seem fine.

AC: We thank the editor for the interest shown on our work and the comments and suggestions.

The manuscript has been revised and unnecessary repetitions have been removed. We expect that with the english revision the meaningless statements have been eliminated.

The authors have tried to make a presentation of the scientific results more concise but we consider that this is a work with many experiments and results, and it is difficult for us to reduce the work without loss of scientific rigor. We know that in the kinetic discussion section there are information relative to general topics of reactivity of alcohols that could be omitted (page 6 lines 28-38 and page 7, lines 1-3 of new manuscript.pdf) because this information is well established and known. The authors consider that given the scope and dissemination of the ACP, more details would help to anyone who was not an expert in the subject to understand better the work. Indeed, at the suggestions of both referees, more information and explications had to be included. However, if you consider that this information is not necessary to understand better the discussion of the kinetic results, we can eliminate it.

The authors would appreciate if you could tell us which parts should be reduced or otherwise should be discussed further. In our opinion we consider that this version is fine, although it is possible that there are still some grammatical errors, that would be eliminated in the final review before publication.

I highly recommend that the authors have a native English-speaking colleague critically (line-by-line) review the manuscript prior to re-submission. This is simply too large a task for a reviewer or editor.

AC: The text of manuscript has been sent to "Proof-Reading-Services.com" for american english revision in order to make a rigorous of /vocabulary/grammar/scientific expressions. In order to do not pay a lot of for the revision, abstract, references, tables, figures and Supplementary material have not been sent to english revision.

EC: A few general theme comments:

***Only one of the compounds included in this study (3,3DM2ButOH) is actually methyl saturated. Therefore, the title and text are in error. This correction would lead to the removal of the misleading MSA acronym.**

AC: All studied alcohols are saturated compounds (they do not have double or triple bonds) and all of them have at least one methyl group, so we considered methyl saturated alcohols to be a good acronym for the studied compounds, that is usual for compounds without multiple bonds. To avoid confusion, the authors have decided to use saturated alcohols generically (SAs)

EC: * The abstract mentions “tentative estimation of molecular yields” and quotes very large ranges of yields, which are not meaningful. The authors need to be more specific regarding the actual yield values and for what initiation reaction they are obtained from. I believe this information is actually available within the manuscript tables, although hard to follow in the text. Or, is this a problem with not having good standards?

AC: The individual molar yield of reaction product obtained for each reaction had not been included in the abstract in order to avoid making it very extensive. The phrase “A tentative estimation of molecular yields has been done obtaining the following ranges (25-60) % for 4-methylcyclohexanone, (40-60) % for 3,3-dimethylbutanal and (40-80) % for 3,3-dimethyl-2-butanone.” has been removed to avoid confusion, and some new sentences have been included.

“The main products detected in the reaction of SAs with Cl atoms (absence/presence of NO_x), OH and NO₃ radicals were: E-4-methylcyclohexanone for the reactions of E-4-methylcyclohexanol, 3,3-dimethylbutanal for the reactions of 3,3-dimethyl-1-butanol and 3,3-dimethyl-2-butanone for the reactions of 3,3-dimethyl-2-butanol”

“In addition, the molar yields of the reaction products were estimated”.

The molar yields of the products are shown in Tables 3-6. In some cases, as 2,2-dimethylpropanal the molar yields could have large uncertainty because the reference spectra used was an FTIR spectra of a similar compound (2-methylpropanal).

EC * There is a long discussion of reactivity observations that have been well-established through the development of structure activity relationships (SARs) by Atkinson and co-workers. The small data set from this work is probably not sufficient to revise our thinking of SARs. The present work needs to be placed in the proper perspective.

AC: In the analysis made of the kinetic results obtained in our work, it is observed that the OH group exerts an activating effect that makes the reactivity of the alcohols greater than that of its alkane homologue, but it is also observed that this activating effect is different according to the oxidant. Indeed, these observations are well established through the development of structure activity relationships (SARs) by Atkinson and co-workers.

The authors do not have the intention of reviewing or changing results of the SARs method. On the contrary, what we want to highlight is that our results are in good agreement with those established by the SARs method in the case of reactions of saturated alcohols with Cl atoms and OH radical and in the case of NO₃ radical with secondary alcohols.

In order to avoid more confusion, the paragraph of page. 7, lines 31-37 (clean manuscript date 26/10/2019) has been replaced by.

“..These results show that the activating effect of the OH group of the SA is less important for the Cl than with the OH, behavior that agrees with that established by the Structure Activity Relationship (SARs) methods (Kwok and Atkinson 1995; Calvert et al. 2011).”

On the other hand, in the case of the nitrate radical our study shows that in addition to the activating effect of the OH group there is also an activating effect of the -CH₂OH group and that

it is necessary to perform more kinetic studies of reactions of primary alcohols with the nitrate radical, in order to establish this factor, since there are currently very few data available in bibliography. This last is indicated in the manuscript as a conclusion.

“..For primary alcohols, the abstraction of a hydrogen atom in β -position could also be important in the reaction with NO_3 radical. Therefore, more kinetic studies for NO_3 radical with primary alcohols are necessary to quantify the effect of the OH group in β -position, (-CH₂OH) and to update the SAR method developed by Kerdouci et al.”

EC:* The introduction contains a great deal of seemingly unnecessary material and background information. Even with that said, it is not made clear why these particular compounds were chosen for study (biofuels?).

AC: The introduction has been reorganized in order to show better the relevance of compounds studied.

So, in the first paragraph the necessity of using biodiesel as alternative to conventional diesel is remarked. In this same paragraph, it is also shown that alcohol-diesel blends are a good alternative. Initially the alcohols used were alcohols of short chain (methanol, ethanol) but some problems were found due to their low cetane number, high latent heat of vaporization and high resistance to auto-ignition. In order to avoid these problems, high alcohols as propanol, n-butanol, isobutanol, n-pentanol and therefore the alcohols studied in this work, could be a good alternative as additives in the diesel blends. A new reference has been included to support this last.

“ Li, F., Yi, B., Song, L., Fu, W., Liu, T., Hu, H., & Lin, Q. Macroscopic spray characteristics of long-chain alcohol-biodiesel fuels in a constant volume chamber. *Proceedings of the Institution of Mechanical Engineers, Part A: JPE*, 232(2), 195–207. <https://doi.org/10.1177/0957650917721336>, 2017.”

In the second paragraph of introduction, a revision of sources of alcohols and data about concentrations found in the atmosphere is shown. The use of alcohols of long chain as additives for biodiesel fuel could imply an important source of these alcohols in the atmosphere. So, it is necessary to evaluate their atmospheric reactivity and to establish the atmospheric impact of these compounds.

In the next paragraph, a revision of the atmospheric reactivity of short and long alcohols is made, it does remark the absence of kinetic data or about reaction products of the alcohols studied in this work. The last paragraph explains the study that has been carried out and why it has been done.

We expect that this reorganization and correction of the introduction section, allows a better compression and better justification of the developed research.

LIST OF THE MAIN CHANGES MADE IN THE MANUSCRIPT.

Apart of the modification due to the English revision, the next modifications have been made at suggestion of the editor in order to more clarity. Pages and lines indicated are related to the manuscript.pdf with date sent of 26_10_2019.

1-Page 1. Lines 1 and 2. The title has been modified. The acronym MSA has been removed, and "methyl saturated alcohols" has been replaced by "saturated alcohols" in all text of the manuscript, tables and figures.

2-Page 1. Line 12. The next sentence has been included:

"These SAs are alcohols that could be used as fuel additives"

3-Page 1. Lines 17-21. The paragraph has been replaced by the following:

"The main products detected in the reaction of SAs with Cl atoms (absence/presence NO_x), OH and NO₃ radicals were: E-4-methylcyclohexanone for the reactions of E-4-methyl-cyclohexanol, 3,3-dimethylbutanal for the reactions of 3,3-dimethyl-1-butanol and 3,3-dimethyl-2-butanone for the reactions of 3,3-dimethyl-2-butanol"

4-Page 1. Line 21. The sentence has been modified including "..of Cl atoms and OH radicals with.."

5-Page 1. Line 23. The next sentence has been included "In addition, the molar yields of the reaction products were estimated".

6-Page 1. Line 23. The sentence has been modified as follows

"The products detected, indicate a hydrogen atom abstraction mechanism at different sites on the carbon chain of alcohol..."

7-Page 1. Line 31 "Therefore, the use of saturated alcohols as additives in diesel-blends should be considered with caution"

8-Pages 2 and 3. The introduction has been reorganized.

9-Page 2. Line 17. A new reference has been included.

10-Page 7. Lines 31-37. The text has been modified as follows:

"These results show that the activating effect of the OH group of the SA is less important for the reaction with Cl atoms than with the OH radical, behavior that agrees with that established by the Structure Activity Relationships (SARs) methods (Kwok and Atkinson 1995; Calvert et al. 2011)."

12-Page 9. Lines 21, 22. The sentence has been removed.

13-Page 9. Lines 28-31. The sentence has been rewritten to more clarity as follows:

"Some of these compounds are products from the reactions of the SAs with oxidants. They can also be formed by decomposition of the employed precursors (Cl₂, CH₃ONO and N₂O₅) and in some cases, by heterogeneous reactions of these precursors with the Pyrex glass reactor walls."

14-Page 10. Lines 32. Part of the sentence has been modified as follows:

“Figure 5A shows the paths that explain the formation of organic compounds (carbonyl, hydroxycarbonyl, etc)..”

15-Page 12. Lines 1 and 2 have been modified to more clarity, as follows:

“..presented in Fig. 6, showing that in the absence of NO_x the profiles of acetone and formaldehyde have two trends. It indicates that these compounds are formed as primary and secondary products.”

16-Page 12. Lines 13-18. The sentence has modified to more clarity, as follows:

“The higher yield of nitrated compounds in the reaction of 3,3DM1ButOH with nitrate radical could indicate an extra formation of nitrated compounds from secondary reactions..”.

17-Page 12. Lines 25-27. The paragraph has rewritten to more clarity, as follows:

“As it can see in Table 4, the estimated molecular yields of 3,3-dimethylbutanal (formed by H atom abstraction at the α -position of 3,3DM1BuOH) are very similar to the one predicted by the SARs method for the Cl and OH reactions.”

18-Page 13. Lines 15-17. The paragraph has rewritten to more clarity, as follows:

“Plots of concentration versus time for formaldehyde, acetone (Fig. S12A) and nitrated compounds in the reactions of Cl in the presence of NO_x (Fig. S12B) show profiles with two trends. This type of profile indicates that formaldehyde and acetone could also be formed by degradation of 3,3-dimethyl-2-butanone (Fig. S13).”

19-Page 13. Lines 17-19. The sentence has been moved to Line 14.

20-Page 13. Line 33. The next sentence has been added to more clarity, as follows:

“It could justify the low estimated molecular yield for 3,3-dimethyl-2-butanone.”

21-Page 14. Lines 3-4. The paragraph has been eliminated

22-Page 15. Lines 30-32 The sentence has been rewritten as follows:

“Therefore, more kinetic studies of the NO₃ radical reaction with primary alcohols are necessary to quantify the effect of the OH group at the β -position ($-\text{CH}_2\text{OH}$) and to update the SARs method developed by Kerdouci et al.”

23-Page 16. Lines 18-20. The sentence has been changed by:

“Therefore, the use of SAs as additives for diesel blends should be controlled, as poor handling could result in high concentrations of these alcohols in the atmosphere.”

24-Page 19. Lines 36-38. A new reference has been included:

“Li, F., Yi, B., Song, L., Fu, W., Liu, T., Hu, H., & Lin, Q. Macroscopic spray characteristics of long-chain alcohol-biodiesel fuels in a constant volume chamber. Proceedings of the Institution of Mechanical Engineers, Part A: JPE, 232(2), 195–207. <https://doi.org/10.1177/0957650917721336>, 2017.”

Atmospheric fate of a series of ~~Methyl~~Saturated ~~a~~Alcohols (MSA): kinetic and mechanistic study

Inmaculada Colmenar^{1,2}, Pilar Martín^{1,2}, Beatriz Cabañas^{1,2}, Sagrario Salgado^{1,2}, Araceli
Tapia^{1,2}, Inmaculada Aranda^{1,2}

¹Universidad de Castilla La Mancha, Departamento de Química Física, Facultad de Ciencias y Tecnologías
Químicas, Avda. Camilo José Cela S/N, 13071 Ciudad Real, Spain

²Universidad de Castilla La Mancha, Instituto de Combustión y Contaminación Atmosférica (ICCA), Camino
Moledores S/N, 13071 Ciudad Real, Spain

Correspondence to: Pilar Martín (mariapilar.martin@uclm.es)

Keywords. ~~Methyl~~Saturated alcohols; additives; ~~biofuel~~; atmosphere; re; re reactivity.

Abstract. The atmospheric fate of a series of ~~Methyl~~Saturated ~~a~~Alcohols (MSAs) has been evaluated through the kinetic and reaction product studies with the main atmospheric oxidants. These SAs are alcohols that could be used as fuel additives. Rate coefficients (in $\text{cm}^3 \text{molecule}^{-1} \text{s}^{-1}$ unit) measured at $\sim 298\text{K}$ and atmospheric pressure (720 ± 20 Torr) were as follows: k_1 (E-4-methyl-cyclohexanol + Cl) = $(3.70 \pm 0.16) \times 10^{-10}$, k_2 (E-4-methyl-cyclohexanol + OH) = $(1.87 \pm 0.14) \times 10^{-11}$, k_3 (E-4-methyl-cyclohexanol + NO_3) = $(2.69 \pm 0.37) \times 10^{-15}$, k_4 (3,3-dimethyl-1-butanol + Cl) = $(2.69 \pm 0.16) \times 10^{-10}$, k_5 (3,3-dimethyl-1-butanol + OH) = $(5.33 \pm 0.16) \times 10^{-12}$, k_6 (3,3-dimethyl-2-butanol + Cl) = $(1.21 \pm 0.07) \times 10^{-10}$ and k_7 (3,3-dimethyl-2-butanol + OH) = $(10.50 \pm 0.25) \times 10^{-12}$. ~~The main detected products~~ detected in the reaction of SAs with Cl atoms (absence/presence of NO_x), OH and NO_3 radicals were: E-4-methylcyclohexanone; for the reactions of E-4-methyl-cyclohexanol, 3,3-dimethylbutanal for the reactions of 3,3-dimethyl-1-butanol and 3,3-dimethyl-2-butanone for the reactions of E-4-methyl-cyclohexanol, 3,3-dimethyl-1-butanol and the reactions of 3,3-dimethyl-2-butanol, respectively with the three oxidants. A tentative estimation of molecular yields has been done obtaining the following ranges (25–60)% for 4-methylcyclohexanone, (40–60)% for 3,3-dimethylbutanal and (40–80)% for 3,3-dimethyl-2-butanone. Other products such as formaldehyde, 2,2-dimethylpropanal and acetone also have also been identified in the reactions of Cl atoms and OH radicals with 3,3-dimethyl-1-butanol and 3,3-dimethyl-2-butanol. In addition, the ~~The~~ molar yields of these reaction products ~~detected~~ were estimated. The products detected, S- indicate a hydrogen atom/hydrogen abstraction mechanism at different sites ~~of on the carbon chain of alcohol~~ the alkyl chain in the case of Cl reactions and a predominant site in the case of OH and NO_3 reactions, confirming the predictions of Structure Activity Relationships (SAR) methods.

Tropospheric lifetimes (τ) of these MSAs have been calculated using the experimental rate coefficients. Lifetimes are in the range of 0.6–2 days for OH reactions, 7–13 days for NO_3 radical reactions and 1–3 months for Cl atoms. In coastal areas the lifetime due to the reaction with Cl decreases to hours. The calculated global tropospheric lifetimes ~~calculated~~, and the polyfunctional compounds detected as reaction products in this work, imply that the ~~Methyl~~Saturated ~~a~~Alcohols could contribute to the formation of ozone and nitrated compounds ~~formation~~ at

1 local, ~~but also~~-regional and even to global scale. Therefore, the use of ~~large~~-saturated alcohols as additives in
2 ~~diesel-blends biofuels~~should ~~must~~ be considered~~taken~~ with caution.

3

4

1 1. Introduction

2 A multitude of scientific studies ~~about~~ combustion emissions confirm that fossil fuels, especially diesel fuel,
3 are the substances mainly responsible for air pollution. The loss of air quality and its consequences ~~for~~ health
4 as well as for global warming are some of the most important problems caused by air pollution (www.iea.org).
5 These consequences have led governments to set restrictive limits ~~on~~ the presence of certain pollutants in the
6 atmosphere, such ~~as~~ the case of particulate matter (PM) (EURO 6) and to develop biofuels (Sikarwara et al.,
7 2017) as alternatives to conventional ones.

8 Biodiesel ~~is~~ obtained from ~~the~~ transesterification ~~process~~ of oils animal or vegetable ~~oils~~ origin. Also, the
9 fermentation of vegetal biomass gives methanol and ~~or~~ ethanol (bioethanol). These lower alcohols have been
10 used as fuels and showing advantages such as the reduction ~~in~~ the smoke, due to the presence of the OH group
11 that increases the oxygen content during the combustion process (Ren et al., 2008; Lapuerta et al., 2010; Sarathy
12 et al., 2014; Sikarwara et al., 2017). However, several studies have shown certain complications in the use of lower
13 alcohols due to their low cetane number, the high latent heat of vaporization and high resistance to auto-ignition
14 (Karabektas and Hosoz, 2009). In order to avoid or to minimize these limitations, alcohol–diesel blends and
15 alcohol–diesel emulsions have been used in diesel engines (Ozsezen et al., 2011). Another alternative is the use
16 of longer-chain alcohols (propanol, n-butanol, isobutanol and n-pentanol) with superior fuel properties than lower
17 alcohols when mixed with diesel fuel (Cheung, et al., 2014; Kumar and Saravanan, 2016; Li et al., 2017).

~~The fact that the use of high alcohols is a good alternative to conventional fuels could suppose an important
18 presence of these alcohols in the atmosphere. Therefore, previously to the massive use, it is necessary to study the
19 reactivity of the large alcohols in atmospheric conditions, in order to establish and to evaluate their atmospheric
20 impact.~~

21 Alcohols are present in the atmosphere from a wide variety of anthropogenic and biogenic sources (Calvert et al.,
22 2011). Methanol, ethanol and isopropanol are some of the main alcohols detected in urban areas such as Osaka
23 and Sao Paulo cities (Nguyen et al., 2001) with concentrations ~~of~~ 5.8–8.2 ppbv and 34.1–176.3 ppbv
24 respectively. Other alcohols, such as E-4-methylcyclohexanol, have been identified in the exhaust gas emissions
25 resulting from burning fuel blends containing 7-% v/v (B7) and 20-% v/v (B20) of soy bean/palm biodiesel (84
26 %–16%) (Lopes et al., 2014). 3,3-dimethyl-1-butanol is a glass-forming material, used as a chemical intermediate
27 in organic syntheses (www.capotchem.com). 3,3-dimethyl-2-butanol is a potential precursor for prohibited
28 chemical weapons such as soman, a nerve agent (Murty et al., 2010). It is also used in conversion of ribose- and
29 glucose- binding proteins into receptors for pinacolyl methyl phosphonic acid (Allert et al. 2004). The fact that the
30 use of long-chain high alcohols as biofuels as biofuels is a good alternative to conventional fuels could be suppose
31 present an additional additional important source presence of these alcohols in the atmosphere. Therefore,
32 previously to their massive use as biofuels, it is necessary to study their reactivity of the large alcohols of
33 long-chain alcohols in atmospheric conditions, in order to establish and to evaluate their their atmospheric impact.

34 In the case of smaller alcohols, the knowledge of ~~their~~ reactivity is well established and indicating that the
35 main degradation process mechanism of saturated alcohols ~~(SAs) in the atmosphere is the reaction with
36 hydroxyl(OH) radicals (OH) during the daytime is mainly initiated by the H abstraction from C-H bond. The H-~~

1 ~~abstraction from the OH group seems to be less favored~~ (Grosjean et al., 1997; Atkinson and Arey, 2003; Atkinson
2 et al., 2006; Calvert et al., 2011; Caravan et al., 2015; Mellouki et al, 2015). ~~According to literature (Atkinson and~~
3 ~~Arey, 2003; Atkinson et al., 2006; Calvert et al., 2011; Caravan et al., 2015; Mellouki et al, 2015)~~, the main
4 degradation route of saturated alcohols in the atmosphere is the reaction with OH radicals during daytime. Kinetics
5 with chlorine atoms (Cl) are expected to be high, therefore reactions with Cl could also be an important degradation
6 route, especially in coastal areas where concentration peaks of Chlorine atoms can be found. Reactions with
7 ozone molecular (O₃) ($k \leq 10^{-20} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$) and nitrate radical (NO₃) ($\sim 10^{-15} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$)
8 are too low to ~~make~~ have a significant contribution to their degradation. However, the determination of the rate
9 coefficients and the reaction products of alcohols with the nitrate radical are also necessary to ~~better~~ understand
10 better the general reactivity of alcohols in the atmosphere, since the reactions with this radical are a source of OH
11 during the night-time (Finlayson-Pitts and Pitts, 2000).

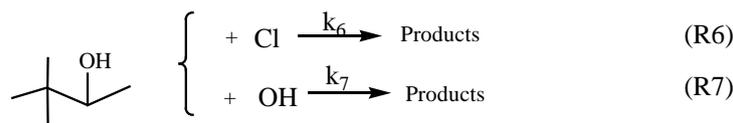
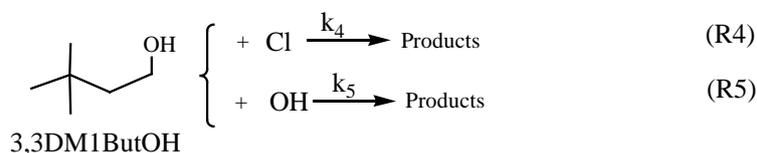
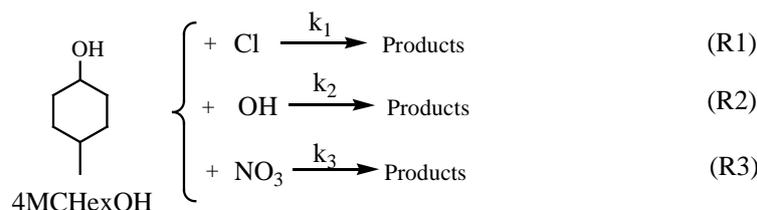
12 Although in ~~the~~ last years some studies ~~into the about~~ reactivity of higher/large alcohols have been made
13 (Ballesteros et al., 2007; Hurley et al., 2009; Andersen et al., 2010; Calvert et al., 2011; Moreno et al., 2012, 2014,
14 Mellouki et al., 2015) the kinetic and mechanistic database is still scarce. In the case of ~~the~~ 3,3-dimethyl-1-butanol
15 and 3,3-dimethyl-2-butanol (~~derivates from of~~ 1-butanol and 2-butanol) there is a lack of information with
16 respect to the diurnal reactivity (Moreno et al., 2014; Mellouki et al., 2015). Regarding ~~to~~ cyclic alcohols, only
17 data concerning the reactivity of chlorine atoms and OH radicals ~~with for~~ cyclohexanol (Bradley et al., 2001;
18 Ceacero-Vega et al., 2012) and OH ~~with for~~ cyclopentanol (Wallington et al., 1988) have been reported.

19 Therefore, in the present work, ~~the studies~~ into the kinetics and the products of gas-phase reactions of some
20 Methyl Saturated Alcohols (MSAs, such as): E-4-methylcyclohexanol (4MCHexOH), 3,3-dimethyl-1-butanol
21 (3,3DM1ButOH) and 3,3-dimethyl-2-butanol (3,3DM2ButOH) with Cl atoms and OH and NO₃ radicals ~~the main~~
22 atmospheric oxidants ~~have~~ been carried out, done These SAs have been chosen for their potential use as
23 biofuels and because their reactivity has not yet been evaluated. So, our work will permit us to in order to complete
24 the kinetic and mechanism database, to improve our knowledge of the atmospheric chemistry of higher/long chain
25 SAs alcohols ~~in special saturated alcohols, and~~ and to assess their environmental chemical impact in the case of
26 their future use as biofuels.

27 **2. Experimental Section**

28 **2.1 Kinetic experiments**

29 The reactions of a series of Methyl Saturated Alcohols (MSAs) with the main atmospheric oxidants have been
30 studied:



1
2

3 Rate coefficients were determined using a relative rate method. This method relies on the assumption that the
4 organic compound (MSA) and the reference compound (R), are removed solely by their reactions with the oxidants
5 (Ox: OH and NO₃ radicals and Cl atoms):



8 ~~Where~~ k_{MSA} and k_{R} are the rate coefficients of the MSA and the reference compound, respectively.

9 On the assumption that the MSA and the reference compound are only consumed by reaction with the oxidants,
10 the kinetic treatment for the reactions expressed by R8 and R9 gives the following relationship;

$$\ln \left(\frac{[\text{MSA}]_0}{[\text{MSA}]_t} \right) = \frac{k_{\text{MSA}}}{k_{\text{R}}} \ln \left(\frac{[\text{R}]_0}{[\text{R}]_t} \right) \quad \text{-----} \quad \text{(1)}$$

12 where $[\text{MSA}]_0$, $[\text{R}]_0$, $[\text{MSA}]_t$, and $[\text{R}]_t$ are the initial concentrations and those at time t for the ~~S~~Methyl Saturated
13 ~~Alcohol~~ and the Rreference compound, respectively. Two reference compounds were used with each oxidant were
14 used to ensure that the reference compound ~~does not have had~~ no influence on overall rate coefficient.

15 According to Eq (1), a plot of $\ln([\text{MSA}]_0/[\text{MSA}]_t)$ versus $\ln([\text{R}]_0/[\text{R}]_t)$ should be a straight line that passes through
16 the origin. The slope of this plot gives the ratio of rate coefficients $k_{\text{MSA}}/k_{\text{R}}$. Therefore, the value of k_{MSA} can be
17 obtained if the rate coefficient k_{R} of the reference compound is known.

18 The experimental systems ~~have been~~ are described in previous works (Tapia et al., 2011; Martin et al., 2013) and
19 only a brief description is given ~~shown~~ here. Kinetic measurements were performed at room temperature (~298
20 K) and atmospheric pressure (720 ± 20 Torr) by employing two separated experimental set-ups: 1) ~~An FTIR~~
21 ~~system formed by~~ 50-L Pyrex® glass reactor was coupled to ~~an~~ the Fourier Transform Infrared ~~Radiation~~
22 spectrometer as a detection technique ("on-line" analysis). Inside of the Pyrex® glass reactor there is a multi-
23 ~~reflect~~ ion system with three mirrors that allows an infrared radiation path of 2.8 ~~200~~ meters. This reactor is
24 known as white cell (Saturn Series Multi-Pass cell). The FTIR spectrometer (Thermo Nicolet 6700) was ~~is~~ equipped
25 with a KBr beam splitter and liquid nitrogen-cooled MCT. Typically, for each spectrum, 60 interferograms were
26 co-added over 98 s and approximately 30 ~~40~~ spectra were recorded per experiment with a spectral resolution of
27 1 cm⁻¹. 2) ~~A~~ Teflon® gas bag reactor of 500 L with sSolid pPhase mMicro eExtraction fiber (SPME) as a

1 ~~fiber pre-concentration sample method, was for sampling followed by analysis on~~ and Gas Chromatography —
2 ~~m~~Mass Spectrometry system with a ~~t~~Time of ~~f~~Flight mass spectrometer analyzer (SPME/GC-TOFMS) (AccuTOF
3 GCv, Jeol) (“off-line” analysis). Samples were collected by exposing a 50/30-~~mm~~ DVB/CAR/PDMS Solid Phase
4 ~~Micro-Extraction fiber~~ (SPME, ~~(~~SUPELCO) for 5 min during the reaction and then thermally desorbed for 15 min
5 at 250 °C in the heated GC injection port. A capillary column (30-~~m~~ × 0.3-~~mm~~ id × 1.0-~~mm~~ film thickness, Tracsil
6 TRB-1701, Teknokroma) was used to separate the compounds. The chromatographic conditions used for the
7 analysis were as follows: injector, 250 °C; interface, 250 °C; ~~even~~ initial oven temperature, 40 °C for 4 min; ramp,
8 30 °C min⁻¹ to 120 °C, held for 6 min; second ramp, 30 °C min⁻¹ to 200 °C, held for 3 min.

