Contributions of Nordic anthropogenic emissions on air pollution and premature mortality over the Nordic region and the Arctic

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Abstract

This modelling study presents the sectoral contributions of anthropogenic emissions in the four Nordic countries; Denmark, Finland, Norway and Sweden, on air pollution levels and the associated health impacts and costs over the Nordic and the Arctic region for the year 2015. The Danish Eulerian Hemispheric Model (DEHM) has been used on a 50 km resolution over Europe in tagged mode in order to calculate the response of a 30% reduction of each emission sector in each Nordic country individually. The emission sectors considered in the study were energy production, residential/commercial heating, industry, traffic, off-road mobile sources, and waste management/ agriculture. In total, 28 simulations were carried out. Following the air pollution modelling, the Economic Valuation of Air Pollution (EVA) model has been used to calculate the associated premature mortality and their costs. Results showed that more than 80% of PM$_{2.5}$ concentrations in the considered four Nordic countries were transported from outside these four countries. The leading emission sector in each country was found to be non-industrial combustion (contributing by more than 60% to the total PM$_{2.5}$ mass), except for Sweden, where industry contributed to PM$_{2.5}$ with a comparable amount as non-industrial combustion. In addition to residential combustion, the next most important source categories were industry, agriculture and traffic. The main chemical constituent of PM$_{2.5}$ concentrations was organic carbon in all countries, which suggested that residential wood burning was the dominant national source of pollution in the Nordic countries. We have estimated the total number of premature mortality cases due to air pollution to be around 4 000 in Denmark and Sweden and around 2 000 in Finland and Norway. These premature mortality cases led to a total cost of 7 billion Euros in the selected Nordic countries. The assessment of the related premature mortality and associated cost estimates suggested that residential combustion, together with industry and traffic, will be the main sectors to be targeted in emission mitigation strategies in the future.
1. Introduction

Air pollution is the world’s largest single environmental health risk (WHO, 2014), estimated to be responsible for 3.7 million premature deaths in 2012 from urban and rural sources worldwide. In Europe, recent results (Andersson et al., 2009; Brandt et al., 2013a; 2013b; Geels et al., 2015; Im et al., 2018a; Liang et al., 2018; Solazzo et al., 2018) show that outdoor air pollution causes ~500 000 premature deaths in Europe. Brandt et al. (2013a) calculated that due to exposure to ambient air pollution, there were around 3.500 premature deaths in 2011 in Denmark alone. Lehtomäki et al. (2018) have recently evaluated that ambient air pollution caused approximately 2000 premature deaths in Finland in 2015. Other studies have made assessments for some of the Nordic countries (Denmark, Sweden and Finland) with estimates ranging from 6500 to 9500 for the year 2000 (Geels et al., 2014; Watkiss et al., 2005, Karvosenoja et al., 2010, respectively). Kukkonen et al. (2018) and Forsberg et al. (2015) have concluded that long-range transported fine particulate matter dominates the health effects in the Scandinavian countries, with largest contribution to long-term effects in Sweden originate from south-western Europe, while the largest contribution to short-term exposure originates from south-eastern Europe (Jönsson et al. 2013).

Air pollution is a transboundary problem covering global, regional, national and local sources, leading to large geographic variability and therefore to large differences in the geographical distribution of human exposure to air pollution (Im et al., 2018a,b). In the Nordic countries, there are large spatial differences in air pollution levels because of long-range transported and polluted air masses especially from the south and east as well as due to the degree of urbanization. There are also local differences depending on wind direction and distance from local emission sources such as road transport, power plants and industry (Brandt et al., 2013a). Furthermore, the widespread use of domestic wood stoves in the Nordic countries represents a special challenge for exposure to air pollution, where e.g. more than a third of the health impacts from Danish emissions are due to smoke from wood stoves. International ship traffic is also a significant source of air pollution and health impacts in highly trafficked areas of the Baltic and North Seas (Brandt et al., 2013b; Jalkanen et al., 2016, Johansson et al., 2017). Based on simulations for the period 1997-2003, Andersson et al. (2009) calculated that Sweden contributed to 1.4% of the European primary PM$_{2.5}$ (PM$_{2.5}$) concentrations while Denmark, Finland and Norway were responsible for 4% of European PM$_{2.5}$. Contribution to secondary inorganic PM$_{2.5}$ (SIA) levels were much smaller (0.5% from Sweden and 1.4% from Denmark, Finland and Norway). They also calculated a death rate increase of 2 and 3% due to exposure to PM$_{2.5}$ and SIA, respectively, in Europe due to emissions from Denmark, Finland, Norway and Sweden.

The external (or indirect) costs to society related to health impacts from air pollution are substantial. In the whole of Europe, the total external costs have been estimated to be approx. 800 billion Euros per year and in Denmark alone the external costs are nearly 4 billion Euro per year (Brandt et al., 2013a). In a more recent study, Im et al. (2018a), using a multi-model ensemble of 14 chemistry transport models (CTM), estimated that ambient air pollution in Europe in 2010 was responsible for 414 000±100 000 premature deaths, leading to a cost of 300 billion Euros. The study also showed that a 20% decrease of anthropogenic emissions in Europe source could avoid 47 000 premature deaths in Europe, while a similar reduction in the U.S. would avoid around 1 000 premature deaths in Europe due to long-range transport.