9 In each independent experiment, the reactants ~~were~~ injected into the reactors from a vacuum line by dragging
10 with a stream of carrier gas used in the reaction. Both reactors ~~were~~ inside ~~of~~ a metallic housing in the walls of
11 which ~~walls~~ there is a rack of actinic lamps (Philips, TL-40-W, Actinic BL, $\lambda_{\text{max}} = 360$ nm). A scheme of the
12 experimental systems is shown in Fig. S1-~~of~~ Supplementary Material.

13 The kinetic experiments, for the Cl and OH reactions, were performed in the 50 L Pyrex® glass reactor coupled
14 to an FTIR spectrometer~~the FTIR system~~. A spectral subtraction procedure was used to derive the concentrations
15 of reactant and reference compounds at time $t = 0$ and time t . The reaction of NO₃ with 4MCHexOH was studied
16 using a 500 L Teflon® reactor ~~of 500 L~~ in order to minimize the wall deposition and dilution effects of consecutive
17 additions of N₂O₅. Chlorine atoms and OH radicals were obtained by photolysis of Cl₂ in N₂ and methyl nitrite,
18 CH₃ONO, in the presence of NO in air. Methyl nitrite, ~~CH₃ONO~~, was synthesized in the laboratory as described
19 elsewhere (Taylor et al., 1980).

20 Nitrate radicals were generated in situ in the dark by the thermal decomposition of N₂O₅ (Atkinson et al., 1984,
21 1988). N₂O₅ was obtained mixing O₃ with an excess of NO₂ (Scott and Davidson, 1958). Prioreviously to the
22 kinetic experiments a series of tests in the dark and photoly~~tic~~ conditions were carried out to evaluate secondary
23 reactions, such as wall depositions and photodegradation ~~processes~~ of all reactants.

24 The range concentrations of reactants employed were as follows: 2-16-ppm of 4MCHexOH, 3-9-ppm of
25 3,3DM1ButOH and 3,3DM2ButOH, 9-30-ppm of Cl₂, 4-13-ppm of 1-butene and 2-methylpropene, 7-14-
26 ppm of propene and cyclohexene, 5-7-ppm of isopropanol, 3-5 ppm of 2-methyl-2-butanol, 26-55-ppm
27 of CH₃ONO, 20-60-ppm of NO, 3-4-ppm of 2-ethyl-1-hexanol, 4-5-ppm of 1-butanol. For reactions of
28 4MCHexOH with nitrate radicals 2-5 additions of N₂O₅ with concentrations between 8-36-ppm were made per
29 each experiment. N₂ and synthetic air were used as bath gases for Cl, NO₃ and OH reactions, respectively.

30 2.2 Product experiments

31 The product study was carried out at room temperature (~ 298 K) and at a pressure of ~~(720 ± 20)~~ Torr of synthetic
32 air employing the two experimental set-ups previously described~~mentioned above~~. In some experiments carried
33 out in the 50-L Pyrex® reactor, a simultaneous identification of products was performed using both detection
34 techniques. For that, one sample of mixing-reaction mixture was taken from this reactor using the SPME and
35 subsequently analyzed with GC-TOFMS. In addition, independent experiments using SPME/GC-TOFMS
36 technique in a 150-L Teflon® reactor were developed. Products analyses were carried out using the same
37 procedure as for the kinetic experiments, without the reference compound, and employing synthetic air as bath
38 gas. On this occasion the heating of the oven was changed slightly in order to get a better separation and to detect

1 the products generated. The temperature ramps of the oven employed in the chromatograph were: 40 °C for 4 min;
2 ramp, 25 °C min⁻¹ to 120 °C, held for 10 min; second ramp, 25 °C min⁻¹ to 200 °C, held for 4 min.

3 The qualitative analysis in the FTIR experiments was carried out using the FTIR library that provided~~s~~ the FTIR
4 spectrophotometer (Aldrich vapor phase sample library,
5 <https://www.thermofisher.com/search/browse/results?customGroup=Spectral+Libraries>) and/or the FTIR
6 database of Eurochamp (<https://data.eurochamp.org/data-access/spectra/>).

7 For the SPME/GC-TOFMS experiments, the NIST webbook (<https://webbook.nist.gov/chemistry/>) and the ~~m~~Mass
8 ~~s~~Spectra database of the instrument were used to identify the products. Calibrated FTIR spectra and SPME/GC-
9 TOFMS chromatograms were used ~~for~~ quantification in those cases where the product was commercially
10 available.

11 The molecular yields of the reaction products were estimated from the slopes of plots of the concentration of
12 ~~formed~~-product ~~formed~~ versus the amounts of MSA ($\Delta[\text{MSA}]$) consumed. In the cases where ~~it was observed~~-an
13 important loss of the reaction product by reaction of the oxidant and/or by photolytic process ~~was observed~~, the
14 concentration of the product was corrected using the formulism of Tuazon et al. (1986) (See S1, ~~Supplementary~~
15 ~~M~~aterial). Range of concentrations of reactants employed were as follows: 2–14 ppm of MSA, 8–31 ppm of
16 Cl₂, 12–57 ppm of NO, 19–66 ppm of CH₃ONO and 6–36 ppm of N₂O₅.

17 Chemicals used were as follows: 4MCHexOH (97%, Aldrich), 3,3DM1ButOH and 3,3DM2ButOH (98%,
18 Aldrich); 1-butene, propene, 2-methyl-2-butanol, isopropanol, 2-methylpropene, 4-methylcyclohexanone and
19 cyclohexene (\geq 99%, Aldrich), 2-ethyl-1-hexanol (\geq 99%, Fluka), 1-butanol (99.8%, Aldrich), 3,3-
20 dimethylbutanal (95%, Aldrich) and 3,3-dimethyl-2-butanone (98%, Aldrich), NO (99%, Praxair), Cl₂ (> 99.8
21 %, Praxair), synthetic ~~a~~Air (Praxair Ultrahigh purity 99.999%) ~~and~~, N₂ (99.999%, Praxair). For N₂O₅ synthesis,
22 N₂O₄ ~~was~~ used (>99.5%) from Fluka, P₂O₅ (98%, such as desiccant) from Fluka and O₃ ~~was~~ synthesized by
23 a generator model TRCE-5000, 5 g_{O3} h⁻¹ OZOGAS.

24 3. Results and ~~D~~iscussion

25 3.1 Kinetic study

26 Preliminary test experiments indicated that dark heterogeneous reactions and photolytic losses of ~~MSAs and~~
27 ~~reference compounds could~~ be considered negligible in our experimental conditions ($k \approx 10^{-6}$ s⁻¹). ~~As~~
28 ~~mentioned above, the kinetic study of Cl atoms and OH radical with the MSA~~ was carried out at room temperature
29 (\sim 298 K) and at 720 ± 20 Torr of N₂ gas ~~for Cl atoms reactions~~ and synthetic air ~~for OH radical~~
30 ~~reactions respectively~~. Nitrate radical experiments were performed using N₂ gas in a 500-L Teflon reactor and
31 employing the ~~system~~-SPME/GC-TOFMS ~~system~~. A number of injections of the unreacted mixture were carried
32 out in order to determine the associated precision ~~of~~with the sampling method to be used in the error analysis
33 (Brauers and Finlayson-Pitts, 1997). The standard deviations (σ) were as follows: 3.7% for 4MCHexOH, 1.7%
34 for 1-butanol and 3.5% for 2-ethyl-1-hexanol.

35 -Figure 1 shows examples of the kinetic data plotted according to Eq (1) for the reactions of ~~MSAs~~ with different
36 atmospheric oxidants.

1 A good correlation was obtained with an intercept close to zero, which indicateds the absence of ~~-other-~~secondary
2 ~~reactionsprocesses~~. From the slopes of the plots (k_{MSA}/k_R) and knowing the values of the rate coefficients for the
3 reference compounds employed (k_R), the value of the absolute rate coefficient for each ~~-methyl~~ saturated alcohol
4 (k_{MSA}) ~~washas been~~ determined.

5 Rate coefficients of ~~reactions of~~ reference compounds; ~~withfor~~ Cl atoms reactions (in 10^{-10} cm³ molecule⁻¹ s⁻¹
6 units) were: 2-methylpropene (3.40 ± 0.28), 1-butene (3.38 ± 0.48), (Ezzel et al., 2002) and propene (2.23 ± 0.31)
7 (Ceacero-Vega et al., 2009); ~~withfor~~ OH radicals~~s-reactions~~ (in 10^{-11} cm³ molecule⁻¹ s⁻¹ units): propene ($2.66 \pm$
8 0.40)~~-(~~Atkinson and Aschman, 1989), cyclohexene (6.77 ± 1.69) (Atkinson and Arey, 2003), isopropanol (0.51
9 ± 0.008) (IUPAC www.iupac-kinetic.ch.cam.ac.uk) and 2-methyl-2-butanol (0.36 ± 0.06) (Jiménez et al., 2005);
10 ~~aAnd withfor~~ NO₃ radicals ~~eaactions~~ (in 10^{-15} cm³ molecule⁻¹ s⁻¹ units): 1-butanol (3.14 ± 0.97) and 2-ethyl-1-
11 hexanol (2.93 ± 0.92) (Gallego-Iniesta et al., 2010).

12 ~~The~~ ratios of the rate coefficients, k_{MSA}/k_R , the absolute rate coefficients and the weighted averages are shown in
13 Table 1. The error of k_{MSA}/k_R ~~wasare~~ given by ~~two2~~ times the statistical deviation calculated from the least ~~-squares~~
14 fit of the plot of Eq. (1). The uncertainties for rate coefficients of MSA ($\sigma_{k_{MSA}}$) were calculated from the
15 uncertainty of slope of plots (σ_{slope}) and the uncertainty of the reference (σ_{k_R}) by using the propagation of
16 uncertainties. The average value of the rate coefficients obtained with different reference compounds and ~~theirts~~
17 associated errors were obtained by weighted average (See footnote, Table 1). ~~The~~ rate coefficients obtained in
18 this work are the first kinetic data reported for these MSAs; therefore, the results obtained cannot be compared
19 with literature values.

20
21 ~~As it has been mentioned in introduction section,~~ it is well established that the gas-phase reaction mechanism of
22 saturated organic compounds (alkanes, alcohols, ethers, etc.) with ~~the~~ atmospheric oxidants (Cl atoms, OH and
23 NO₃ radicals) are initiated "via" hydrogen atom abstraction from the organic compound to form a stable molecule
24 and an alkyl radical (Finlayson-Pitts and Pitts, 2000; Atkinson and Arey, 2003; Calvert et al., 2011; Ziemann and
25 Atkinson, 2012). The presence of ~~the OHhydroxyl~~ group in saturated alcohols implies two types of hydrogens that
26 can be abstracted; hydrogen bonded to a carbon (C-H) of the main chain or of an alkyl substituent and hydrogen
27 bonded to oxygen of ~~OHhydroxyl~~ group (~~-OH~~). Two literature reviews ~~of theabout~~ reactivity of saturated alcohols
28 (Calvert et al., 2011; Mellouki et al., 2015) conclude that:

29 1 ~~->~~ ~~The~~ reactions of aliphatic alcohols with atmospheric oxidants proceed mainly by H atom abstraction from
30 various C-H groups in the alkyl chain, ~~being the abstraction of~~ H atom ~~abstraction~~ from the O-H ~~group being~~
31 negligible;

32 2 ~~->~~ Rate coefficients for the reactions of Cl, OH and NO₃ are higher for alcohols than those of the corresponding
33 alkanes due to the activating effect of the OH group. This effect is extended over about ~~four4~~ carbon atoms (Nelson
34 et al., 1990); ~~and: As will be discussed below, the activating effect of the OH group depends on the oxidant.~~

35 3 ~~->~~ The attack percentage of a radical to the different sites of the alcohol (α , β , γ and δ) depends on the oxidant,
36 ~~the~~ structure of saturated alcohol, ~~the~~ type and numbers of substituents, and temperature; (Moreno et al., 2012,
37 2014; McGillen et al., 2013, 2016).

38 In order to ~~verifyheck~~ these remarks, the reactivity of the ~~Methyl SAsaturated Alcohols~~ studied in this work
39 ~~werehas been~~ analyzed and discussed ~~by comparing-a)~~ the rate coefficients of the MSAs obtained ~~with for the~~

1 different oxidants; ~~b) comparing~~ the rate coefficients of the ~~MSAs~~ and the rate coefficients of their homologous
2 alkanes available in ~~literature, bibliography~~ and ~~comparing e)~~ the rate coefficients obtained in the reaction of the
3 same oxidant but with different alcohols. The data used ~~in the to-comparisone~~ are summarized in Table S1 in
4 ~~Supplementary Mmaterial~~.

5 From the analysis of all data ~~shown in Table S1~~, it can be observed that:

6 ~~a)→)~~—The trend in the reactivity of ~~MSAs~~ in relation to the different oxidants is the same that ~~the~~ observed for
7 other saturated alcohols: k_{Cl} ($k \approx 10^{-10}$) > k_{OH} ($k \approx 10^{-11}$) >> k_{NO_3} ($k \approx 10^{-15}$), (k in $\text{cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$
8 units). This behavior could be explained ~~by~~ considering the geometry and ~~the~~ electronic density of each oxidant,
9 together with ~~the~~ kinetic ~~c~~Collision ~~t~~Theory. As ~~the~~ Cl atom has spherical ~~electronic density~~ distribution ~~of its~~
10 ~~density, any, for the collision any~~ orientation is ~~suitable~~adequate, in addition, the Cl atoms presents ~~lowess~~ steric
11 hindrance. ~~So Then, comparatively~~ the Cl reaction is ~~comparatively~~ less selective and faster, with values for the
12 rate coefficients, k , in the collision limit. However, the OH radical presents an asymmetric electron density located
13 mostly over its oxygen atom. Therefore, for the OH reaction the oxygen of ~~the~~ OH radical, must be specifically
14 oriented ~~toward~~ the hydrogen of the ~~MSA~~ that will be abstracted. The electronic density of nitrate radical is
15 distributed around the three oxygens, which implies several appropriate orientations. ~~However, but sinceas~~ the
16 nitrate radical has a non-linear structure, the steric hindrance is much ~~greaterbigger~~ than for the OH ~~radical,~~
17 ~~whichand it~~ reduces the reactivity of ~~NO₃~~ in relation to th~~atose~~ of OH ~~radical~~.

18 ~~b)→)~~ The rate coefficient for the reaction of 4MCHexOH with Cl atoms is similar to the rate coefficient of its
19 homologous alkane (E-1,4-dimethylcyclohexane): $k_{4MCHexOH+Cl} = 37.0 \times 10^{-11} \approx k_{E-1,4-dimethylcyclohexane+Cl}$
20 $= 36.3 \times 10^{-11}$. In the case of the reaction with OH radical, the rate coefficient of 4MCHexOH is 1.5 times higher
21 than ~~with~~ E-1,4-dimethylcyclohexane; $k_{4MCHexOH+OH} = 18.7 \times 10^{-12} > k_{E-1,4-dimethylcyclohexane+OH} = 12.1 \times 10^{-12}$
22 (Table S1). These results show that the activating effect of ~~the hydroxyl group (-OH group)~~ of the ~~MSA~~ is less
23 important for the ~~reaction with~~ Cl atoms than ~~with the~~ OH radical, ~~behavior that agrees with that established by~~
24 ~~the Structure Activity Relationships (SARs) methods. The activating effect of hydroxyl group of the alcohols was~~
25 ~~quantified by different authors (Kwok and Atkinson 1995; Kerdouei et al., 2010; Calvert et al. 2011), taking into~~
26 ~~account the available kinetic data reported in bibliography, obtaining the factor of reactivity for the hydroxyl group,~~
27 ~~F(-OH)). This factor of reactivity is different for each oxidant, 1.18 for Cl reaction, 2.35 for reaction with OH~~
28 ~~(Calvert et al. 2011) and 18 for NO₃ reaction (Kerdouei et al., 2010). There are no data of rate coefficient datas~~
29 for the reactions ~~with NO₃ radical~~ of the ~~alkanes~~ homologous ~~to alkanes of~~ the ~~MSAs~~ studied in this work ~~with~~
30 ~~NO₃ radical~~, and therefore it ~~was~~is not possible to check ~~out~~ the effect of ~~OH hydroxyl~~ group in the ~~reactivity of~~
31 ~~NO₃ reaction with NO₃~~. However, according to the ~~SAR method developed by factor of reactivity obtained by~~
32 Kerdouci et al. (2010) for the reactions of alcohols with NO_3 , this effect is ~~greaterhigher~~ than ~~in~~ the corresponding
33 ~~to~~ Cl and OH reactions.

34 ~~c)→)~~—The activating effect of the ~~chain~~ length ~~chain~~ in the reactivity of alcohols is ~~also more differentevident in~~
35 ~~for the Cl and OH reactions than OH reactions~~ (See Table S1). Furthermore, if the rate coefficients of 3-methyl-
36 1-butanol (3M1ButOH) and 3,3DM1ButOH with Cl and OH ~~reactions~~ are compared, ~~it can be observed~~ a slight
37 increase ~~in the of~~ rate coefficient for Cl reaction ~~can be observed~~ ($k_{3M1ButOH+Cl} = 25.0 \times 10^{-11}$; $k_{3,3DM1ButOH+Cl} = 26.9$
38 $\times 10^{-11}$) ~~together withand~~ an important decrease ~~in of~~ the rate coefficient for ~~the~~ OH reactions ($k_{3M1ButOH+OH} = 14$
39 $\times 10^{-12}$; $k_{3,3DM1ButOH+OH} = 5.33 \times 10^{-12}$). This behavior could be explained by the different order of reactivity

1 ~~of~~between the two oxidants. ~~So, For~~ Cl atom, more reactive (k in the order of 10^{-10} $\text{cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$) but less
2 selective than OH, an increase in the chain of the-length-~~chain~~ or in the number of methyl groups in the SAs implies
3 more hydrogens available to be abstracted and therefore an increase in~~ef~~ the rate coefficient. However, for OH
4 radicals, less reactive (k in the order of 10^{-11} – 10^{-12} $\text{cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$) and more selective than Cl, the attack for
5 H-abstraction will be carried out at ~~in~~-a specific position in the SA~~place~~, so an increase in~~ef~~ the chain length of the
6 alcohol does~~has~~ not have a significant~~tative~~ effect on~~to~~ the reactivity; ~~even~~-the presence of a second methyl group
7 even disfavors the reaction, probably due to ~~the~~-steric hindrance near ~~to~~-the attack position.

8 In addition, as can be seen in Table S1, the position of the OHhydroxyl group of SA has a different effect on the
9 reactivity depending on the oxidant. In the case of the Cl atom reactions, the rate coefficients for primary alcohols
10 (1-propanol, 1-butanol, 1-pentanol, 3-methyl-1-butanol and 3,3-dimethyl-1-butanol) are higher than the tho~~ese~~
11 ~~of~~-the secondary alcohols (2-propanol, 2-butanol, 2-pentanol, 3-methyl-2-butanol and 3,3-dimethyl-2-butanol)
12 contrary to the OH and NO_3 radical reactions. This ~~fact~~ indicates that in the reaction of Cl atoms the formation of
13 the most stable radical seems to have less importance in the reactivity than the number of hydrogens in the α -
14 position available to be abstracted.

15 All these observations (a, b and c) could imply a different mechanism for~~in~~ the hydrogen abstraction process for
16 Cl atoms versus the OH and NO_3 -radicals. Nelson et al. (1990) and Smith and Ravishankara (2002) indicate the
17 possible formation of an intermediate adduct between the OH radical and the oxygen of the OHhydroxyl group
18 via hydrogen bond-that will imply a specific orientation. Theoretical studies found in the literature~~bibliography~~
19 show this different hydrogen-abstraction process in the reaction of saturated alcohols with Cl atoms (Garzon et
20 al., 2006) and OH radical (Moc and Simmie, 2010). There are not enough rate coefficient data for the reaction of
21 SAs with NO_3 radicals to establish conclusions about the mechanism.

22 Different mechanisms in the hydrogen atom abstraction process for each oxidant will imply different product
23 distributions and molecular yields, as ~~it~~-will be shown below, in the section on~~f~~ product and mechanism and
24 mechanistic-study.

25 3.1.1 Estimation of rate coefficients

26 In order to estimate the rate coefficients of the reactions of organic compounds with the atmospheric oxidants, a
27 multitude of methods have been proposed (Vereecken et al., 2018). The most popular and widely used is the SAR
28 method developed initially by Kwok and Atkinson (1995) to estimate the rate coefficients at room temperature for
29 gas-phase reactions with the OH radical. This method has been updated for OH reactions (Jenkin et al., 2018) and
30 extended to reaction with NO_3 (Kerdouci et al., 2010, 2014) and Cl (Calvert et al., 2011; Poutsma 2013)-reactions.
31 The EPA (United States Environmental Protection Agency) has developed the EPI SuiteTM-Estimation Program
32 Interface that allows ~~to~~-estimation of the rate coefficients s for the reactions s of the OH radical with~~and~~ organic
33 compounds using the AOPWIN v1.92 program.

34 -In our work the rate coefficients of MSAs with the three oxidants have been estimated using the SARs method
35 (See S2, Ssupplementary Mmaterial). The results are shown in Table 2.

36 The estimated values of ~~estimated~~-rate coefficients agree with experimental data, with ratios $k_{\text{exp}}/k_{\text{SAR}}$ ratios
37 between 0.9 and 1.28, except for the case of 3,3DM1ButOH and NO_3 radical, with which showed a $k_{\text{exp}}/k_{\text{SAR}}$ ratio
38 of 3.24. In general, the SARs method when applied to alcohols predicts better rate coefficients for the Cl atoms

1 and the OH radical than for the NO₃ radical, especially for primary alcohols. It is important to note that the kinetic
2 database for the NO₃ reactions is more limited than for Cl and OH reactions, so the estimated rate coefficient for
3 NO₃ radical should be treated with caution (Kerdouci et al. 2010, 2014; Calver et al., 2011).

4 It is known that organic compounds that which reacts in the same way with different atmospheric oxidants, present
5 a correlation between their rate coefficients. In this sense, over along the years, different correlations have been
6 proposed to that allow to estimation of the unknown rate coefficient when the other one is known (Wayne, 1991,
7 2000; Atkinson, 1994; Calvert et al., 2011; Gallego-Iniesta et al., 2014). The Correlations logk_{Cl} - logk_{OH} and
8 logk_{NO₃} - logk_{OH} correlations have been built for a set of alcohols, ethers and saturated alcohols by Calvert et al.,
9 (2011) obtaining the following relationships:

$$\log(k_{\text{Cl}}/\text{cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}) = 0.634 \times \log(k_{\text{OH}}/\text{cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}) - 2.71 \quad (r^2 = 0.72) \quad (2)$$

$$\log(k_{\text{NO}_3}/\text{cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}) = 1.11 \times \log(k_{\text{OH}}/\text{cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}) - 2.42 \quad (r^2 = 0.66) \quad (3)$$

13 These equations have been used to estimate the rate coefficients of the reactions of MSAs with Cl and NO₃ nitrate
14 radical using the experimental rate coefficients measured in this work for OH reactions. The estimated rate
15 coefficients, k_{log}, according to with Eqs (2) and Eq (3), and the ratios (k_{exp}/k_{log}), are also shown in Table 2. This
16 estimation method obtains slightly better rate coefficient for 3,3DM1ButOH + NO₃ reaction (k_{exp}/k_{log} = 1.53) than
17 the SAR method (k_{exp}/k_{SAR} = 3.24). However, for Cl reactions the ratios k_{exp}/k_{log} ratios are in the range of 0.6 -
18 1.97, indicating that the Eq (2) predicts worse the rate coefficients worse than SAR method. Again, this fact could
19 be due to the different reaction mechanism reaction in the H-abstraction process for the Cl and OH reactions. As
20 Such as it has been indicated above, to apply these relationships both oxidants must react according to the same
21 mechanism. It is important to indicate that in the case of Cl reactions, other effects, such as thermochemistry and
22 the polar effect, must be considered to estimate the rate coefficients for hydrogen abstraction reactions (Poutsma,
23 2013).

24 3.2 Product and Mechanistic Study

25 A product study of the reaction of 4MCHexOH, 3,3DM1ButOH and 3,3DM2ButOH with chlorine atoms in the
26 absence/presence of NO_x, hydroxyl and nitrate radicals has been performed. by employing the two experimental
27 set ups mentioned above (FTIR and SPME/GC-TOFMS). IR absorption bands of HCl, CO₂, CO, HNO₃, N₂O,
28 NO₂, HCOOH, HCOH, ClNO, ClNO₂ and CH₃NONO₂ were observed in the FTIR experiments. Some of these
29 compounds are products from the reactions of the SAs with oxidants. They can also be formed bands are due by to
30 the decomposition of the employed precursors employed (Cl₂, CH₃NONO₂ and N₂O₅) and in some cases, are due
31 by to heterogeneous reactions of these precursors with the walls of the Pyrex® glass reactor walls. The formation
32 of O₃ and N₂O₅ have also been observed after long large reaction times for the reactions of the MSAs with Cl
33 atoms in the presence of +NO_x, due to the high concentration of the NO₂ in the medium of reaction medium and
34 their presence of radiation. Quantitative analysis was carried out by linear subtraction of a spectrum's absorption
35 bands and the peak areas of GC chromatograms by with the use of calibrated spectra and reference chromatograms.
36 The experimental conditions and molecular yields of the main products formed in the reactions of MSAs and
37 analyzed by FTIR and SPME/GC-TOFMS techniques are given in Tables 3-6. Molecular yields could be affected

1 ~~by~~ large errors associated with the SPME sampling method and due to the presence of interfering IR absorption
2 bands, mainly associated with precursors of the OH and NO₃ radicals or by nitrated compounds formed.

3 3.2.1. ~~4MCHexOH~~

4 E-4-methylcyclohexanone was identified in the reactions with Cl₂ (~~absence/presence of NO_x~~)Cl+NO, OH+NO
5 and NO₃+NO₂. An example of the product spectra obtained ~~by in the FTIR system~~ is shown in Fig. 2. Formation
6 of ~~the~~ E-4-methylcyclohexanone was confirmed by introducing a sample of the commercial product (spectrum
7 (e)). A set of experiments using the SPME ~~as sampling method and the GC-TOFMS as detection technique system~~
8 were also carried out for the reactions of 4MCHexOH with Cl atoms and OH and NO₃ radicals. An example of the
9 chromatogram obtained for the reaction of 4MCHexOH with chlorine atoms is shown in Fig. 3. In all the studied
10 reactions, formation of a product peak at 10.35 min was observed. The peak (B) shown ~~ned~~ in Fig. 3 was assigned
11 to E-4-methylcyclohexanone and confirmed by comparing with the retention time and MS ~~spectrum~~ of a
12 commercial sample. In the reactions with ~~Chlorine~~ atoms (absence/presence of NO_x) and OH radical two
13 additional peaks at 19.80 min (C) and 20.25 min (D) were observed.

14 The time-concentration profiles of 4MCHexOH and E-4-methylcyclohexanone obtained by FTIR ~~analysis~~ for
15 the reaction with ~~Chlorine~~ atoms in ~~the~~ presence of NO_x ~~are~~ shown ~~in~~ Fig. S2. The concentrations of E-4-
16 methylcyclohexanone, corrected according to Eq_s (S1), (S2) and (S3), were plotted versus the amounts of
17 4MCHexOH consumed in order to obtain the yield of 4-methylcyclohexanone from the slope. An example of the
18 ~~plots~~ obtained ~~plots~~ is shown in Fig. 4. Molecular yields, Y (%), of E-4-methylcyclohexanone obtained in all
19 experiments are listed in Table 3. Based on the average molecular yield of E-4-methylcyclohexanone, the carbon
20 balance ~~was~~ below to 50-% for reactions with Cl and OH radical and ~60-% for NO₃.

21 Residual spectra after subtraction of ~~the~~ E-4-methylcyclohexanone show IR absorption bands compatible with the
22 presence of hydroxy-carbonyl compounds (~1750, 1720 ~~and~~, 1060 ~~cm⁻¹~~) and nitrated organic compounds
23 (RONO₂ ~1660, 1264 and 862 ~~cm⁻¹~~, and/or ROONO₂ ~1720, 1300 and 760 ~~cm⁻¹~~) (See residual spectra, Fig.
24 S3 in ~~S~~supplementary ~~M~~aterial). The amount of nitrated compounds was estimated using the average integrated
25 absorption coefficient of 1.2×10^{-17} cm molecule⁻¹ of similar compounds corresponding to the IR range 1260-
26 1305 ~~cm⁻¹~~ (Tuazon and Atkinson,1990). The calculated yields of RONO₂ were 20-% and 60-% for Cl (~~in the~~
27 ~~presence of +NO_x~~) and NO₃ reactions, respectively. A yield of 10-% of nitrated compounds was estimated for the
28 reaction with OH radical. This lower yield could be due to fact that the NO_x, presents in the reaction medium,
29 reacts faster with the CH₃O* radical, (~~formed by in the reaction of photolysis of CH₃ONO₂ with NO~~) than others
30 alkoxy radicals. Table 6 shows a summary of the average yields of reaction products quantified for 4MCHexOH
31 reactions.

32 Considering the products detected here and the ~~those~~ detected in the study of Bradley et al., (2001) relative to
33 cyclohexanol with OH radical reactions, a degradation mechanism for 4MCHexOH with the atmospheric oxidants
34 has been proposed. Figure 5A shows the paths ~~that~~ explain the formation of ~~organic compounds (carbonyl, or~~
35 ~~hydroxy-carbonyl, etc) compounds,~~ and Figure 5B shows an example of a path to explain the formation of nitrated
36 organic compounds (ROONO₂ and RONO₂). Similar nitrated compounds could be formed by routes II-IV. The
37 abstraction of hydrogen atoms ~~at their α-α~~ position with respect to ~~OHaleohols~~ group (route I) followed by the

1 addition of oxygen, the formation of a peroxy radical and the fast decomposition of this radical explains the
2 formation of E-4-methylcyclohexanone. Based on the molecular yield obtained for E-4-methylcyclohexanone for
3 each oxidant (See Table 6), this route represents ~~~25/30%~~, ~~~40%~~ and ~~~60%~~ of the ~~reaction mechanism of~~
4 4MCHexOH with Cl ~~(absence/presence of NO_x), and Cl + NO~~, OH and NO₃, respectively. Percentages are two-
5 ~~foldtimes~~ higher than ~~those predicted by the SARs method-prediction~~ in the case of the Cl atoms-reactions and 1.3-
6 and 1.5-~~fold-times~~ lower for the OH and NO₃ reactions, respectively. These data should be taken with caution,
7 since they could ~~includemply~~ many sources of error.

8 Apart from E-4-methylcyclohexanone, other carbonyl and hydroxy-carbonyl compounds could be formed by
9 routes II, III and IV. The presence of ~~these types is kind~~ of compounds has ~~ve~~ been observed in the reactions with
10 Cl and OH. According to the ~~e~~Electron ~~i~~onization ~~m~~Mass ~~s~~Spectra (EI-MS) (Fig. S4, ~~S~~supplementary ~~M~~aterial)
11 ~~an assignation~~ of peaks (C) and (D) shown in Fig. 3, ~~an assignation~~ to 2-hydroxy-5-methyl-cyclohexanone, 5-
12 hydroxy-2-methylc-eyclohexanone and/or 3-methyl-1,6-hexanedial has been proposed. However, according to the
13 atmospheric reactivity (Finlayson and Pitts, 2000; Calvert et al., 2011; Ziemann and Atkinson, 2012) and the study
14 of Bradley et al., (2001), the compound that would be expected is ~~the~~ 3-methyl-1,6-hexanedial, which ~~comearisess~~
15 from the decomposition of the alkoxy radical formed in route II. However, confirmation was not possible since
16 these compounds ~~awere~~ not commercially available. The detection ~~about 9%~~ of HCOH ~~about 9%~~ in the reaction
17 with Cl atoms indicates that the elimination of the methyl group in route IV is ~~minority~~.

18 In the case of nitrate radical, ~~the only carbonyl compound detected was~~ E-4-methylcyclohexanone ~~was detected~~
19 ~~as carbonyl compound~~, suggesting that ~~the~~ route I may be the dominant pathway for this radical. The large
20 difference between the yields of E-4-methylcyclohexanone obtained using the SPME/GC-TOFMS system (~75
21 %) or the FTIR (35-%) could be due to the different ~~methods of adding way to add~~ the precursor ~~to the two in both~~
22 reactors (small aliquots of N₂O₅ in the Teflon® reactor versus one large addition in the Pyrex® ~~glass~~ reactor). This
23 procedure causes a lower initial concentration of nitrated inorganic species (NO₃, NO₂, HNO₃) in reactor of 150 ~~L~~
24 than in ~~of 50 L reactor-of 50 L~~, favoring the formation of carbonyl compounds instead of nitrated organic
25 compounds. Taking into account, the yields of E-4-methylcyclohexanone and the nitrated compounds for the NO₃
26 reaction using FTIR ~~analysis~~, a total carbon balance of 100% is obtained (See Table 6).

27 3.2.2. ~~-3,3DM1ButOH~~

28 ~~For the reaction of 3,3DM1ButOH with the three atmospheric oxidants. Following the same procedure as above,~~
29 3,3-dimethylbutanal was identified as the main reaction product, ~~in the reaction of 3,3DM1ButOH with the three~~
30 ~~atmospheric oxidants.~~ Figure S5A shows the FTIR spectra obtained for the reactions of 3,3DM1ButOH with Cl
31 ~~(absence/presence of NO_x), Cl + NO~~, OH and NO₃ ~~after subtraction~~.

32 Residual FTIR spectra after subtraction of 3,3-dimethylbutanal (Fig. S5B), the SPME/GC-TOFMS
33 chromatograms (Fig. S6) and the EI-MS spectra (Fig. S7), show that other reaction products such as carbonyl,
34 hydroxy-carbonyl and nitrated compounds are formed. These reaction products could be formaldehyde, 2,2-
35 dimethylpropanal, glycolaldehyde, acetone, and peroxy-3,3-dimethyl-butyryl nitrate (P33DMBN)
36 (CH₃)₃CCH₂C(O)OONO₂. These compounds can be formed as primary products (See Fig. 6) and/or secondary
37 products from the degradation of 3,3-dimethylbutanal (See Fig. S8). The SPME/GC-TOFMS chromatograms show

1 common peaks for the three oxidants, but the numbers of peaks and their distribution are very different, especially
2 for OH reactions. In the case of the SPME/GC-TOFMS system, a set of experiment using Field Ionization was
3 carried out in order to help us to establish the identification of reaction products.

4 Time-concentration profiles of 3,3DM1ButOH, 3,3-dimethylbutanal and those reaction products positively
5 identified by FTIR analysis were made in order to establish test whether if the profiles correspond to with a primary
6 or secondary reaction products. An example of the reactions with Chlorine atoms in the absence and presence of
7 NO_x is presented shown in Fig. 6, showing observing that in the absence of NO_x the profiles of acetone and
8 formaldehyde have show two trends. It profiles indicates a typical profile of secondary reactions. that these
9 compounds are formed as primary and secondary products. This profile is clearly observed for nitrated compounds
10 in the reaction with Cl atoms in the presence of NO_x (Fig. 6B).

11 A Commercial sample of 3,3-dimethylbutanal was used to estimate the molecular yields in both experimental
12 systems. These molecular yields are shown in the Table 4. The yields of acetone and formaldehyde were
13 calculated using an FTIR reference spectrum of a commercial sample and FTIR reference spectra from Eurochamp
14 database (<https://data.eurochamp.org/data-access/spectra/>), respectively. An FTIR reference spectrum of 2-
15 methylpropanal (from this same database) was has been used to estimate the molecular yield of 2,2-
16 dimethylpropanal.

17 The amounts of 3,3-dimethylbutanal formed were corrected by their reactions with Cl atoms, and with OH and
18 NO₃ radicals as is described previously, using the rate coefficients available in literature bibliography or for
19 reactions of structurally similar compounds (see footnote, Table 4). Estimated yields of formaldehyde, acetone,
20 2,2-dimethylpropanal and nitrated compounds are summarized in Table 6 along with an average yield of 3,3-
21 dimethylbutanal. The higher yield of nitrated compounds in the reaction of 3,3DM1ButOH with nitrate radical
22 could indicate an extra formation of nitrated compounds from secondary products reactions (See Fig. S8). The total
23 of carbon yields (nitrated compounds were have not been accounted for) of 60%, 81% and 36% have been justified
24 for Cl (in the absence and presence of NO), OH and NO₃ reactions respectively. It, but must it be noted that there
25 are reaction products that could not be quantified as dihydroxy carbonyl or hydroxy-carbonyl compounds in the
26 reactions with Cl atoms atoms reactions in the absence of NO_x and primary nitrated compounds in the reactions
27 with Cl atoms in the presence of NO_x+NO and with NO₃ radical reactions.

28 This work is the first study of the products of reaction products of 3,3DM1ButOH with the atmospheric oxidants,
29 so there is no other any study with which to compare. Figure 7 shows the reaction mechanism proposed based on
30 the literature studies of about saturated alcohols reactions with Cl atoms and the OH radical (Cavalli et al., 2002;
31 Hurley et al., 2009; McGillen et al., 2013; Welz et al., 2013) and considering the reaction products identified in
32 this our work. Table S2, in Supplementary Material, shows a summary of the reaction products, either proposed
33 in this mechanism observed or tentatively identified in proposed in this mechanism of the reactions of
34 3,3DM1ButOH with the atmospheric oxidants.

35 As it can see in Table 4, the Estimated molecular yields of 3,3-dimethylbutanal (formed by H atom -a Abstraction
36 at the α -position of 3,3DM1BuOH) are very similar to the one predicted by the SARs method for the Cl (40%
37 without NO and 43% with NO) and OH reactions. (40-43%, 60%) are very similar to the one predicted by
38 the SAR method (40% for Cl and 66% for OH respectively). In the case of NO₃ radical a large difference between
39 both yields was are observed (36% estimated in this work, 86% predicted by SAR method). This discrepancy

could be explained by the fact that the SAR method applied to the NO₃ radical reaction with primary alcohols (Kerdouci et al., 2010, 2014) underestimates the attack of NO₃ at their β-position, since it does not consider the possible activating effect of the -CH₂OHhydroxy group. (See S2,Supplementary Material)-joined to -CH₂, (-CH₂OH)). This could also explain the large difference observed between the estimated and measured rate coefficients of the reaction of 3,3DM1ButOH with NO₃ as was shown in Table 2. On the other hand, as it has been discussed above, the volume of the reactor could also have an influence on formation of 3,3-dimethylbutanal. According to the molecular yields of the products quantified and/or observed in the SPME/GC-TOFMS chromatograms, it can be concluded that for the OH radical reaction, the route I (attack at their α-position) seems to be the main reaction route. For Cl atoms the three routes can occur to a significant extent with a high percentage. The major molecular yields of formaldehyde and acetone (route III) in the reactions with Cl atoms in the presence of NO_x versus those of Cl atoms reactions in the absence of NO_x could indicate that in the absence of NO_x the self reactions of peroxy radicals (RO₂) reaction “via the” molecular pathway channel is more favored than “via” the radical pathway channel. For the NO₃ radical, routes I and II (attack at their β-position) with the formation of nitrated compounds seem to be the major unique routes.

3.2.3. -3,3DM2ButOH

The analysis of the FTIR spectra obtained for the reactions of 3,3DM2ButOH with Cl atoms, in the presence and absence of NO_x, OH radical and NO₃ radical shows the formation of 3,3-dimethyl-2-butanone as a main product (see Fig. S9). Other compounds, such as formaldehyde, acetone, 2,2-dimethylpropanal and pPeroxy aAcetyl nNitrate (PAN), were have also been observed. The residual FTIR spectra after subtraction of all known IR bands, again shows the presence of carbonyl compounds (IR bands absorption in the range of 1820–1700 cm⁻¹); hydroxy compounds (1060–1040 cm⁻¹) in the reaction with Cl atoms in the absence of NO_x, and also nitrated compounds (RONO₂; 1650, 1305–1260, 890 cm⁻¹) in the reaction with Cl in the presence of NO_x + NO and NO₃ radical (Fig. S9C). The presence in the residual an FTIR spectra of an IR absorption band around 1800 cm⁻¹ for the reaction with Cl atoms at longer reaction times could be due to the formation of chlorine compounds by reaction of 3,3-dimethyl-2-butanone with Cl₂ (Ren et al., 2018) or the formation of cyclic compounds such as hydrofurans. The SPME/GC-TOFMS chromatograms and MS spectra (Fig. S10 and S11) confirm the presence of other reaction products apart from 3,3-dimethyl-2-butanone in the case of Cl (absence/presence of NO_x), Cl + NO and NO₃ reactions. Only one significant peak is observed in chromatograms of obtained for the OH reactions. The estimated molecular yields of 3,3-dimethyl-2-butanone for all individual experiments are given in Table 5, where the measured concentrations have been corrected for secondary reactions. Acetone, formaldehyde, 2,2-dimethylpropanal, nitrated compounds and acetaldehyde have were also been quantified. Plots of concentration versus time for formaldehyde, acetone (Fig. S12A) and nitrated compounds in the reactions of Cl in the presence of NO_x (Fig. S12B) show typical profiles with two trends. This type of profile indicates that of secondary reactions for formaldehyde, acetone (Fig. S12A) and nitrated compounds in the reactions of Cl with NO (Fig. S12B). formaldehyde and acetone These compounds could also be formed by degradation of 3,3-dimethyl-2-butanone (Fig. S13). The estimated molecular yields of 3,3 dimethyl 2 butanone for all individual experiments are given in Table 5, where the measured concentrations have again been corrected again for secondary reactions. Table 6 summarizes the molecular yields of all quantified products.

1 ~~A~~Total carbon yields of ~60% ~~/~~100%, 90% and 60% have been accounted for Cl (in the absence and presence
2 of NO_x), OH and NO₃ reactions, respectively (See Table 6). It is important to note that in the case of the reaction
3 of Cl atoms in the absence of ~~without~~ NO_x, where the total carbon yield ~~was~~ lower than 100%, there ~~were~~
4 many reaction products that could not be quantified, such as ~~dihydroxy-carbonyl~~ and/or hydroxy-carbonyl
5 compounds. In the reaction ~~with~~ of NO₃ radical, due to our experimental conditions, ~~significant number~~ ~~an~~
6 ~~important amount~~ of primary nitrated compounds ~~was~~ expected to be formed (Fig. S9C).

7 A mechanism of hydrogen ~~atom~~ abstraction ~~at~~ different positions ~~on~~ the carbon chain has been proposed for the
8 reaction of 3,3DM2ButOH with Cl, OH and NO₃ ~~reactions~~. The mechanism is shown in Figure 8. Table S3, in
9 ~~S~~supplementary ~~M~~aterial, shows a summary of the reaction products ~~either observed or tentatively identified~~
10 proposed ~~by~~ in this mechanism ~~observed or tentatively identified in~~ of the reactions of 3,3DM2ButOH with the
11 atmospheric oxidants.

12 Molecular yields of 3,3-dimethyl-2-butanone obtained in this work imply ~~a~~ percentages of attack of the oxidant ~~at~~
13 ~~the~~ ~~n~~-~~α~~-~~position~~ (route II) of: 43% ~~/~~and 44%, in the case of ~~Cl~~ chlorine atom (absence/presence of NO_x ~~without~~
14 ~~and with NO respectively~~); 81% for ~~the~~ OH radical and 58% for ~~the~~ NO₃ radical. Percentages are very similar to
15 ~~those~~ at predicted by SARs ~~method~~ except for ~~the~~ NO₃ radical (See Table 5). High NO₂ concentration present in
16 the reactions ~~with~~ NO₃ radical would ~~greatly~~ ~~highly~~ favor the formation of nitrated compounds ~~over~~ ~~versus~~ 3,3-
17 dimethyl-2-butanone. ~~It could justify the low estimated molecular yield for 3,3- dimethyl-2-butanone.~~

18 The main reaction products observed in the reaction ~~with~~ of Cl atoms in ~~the~~ presence of NO_x (3,3-dimethyl-2-
19 butanone, formaldehyde, 2,2-dimethylpropanal, acetone ~~and~~, acetaldehyde) confirm ~~that~~ ~~the~~ ~~attack~~ of Cl atoms
20 ~~could~~ ~~attack~~ ~~in~~ ~~at~~ other sites (apart ~~from~~ ~~the~~ ~~of~~ ~~α~~-~~position~~) ~~with~~ ~~an~~ ~~important~~ ~~percentage~~. Based on the estimated
21 molecular yield of acetone, ~~the~~ ~~attack~~ ~~at~~ ~~the~~ ~~in~~ ~~δ~~-~~position~~ with abstraction of hydrogen ~~atom~~ ~~from~~ ~~of~~ methyl groups
22 (route III) could be ~58%, and based on the estimated molecular yield ~~offer~~ 2,2-dimethylpropanal, the attack ~~in~~
23 ~~at~~ ~~the~~ ~~β~~-~~position~~ (route I) could account ~~with~~ ~~a~~ ~~for~~ 10%. These data agree with the SAR predictions for Cl atom
24 reactions. On the other hand, the major molecular yields of acetone, formaldehyde and acetaldehyde (route III) in
25 the reaction ~~with~~ of Cl atoms in ~~the~~ presence of NO_x ~~rather~~ than in ~~its~~ absence ~~of~~ ~~NO~~ could indicate that in ~~the~~
26 presence of NO_x the ~~self-reactions of peroxy radicals self reactions (RO₂)~~ “via” ~~the~~ molecular ~~pathway~~ ~~channel~~ is
27 negligible. The lower yield ~~(17%)~~ ~~estimated~~ ~~(17%)~~ ~~by~~ ~~of~~ acetaldehyde versus 58% of its coproduct (acetone) is
28 due to its fast degradation by ~~Cl atoms~~ reaction with ~~Cl atoms with the~~ formation ~~of Peroxy acetyl nitrates PAN~~
29 ~~as it has been~~ observed in the FTIR experiments (See Fig. S9B). ~~Moreover, the molecular yield of acetone could~~
30 ~~be overestimated due to secondary reactions as can be seen in Figs. S12 and FigS13.~~

31 4. Atmospheric Implications

32 ~~The~~ ~~p~~ollutants in the atmosphere, ~~can~~ ~~could~~ create serious environmental problems, such ~~a~~ photochemical smog,
33 acid rain and degradation of the ozone layer (Finlayson-Pitts and Pitts, 2000). So, it is important to evaluate the
34 parameters that help us to ~~understand~~ ~~know~~ the impact of the presence of these compounds in the atmosphere.
35 These parameters are; the time that such compounds remains in the atmosphere, their ~~g~~lobal ~~w~~arming
36 ~~p~~otential (GWP) and their ~~mechanisms of~~ degradation, ~~mechanisms~~ in order to estimate the atmospheric effect
37 of products formed.

The first important parameter of the environmental impact of ~~an~~ ~~O~~xygenated ~~v~~olatile ~~o~~rganic ~~c~~ompounds in the atmosphere, is the global lifetime, τ_{global} , which considers all the degradation processes ~~that~~ ~~which~~ could ~~affect~~ ~~suffer~~ these compounds in the ~~t~~roposphere. This parameter can be obtained from the sum of the individual sink processes such as reactions initiated by OH and NO₃ radicals, Cl atoms, and O₃ molecules; photolysis and dry and wet deposition, Eq (4):

$$\tau_{global} = \left[\frac{1}{\tau_{OH}} + \frac{1}{\tau_{Cl}} + \frac{1}{\tau_{NO_3}} + \frac{1}{\tau_{O_3}} + \frac{1}{\tau_{photolysis}} + \frac{1}{\tau_{other\ processes}} \right]^{-1} \quad (4)$$

The ~~T~~ropospheric lifetime (τ) of 4MCHexOH, 3,3DM1ButOH and 3,3DM2ButOH for each process have been estimated ~~by~~ considering Eqs (4) and (5):

$$\tau = \frac{1}{k_{Ox}[Ox]} \quad (5)$$

where k_{Ox} and [Ox] are the rate coefficients ~~s~~ obtained in this work for each oxidant and ~~the~~ typical atmospheric concentrations ~~s~~ of the oxidants Cl, OH and NO₃, respectively. Concentrations employed were as follows: for 24 ~~hours~~ average: 1×10^3 atoms cm⁻³ (Platt and Janssen, 1995) for ~~C~~hlorine atoms, 12-~~hours~~ average day-time concentration of 1×10^6 radicals cm⁻³ for OH (Prinn et al., 2001) and 5×10^8 radicals cm⁻³ for NO₃ ~~radicals~~ (Atkinson, 2000), and a peak concentration of ~~C~~hlorine atoms of 1.3×10^5 atoms cm⁻³ in the coastal marine boundary layer at dawn (Spicer et al., 1998). Reactions with O₃ and photolysis are negligible loss processes for this ~~type~~ ~~kind~~ of compounds (Mellouki et al., 2015). Other processes are referred to ~~as~~ ~~dry~~ ~~and~~ wet deposition. ~~For~~ ~~To~~ ~~estimate~~ ~~ing~~ the lifetime associated with wet deposition, Eq (6) proposed by (Chen et al. 2003) ~~was~~ ~~has~~ ~~been~~ used:

$$\tau_{wet} = \frac{H_{atm}}{v_{pm}RTk_H} \quad (6)$$

~~w~~Where k_H is the Henry's law constant, H_{atm} is the height ~~of~~ ~~in~~ the troposphere, taking a value of 630-~~m~~, v_{pm} is the average precipitation rate for Ciudad Real (Spain) (402-~~mm~~ ~~yr~~ ~~ear~~) (www.aemet.es), R is the gases constant and T is the temperature, considered ~~as~~ ~~to~~ ~~be~~ constant and equal to 298 K. In ~~the~~ ~~literature~~ ~~bibliography~~ there ~~are~~ ~~is~~ only data of ~~the~~ ~~constant~~ ~~of~~ ~~Henry's~~ ~~constant~~ for 3,3DM2ButOH (5.6×10^{-1} mol m⁻³ Pa⁻¹) (Sander, 2015). Comparing the available data for similar compounds ~~it~~ ~~has~~ ~~been~~ ~~used~~ ~~an~~ approximated values ~~s~~ of K_H of 3-~~mol~~ ~~m~~⁻³ Pa⁻¹ and 0.4-~~mol~~ ~~m~~⁻³ Pa⁻¹ for 4MCHexOH and 3,3DM1ButOH, respectively ~~have~~ ~~been~~ ~~used~~.

~~The~~ ~~calculated~~ ~~l~~ifetimes ~~e~~alculated of the three ~~studied~~ alcohols ~~studied~~ in this work are shown in ~~the~~ Table 7. It can be seen that the dominant tropospheric loss process for the three alcohols is clearly their reaction with OH radicals followed by their reaction with NO₃ radicals at night. However, in places where there is a peak concentration of chlorine atoms (coastal areas) the reaction of these alcohols with chlorine atoms may compete with ~~that~~ ~~with~~ OH radicals ~~as~~ ~~becoming~~ their main degradation process.

The global lifetime of the three alcohols is of the order of ~ 1 -~~2~~ days, indicating that these compounds will probably be degraded near their sources. These global lifetimes also indicate that ~~MSAs~~ do not ~~make~~ ~~have~~ a significant contribution to ~~the~~ radiative forcing of climate change (Mellouki et al., 2015), which is supported through the estimation of their GWP values. For ~~a~~ time-horizon of 20 years, the ~~values~~ ~~estimated~~ ~~values~~ ~~are~~ ~~have~~ ~~been~~: 8.33×10^{-4} , 1.78×10^{-2} and 5.80×10^{-3} for 4MCHexOH, 3,3DM1ButOH and 3,3DM2ButOH respectively,

1 which are very low. So, these compounds will only have an important impact in the troposphere at a local or
2 regional level.

3 Their degradation products (mostly carbonyl-containing compounds and nitro-compounds in polluted areas) must
4 be considered. Thus, the nitrated compounds generated can act as NO_x reservoir species especially during the
5 night (Altshuler, 1993) and could have an influence at the global scale. Moreover, since 4MCHexOH,
6 3,3DM1ButOH and 3,3DM2ButOH react quickly with Chlorine atoms and OH radicals, their contribution to the
7 formation of photochemical smog might be important. For that reason, the contribution of these three alcohols to
8 the formation of smog was estimated by obtaining the average ozone production during 99-% of their reactions
9 with OH radical, using the equation indicated by Dash and Rajakumar (2013). The values obtained were 3.24, 0.90
10 and 1.69 ppm for 4MCHexOH, 3,3DM1ButOH and 3,3DM2ButOH, respectively. These values suggest that these
11 compounds may be a potential generators of tropospheric ozone and could contribute significantly to the formation
12 of photochemical smog (depending on their concentrations in the atmosphere).

13 5. Conclusions

14 The main conclusions that have been obtained with from the present study, are the following:

15 -- The kinetic and product study support that: 1 -- The atmospheric degradation mechanism for MSAs, and
16 possibly for other the rest of unstudied saturated alcohols, proceeds by abstraction of the a hydrogen atom bonded
17 to a carbon rather than a instead of hydrogen atoms bonded to the oxygen atom of the alcohol group. and. 2 -- The
18 reaction mechanism in the H atom-abstraction process depends on the oxidant. Chlorine atoms abstract any type
19 of alkyl hydrogen (α , β , δ) from SAsaturated alcohols with a high percentage, compared to the OHhydroxyl
20 radical and the NO₃ nitrate radicals. The OH and NO₃ radicals abstract mainly the hydrogen atom at the α -
21 position, if the saturated alcohols are secondary. For primary alcohols, the abstraction of a hydrogen atom at their
22 β -position could be also be important in the reaction with NO₃ radical. Therefore, more kinetic studies of the for
23 NO₃ radical reaction with primary alcohols are necessary to quantify the effect of the OH group at the β -position
24 ($-CH_2OH$) and to update the SAR method developed by Kerdouci et al., and to quantify the effect of the OH group
25 at their β position, ($-CH_2OH$).

26 -- Theoretical ab-initio studies of the reactions of MSAs with atmospheric oxidants should be performed in
27 order to obtain more information about their reaction mechanisms in the H atom-abstraction process.

28 -- The atmospheric conditions determine the reaction mechanism and therefore the reaction products obtained in
29 the degradation of methyl saturated alcohols SAs. So, in polluted environments with high concentrations of NO_x,
30 the peroxy radicals (RO_2) reacts mainly with NO to form the alkoxy radical instead of molecular compounds. In
31 these conditions, nitrated organic compounds (RONO₂) are formed as well as apart from polyfunctional organic
32 compounds. Also, when the concentration of NO₂ is higher than that of NO concentration, ozone is formed. In a
33 clean atmosphere, as in the case of the experiments with Cl atoms in the absence of NO_x, the reaction products
34 are different because of peroxy radicals (RO_2) could react mainly "via" a self-reaction molecular pathway channel
35 instead of to "via" a self-reaction radical pathway with formation of dihydroxy and hydroxycarbonyl
36 compound channel.

1 ~~--~~The unquantified ~~counted~~ polyfunctional organic compounds could explain the low carbon balance obtained in
2 the Cl or NO₃ reactions. The carbon balance must be taken with caution since the calculated molecular yields have
3 a high degree of uncertainty.

4 ~~--~~Calculated lifetimes for ~~methy~~ saturated alcohols (in the order of ~1 day) imply that these compounds are
5 pollutants at a local ~~--~~ regional scale, but it is also important to indicate that MSAs are sources of stable nitrated
6 compounds (ROONO₂), depending on environmental conditions, that can travel ~~to~~ large distances from their
7 sources and contribute ~~ing~~ to form ozone in clean areas, for example in forest or rural areas.

8 ~~--~~The main products ~~of coming from~~ the degradation of the MSAs, aldehydes and ketones, develop a very
9 important secondary chemistry with the formation of products of special relevance, such as the PAN observed in
10 the degradation of 3,3-dimethyl-2-butanol. More experiments should be carried out ~~one~~ using other detection
11 techniques, in order to evaluate the formation of secondary ~~S~~organic ~~Q~~aerosol (SOA) ~~A~~ because it is well known
12 that polyfunctional organic compounds are important SOA precursors.

13 ~~--~~From the environmental point of view, this work shows that the degradation of MSAs is an important source of
14 pollutants in the atmosphere ~~of with~~ greater or lesser impact depending on the environmental conditions and the
15 quantities of these alcohols present in the atmosphere. Therefore, the use of MSAs as additives in the production
16 of biofuels ~~as biofuels~~ should be controlled, ~~avoiding that a bad as poor~~ handling ~~could result~~ involves in high
17 concentrations of these alcohols in the atmosphere.

18 ~~--~~The ~~r~~Rate coefficients and reaction products ~~measured reported~~ in this work are the first available data, so this
19 work contributes to a better understanding of the atmospheric chemistry of oxygenated compounds, expanding
20 the kinetic and mechanistic database, and ~~additionally it~~ contributes to develop ~~ing~~ or to improve ~~ing~~ predictive ~~veon~~
21 models ~~that which~~ help us to avoid or mitigate the effects of climate change or air quality. However, kinetic
22 experiments in the tropospheric temperature range are necessary to obtain more information about the reaction
23 mechanism and to extrapolate the ~~data of~~ rate coefficients ~~data~~ to other typical atmospheric conditions and thus be
24 able to better establish the atmospheric impact of these ~~se~~ alcohols.

25 6. Supplementary material.

26 Attached in a separated file.

27 7. Author contribution

28 Salgado S. and Martín P. designed the experiments. Cabañas B. is the leader of the group and the responsible to
29 control the research and got the financial support for the project leading to this publication. Colmenar I. carried
30 out the experiments of 4MCHexOH. Tapia A. carried out the kinetic experiments of 3,3DM1ButOH and
31 3,3DM2ButOH and Aranda I. carried out the product experiments of 3,3DM1ButOH and 3,3DM2ButOH. Martín
32 P. supervised all analysis of data and prepared the manuscript with contributions from all co-authors.

33 8. Competing interests

34 The authors declare that they have no conflict of interest.

35 9. Acknowledgment

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3 **10. References**

- 4 -Allert, M., Rizk, S. S., Looger, L. L., Hellinga, H. W., and Wells, J. A Computational Design of Receptors for
5 an Organophosphate Surrogate of the Nerve Agent Soman. *Proc. Natl. Acad. Sci.* 101, 21. 7907-7912,
6 <https://doi.org/10.1073/pnas.0401309101>, 2004.
- 7 -Altshuller A. P. PANs in the Atmosphere, *Air & Waste*, 43:9, 1221-1230,
8 <https://doi.org/10.1080/1073161X.1993.10467199>, 1993.
- 9 -Andersen, V. F., Wallington, T. J., Nielsen, O. J. Atmospheric Chemistry of i-Butanol. *J. Phys. Chem. A.*, 114,
10 12462–12469, <https://doi.org/10.1021/jp107950d>, 2010.
- 11 -AOPWIN, v1.92. ©2000 U.S. Environmental Protection Agency.
- 12 -Aschmann, S. M. Arey, J. and Atkinson R. Kinetics and Products of the Reactions of OH Radicals with 4,4-
13 Dimethyl-1-pentene and 3,3-Dimethylbutanal at 296 ± 2 K. *J. Phys. Chem. A*, 114, 5810–5816,
14 <https://doi.org/10.1021/jp101893g>, 2010.
- 15 -Atkinson, R. Gas-phase tropospheric chemistry of organic compounds. *J. Phys. Chem. Ref. Data*. Monograph, 1-
16 216, 1994.
- 17 -Atkinson, R. Atmospheric chemistry of VOCs and NO(x). *Atmos. Environ.* 34(12-14), 2063-2101,
18 [https://doi.org/10.1016/S1352-2310\(99\)00460-4](https://doi.org/10.1016/S1352-2310(99)00460-4), 2000.
- 19 -Atkinson, R. Kinetics of the gas-phase reactions of OH radicals with alkanes and cycloalkanes. *Atmos. Chem.*
20 *Phys.* 3, 2233-2307, <https://doi.org/10.5194/acp-3-2233-2003>, 2003.
- 21 -Atkinson, R. and Arey J. Atmospheric degradation of volatile organic compounds. *Chem Rev* 103:4605–4638,
22 <https://doi.org/10.1021/cr0206420>, 2003.
- 23 -Atkinson, R., Aschmann, S.M., Pitts Jr., J.N. Rate constants for the gas-phase reactions of the NO₃ radical with a
24 series of organic compounds at 296 ± 2 K. *J. Phys. Chem.* 92, 3454-3457, <https://doi.org/10.1021/j100323a028>,
25 1988.
- 26 -Atkinson, R., Aschmann, S.M., Pitts Jr. J.N. Rate Constants for the Gas-Phase Reactions of the OH Radical with
27 a Series of Aromatic Hydrocarbons at 296 ± 2 K. *Int. J. Chem. Kinet.* 21, 355-365,
28 <https://doi.org/10.1002/kin.550210506>, 1989.
- 29 -Atkinson, R., Baulch, D. L., Cox, R. A., Crowley, J. N., Hampson, R. F., Hynes, R. G., Jenkin, M. E., Rossi, M.
30 J., Troe, J., and IUPAC Subcommittee. Evaluated kinetic and photochemical data for atmospheric chemistry:
31 Volume II – gas phase reactions of organic species, *Atmos. Chem. Phys.*, 6, 3625-4055, [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-6-3625-2006)
32 [6-3625-2006](https://doi.org/10.5194/acp-6-3625-2006), 2006.
- 33 -Atkinson, R., Plum, C.N., Carter, W.P.L., Winer, A.M., Pitts Jr. J.N. Rate constants for the gas-phase reactions
34 of nitrate radicals with a series of organics in air at 298 ± 1 K. *J. Phys. Chem.* 88, 1210-1215,
35 <https://doi.org/10.1021/j150650a039>, 1984.
- 36 -Ballesteros, B., Garzón, A., Jiménez, E., Notario, A., Albaladejo, J. Relative and absolute kinetic studies of 2-
37 butanol and related alcohols with tropospheric Cl atoms *J. Phys. Chem. Chem. Phys.* 9 (10), 1210-1218,
38 <https://doi.org/10.1039/B614531K>, 2007.

1 -Bradley, W.R., Wyatt, S.E., Wells, J.R., Henley, M.V., Graziano, G.M. The Hydroxyl Radical Reaction Rate
2 Constant and Products of Cyclohexanol. *Int. J. Chem. Kinet.*, 33, 108-117, [https://doi.org/10.1002/1097-
3 4601\(200102\)33:2<108::AID-KIN1002>3.0.CO;2-%23](https://doi.org/10.1002/1097-4601(200102)33:2<108::AID-KIN1002>3.0.CO;2-%23), 2001.

4 -Brauers, T., Finlayson-Pitts, B.J. Analysis of relative rate measurements. *Int. J. Chem. Kinet.* 29 (9), 665-672,
5 [https://doi.org/10.1002/\(SICI\)1097-4601\(1997\)29:9<665::AID-KIN3>3.0.CO;2-S](https://doi.org/10.1002/(SICI)1097-4601(1997)29:9<665::AID-KIN3>3.0.CO;2-S), 1997.

6 -Calvert, J.G., Mellouki, A., Orlando, J.J., Pilling, M.J., Wallington, T.J. The mechanisms of atmospheric
7 oxidation of the oxygenates. Oxford University Press, New York, 2011.

8 -Caravan, R. L., Shannon, R. J., Lewis, T., Blitz, M. A., and Heard, D. E. Measurements of Rate Coefficients for
9 Reactions of OH with Ethanol and Propan-2-ol at Very Low Temperatures. *J. Phys. Chem. A*, 119, 7130–7137,
10 <https://doi.org/10.1021/jp505790m>, 2015.

11 -Cavalli, F., Geiger, H., Barnes, I., Becker, K. H. FTIR Kinetic, Product, and Modeling Study of the OH-Initiated
12 Oxidation of 1-Butanol in Air. *Environ. Sci. Technol.*, 36, 1263–1270, <https://doi.org/10.1021/es010220s>, 2002.

13 -Ceacero-Vega, A. Ballesteros, B., Albaladejo, J., Bejan, I., and Barnes, I. Temperature dependence of the gas-
14 phase reactions of Cl atoms with propene and 1-butene between 285 < T < 313 K. *Chem. Phys. Lett.* 484(1-3):10–
15 13, <https://doi.org/10.1016/j.cplett.2009.10.080>, 2009.

16 -Ceacero-Vega, A.A., Ballesteros, B., Bejan, I., Barnes, I., Jiménez, E., Albaladejo, J. Kinetics and Mechanisms
17 of the Tropospheric Reactions of Menthol, Borneol, Fenchol, Camphor, and Fenchone with Hydroxyl Radicals
18 (OH) and Chlorine Atoms (Cl). *J. Phys. Chem. A*, 116, 4097-4107, <https://doi.org/10.1021/jp212076g>, 2012.

19 -Chen, L., Takenaka N., Bandow, H. Maeda, Y. Henry's law constants for C2–C3 fluorinated alcohols and their
20 wet deposition in the atmosphere. *Atmos. Environ.* 37, 34,4817-4822,
21 <https://doi.org/10.1016/j.atmosenv.2003.08.002>, 2003.

22 -Cheung C.S., Huang Z. Effect of n-pentanol addition on the combustion, performance and emission characteristics
23 of a direct-injection diesel engine. *Energy*; 70:172–80, <https://doi.org/10.1016/j.energy.2014.03.106>, 2014.

24 -D'Anna, B., Andresen, W., Gefen, Z., Nielsen, C.J. 2001. Kinetic study of OH and NO₃ radical reactions with 14
25 aliphatic aldehydes. *Phys. Chem. Chem. Phys.* 3, 15, 3057-3063, <https://doi.org/10.1039/B103623H>, 2001.

26 -Dash, M.R. and Rajakumar, B. Experimental and theoretical rate coefficients for the gas phase reaction of β-
27 Pinene with OH radical. *Atmos. Environ.*, 79, 161-171, <https://doi.org/10.1016/j.atmosenv.2013.05.039>, 2013.

28 -Ezzel, M.J., Wang, W., Ezell, A.A., Soskin, G., Finlayson- Pitts, B.J. **2002**. Kinetics of reactions of chlorine atoms
29 with a series of alkenes at 1 atm and 298 K: structure and reactivity. *Phys. Chem. Chem. Phys.*, 1, 5813-5820,
30 <https://doi.org/10.1039/B207529F>, 2002.

31 -Farrugia, L.N., Bejan, I., Smith, S.C., Medeiros, D.J., Seakins, P.W. Revised structure activity parameters derived
32 from new rate coefficient determinations for the reactions of chlorine atoms with a series of seven ketones at 290
33 K and 1 atm. *Chem. Phys. Lett.* 640 87–93, <https://doi.org/10.1016/j.cplett.2015.09.055>, 2015.

34 -Finlayson-Pitts, B. J. and Pitts, J.N. Chemistry of the Upper and Lower Atmosphere: Theory, Experiments, and
35 Applications” Academic Press, San Diego, 2000.

36 -Gallego-Iniesta, M.P., Moreno, A., Martín, P., Tapia, A., Cabañas, B., Salgado. M.S. Reactivity of 2-ethyl-1-
37 hexanol in the atmosphere. *Phys. Chem. Chem. Phys.* 12, 3294-3300, <https://doi.org/10.1039/B923899A>, 2010.

38 -Gallego-Iniesta, M.P., Cabañas, B., Salgado. M.S. Martinez, E., Martin, P. Estimation of gas-phase rate
39 coefficients for the reactions of a series of α,β–unsaturated esters with OH, NO₃, O₃ and Cl. *Atmos. Environ.* 90,
40 133-145, <https://doi.org/10.1016/j.atmosenv.2014.03.036>, 2014.

1 -Garzón, A., Cuevas, C.A., Ceacero, A.A., Notario, A., Albaladejo, J. and Fernández-Gómez, M. Atmospheric
2 reactions $\text{Cl} + \text{CH}_3 - (\text{CH}_2)_n - \text{OH}$ ($n=0-4$): A kinetic and theoretical study. *J. Chem. Phys.* 125., 104305-104315,
3 <https://doi.org/10.1063/1.2244556>, 2006.

4 -Grosjean, D. Atmospheric chemistry of alcohols. *J. Braz. Chem. Soc.* 8, 433-442,
5 <http://dx.doi.org/10.1590/S0103-50531997000500002>, 1997.

6 -Herath, N.T., Orozco, I., Clinch, E.C., Marshall, P. Relative Rate Studies of the Reactions of Atomic Chlorine
7 with Acetone and Cyclic Ketones. *Int. J. Chem. Kinet.* 50, 41-46, <https://doi.org/10.1002/kin.21138>, 2018.

8 -Hurley, M. D., Wallington, T. J., Laursen, L., Javadi, M. S., Nielsen, O. J., Yamanaka, T., Kawasaki, M.
9 Atmospheric Chemistry of n-Butanol: Kinetics, Mechanisms, and Products of Cl Atom and OH Radical Initiated
10 Oxidation in the Presence and Absence of NO_x. *J. Phys. Chem. A*, 113, 7011–7020,
11 <https://doi.org/10.1021/jp810585c>, 2009.

12 -Jenkin, M.E., Valorso, R., Aumont, B., Rickard, A.R., Wallington, T.J. Estimation of rate coefficients and
13 branching ratios for gas-phase reactions of OH with aliphatic organic compounds for use in automated mechanism
14 construction. *J. Atmos. Chem. Phys.* 18, 9297-9328, <https://doi.org/10.5194/acp-18-9297-2018>, 2018.

15 -Jiménez, E., Lanza, B., Garzón, A., Ballesteros, B., Albaladejo, J. Atmospheric Degradation of 2-Butanol, 2-
16 Methyl-2-butanol, and 2,3-Dimethyl-2-butanol: OH Kinetics and UV Absorption Cross Sections. *J. Phys. Chem.*
17 *A*, 109, 10903-10909, <https://doi.org/10.1021/jp054094g>, 2005.

18 -Karabektas, M. and Hosoz, M. Performance and emission characteristics of a diesel engine using isobutanol–
19 diesel fuel blends, *Renew Energy*, 34 (6), 1554-1559, <https://doi.org/10.1016/j.renene.2008.11.003>, 2009.

20 -Kerdouci, J., Picquet-Varrault, B., Doussin, J.F. Prediction of Rate Constants for Gas-Phase Reactions of Nitrate
21 Radical with Organic Compounds: A New Structure–Activity Relationship. *Chem. Phys. Chem.*, 11, 3909-3920,
22 <https://doi.org/10.1002/cphc.201000673>, 2010.

23 -Kerdouci, J., Picquet-Varrault, B., Doussin, J.F. Structure–activity relationship for the gas-phase reactions of NO₃
24 radical with organic compounds: Update and extension to aldehydes. *Atmos. Environ.* 84, 363-372,
25 <https://doi.org/10.1016/j.atmosenv.2013.11.024>, 2014.

26 -Kumar., B.R. and Saravanan, S. Use of higher alcohol biofuels in diesel engines: A review. *Renew. Sust. Energ.*
27 *Rev.* 60, 84–115, <https://doi.org/10.1016/j.rser.2016.01.085>, 2016.

28 -Kwok, E.S.C. and Atkinson R. Estimation of Hydroxyl Radical Reaction Rate Constants For Gas-Phase Organic
29 Compounds Using A Structure-Reactivity Relationship: An Update. *Atmos. Environ.*, 29, 1685-1695,
30 [https://doi.org/10.1016/1352-2310\(95\)00069-B](https://doi.org/10.1016/1352-2310(95)00069-B), 1995.

31 -Lapuerta, M., García-Contreras, R., Campos-Fernández, J., and Dorado P. Stability, lubricity, viscosity, and cold-
32 flow properties of alcohol–diesel blends, *Energy Fuels*, 24, 4497-4502, <https://doi.org/10.1021/ef100498u>, 2010.

33 -Li, F., Yi, B., Song, L., Fu, W., Liu, T., Hu, H., & Lin, Q. Macroscopic spray characteristics of long-chain alcohol-
34 biodiesel fuels in a constant volume chamber. *Proceedings of the Institution of Mechanical Engineers, Part A:*
35 *JPE*, 232(2), 195–207. <https://doi.org/10.1177/0957650917721336>, 2017.

36 -Lopes, M., Serrano, L., Ribeiro, I., Cascao, P., Pires, N. Emissions characterization from EURO 5 diesel/biodiesel
37 passenger car operating under the new European driving cycle. *Atmos. Environ.* 84, 339-348,
38 <https://doi.org/10.1016/j.atmosenv.2013.11.071>, 2014.

1 - Martin, P., Cabañas, B., Colmenar, I., Salgado, M.S., Villanueva, F., Tapia, A. Reactivity of E-butenedial with
2 the major atmospheric oxidants. *Atmos. Environ.* 70, 351-360, <https://doi.org/10.1016/j.atmosenv.2013.01.041>,
3 2013

4 - McGillen, M. R., Baasandorj, M., Burkholder, J. B. Gas-Phase Rate Coefficients for the OH + n-, i-, s-, and t-
5 Butanol Reactions Measured Between 220 and 380 K: Non-Arrhenius Behavior and Site- Specific Reactivity. *J.*
6 *Phys. Chem. A*, 117, 4636–4656, <https://doi.org/10.1021/jp402702u>, 2013.

7 -McGillen, M.R., Tyndall, G.S., Orlando, J.J, Pimentel, A.S., Medeiros, D.J., and Burkholder J.B. Experimentally
8 Determined Site-Specific Reactivity of the Gas-Phase OH and Cl + i-Butanol Reactions Between 251 and 340 K.
9 *J. Phys. Chem. A*, 120, 9968–9981, <https://doi.org/10.1021/acs.jpca.6b09266>, 2016.

10 -Mellouki, A., Wallington, T. J., Chen, J. Atmospheric chemistry of oxygenated volatile organic compounds:
11 impacts on air quality and climate. *Chem Rev.*, 115:3984–4014, <https://doi.org/10.1021/cr500549n>, 2015.

12 -Moc, J and Simmie, J.M. Hydrogen Abstraction from n-Butanol by the Hydroxyl Radical: High Level Ab Initio
13 Study of the Relative Significance of Various Abstraction Channels and the Role of Weakly Bound Intermediates.
14 *J. Phys. Chem. A*, 114, 5558–5564, <https://doi.org/10.1021/jp1009065>, 2010.

15 -Moreno, A., Salgado, S., Martin, P., Martinez, E., and Cabañas, B. Kinetic Study of the Gas Phase Reactions of
16 a Series of Alcohols with the NO₃ Radical. *J. Phys. Chem. A*, 116, 42, 10383-10389,
17 <https://doi.org/10.1021/jp305682h>, 2012.

18 -Moreno, A., Salgado, S., Taccone, R., Martín, P., Cabañas, B. Atmospheric degradation of saturated alcohols:
19 room temperature rate coefficients for NO₃ radical reactions. *Atmos. Environ.*, 96, 229-235,
20 <https://doi.org/10.1016/j.atmosenv.2014.07.037>, 2014.

21 -Murty, M. R. V. S, Prasada, R. N., Prabhakar, S., and Vairamani, M. Chemical ionization mass spectral analysis
22 of pinacolyl alcohol and development of derivatization method using p-tolyl isocyanate. *Anal Methods* 2:1599–
23 1605, <https://doi.org/10.1039/C0AY00346H>, 2010.

24 -Nelson, L., Rattigan, O., Neavyn, R., Sidebottom, H., Treacy, J., Nielsen, O. J. Absolute and Relative Rate
25 Constants for the Reactions of Hydroxyl Radicals and Chlorine Atoms with a Series of Aliphatic Alcohols and
26 Ethers at 298 K. *Int. J. Chem. Kinet.*, 22, 1111– 1126. <https://doi.org/10.1002/kin.550221102>, 1990.

27 -Nguyen, H.T., Takenaka, N., Bandow, H., Maeda, Y., Oliva, S.T., Botelho, M.M. and Tavares T.M. Atmospheric
28 alcohols and aldehydes concentrations measured in Osaka, Japan and in Sao Paulo, Brazil. *Atmos. Environ.*, 35,
29 3075-3083. [https://doi.org/10.1016/S1352-2310\(01\)00136-4](https://doi.org/10.1016/S1352-2310(01)00136-4), 2001.

30 -Ozsezen, N. A., Turkcan, A., Sayin, C. and Canakci, M. Comparison of performance and combustion parameters
31 in a heavy-duty diesel engine fueled with iso-butanol/diesel fuel blends. *Energy. Explor. Exploit*, 29, 525-541,
32 <https://doi.org/10.1260/0144-5987.29.5.525>, 2011.

33 -Poutsma, M. L. Evolution of Structure–Reactivity Correlations for the Hydrogen Abstraction Reaction by
34 Chlorine Atom *J. Phys. Chem. A*, 117(4), 687–703. <https://doi.org/10.1021/jp310970t>, 2013.

35 -Platt U. and Janssen C. 1995. Observation and role of the free radicals NO₃, ClO, BrO and IO in the troposphere.
36 *Faraday Discussions.* <https://doi.org/100:175-198,10.1039/FD9950000175>, 1995.

37 -Prinn, R.G., Huang, J., Weiss, R.F., Cunnold, D.M., Fraser, P.J., Simmonds, P.G., McCulloch, A., Harth, C.,
38 Salameh, P., O'Doherty, S., Wang, R.H.J., Porter, L., Miller, R.B. Evidence for Substantial Variations of
39 Atmospheric Hydroxyl Radicals in the Past Two Decades *Science* 292, 1882-1888,
40 <https://doi.org/10.1126/science.1058673>, 2001.

1 -Ren, Y., Huang, Z., Miao, H., Di, Y., Jiang, D., Zeng, K., Liu, B., Wang, X. Combustion and emissions of a DI
2 diesel engine fuelled with diesel-oxygenate blends, *Fuel*, 87, 2691-2697,
3 <https://doi.org/10.1016/j.fuel.2008.02.017>, 2008.

4 -Ren Y. Wang J. Grosselin B. Daele V. and Mellouki A. Kinetic and product studies of Cl atoms reactions with a
5 series of branched ketones. *J. Environ. Science*. 71, 271-282, <https://doi.org/10.1016/j.jes.2018.03.036>, 2018.

6 -Sander, R. Compilation of Henry's law constants (version 4.0) for water as solvent *Atmos. Chem. Phys*, 15, 8,
7 4399-4981, <https://doi.org/10.5194/acp-15-4399-2015>, 2015.

8 -Sarathy, M., Obwald P., Hansen, N., and Kohse-Höinghaus, K. Alcohol combustion chemistry. *Prog. Energy*
9 *Combust. Sci.* 44. 40-102, <https://doi.org/10.1016/j.pecs.2014.04.003>, 2014.

10 -Schott, G., and Davidson, N. Shock Waves in Chemical Kinetics: The Decomposition of N₂O₅ at High
11 Temperatures, *J. Am. Chem. Soc.*, 80, 1841-1853, <https://doi.org/10.1021/ja01541a019>, 1958.

12 -Sikarwara, V.S, Zhao M., Fennell, P.S., Shah, N., Anthony, E.J. Progress in biofuel production from gasification.
13 *Prog. Energ. Combust.* 61, 189-248, <https://doi.org/10.1016/j.pecs.2017.04.001>, 2017.

14 -Smith, I. W. M., Ravishankara, A. R. J. Role of Hydrogen-Bonded Intermediates in the Bimolecular Reactions
15 of the Hydroxyl Radical. *Phys. Chem. A*, 106, 4798-4807, <https://doi.org/10.1021/jp014234w>, 2002.

16 -Spicer, C., Chapman, E.G., Fynallysson-Pitts, B.J., Plastridge, R.A., Hybbe, J.M., Fast, J.D., Berkowitz, C.M.
17 Unexpectedly high concentrations of molecular chlorine in coastal air *Nature*, 394, 353-356,
18 <https://doi.org/10.1038/28584>, 1998.

19 -Tapia, A., Villanueva, F., Salgado, S., Cabañas, B., Martinez, E. and Martin, P. Atmospheric degradation of 3-
20 methylfuran: kinetic and products study. *Atmos. Chem. Phys.*, 2011, 11, 3227-3241. [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-11-3227-2011)
21 [11-3227-2011](https://doi.org/10.5194/acp-11-3227-2011).

22 -Taylor, W.D., Alston, T.D., Moscato, M.J., Fazekas, G.B., Kozlowski, R., Takacs, G.A Atmospheric
23 photodissociation lifetimes for nitromethane, methyl nitrite, and methyl nitrate. *Int. J. Chem. Kinet.* 12, 231-240,
24 <https://doi.org/10.1002/kin.550120404>, 1980.

25 -Thevenet, R., Mellouki, A., Bras, G. L. Kinetics of OH and Cl Reactions with a Series of Aldehydes. *Int. J. Chem.*
26 *Kinet.*, 32, 676-685, [https://doi.org/10.1002/1097-4601\(2000\)32:11<676::AID-KIN3>3.0.CO;2-V](https://doi.org/10.1002/1097-4601(2000)32:11<676::AID-KIN3>3.0.CO;2-V), 2000.

27 -Tuazon E. C. and Atkinson, R. A Product Study of the Gas-Phase Reaction of Isoprene with the OH Radical in
28 the Presence of NO_x. *Int. J. Chem. Kinet.* 22, 1221-1236, <https://doi.org/10.1002/kin.550221202>, 1990.

29 -Tuazon E. C., Leod, H.M, Atkinson, R. and Carter W.P.L. α-Dicarbonyl Yields from the NO_x Air Photooxidations
30 of a Series of Aromatic Hydrocarbons in Air. *Environ. Sci. Technol.*, 20, 4, 383-387,
31 <https://doi.org/10.1021/es00146a010>, 1986.

32 -Vereecken, L., Aumont, B., Barnes, I., Bozzelli, J.W., Goldman, M.J., Green, W.H., Madronich, S., McGillen,
33 M.R., Mellouki, A., Orlando, J. J., Picquet-Varrault, B., Rickard, R., Stockwell, W. R., Wallington, T.J., Carter,
34 W.P.L. Perspective on Mechanism Development and Structure-Activity Relationships for Gas-Phase Atmospheric
35 Chemistry. *Int. J. Chem. Kinet.* 50, 435-469, <http://dx.doi.org/10.1002/kin.21172>, 2018.

36 -Wallington, T.J. and Kurylo, M.J. Flash Photolysis Resonance Fluorescence Investigation of the Gas-Phase
37 Reactions of OH Radicals with a Series of Aliphatic Ketones over the Temperature Range 240-440 K. *J. Phys.*
38 *Chem.*, 91, 19, 5050-5054, <https://doi.org/10.1021/j100303a033>, 1987.

1 -Wallington, T.J., Skwes, L.M., Siegl, W.O, Wu. C., Japar, S.M. Gas phase reaction of Cl atoms with a series of
2 oxygenated organic species at 295 K, *Int. J. Chem. Kinet.*, 20, 867-875, <https://doi.org/10.1002/kin.550201105>,
3 1988.

4 -Wayne R.P., Barnes I., Biggs P., Burrows J.P., Canosa-Mas C.E., Hjorth J., LeBras G., Moortgat G.K., Perner
5 D., Poulet G., Restelli G., Sidebottom H. The nitrate radical: Physics, chemistry, and the atmosphere. *Atmos.*
6 *Environ.*, 25A, 1-203, [https://doi.org/10.1016/0960-1686\(91\)90192-A](https://doi.org/10.1016/0960-1686(91)90192-A), 1991.

7 -Wayne, R.P. Chemistry of Atmospheres. Oxford University Press, New York, 2000.

8 -Welz, O., Klippenstein, S., O., J., Harding, L. B., Taatjes, C. A., Zádor, J. Unconventional Peroxy Chemistry in
9 Alcohol Oxidation: The Water Elimination Pathway. *J. Phys. Chem. Lett.*, 43, 350-354,
10 <https://doi.org/10.1021/jz302004w>, 2013.

11 -Ziemann P., and Atkinson R. Kinetics, products, and mechanisms of secondary organic aerosol formation. *Chem*
12 *Soc Rev.* 41(19), 6582-6605, <https://doi.org/10.1039/C2CS35122F>, 2012.

13 -www.aemet.es
14 -www.capotchem.com
15 -www.iea.org. Energy and Air Pollution. World Energy Outlook 2016 Special Report Workshop International
16 Energy Agency.
17 -www.iupac-kinetic.ch.cam.ac.uk IUPAC Subcommittee on Gas Kinetic Data Evaluation.
18 -www.thermofisher.com/search/browse/results?customGroup=Spectral+Libraries)
19 -webbook.nist.gov/chemistry/
20 -<https://data.eurochamp.org/data-access/spectra/>
21

1 **Table 1.** Rate coefficient ratios, absolute rate coefficients and average rate coefficients for the reactions of a series
2 of MSAs with Cl atoms and OH and NO_3 radicals at 298 K and 720 ± 20 Torr of pressure. Rate coefficients, k , in
3 $\text{cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$.

Reaction	Reference	$(k_{\text{MSA}}/k_{\text{R}}) \pm 2\sigma$	$(k_{\text{MSA}} \pm 2\sigma) / 10^{-10}$	$(\bar{k}_{\text{MSA}} \pm 2\sigma) / 10^{-10}$
3,3DM1ButOH + Cl	1-butene	0.85 ± 0.03	2.89 ± 0.42	2.69 ± 0.16
		0.79 ± 0.02	2.68 ± 0.38	
	0.76 ± 0.02	2.58 ± 0.37		
	Propene	1.18 ± 0.02	2.63 ± 0.37	
		1.21 ± 0.03	2.70 ± 0.38	
1.22 ± 0.03	2.71 ± 0.38			
3,3DM2ButOH + Cl	1-butene	0.42 ± 0.01	1.42 ± 0.21	1.21 ± 0.07
		0.35 ± 0.01	1.17 ± 0.17	
	0.41 ± 0.01	1.38 ± 0.20		
	Propene	0.48 ± 0.01	1.08 ± 0.15	
		0.50 ± 0.02	1.12 ± 0.16	
0.56 ± 0.03	1.26 ± 0.19			
4MCHexOH + Cl	2-methylpropene	1.08 ± 0.03	3.69 ± 0.32	3.70 ± 0.16
		1.16 ± 0.02	3.95 ± 0.33	
		0.98 ± 0.05	3.35 ± 0.32	
	1-butene	1.14 ± 0.03	3.86 ± 0.56	
		1.12 ± 0.03	3.78 ± 0.55	
1.15 ± 0.04	3.90 ± 0.57			
Reaction	Reference	$(k_{\text{MSA}}/k_{\text{R}}) \pm 2\sigma$	$(k_{\text{MSA}} \pm 2\sigma) / 10^{-12}$	$\bar{k}_{\text{MSA}} \pm 2\sigma / 10^{-12}$
3,3DM1BuOH + OH	Isopropanol	1.00 ± 0.04	5.09 ± 0.20	5.33 ± 0.16
		1.13 ± 0.09	5.78 ± 0.47	
	1.12 ± 0.08	5.72 ± 0.40		
	2-methyl-2-butanol	1.60 ± 0.09	5.78 ± 1.01	
		1.57 ± 0.08	5.65 ± 1.00	
1.61 ± 0.09	5.79 ± 1.02			
3,3DM2BuOH + OH	Isopropanol	2.33 ± 0.09	11.90 ± 0.48	10.50 ± 0.25
		2.05 ± 0.08	10.50 ± 0.45	
	1.95 ± 0.08	9.95 ± 0.43		
	2-methyl-2-butanol	2.39 ± 0.09	8.61 ± 1.50	
		2.92 ± 0.09	10.50 ± 1.78	
2.25 ± 0.09	8.12 ± 1.34			
4MCHexOH + OH	Propene	0.64 ± 0.01	17.10 ± 2.59	18.70 ± 1.42
		0.76 ± 0.03	20.30 ± 3.19	
	0.76 ± 0.02	20.40 ± 3.10		
	Cyclohexene	0.27 ± 0.01	18.20 ± 4.55	
		0.27 ± 0.01	18.40 ± 4.62	
0.27 ± 0.01	18.00 ± 4.46			
Reaction	Reference	$(k_{\text{MSA}}/k_{\text{R}}) \pm 2\sigma$	$(k_{\text{SANO}_3} \pm 2\sigma) / 10^{-15}$	$\bar{k}_{\text{SANO}_3} \pm 2\sigma / 10^{-15}$
4MCHexOH + NO_3	1-butanol	1.08 ± 0.12	3.39 ± 1.11	2.69 ± 0.37
		1.81 ± 0.15	5.70 ± 1.82	
	0.79 ± 0.07	2.51 ± 0.80		
	2-ethyl-1-hexanol	0.71 ± 0.10	2.08 ± 0.72	
		1.00 ± 0.10	2.93 ± 0.96	
0.86 ± 0.08	2.52 ± 0.82			

4 ^aThe uncertainties for rate coefficients of MSAs (σ_{MSA}) were calculated from the uncertainty of slope of plots
5 (σ_{slope}) and the uncertainty of the reference (σ_{KR}) by using the propagation of uncertainties. ^bWeighted average
6 according to the equation $(w_1k_1 + w_2k_2 + \dots) / (w_1 + w_2 + \dots)$; ($w_i = 1/\sigma_i^2$). The uncertainty of weighted average (σ) was
7 given by $(1/w_1 + 1/w_2 + \dots)^{-0.5}$

1 **Table 2.** Estimated and experimental rate coefficients (k_{SAR} , k_{log} and k_{exp}) for the reaction of **MSAs** with atmospheric
 2 oxidants and ratio of rate coefficients (k_{exp}/k_{SAR} and k_{exp}/k_{log}), $k_{Cl}/10^{-11}$, $k_{OH}/10^{-12}$ and $k_{NO_3}/10^{-15}$ in $\text{cm}^3 \text{ molecule}^{-1}$
 3 s^{-1} units.

4

	4MCHexOH			3,3DM1ButOH			3,3DM2ButOH		
	k_{Cl}	k_{OH}	k_{NO_3}	k_{Cl}	k_{OH}	k_{NO_3}	k_{Cl}	k_{OH}	k_{NO_3}
k_{exp}	37.0 ^a	18.7 ^a	2.69 ^a	26.9 ^a	5.33 ^a	1.78 ^b	12.1 ^a	10.5 ^a	3.4 ^b
k_{SAR}	34.2 ^c	19.2 ^c	2.27 ^c	21.0 ^c	6.08 ^c	0.55 ^c	15.2 ^c	9.16 ^c	3.86 ^c
k_{exp}/k_{SAR}	1.08	0.97	1.18	1.28	0.88	3.24	0.80	1.15	0.88
k_{log}	30.7 ^d	-	4.69 ^d	13.9 ^d	-	1.16 ^d	21.4 ^d	-	3.19 ^d
k_{exp}/k_{log}	1.21	-	0.57	1.94	-	1.53	0.57	-	1.07

5 ^aData obtained in this work

6 ^bData obtained by Moreno et al., 2014

7 ^cSee S2 **S**supplementary **M**aterial

8 ^dEstimated using the correlations $\log_{Cl}-\log_{OH}$ (Eq. (2)) and $\log_{NO_3}-\log_{OH}$ (Eq. (3)) described by Calvert et al, 2011.

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1 **Table 3.** Experimental conditions and molecular yields of E-4-methylcyclohexanone for the reaction
 2 of 4MCHexOH with atmospheric oxidants.

MSA	Oxidant	Exp	[MSA] (ppm)	[Precursor] (ppm)	[NO] (ppm)	Carbonyl compound ^dYield (%)	Technique	Average^f (%)	SAR Yield (%)
4MCHexOH	Cl ^a	1	3	21	-	24.8±0.9	FTIR	25.2±1.9	14
		2	8	22	-	23.8±0.6	FTIR		
		3	13	16	-	27.5±0.2	SPME/GC- TOFMS ^e		
	Cl ^a + NO	1	11	23	30	30.4±0.9	FTIR	29.5±0.7	
		2	5	25	19	30.0±0.6	FTIR		
		3	7	13	12	31.6±1.3	SPME/GC- TOFMS ^e		
	OH ^b	1	7	36	23	35.1±1.3	FTIR	40.2±5.4	53
		2	13	31	29	38.2±1.5	FTIR		
		3	11	28	28.5	47.8±0.4	FTIR		
		4	6	19	12	39.8±0.9	SPME/GC- TOFMS ^e		
	NO ₃ ^c	1	3	6	-	56.8±11.4	SPME/GC- TOFMS ^e	58.0±23.5	86
		2	6	34	-	88.3±7.0	SPME/GC- TOFMS ^e		
		3	4	30	-	77.1±4.6	SPME/GC- TOFMS ^e		
		4	4	21	-	34.6±0.5	FTIR		
		5	7	10	-	33.4±0.6	FTIR		

3 ^a Rate coefficient k (in cm³ molecule⁻¹ s⁻¹ unit) used to correct the concentration of E-4-methylcyclohexanone by
 4 loss with the reaction of Cl atoms = 11.2 × 10⁻¹¹ (data of 2-methylcyclohexanone and Cl atoms (Herath et al.,
 5 2018)). Photolysis rate coefficient estimated for E-4-methylcyclohexanone under our experimental conditions, k_p
 6 = 5 × 10⁻⁵ s⁻¹

7 ^b Rate coefficient k (in cm³ molecule⁻¹ s⁻¹ unit) used to correct the concentration of E-4-methylcyclohexanone by
 8 loss with the reaction of OH radical = 13.7 × 10⁻¹² (estimated using AOPWIN, v1.92). Photolysis rate coefficient
 9 estimated for E-4-methylcyclohexanone under our experimental conditions, k_p = 5 × 10⁻⁵ s⁻¹

10 ^c Rate coefficient k (in cm³ molecule⁻¹ s⁻¹ unit) used to correct the concentration of E-4-methylcyclohexanone by
 11 loss with the reaction of NO₃ radical = 2.28 × 10⁻¹⁶ (estimated using SAR method, Kerdouci et al., 2014)

12 ^d Indicated errors are the associated error to the slope of plots obtained in the least squares analysis

13 ^e Experiment using a Teflon gas bag of 150L

14 ^f Standard deviations 1σ

15

1 **Table 4.** Experimental conditions and molecular yields of 3,3-dimethylbutanal for the reaction of
 2 3,3DM1ButOH with atmospheric oxidants.

MSA	Oxidant	Exp	[MSA] (ppm)	[Precursor] (ppm)	[NO] (ppm)	Carbonyl compound Yield (%)	Technique	Average ^g (%)	SAR Yield (%)			
3,3DM1ButOH	Cl ^a	1	11	24	-	40.3±0.2 41.8±4.6	FTIR SPME/GC- TOFMS ^d	39.4±15.0				
		2	2.6	8	-	19.6±0.5	SPME/GC- TOFMS ^e					
		3	6	25	-	55.9±1.7	FTIR					
	Cl ^a + NO	1	10	21	21	61.6±3.4 34.7±4.4	FTIR SPME/GC- TOFMS ^d	43.3±17.7	40			
		2	4	9	8	23.0±4.2	SPME/GC- TOFMS ^e					
	OH ^b	3	10	25	25	48.8±0.6	FTIR	62.2±15.0	66			
		1	10	60	36	82.1±4.2 40.8±2.7	FTIR SPME/GC- TOFMS ^d					
		2	7	35	57	67.4±1.4	FTIR					
		3	11	28	55	61.9±0.9	FTIR					
		4	11	29	30	59.1±3.8	FTIR					
		NO ₃ ^c	1	11	36	-	29.2±0.5 53.9 ^h			FTIR SPME/GC- TOFMS ^d	36.2±14.6	86
			2	11	32	-	26.5±1.6			FTIR		

3 ^a Rate coefficient k (in cm³ molecule⁻¹ s⁻¹ unit) used to correct the concentration of 3,3-dimethylbutanal by loss
 4 with the reaction of Cl atoms = 1.7 × 10⁻¹⁰ (data of iso-Butyraldehyde and Cl atoms (Thevenet et al., 2000)).
 5 Photolysis rate coefficient estimated for 3,3-dimethylbutanal under our experimental conditions, k_p = 1 × 10⁻⁴ s⁻¹
 6

7 ^b Rate coefficient k (in cm³ molecule⁻¹ s⁻¹ unit) used to correct the concentration of 3,3-dimethylbutanal by loss
 8 with the reaction of OH radical = 2.73 × 10⁻¹¹ (Aschmann et al., 2010). Photolysis rate coefficient estimated for
 9 3,3-dimethylbutanal under our experimental conditions, k_p = 1 × 10⁻⁴ s⁻¹

10 ^c Rate coefficient k (in cm³ molecule⁻¹ s⁻¹ unit) used to correct the concentration of 3,3-dimethylbutanal by loss
 11 with the reaction of NO₃ radical = 1.27 × 10⁻¹⁴ (D'Anna, 2001).

12 ^d Experiment using a FTIR gGas cCell of 50 L

13 ^e Experiment using a Teflon gas bBag of 150 L

14 ^f Indicated errors are the associated error to the slope of plots obtained in the least square analysis

15 ^g Standard deviations 1σ

16 ^h Yield estimated using only one data

17

1 **Table 5.** Experimental conditions and molecular yields of 3,3-dimethyl-2-butanone for the reaction
 2 of 3,3DM2ButOH with atmospheric oxidants

MSA	Oxidant	Exp	[MSA] (ppm)	[Precursor] (ppm)	[NO] (ppm)	Carbonyl compound ^fYield (%)	Technique	Average^g (%)	SAR Yield (%)		
3,3DM2ButOH	Cl ^a	1	14	31	-	42.8±0.7 45.2±1.1	FTIR SPME/GC- TOFMS ^d	43.2±1.8	40		
		2	2.3	8	-	41.7±3.2	SPME/GC- TOFMS ^e				
	Cl ^a + NO	1	14	28	20	36.7±5.0 49.6±4.5	FTIR SPME/GC- TOFMS ^d	44.2±7.4			
		2	3	7	6	39.0±6.1	SPME/GC- TOFMS ^e				
	OH ^b	3	8	28	27	51.5±3.9	FTIR	80.7±6.5		91	
		1	8	55	42	82.8±3.1 71.2±2.6	FTIR SPME/GC- TOFMS ^d				
	NO ₃ ^c		2	5	66	36	85.4±5.8	FTIR		58.0±10.9	99
			3	11	28	29	83.6±3.0	FTIR			
		1	12	30	-	66.7±2.05 45.9±1.6	FTIR SPME/GC- TOFMS ^d				
		2	9	30	-	61.5±1.4	FTIR				

3 ^a Rate coefficient k (in cm³ molecule⁻¹ s⁻¹ unit) used to correct the concentration of 3,3-dimethyl-2-butanone by
 4 loss with the reaction of Cl atoms = 4.8 × 10⁻¹¹ (Farrugia et al., 2015). Photolysis rate coefficient estimated for
 5 3,3-dimethyl-2-butanone under our experimental conditions, k_p = 7 × 10⁻⁵ s⁻¹

6 ^b Rate coefficient k (in cm³ molecule⁻¹ s⁻¹ unit) used to correct the concentration of 3,3-dimethyl-2-butanone by
 7 loss with the reaction of OH radical = 1.21 × 10⁻¹² (Wallington and Kurylo., 1987). Photolysis rate coefficient
 8 estimated for 3,3-dimethyl-2-butanone under our experimental conditions, k_p = 7 × 10⁻⁵ s⁻¹

9 ^c No corrected

10 ^d Experiment using a FTIR **g**Gas **c**Cell of 50 L

11 ^e Experiment using a Teflon gas **b**Bag of 150 L

12 ^f Indicated errors are the associated error to the slope of plots obtained in the least squares analysis

13 ^g Standard deviations 1σ

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1 **Table 6:** Summary of molecular yields (%) of reaction products identified in the reaction of **MSAs**
 2 with atmospheric oxidants and the total carbon balance (%).

Product	MSA			
	4MCHexOH			
	Cl	Cl + NO	OH	NO₃
E-4-methylcyclohexanone¹	25.2 ± 1.9	29.5 ± 0.7	40.2 ± 5.4	58.0 ± 23.5
HCOH²	9	-	-	-
Nitrated compounds	-	20	10	60
Total Carbon³	26	50	50	~100
	3,3DM1ButOH			
	Cl	Cl + NO	OH	NO₃
3,3-dimethylbutanal¹	39.4 ± 15.0	43.3 ± 17.7	62.2 ± 15.0	36.2 ± 14.6
HCOH²	10	22	-	-
2,2-dimethylpropanal²	22	8	23	-
Acetone²	5	17	-	-
Nitrated compounds	-	40 ⁵	35 ⁶	200 ⁷
Total Carbon^{3,4}	61	62	81	36
	3,3DM2ButOH			
	Cl	Cl + NO	OH	NO₃
3,3-dimethyl-2-butanone¹	43.2 ± 1.8	44.2 ± 7.4	80.7 ± 6.5	58.0 ± 10.9
HCOH²	10	64	-	-
2,2-dimethylpropanal²	14	10	14	-
Acetone²	3	58	-	-
Acetaldehyde²	-	17	-	-
Nitrated compounds	-	30	20	120
Total Carbon^{3,4}	~60	98	93	58

3 ¹Average Tables 3-5; ²Molecular yield obtained in earlier step of the reaction-;

4 ³ $Total\ Carbon\ (\%) = \sum_1^i \left(\frac{n^o\ of\ carbon\ of\ product_i}{n^o\ of\ carbon\ of\ MSA} \times molar\ yield_i(\%) \right)$

5 ⁴~~Nitrate compounds were not accounted for~~Without accounting nitrated compounds; ⁵From analysis of the
 6 experiment number 3 for the reaction of Cl+ in the presence of NO_x; ⁶From average of experiments number 2, 3
 7 and 4 for the reaction with OH; ⁷From analysis of the experiments number 1 and 2 for reaction with NO₃.

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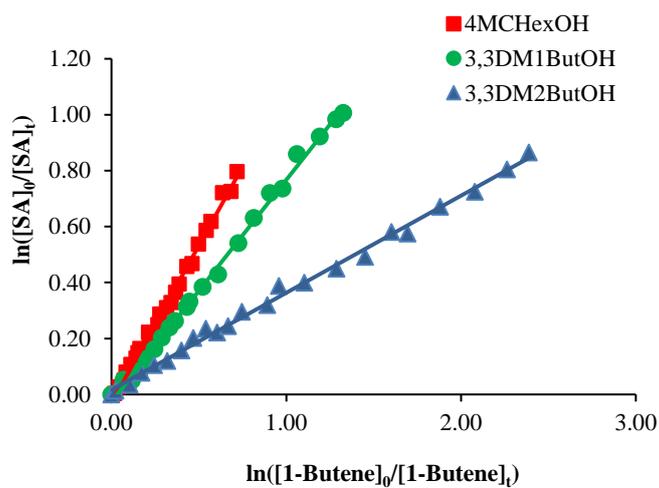
Table 7. Lifetimes of 4MCHexOH, 3,3DM1ButOH and 3,3DM2ButOH.

	τ_{OH} (days)	τ_{Cl}^a (days)	τ_{Cl}^b (days)	τ_{NO_3} (days)	τ_{wet} (years)	τ_{global}^a (days)
4MCHexOH	0.62	31.28	0.24	8.61	~2.1	0.58
3,3DM1ButOH	2.17	43.03	0.33	13 ^c	~15.8	1.78
3,3DM2ButOH	1.10	95.65	0.74	6.73 ^c	11.3	0.94

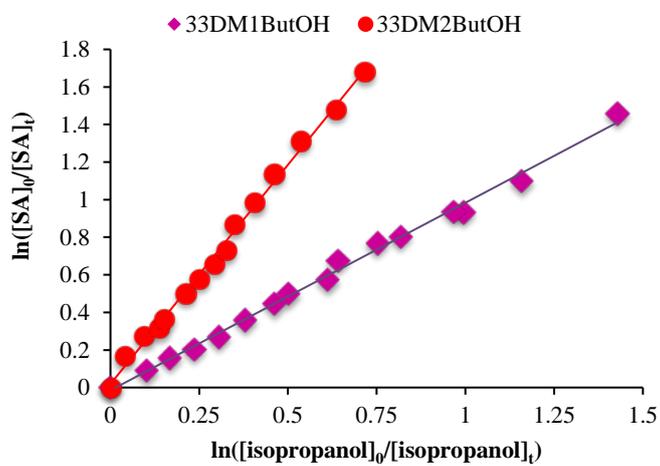
6 ^aDetermined with the 24 hours average of chlorine atoms.
7 ^bDetermined with the peak concentration of chlorine atoms.
8 ^cDetermined using the rate coefficient obtained by Moreno-A. et al., 2014.

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1 A)



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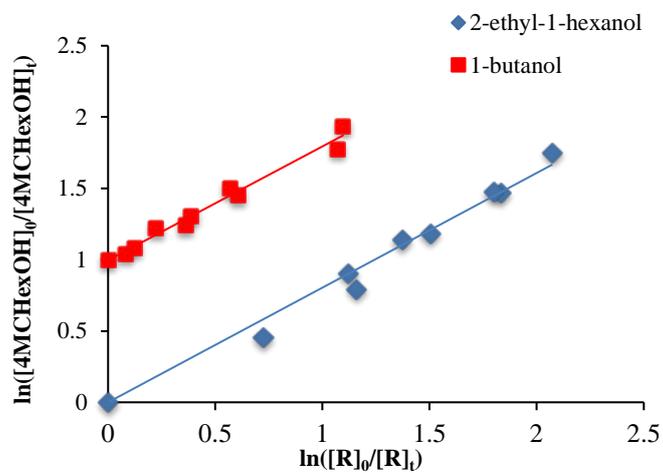


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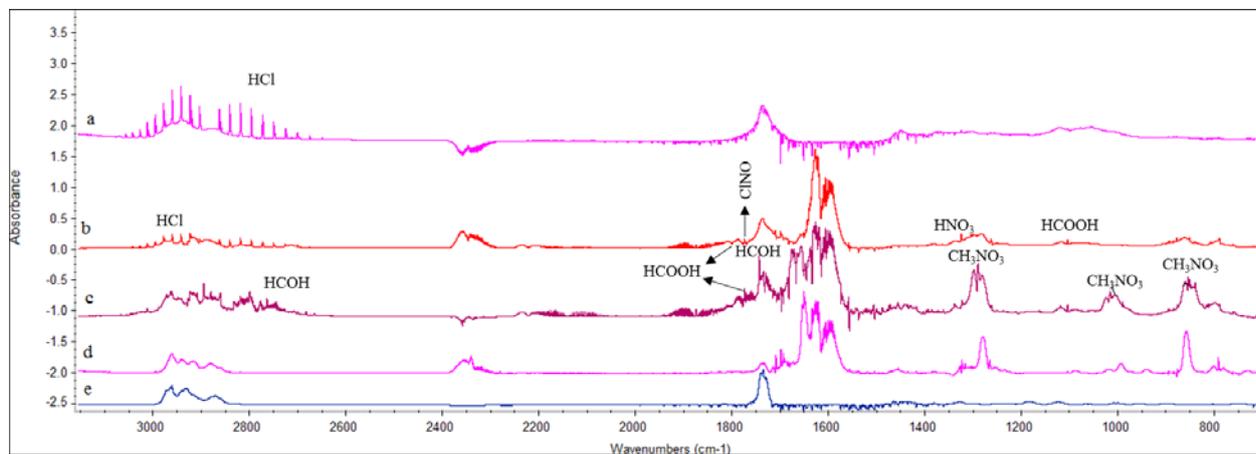
5 C)



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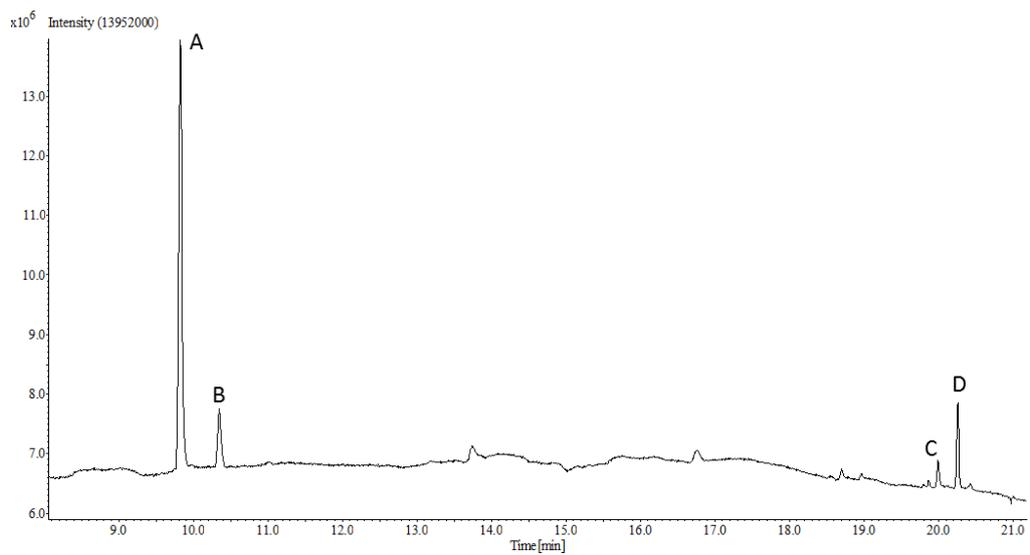
1 Fig.1: Relative rate plots for the reaction of (A) ~~MSAs~~ with chlorine atoms employing 1-butene as a reference
2 compound (B) 3,3-dimethylbutanols and OH radical with isopropanol as a reference compound and (C)
3 4MCHexOH and NO₃ with two reference compounds. Data for 1-butanol have been vertically displaced for clarity.

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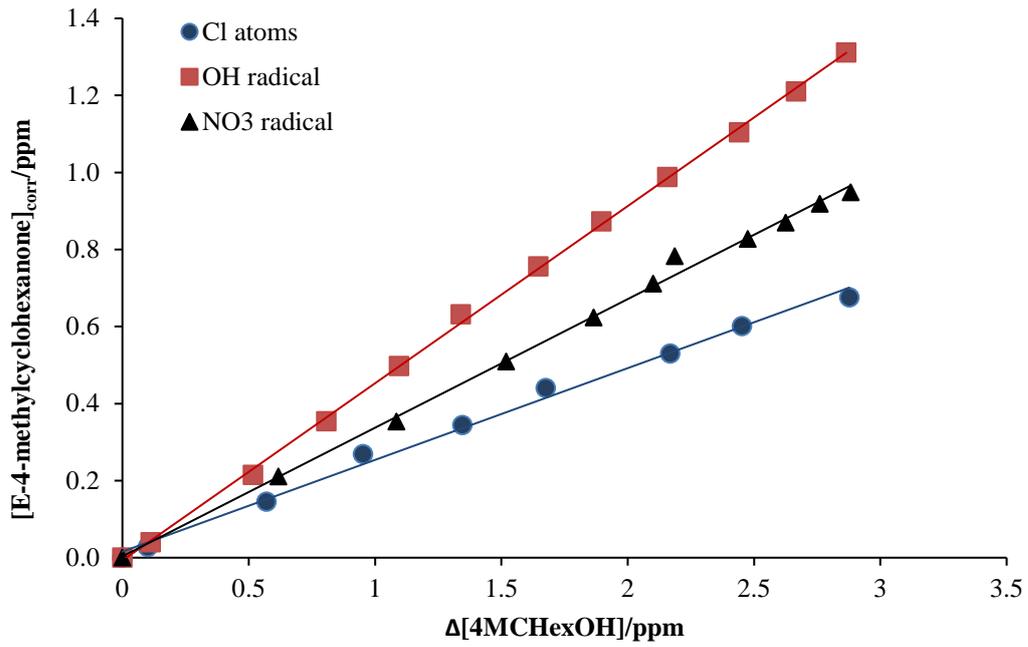
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Fig. 2: Product spectra for reaction of 4MCHexOH with: (a) chlorine atoms at 10 min (x 2 to clarify), (b) Cl atoms in presence of ~~and~~ NO_x at 7 min. (c) OH radical at 40 min and (d) NO₃ radical at 32 min. (e) Spectrum of 4-methylcyclohexanone commercial sample.



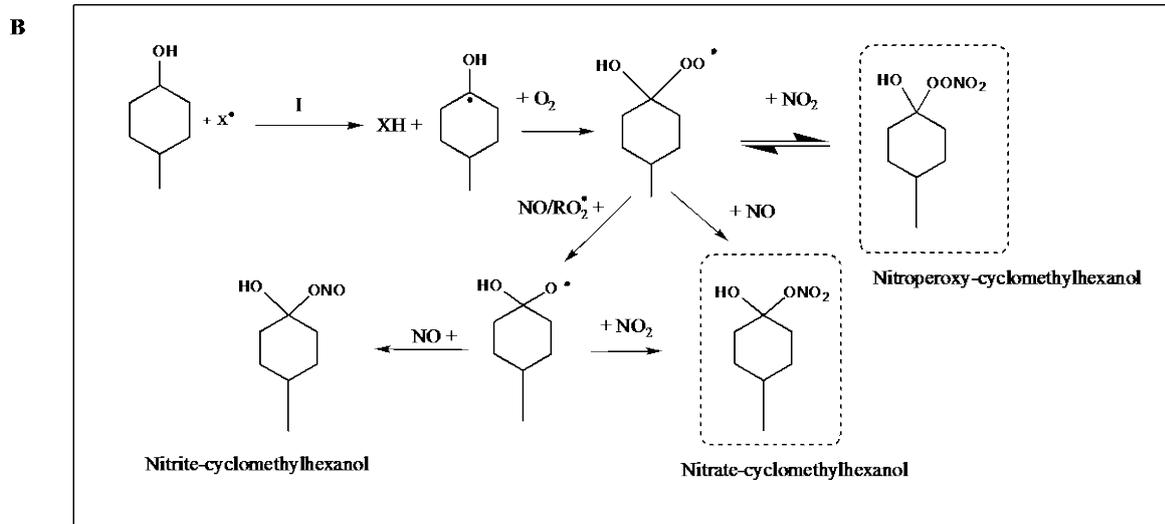
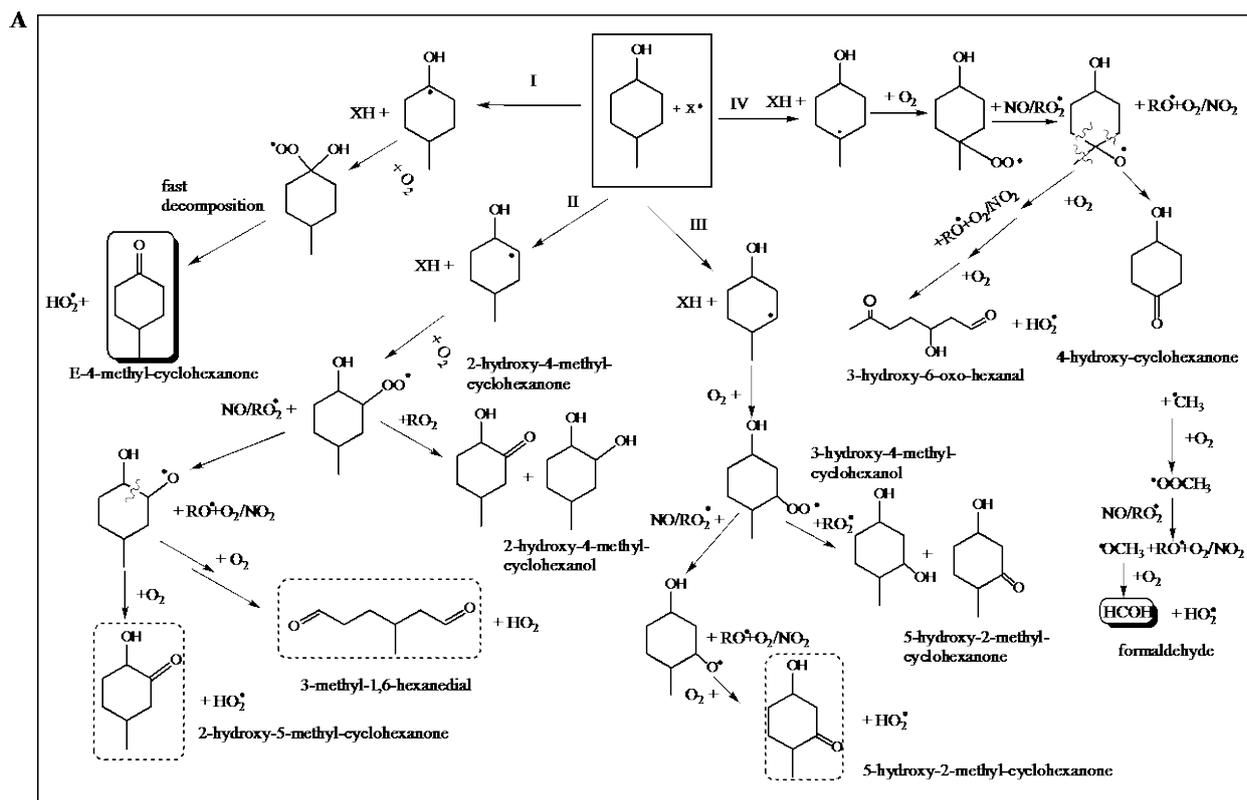
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Fig. 3: SPME/GC-TOFMS chromatogram for the reaction of 4MCHexOH with Cl atoms after 15 minutes of reaction. Peak (A) 4MCHexOH. Peak (B) E-4-methylcyclohexanone. Peaks (C) and (D) reaction products.



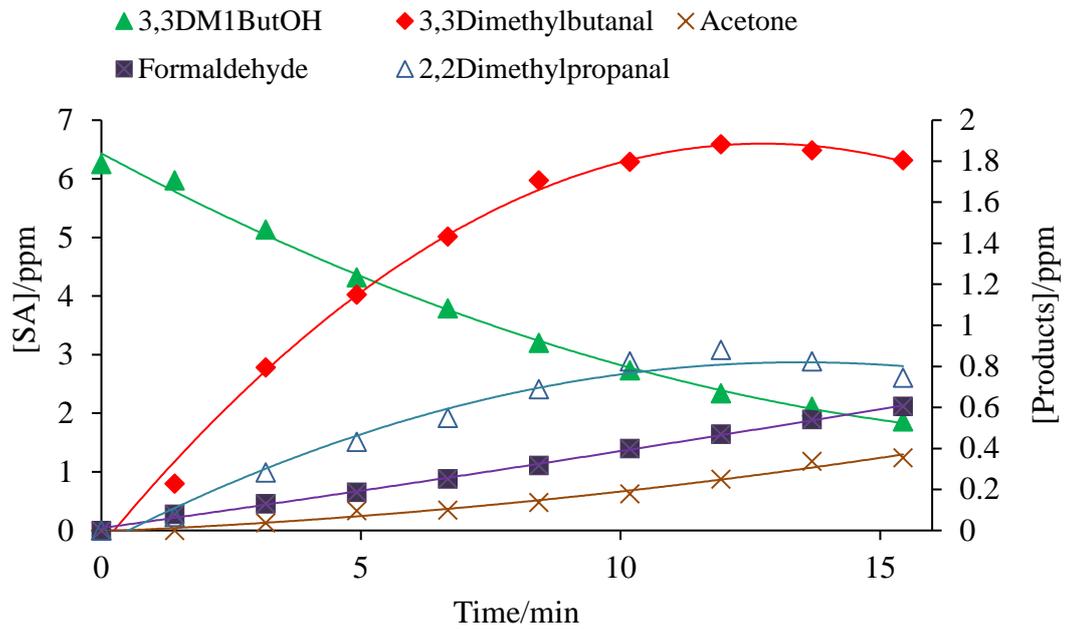
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Fig. 4: Plots of corrected concentration of E-4-methylcyclohexanone ($[E-4-methylcyclohexanone]_{corr}$) against 4MCHexOH consumed ($\Delta[4MCHexOH]$) for Cl atoms (in the absence of NO_x) and OH and NO₃ radical reactions.



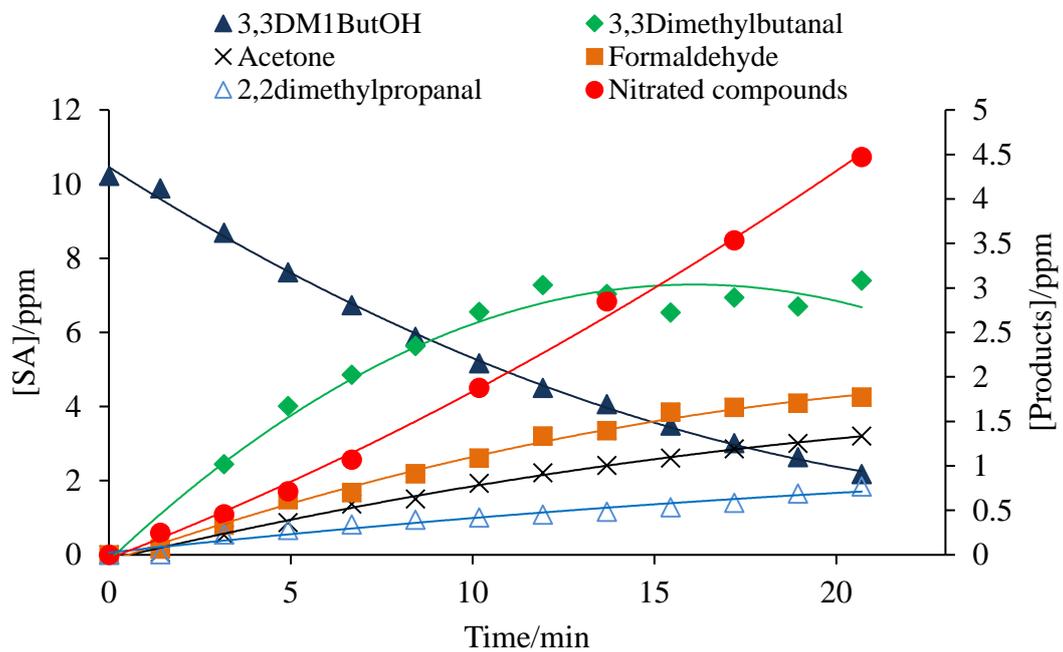
5 Fig. 5: Reaction mechanism for the degradation of 4MCHexOH with X (Cl atom, OH and NO₃ radicals). (A)
 6 Mechanism for the formation of carbonyl compounds, (B) Mechanism for the formation of nitrated compounds.
 7 Compounds marked with solid line are positively identified. Compounds marked with ~~dotted shaded~~ lines are not
 8 positively identified.

1 A)



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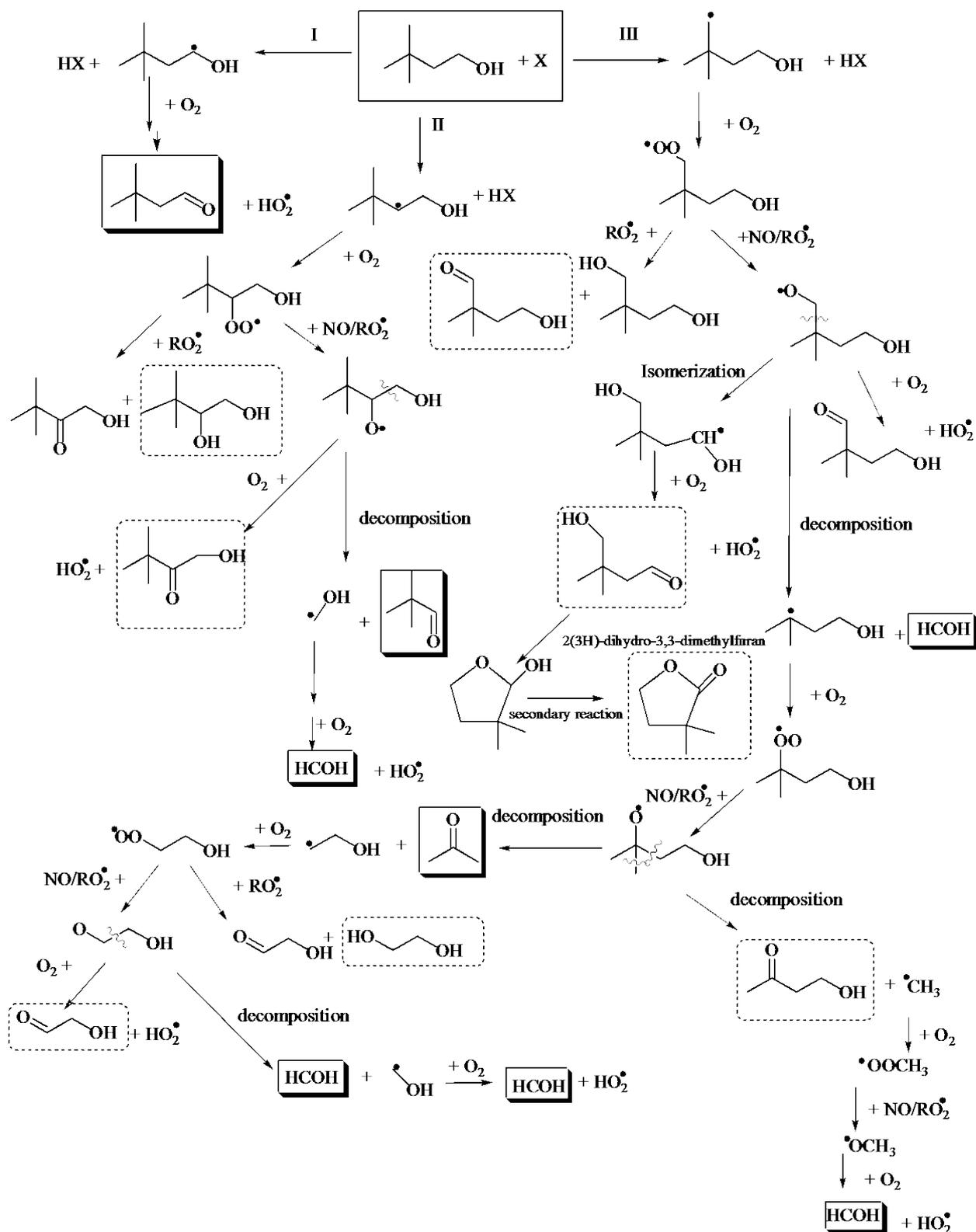
3 B)



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5 Fig. 6: Concentration-time profiles of MSA (3,3DM1ButOH) and reaction products formed in the reaction of
6 3,3DM1ButOH with Cl atoms in the absence (A) and in the presence of NO_x (B).

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3 Fig. 7: Reaction mechanism for the degradation of 3,3DM1ButOH with X (Cl atom, OH and NO₃ radical).
 4 Mechanism for the formation of carbonyl compounds. Compounds marked with solid line are positively identified.
 5 Compounds marked with shaded-dotted lines are not positively identified.

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Supplementary material

S1. Tuazon formalism

Formulism of Tuazon et al. (1986). This procedure is described in A1 supplementary information.



Where Y is the yield of the primary product (product1) from the oxidation of the methyl-saturated alcohols.

If we assume that the concentration of oxidant was constant during the irradiation period, then:

$$[\text{MSA}]_{t_2} = [\text{MSA}]_{t_1} \cdot e^{-(k_s \cdot [\text{Ox}])(t_2 - t_1)} \quad (\text{S1})$$

From Eq (S1) it is possible to calculate the concentration of oxidant in the system.

Using the relationship obtained by Tuazon et al. (1986) the corrected concentration of the reaction products can be calculated as follows:

$$[\text{product1}]_{t_2} = [\text{product1}]_{t_1} \cdot e^{-(k_p \cdot [\text{Ox}] + k_1)(t_2 - t_1)} + \frac{Y_{t_1 - t_2} \cdot [\text{MSA}]_{t_1} \cdot k_s [\text{Ox}]}{\{(k_p - k_s) \cdot [\text{Ox}] + k_1\}} \cdot [e^{-(k_s \cdot [\text{Ox}])(t_2 - t_1)} - e^{-(k_p \cdot [\text{Ox}] + k_1)(t_2 - t_1)}] \quad (\text{S2})$$

Where $[\text{MSA}]_{t_1}$, $[\text{product1}]_{t_1}$ and $[\text{MSA}]_{t_2}$, $[\text{product1}]_{t_2}$ are the observed concentrations of the methyl saturated alcohol and the products at times t_1 and t_2 respectively, and $Y_{t_1 - t_2}$ is the yield of formation of the individual products on the period of time $(t_1 - t_2)$.

By means of equations (S1) and (S2), $Y_{t_1 - t_2}$ can be calculated. The concentration of the reaction products, corrected for the reaction with radical and other processes of loss is given by:

$$[\text{product1}]_{t_2}^{\text{corr}} = [\text{product1}]_{t_1}^{\text{corr}} + Y_{t_1 - t_2} \cdot ([\text{MSA}]_{t_1} - [\text{MSA}]_{t_2}) \quad (\text{S3})$$

Where $[\text{product1}]_{t_1}^{\text{corr}}$ and $[\text{product1}]_{t_2}^{\text{corr}}$ are the concentrations of the product1 corrected at times t_1 and t_2 respectively:

The value of k_s has been obtained in this work. The value of k_p is taken from the bibliography or estimates with SAR method and k_1 is obtained from experiments carried out in the laboratory, taken from bibliography or estimated with the best fit of experimental data.

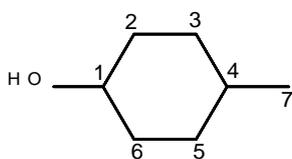
S2. Structure-Activity Relationship (SAR) method

SAR method allows to estimate a rate coefficient of an organic compound from its structure. The only possibility of the reaction of the studied compounds in this work with the atmospheric oxidants is the abstraction of an hydrogen atom. Consequently, the estimated rate coefficients of **MSAs** are obtained from the sum of the rate coefficients for the H-atom abstraction from the primary (k_{prim} (CH_3)), secondary (k_{sec} ($-\text{CH}_2-$)) and tertiary (k_{tert} ($-\text{CH}<$)) groups and from the alcohol (k_{OH} ($-\text{OH}$)) group, taking into account the influence of the substituents attached to these groups, through substituent factors $F(X)$, $F(Y)$ and $F(Z)$ (Equation S4).

$$k_{\text{abs}} = \sum k_{\text{prim}}F(X) + \sum k_{\text{sec}}F(X)F(Y) + \sum k_{\text{tert}}F(X)F(Y)F(Z) + \sum k_{\text{OH}} \quad (\text{S4})$$

At 298K rate coefficients for H-atom abstraction (in units of $\text{cm}^3\text{molecule}^{-1}\text{s}^{-1}$) and the reactivity factor for the reaction with OH are $k_{\text{prim}} = 1.36 \times 10^{-13}$; $k_{\text{sec}} = 9.34 \times 10^{-13}$; $k_{\text{tert}} = 1.94 \times 10^{-12}$ and $k_{\text{OH}} = 1.4 \times 10^{-13}$; $F(\text{CH}_3) = 1$; $F(-\text{CH}_2-) = F(-\text{CH}<) = F(>\text{C}<) = 1.23$ and $F(-\text{OH}) = 3.5$ from AOPWIN. The parameters for the reaction with Cl atoms are $k_{\text{prim}} = 2.84 \times 10^{-11}$; $k_{\text{sec}} = 8.95 \times 10^{-11}$; $k_{\text{tert}} = 6.48 \times 10^{-11}$ (in units of $\text{cm}^3\text{molecule}^{-1}\text{s}^{-1}$); $F(\text{CH}_3) = 1$; $F(-\text{CH}_2-) = F(-\text{CH}<) = 0.8$ and $F(-\text{OH}) = 1.18$ from Calvert et al. 2011. By last, the parameters used for the reaction with NO_3 radicals are $k_{\text{prim}} = 1 \times 10^{-18}$; $k_{\text{sec}} = 2.56 \times 10^{-17}$; $k_{\text{tert}} = 1.05 \times 10^{-16}$ and $k_{\text{OH}} = 2 \times 10^{-17}$ (in units of $\text{cm}^3\text{molecule}^{-1}\text{s}^{-1}$); $F(\text{CH}_3) = 1$; $F(-\text{CH}_2-) = 1.02$; $F(-\text{CH}<) = 1.61$; $F(>\text{C}<) = 2.03$ and $F(-\text{OH})=18$ from Kerdouci et al. 2010, 2014.

The calculations for 4-methyl-cyclohexanol are the following:



$$k_1 = k_{\text{tert}} \times F(-\text{OH}) \times F(-\text{CH}_2-) \times F(-\text{CH}_2-)$$

$$k_2 = k_3 = k_5 = k_6 = k_{\text{sec}} \times F(-\text{CH}<) \times F(-\text{CH}_2-)$$

$$k_4 = k_{\text{tert}} \times F(-\text{CH}_3) \times F(-\text{CH}_2-) \times F(-\text{CH}_2-)$$

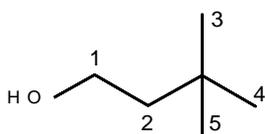
$$k_7 = k_{\text{prim}} \times F(-\text{CH}<)$$

$$k_{4\text{MCHexOH}+\text{Cl}} = 3.42 \times 10^{-10} \text{ cm}^3\text{molecule}^{-1}\text{s}^{-1}$$

$$k_{4\text{MCHexOH}+\text{OH}} = 1.92 \times 10^{-11} \text{ cm}^3\text{molecule}^{-1}\text{s}^{-1}$$

$$k_{4\text{MCHexOH}+\text{NO}_3} = 2.27 \times 10^{-15} \text{ cm}^3\text{molecule}^{-1}\text{s}^{-1}$$

The calculations for 3,3-dimethyl-1-butanol are the following:



$$k_1 = k_{\text{sec}} \times F(-\text{OH}) \times F(-\text{CH}_2-)$$

$$k_2 = k_{\text{sec}} \times F(>\text{C}<) \times F(-\text{CH}_2-)$$

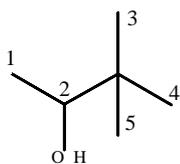
$$k_3 = k_4 = k_5 = k_{\text{prim}} \times F(>\text{C}<)$$

$$k_{3,3\text{DM1ButOH}+\text{Cl}} = 2.10 \times 10^{-10} \text{ cm}^3\text{molecule}^{-1}\text{s}^{-1}$$

$$k_{3,3\text{DM1ButOH}+\text{OH}} = 6.08 \times 10^{-12} \text{ cm}^3\text{molecule}^{-1}\text{s}^{-1}$$

$$k_{3,3\text{DM1ButOH}+\text{NO}_3} = 0.55 \times 10^{-15} \text{ cm}^3\text{molecule}^{-1}\text{s}^{-1}$$

The calculations for 3,3-dimethyl-2-butanol are the following:



$$k_1 = k_{\text{prim}} \times F(-\text{CH} <)$$

$$k_2 = k_{\text{tert}} \times F(>\text{C} <) \times F(-\text{CH}_3) \times F(-\text{OH})$$

$$k_3 = k_4 = k_5 = k_{\text{prim}} \times F(>\text{C} <)$$

$$k_{3,3\text{DM}2\text{ButOH}+\text{Cl}} = 1.52 \times 10^{-10} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$$

$$k_{3,3\text{DM}2\text{ButOH OH}} = 9.16 \times 10^{-12} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$$

$$k_{3,3\text{DM}2\text{ButOH}+\text{NO}_3} = 3.86 \times 10^{-15} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$$

Tables

Table S1: Rate coefficients at 298 K for Cl, OH and NO₃ reactions with alkanes, saturated alcohols and saturated compounds with a similar structure of [MSAs studied in- this work](#). k in cm³ molecule⁻¹ s⁻¹ unit.

Compound	k _{Cl} ×10 ¹¹	k _{OH} ×10 ¹²	k _{NO₃} ×10 ¹⁵
Alkanes			
Methane	0.01	0.0064	<0.001 ^f
Ethane	5.70 ^a	0.24 ^f	<0.01 ^f
Propane	12.7 ^a	1.09 ^f	<0.07 ^f
Butane	19.4 ^a	2.36 ^f	0.046 ^f
Pentane	25.0 ^a	3.80 ^f	0.087 ^f
Hexane	30.5 ^a	5.20 ^f	0.110 ^f
Heptane	36.5 ^a	6.76 ^f	0.150 ^f
Octane	40.9 ^a	8.11 ^{kf}	0.190 ^f
Nonane	46 ^b		0.220 ^b
Decane	52.7 ^b		0.260 ^b
2-Methylpropane	13.0 ^a	2.12 ^f	0.106 ^f
2-Methylbutane	19.6 ^a	3.60 ^f	0.162 ^f
2-Methylpentane	25.8 ^a	5.2 ^f	0.180 ^f
3-Methylpentane	26.9 ^b	5.2 ^f	0.220 ^f
2-Methylhexane	31.2 ^a	6.72 ^e	-
2,3-Dimethylbutane	20 ^a	5.78 ^f	0.44 ^f
2,2-Dimethylbutane	-	22.3 ^f	-
2,2,4-trimethylpentane	22.5 ^a	3.34 ^f	0.09 ^f
Cyclopentane	32.6 ^c	4.97 ^f	-
Cyclohexane	33 ^d	6.97 ^f	0.14 ^f
Methylcyclohexane	35.1 ^e	9.5 ^e	-
Trans-1,4-dimethylcyclohexane	36.3 ^e	12.1 ^e	-
Ethylcyclohexane	38.8 ^e	11.8 ^e	-
Saturated Alcohols			
Methanol	5.5	0.9	0.13
Ethanol	10	3.2	<2
1-propanol	16	5.8	<2.1
2-propanol	8.7	5.1	1.4
1-butanol	22	8.5	1.87 ^g
2-butanol	12 ^h	8.7	2.51 ^q
1-pentanol	24 ⁱ	11 ⁱ	-
2-pentanol	22 ⁱ	11.8 ⁱ	-
3-pentanol	20 ⁱ	13 ⁱ	-
1-hexanol	31 ⁱ	13 ⁱ	-
2-hexanol	-	12 ⁱ	-

1-heptanol	37 ⁱ	13 ⁱ	-
4-heptanol	-	-	<6.2
1-octanol	42 ⁱ	13 ⁱ	-
2-methyl-1-propanol	20.6 ^j	11.4 ^j	-
2-methyl-2-propanol	3.26 ^k	1.07 ^k	-
2-ethyl-1-hexanol	18.8 ^l	11.3 ^l	<u>2.93^l</u>
2-methyl-1-butanol	-	-	2.32 ^g
2-methyl-2-butanol	7 ^h	3.64 ⁿ	1.57 ^q
3-methyl-1-butanol	25 ⁱ	14 ⁱ	2.09 ^g
3-methyl-2-butanol	11.7 ^h	12.5 ^o	3.06 ^q
2-methyl-2-pentanol	-	7.1 ⁱ	-
4-methyl-2-pentanol	-	17 ⁱ	-
2,2-dimethyl-1-propanol	-	5.5 ⁱ	-
3,3-dimethyl-1-butanol	26.9	5.33	1.78 ^g
3,3-dimethyl-2-butanol	12.1	10.50	3.44 ^g
2,3-dimethyl-2-butanol	10.3 ^h	9.1 ⁿ	3.64 ^q
2,4-dimethyl-2-pentanol	-	11 ⁱ	-
3,5-dimethyl-3-hexanol	-	13 ⁱ	-
cyclopentanol	-	10.7 ^k	-
cyclohexanol	32.1 ^m	19.0 ^p	-
4MCHexOH	37.0	18.7	2.69

Data obtained from IUPAC Subcommittee on Gas Kinetics Data Evaluation and the following references:

^aHooshiyar and Niki, 1995; ^bCalvert et al., 2015; ^cWallington et al., 1989; ^dCalvert et al., 2008; ^eBegan et al., 2018; ^fAtkinson, 2003; ^gMoreno et al., 2014; ^hBallesteros et al., 2007; ⁱCalvert et al., 2011; ^jAndersen et al., 2010; ^kWallington et al., 1988; ^lGallego-Iniesta et al., 2010; ^mCeacero-Vega et al., 2012; ⁿJiménez et al., 2005; ^oMellouki et al., 2004; ^pBradley et al., 2001; ^qMoreno 2012.

Table S2. A summary of the reaction products observed or tentative identified in the reactions of 3,3DM1ButOH with the atmospheric oxidants.

Reaction μ Product of reaction mechanism	FTIR				GC-MS				t_R /min
	Cl	Cl + NO	OH	NO ₃	Cl	Cl + NO	OH	NO ₃	
Route I									
3,3-dimethylbutanal	X	X	X	X	X	X	X	X	6.00
Route II									
1-hidroxy-3,3-dimethyl-2- butanone	-	-	-	-	-	-	-	-	-
3,3-dimetil-1,2-butanodiol	-	-	-	-	-	-	-	-	-
2,2-dimethylpropanal	X	X	X	-	/	-	-	-	8.61?
Formaldehyde	X	X	X*	-	-	-	-	-	-
Route III									
4-hidroxy-2,2- dimethylbutanal	-	-	-	-	-	-	-	-	-
2,2-dimethyl-1,4- butanodiol	-	-	-	-	-	-	-	-	-
4-hidroxy-3,3- dimethylbutanal	-	-	-	-	-	-	-	-	-
Cyclic alcohols	-	-	-	-	-	-	-	-	-
Cyclic carbonyls (2(3H)dihydro,3,3- dimethylfuran)	-	-	-	-	/	/	/	/	13.17
4-hidroxy-2-butanone	-	-	-	-	-	-	-	-	-
Acetone	X	X	-	-	-	/	-	-	2.14
Glycolaldehyde	-	/	-	-	-	/	/	/	5.05
1,2-etanodiol	-	-	-	-	-	-	-	-	-
Formaldehyde	X	X	X*	-	-	-	-	-	-

The compounds with X have been positively identified with reference FTIR or MS spectrum. Those that have / have been probably identified through the analysis of the IR bands of the residual spectra and Mass Spectra using as a source of ionization Electron Ionization (EI) and Field Ionization (FI) source that allows us to know the molecular ion. X* Formaldehyde in the reaction of OH, is formed by decomposition of the precursor, so it is observed in the reaction, but it cannot know how much come from the reaction of 3,3DM1ButOH and OH

Table S3. A summary of the reaction products observed or tentative identified in the reactions of 3,3DM2ButOH with the atmospheric oxidants.

Product propose in the mechanism	FTIR				GC-MS				Retention time (min)
	Cl	Cl + NO	O H	NO ₃	Cl	Cl + NO	OH	NO ₃	
Route I									
3,3-dimethyl-1,2- butanediol	-	-	-	-	-	-	-	-	-
3,3-dimethyl-2- hydroxybutanal	-	-	-	-	-	-	-	-	-
Formaldehyde	X	X	X*	-	-	-	-	-	-
2,2- dimethylpropanal	X	X	X	-	-	-	/	/	5.39
Route II									
3,3-dimethyl-2- butanone	X	X	X	X	X	X	X	X	6.04
Route III									
2,2-dimethyl-1,3- butanediol	-	-	-	-	/	-	-	-	13.46
2,2-dimethyl-3- hydroxybutanal	-	-	-	-	-	-	-	-	-
Formaldehyde	X	X	X*	-	-	-	-	-	-
3-hydroxybutanone (Acetoin)	-	-	-	-	-	-	-	-	-
Acetone	X	X	-	-	/	/	-	-	2.16
Acetaldehyde	-	X	-	-	-	-	-	-	-

The compounds with X have been positively identified in denoted equipment. Those that have / have been probably identified through the analysis of the IR bands of the residual spectra and the analysis of the mass spectra using as a source of ionization by electron ionization and field ionization source that allows us to know the molecular ion. X* Formaldehyde is a product, but it is also formed by decomposition of the precursor in the reaction with OH radical, and it is not possible to differentiate which comes from the reaction or from the decomposition process.

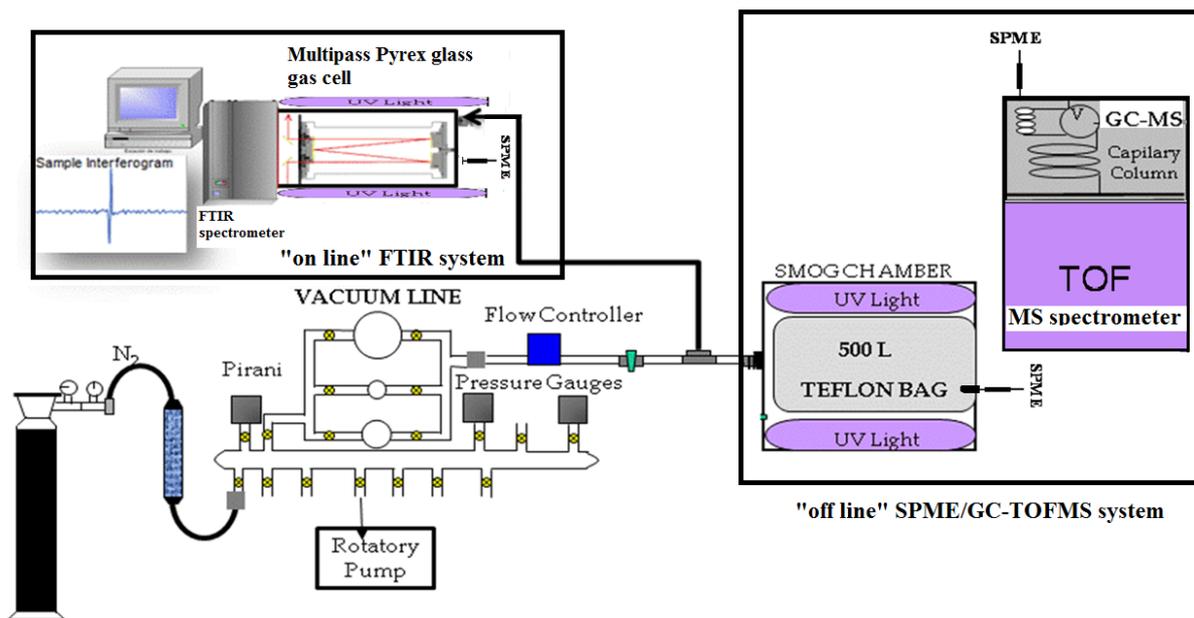


Fig.S1 Schematic diagram of experimental system: FTIR and SPME/GC-TOFMS.

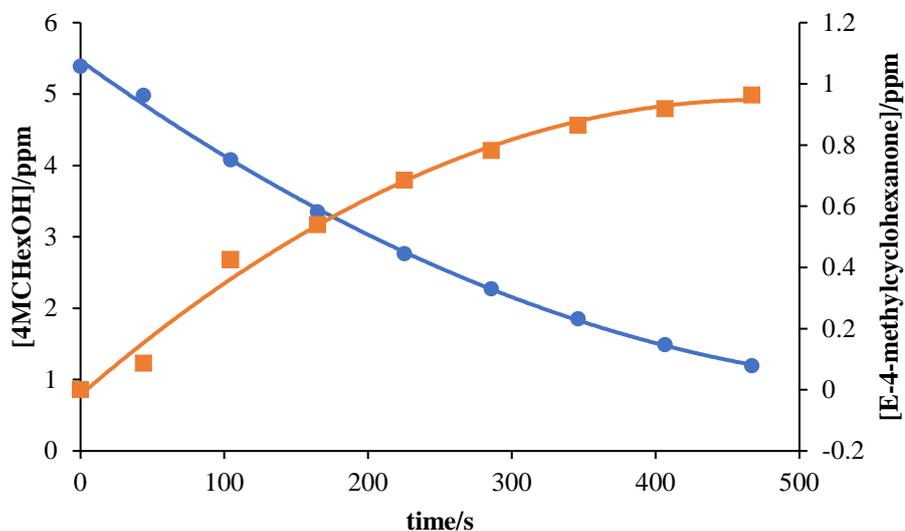


Fig. S2. Time-concentration profiles of 4MCHexOH ● and 4-methylcyclohexanone ■, for the reaction of 4MCHexOH with chlorine atoms in the presence of NO_x obtained by FTIR.

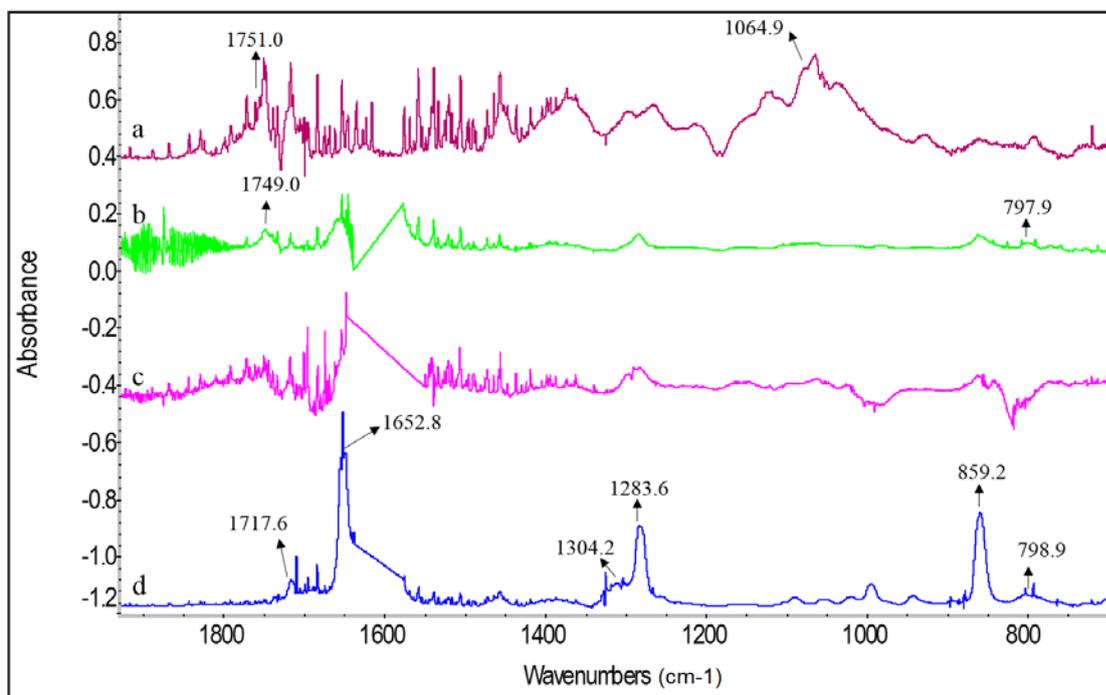


Fig. S3. Residual FTIR spectra obtained in the reaction of 4MCHexOH with (a) Cl atoms (a), (b) Cl+NO (e), HOH radicals, (c) Cl atoms in the presence of NO, and (d) NO₃ radicals (e). The gap in the range of 1650-1590 cm⁻¹ corresponds to NO₂ absorption.

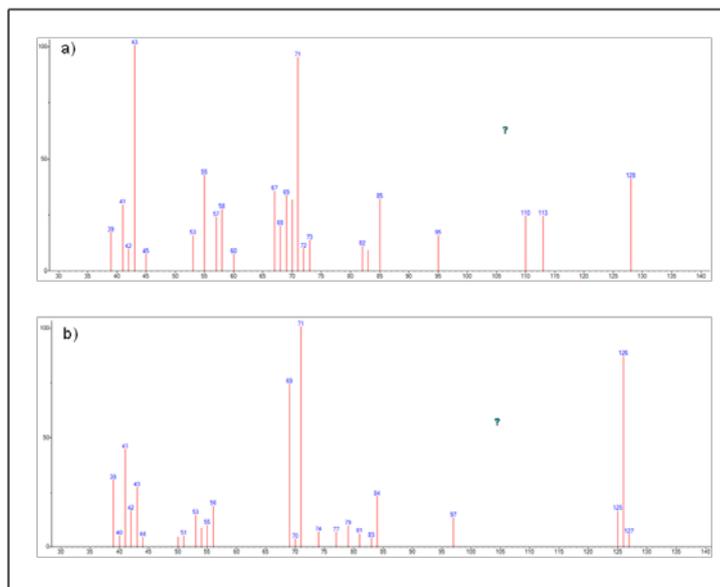


Fig. S4. EI MS spectra of peak C (a) and D (b) observed at 20 and 23 min of retention time in the reaction of 4MCHexOH with Cl atoms. Tentatively assigned to 2-hydroxy-5-methyl cyclohexanone, 5-hydroxy-2-methyl-cyclohexanone and or 3-methyl-1,6-hexanedial.

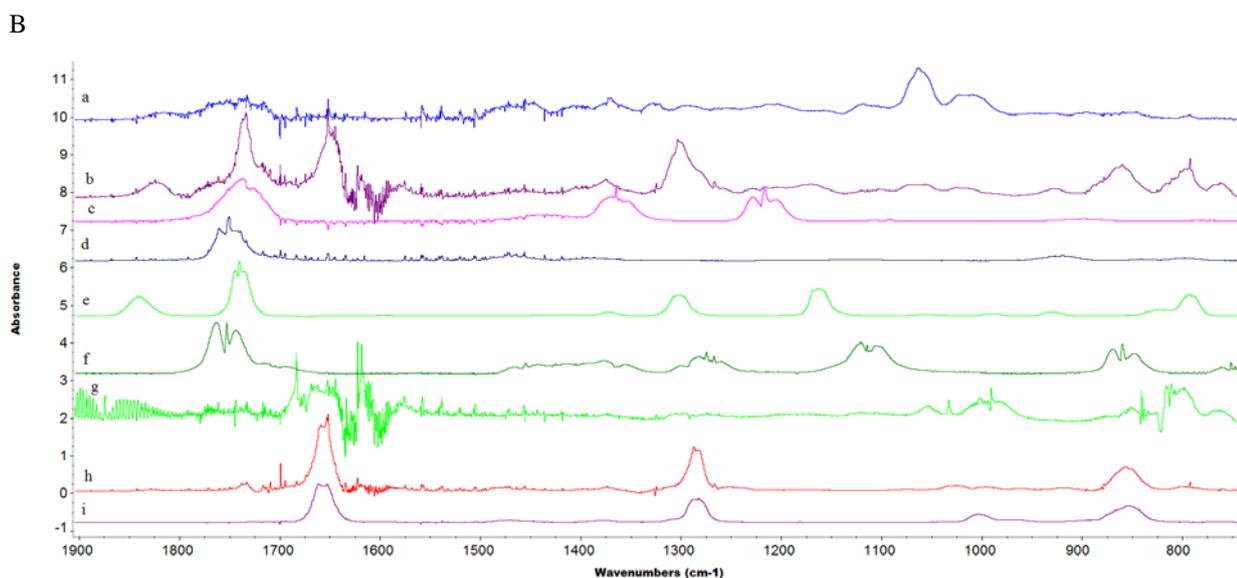
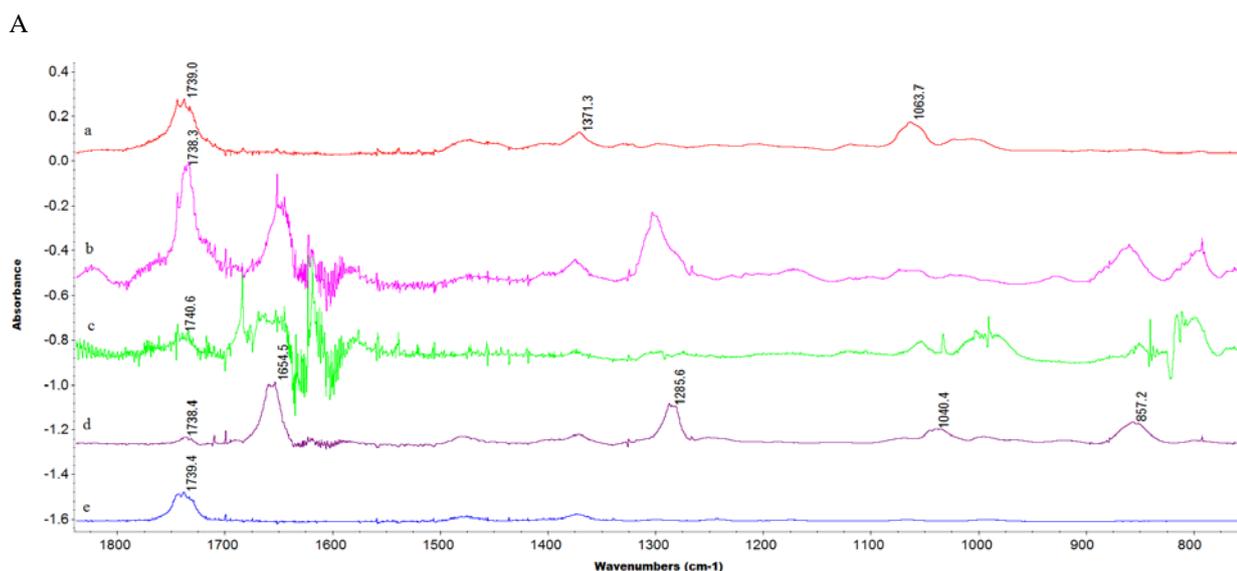


Fig. S5. **A)** Residual FTIR spectra obtained in the reaction of 3,3DM1ButOH with **(a) Cl atoms** ~~(a)~~, **(b) Cl atoms in the presence of +NO** ~~(b)~~, **(c) OH radicals** and **(d) -NO₃ radicals** ~~(d)~~. The IR absorption bands subtracted were: 3,3DM1ButOH, HCl, ClNO₂, ClNO, HCOH, HCOOH, HONO, NO₂, NO, N₂O and peroxy nitric acid (for Cl **in the absence and presence of NO** and **Cl + NO** reactions); N₂O₅, HNO₃, NO₂ (for NO₃ reactions) and HCOH, HCOOH, HNO, NO₂, CH₃ONO and CH₃ONO₂ (for OH reactions). **(e)** FTIR reference spectrum of 3,3-dimethylbutanal. **B)** Residual FTIR spectra: **(a) Cl atoms** ~~(a)~~, **(b) Cl atoms in the presence of +NO₂** ~~(b)~~, **(c) OH radicals** ~~(c)~~ and **(d) NO₃ radicals** ~~(d)~~ without 3,3-dimethylbutanal. Reference spectra **(e) of acetone** ~~(e)~~ from a commercial sample; **(f) 2-methylpropanal** ~~(f)~~; **(g) PAN** ~~(g)~~; **(h) Glycolaldehyde** ~~(h)~~; and **(i) isobutyl nitrate** ~~(i)~~ from Eurochamp 2020 database.

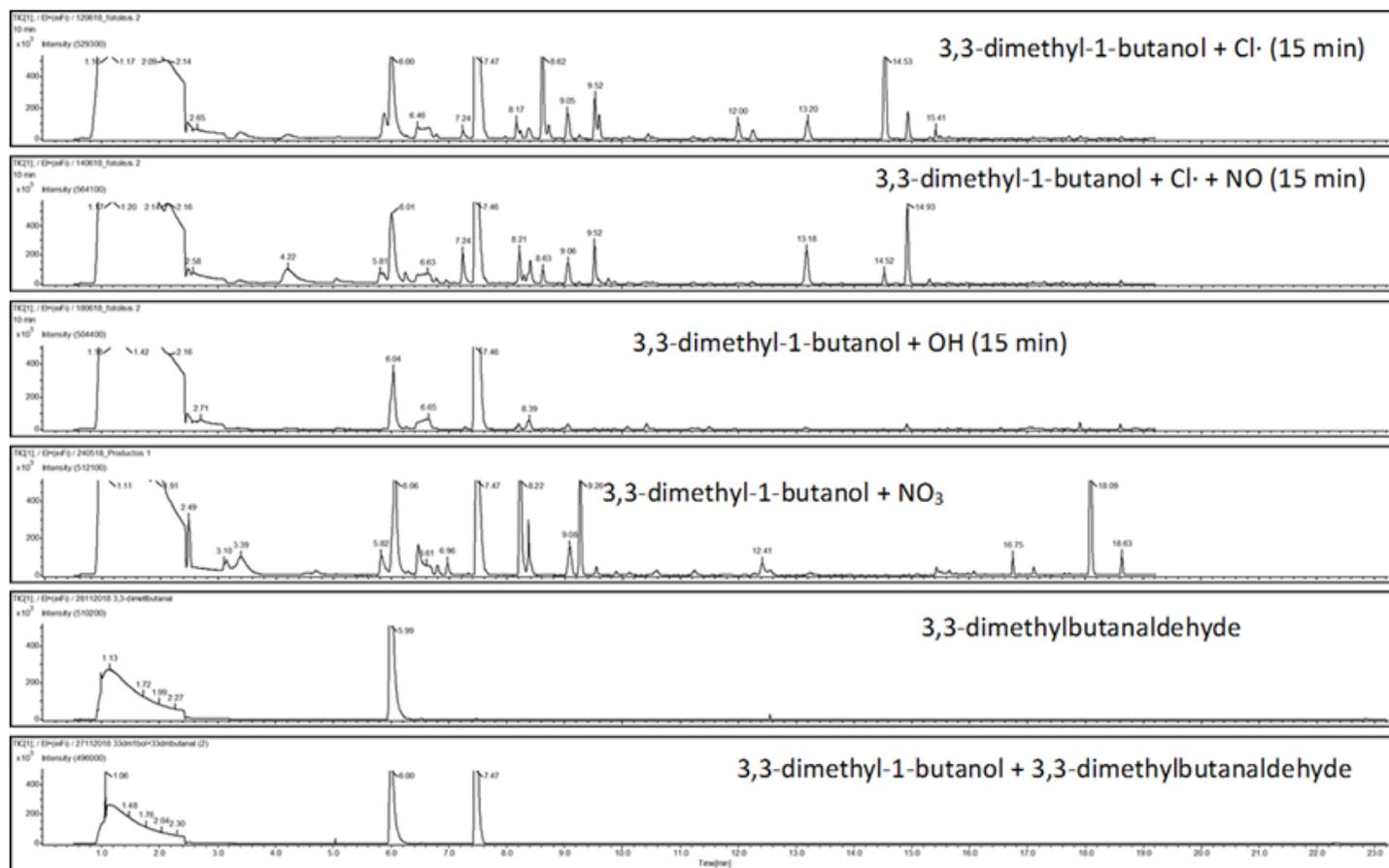


Fig. S6. SPME/GC-TOFMS chromatograms obtained for the reaction of 3,3DM1ButOH with Cl atoms, Cl atoms in the presence of NO, OH radicals and NO₃ radicals (30 min) and reference chromatograms of 3,3DM1ButOH and 3,3-dimethylbutanaldehyde.

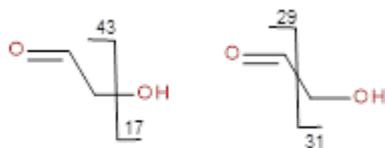
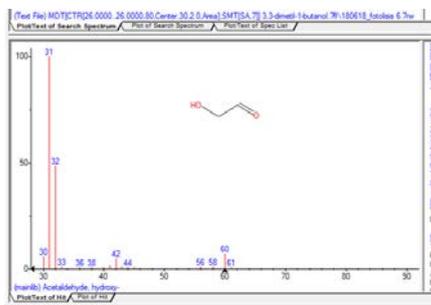
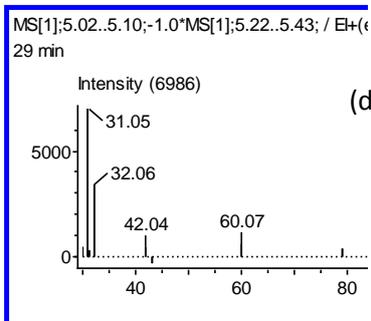
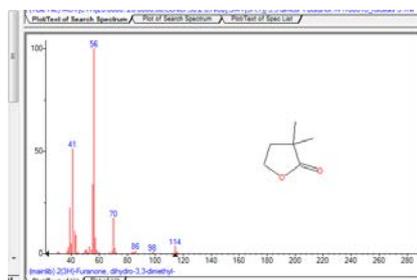
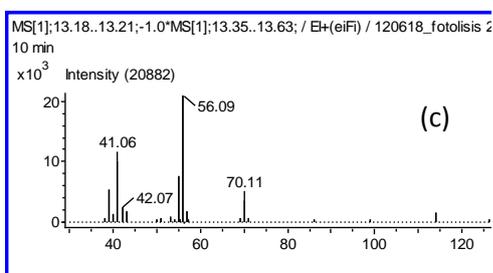
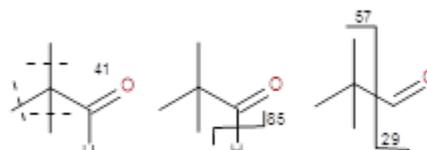
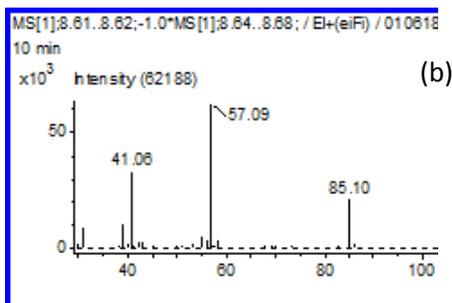
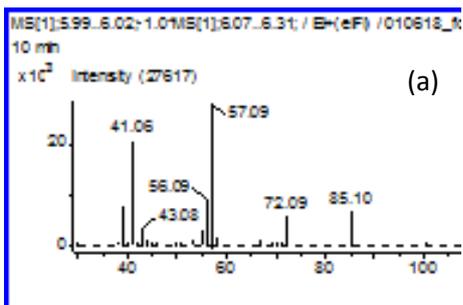


Fig. S7. EI-MS spectra of the peaks of chromatograms shown in Fig. S6 obtained for the reaction of 3,3DM1ButOH with Cl atoms, Cl atoms in the presence of NO+NO, HOH radicals and NO₃ radicals. (a) t_R = 6.00 min; (b) t_R = 8.61 min; (c) t_R = 13.17 min; (d) t_R = 5.05 min).

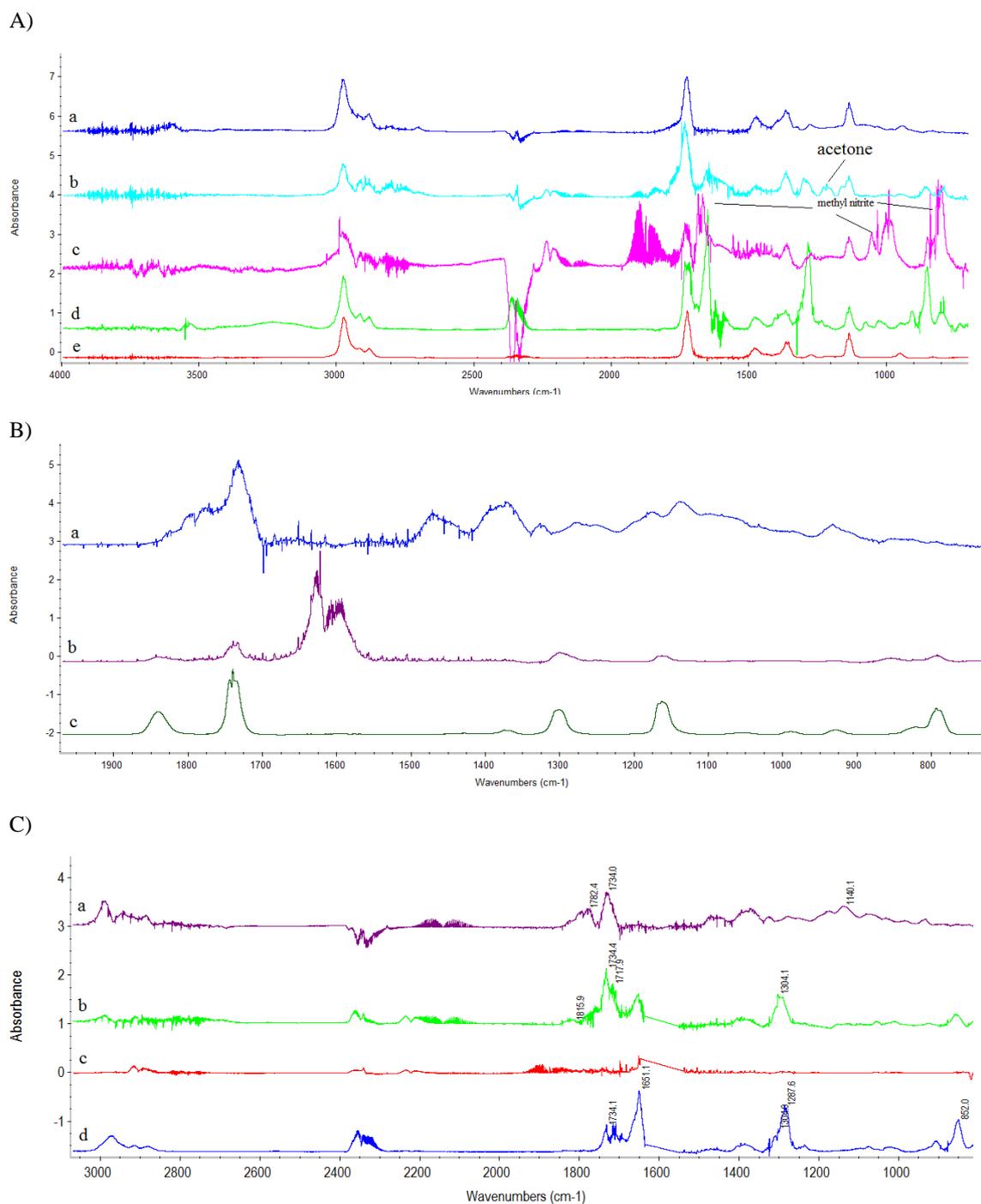


Fig. S9. A) FTIR spectra obtained in the reaction of 3,3DM2ButOH with (a) Cl atoms(a), (b) Cl atoms in the presence of +NO₂(b), (c) OH radicals and (d) e-NO₃ radicals(d) at 5 minutes of reaction. (e) FTIR reference spectrum of 3,3-dimethyl-2-butanone. B) FTIR spectra obtained in the reaction of 3,3-dimethyl-2-butanol with (a) Cl atoms(a), (b) Cl atoms in the presence of +NO(b), 25 minutes and 35 minutes of reactions respectively. (c) IR PAN spectrum. C) Residual FTIR spectra after subtraction of all known bands. (a) Cl atoms(a), (b) Cl atoms in the presence of +NO(b), (c) HOH radicals(e) and (d) NO₃ radicals(d).

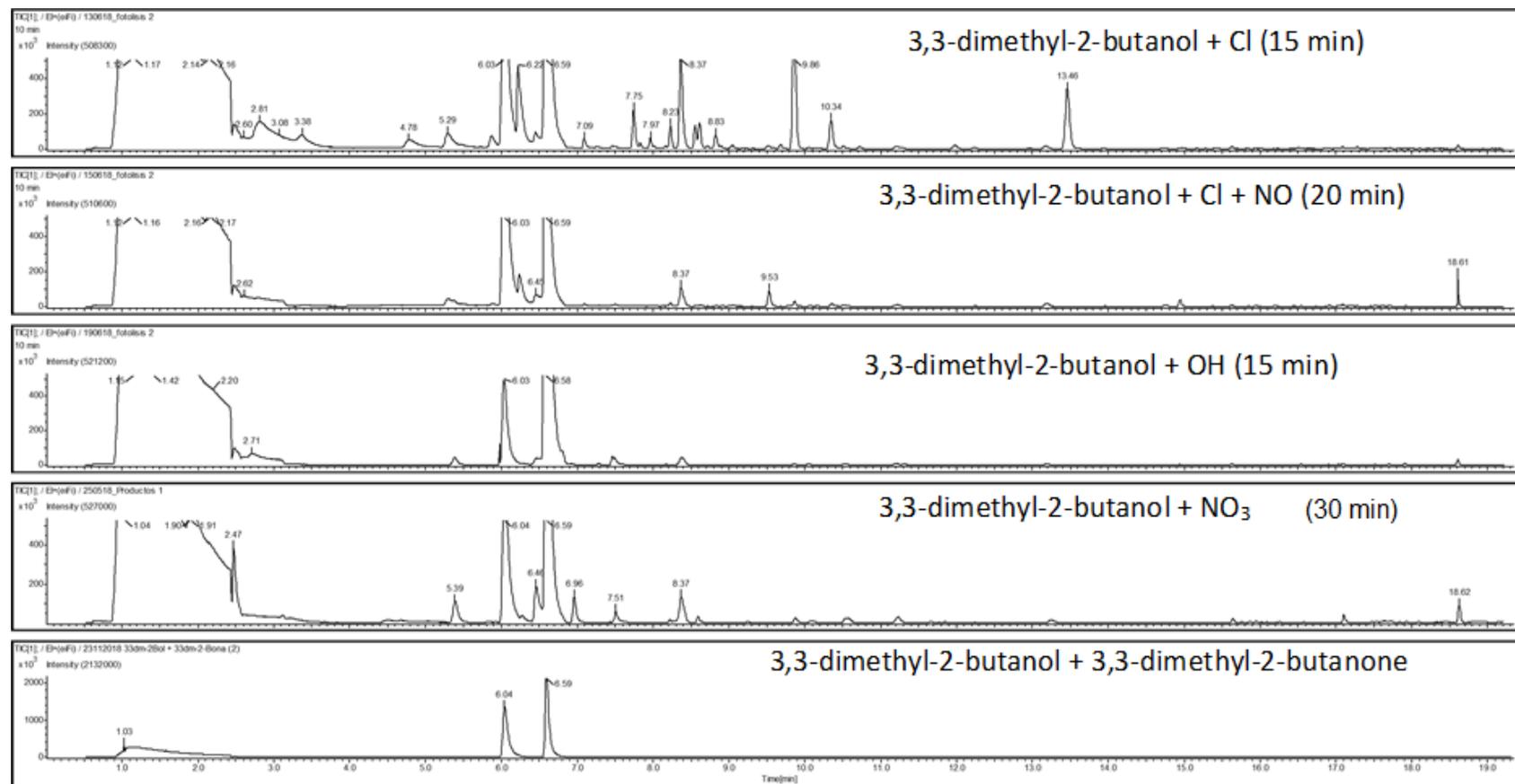
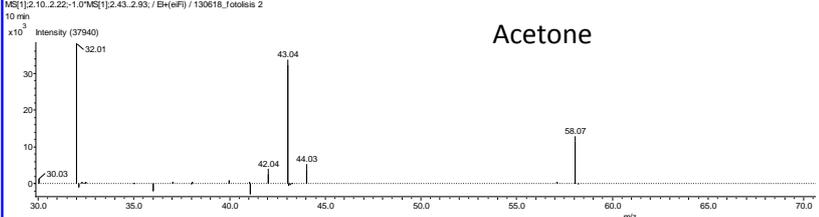
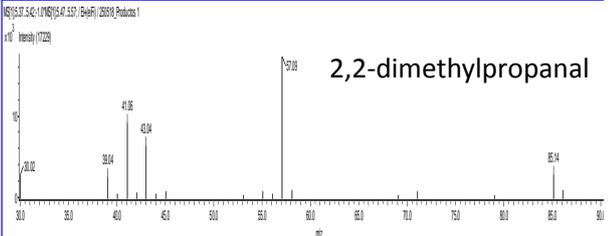
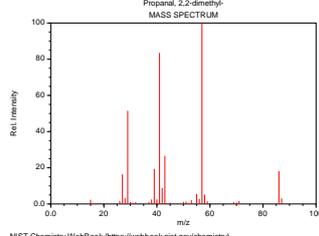
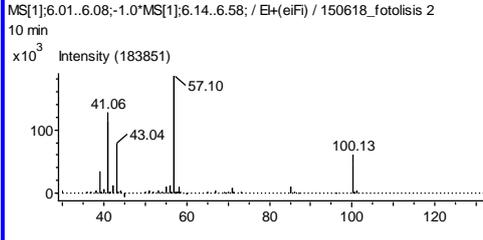
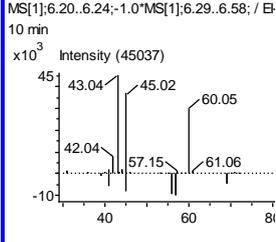
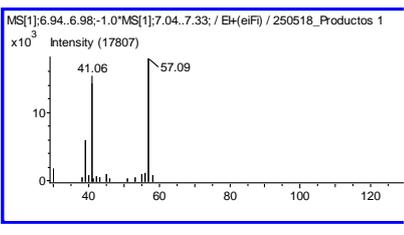
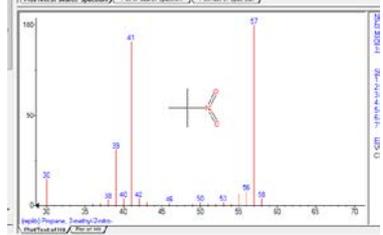
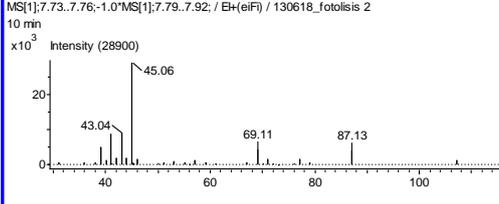
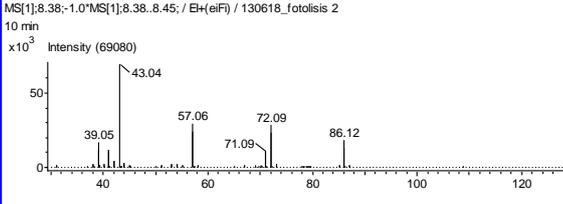


Fig. S10. SPME/GC-TOFMS chromatograms obtained for the reaction of 3,3DM2ButOH with Cl atoms, Cl atoms in the presence of +NO , HOH and NO₃ radicals and reference chromatograms of 3,3DM2ButOH and 3,3-dimethyl-2-butanone.

t_R (min)	EI MS
2.16	<p>MS[1]:2.10..2.22;-1.0*MS[1]:2.43..2.93; / E+(eF) / 130618_fotolisis 2</p> <p>10 min</p> <p>Intensity (37940)</p> <p>Acetone</p> 
5.39 ?	<p>MS[5]:7.54;-1.0*MS[5]:7.57; (E+eF) / 250518_Productos 1</p> <p>10 min</p> <p>Intensity (1722)</p> <p>2,2-dimethylpropanal</p>   <p>Propanal, 2,2-dimethyl- MASS SPECTRUM</p> <p>Rel. Intensity</p> <p>m/z</p> <p>NIST Chemistry WebBook (https://webbook.nist.gov/chemistry)</p>
6.03	<p>MS[1]:6.01..6.08;-1.0*MS[1]:6.14..6.58; / E+(eF) / 150618_fotolisis 2</p> <p>10 min</p> <p>Intensity (183851)</p> <p>3,3-dimethyl-2-butanone</p> 
6.22	<p>MS[1]:6.20..6.24;-1.0*MS[1]:6.29..6.58; / E-</p> <p>10 min</p> <p>Intensity (45037)</p> <p>Acetic Acid (SI 80%)</p> 
6.46	SPME
6.96	<p>MS[1]:6.94..6.98;-1.0*MS[1]:7.04..7.33; / E+(eF) / 250518_Productos 1</p> <p>10 min</p> <p>Intensity (17807)</p> <p>Nitrated compound</p>  
7.75	<p>MS[1]:7.73..7.76;-1.0*MS[1]:7.79..7.92; / E+(eF) / 130618_fotolisis 2</p> <p>10 min</p> <p>Intensity (28900)</p> 
8.37	<p>MS[1]:8.38;-1.0*MS[1]:8.38..8.45; / E+(eF) / 130618_fotolisis 2</p> <p>10 min</p> <p>Intensity (69080)</p> 

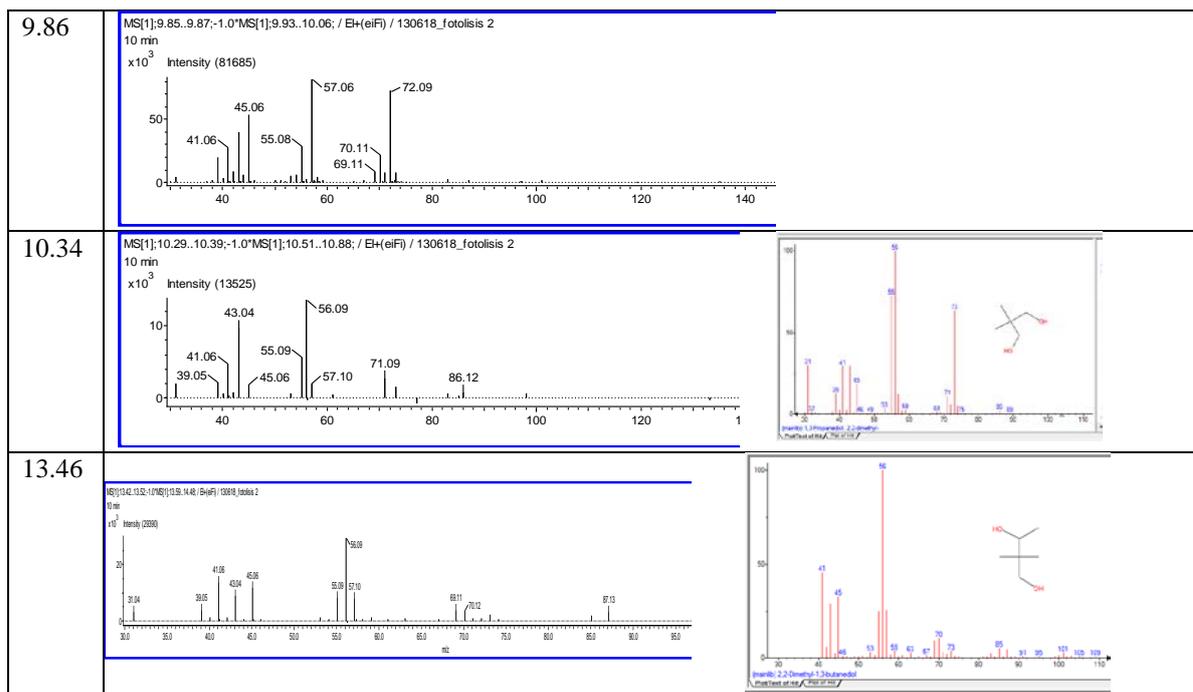
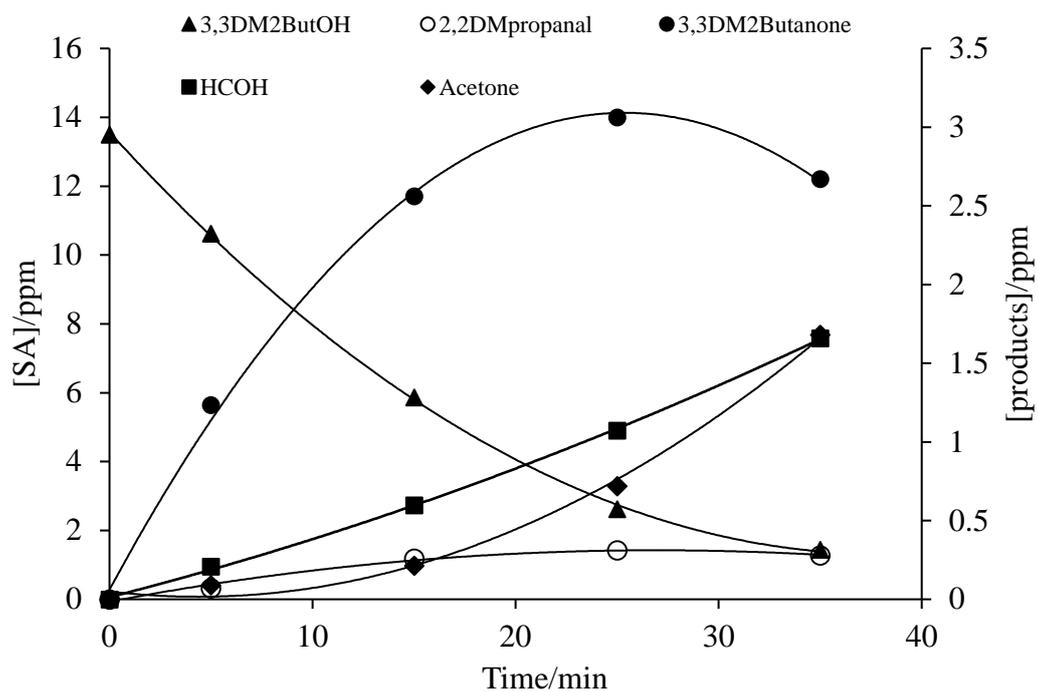


Fig. S11. EI-MS spectra of the peaks of chromatograms shown in Fig. S10 obtained for the reaction of 3,3DM2ButOH with Cl atoms, Cl atoms in the presence of +NO, HOH and NO₃ radicals.

A)



(B)

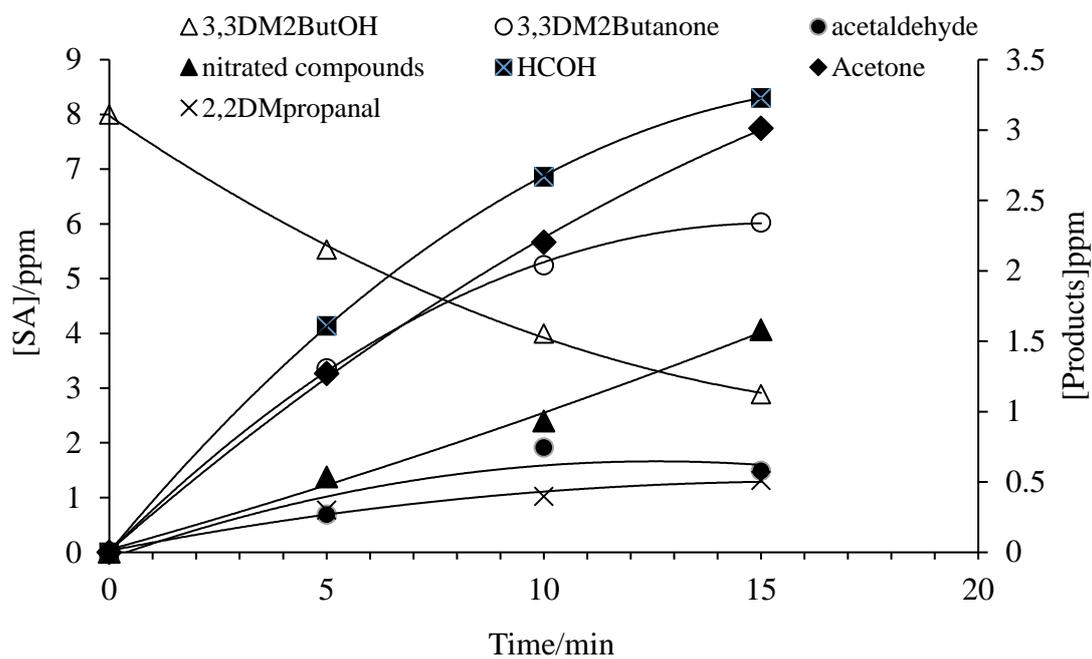


Fig. S12. Concentration-time profiles of 3,3DM2ButOH and reaction products formed for the reaction of 3,3DM2ButOH with Cl atoms in the absence (A) and the presence of NO (B).

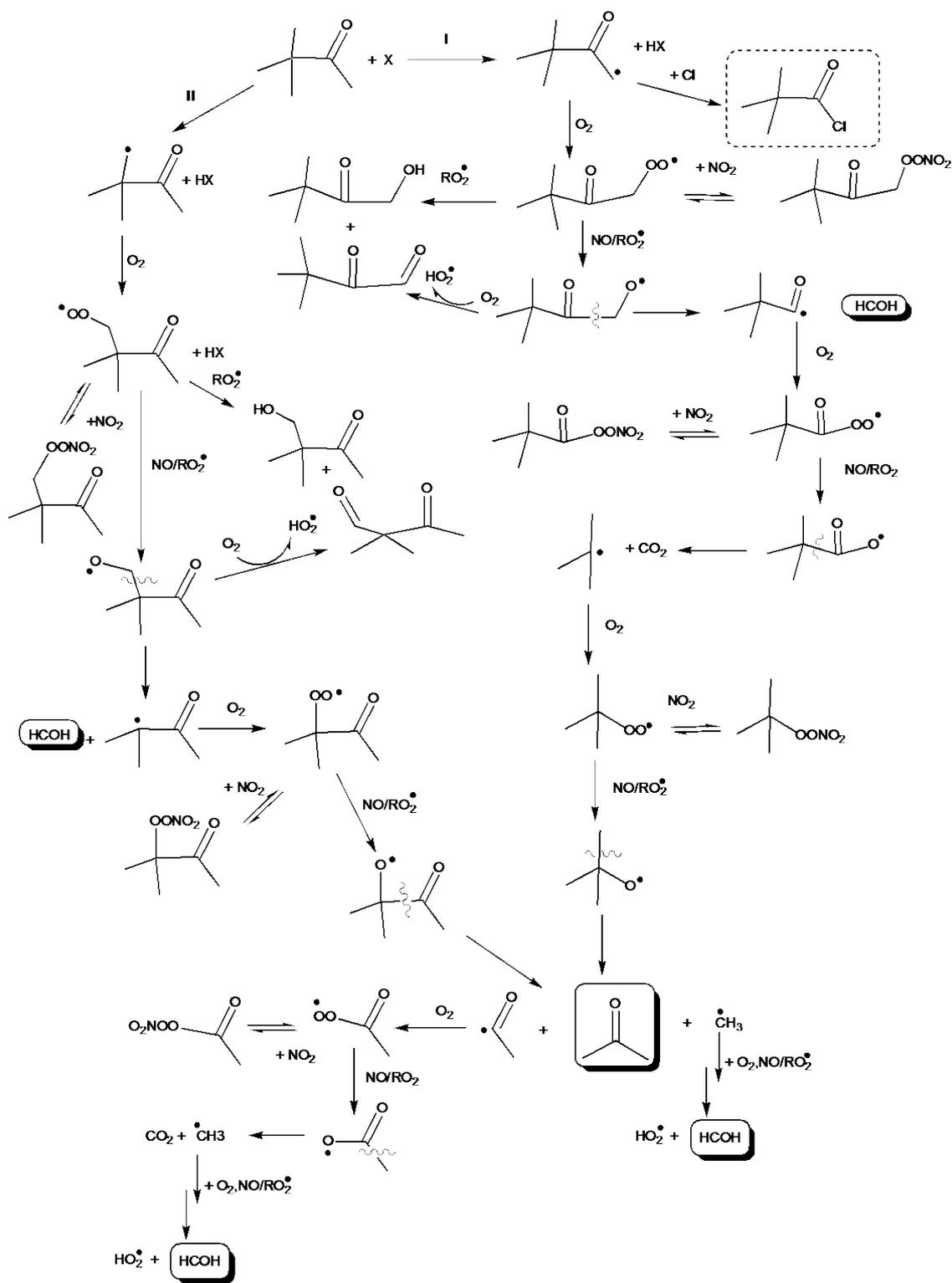


Fig. S13. Reaction mechanism for degradation of 3,3-dimethyl-2-butanone with the atmospheric oxidants in presence of NO_x.

References

- Andersen, V. F., Wallington, T. J., Nielsen, O. J. Atmospheric Chemistry of i-Butanol. *J. Phys. Chem. A.*, 114, 12462–12469, <https://doi.org/10.1021/jp107950d>, 2010.
- Aschmann, S. M. Arey, J. and Atkinson R. Kinetics and Products of the Reactions of OH Radicals with 4,4-Dimethyl-1-pentene and 3,3-Dimethylbutanal at 296 ± 2 K. *J. Phys. Chem. A*, 114, 5810–5816, <https://doi.org/10.1021/jp101893g>, 2010.
- Atkinson, R. Kinetics of the gas-phase reactions of OH radicals with alkanes and cycloalkanes. *Atmos. Chem. Phys.* 3, 2233-2307, <https://doi.org/10.5194/acp-3-2233-2003>, 2003.
- Bejan, I.G., Winiberg, F.A.F., Mortimer, N., Medeiros, D.J., Brumby, C.A., Orr, S.C., Kelly, J., Seakins, P.W. Gas-phase rate coefficients for a series of alkyl cyclohexanes with OH radicals and Cl atoms. *Int. J. Chem. Kinet.* 50(8),544-555, <https://doi.org/10.1002/kin.21179>, 2018.
- Bradley, W.R., Wyatt, S.E., Wells, J.R., Henley, M.V., Graziano, G.M. The Hydroxyl Radical Reaction Rate Constant and Products of Cyclohexanol. *Int. J. Chem. Kinet.*, 33, 108-117, [https://doi.org/10.1002/1097-4601\(200102\)33:2<108::AID-KIN1002>3.0.CO;2-%23](https://doi.org/10.1002/1097-4601(200102)33:2<108::AID-KIN1002>3.0.CO;2-%23), 2001 .
- Calvert, J.G., Derwent, R.G. Orlando, J.J., Tyndall, G.S., Wallington, T.J. Mechanisms of atmospheric oxidation of the alkanes. Oxford University Press, New York, 2008.
- Calvert, J.G., Mellouki, A., Orlando, J.J., Pilling, M.J., Wallington, T.J. The mechanisms of atmospheric oxidation of the oxygenates. Oxford University Press, New York, 2011.
- Calvert, G., Orlando, J.J., Stockwell, W.R., Wallington, T.J. The Mechanisms of Reactions Influencing Atmospheric Ozone. Oxford University Press, New York, 2015.
- Ceacero-Vega, A.A., Ballesteros, B., Bejan, I., Barnes, I., Jiménez, E., Albaladejo, J. Kinetics and Mechanisms of the Tropospheric Reactions of Menthol, Borneol, Fenchol, Camphor, and Fenchone with Hydroxyl Radicals (OH) and Chlorine Atoms (Cl). *J. Phys. Chem., A*, 116, 4097-4107, <https://doi.org/10.1021/jp212076g>, 2012.
- Gallego-Iniesta, M.P., Moreno, A., Martín, P., Tapia, A., Cabañas, B., Salgado. M.S. Reactivity of 2-ethyl-1-hexanol in the atmosphere. *Phys. Chem. Chem. Phys.* 12, 3294-3300, <https://doi.org/10.1039/B923899A>, 2010.
- Hooshiyar and Niki. Rate constants for the gas-phase reactions of Cl atoms with C2–C8 alkanes at T = 296 K. *Int. J. Chem. Kinet.*, 27, 1197-1206, <https://doi.org/10.1002/kin.550271206>, 1995.
- Jiménez. E., Lanza, B., Garzón, A., Ballesteros, B., Albaladejo, J. Atmospheric Degradation of 2-Butanol, 2-Methyl-2-butanol, and 2,3-Dimethyl-2-butanol: OH Kinetics and UV Absorption Cross Sections. *J. Phys. Chem. A*.109, 10903-10909, <https://doi.org/10.1021/jp054094g>, 2005.
- Mellouki, A., Oussar, F., Lun, X., Chakir, A. Kinetics of the reactions of the OH radical with 2-methyl-1-propanol, 3-methyl-1-butanol and 3-methyl-2-butanol between 241 and 373 K. *Phys Chem. Chem. Phys.* 6, 2951-2955, <https://doi.org/10.1039/B316514K>, 2004.
- Moreno, A., Salgado, S., Martin, P., Martinez, E., and Cabañas, B. Kinetic Study of the Gas Phase Reactions of a Series of Alcohols with the NO₃ Radical. *J. Phys. Chem. A*, 116, 42, 10383-10389, <https://doi.org/10.1021/jp305682h>, 2012.

- Moreno, A., Salgado, S., Taccone, R., Martín, P., Cabañas, B. Atmospheric degradation of saturated alcohols: room temperature rate coefficients for NO₃ radical reactions. *Atmos. Environ.*, 96, 229-235, <https://doi.org/10.1016/j.atmosenv.2014.07.037>, 2014.
- Wallington, T.J., Skewes, L.M., Siegl, W.O.; Wu, C., Japar, S.M. Gas phase reaction of Cl atoms with a series of oxygenated organic species at 295 K, *Int. J. Chem. Kinet.*, 20, 867-875, <https://doi.org/10.1002/kin.550201105>, 1988.
- Wallington, T.J., Skewes, L.M., Siegl, W.O., A relative rate study of the reaction of chlorine atoms with a series of chloroalkanes at 295 K. *J. Phys. Chem.*, 93,9, 3649-3651. <https://doi.org/10.1021/j100346a054>, 1989.