The Nordic countries are generally characterized with low air pollution levels compared to the rest of Europe. However, there are still large impacts of air pollution on human health and climate in the...
region itself (Brandt et al., 2013a; Forsberg et al., 2015), as well as over the Arctic (Sand et al., 2015). The Task Force on Short Lived Climate Forcers of the Arctic Council reported that black carbon (BC) emission sources within Arctic Council nations generally have a greater impact on climate change per unit of emissions compared to sources outside of the Arctic (Arctic Council, 2011). The report also states that measures aimed at decreasing these emissions will have positive health effects for communities exposed to air pollution. In a recent study, Sand et al. (2015) showed that although the largest Arctic warming source is from Asian emissions, the Arctic is most sensitive, per unit mass emitted, to SLFCs emissions from a small number of activities within the Arctic nations themselves.

The aim of the study is to quantify the contributions of the main emission sectors in each of the Nordic countries to air pollutant levels and their impacts on premature mortality and associated costs in the Nordic region and the Arctic, in order to identify the emission sectors that should be targeted for mitigation to decrease the air pollution and exposure levels in the Nordic countries. In order to achieve this, we have coupled the Danish Eulerian Hemispheric Model (DEHM) to the Economic Valuation of Air Pollution (EVA) model and conducted a number of perturbation simulations targeting different emission sectors in the four Nordic countries; Denmark, Finland, Norway and Sweden, for the year 2015. The models and perturbation simulations are described in Section 2, the model evaluation against surface measurements in the Nordic countries are presented in Section 3.1, the contributions of sectoral emissions on the air pollution levels in the Nordic region and the Arctic are presented in Section 3.2., and the health impacts and associated costs are presented in Section 3.3. Conclusions are given in Section 4.

2. Materials and methods

2.1. DEHM

DEHM model was originally developed mainly to study the transport of SO2 and SO4 to the Arctic (Christensen 1997), but has been extended to different applications during the last decades. It has been extensively documented in Brandt et al. (2012) and evaluated in several intercomparison studies (e.g. Solazzo et al., 2012 a,b; Solazzo et al., 2017; Im et al., 2018a,b) and recently joined the suit of operational models in the Copernicus Atmospheric Monitoring System (CAMS) to provide regional forecasts of air pollution over Europe. The DEHM model uses a 150 km × 150 km spatial resolution over the Northern Hemisphere, then nests to 50 km × 50 km resolution over Europe, extending up to 100 hPa through 29 vertical levels, with the first layer height of approximately 20 m. The meteorological fields were simulated by the Weather Research and Forecast Model (WRF, Skamarock et al., 2008) setup with identical domains and resolution. The gas-phase chemistry module includes 58 chemical species, 9 primary particles and 122 chemical reactions (Brandt et al., 2012). The model also describes atmospheric transport and chemistry of lead, mercury, CO2, as well as POPs. Secondary organic aerosols (SOA) are calculated using the Volatility Base System (VBS: Bergstrom et al., 2012).

In the current study, the DEHM model used anthropogenic emissions from the EDGAR-HTAP database and biogenic emissions are calculated online based on the MEGAN model. The total emission per country for the different pollutants are presented in Table 1. The sectoral distributions of emissions in each country are presented in Fig.1. As seen in the Table 2, most SNAP (Selected Nomenclature for Air Pollutants; CEIP, 2019) sectors are considered individually, while some are merged in order to reduce the computational costs. All sectors in relation to industrial activities...
As seen in Fig. 1, non-industrial combustion (orange bars), where residential combustion dominates, stands out as a major source contributing to CO and PM emissions while industry (grey bars) (Table 2) is the largest source of NMVOCs, NOx and SOx. Traffic (yellow bars) also contributes significantly to CO and NOx. The largest source of NH3 is from agriculture and waste management, as seen in the ‘Others’ (green bars) (Table 2).

Table 1. Total pollutant emissions in the Nordic countries (in Gg) in 2015.

<table>
<thead>
<tr>
<th></th>
<th>CO</th>
<th>NH3</th>
<th>NMVOC</th>
<th>NOx</th>
<th>SO2</th>
<th>PM10</th>
<th>PM2.5</th>
</tr>
</thead>
<tbody>
<tr>
<td>DK</td>
<td>251</td>
<td>75</td>
<td>106</td>
<td>102</td>
<td>9</td>
<td>31</td>
<td>20</td>
</tr>
<tr>
<td>FI</td>
<td>302</td>
<td>31</td>
<td>85</td>
<td>128</td>
<td>41</td>
<td>31</td>
<td>19</td>
</tr>
<tr>
<td>NO</td>
<td>378</td>
<td>28</td>
<td>155</td>
<td>133</td>
<td>16</td>
<td>35</td>
<td>27</td>
</tr>
<tr>
<td>SE</td>
<td>413</td>
<td>54</td>
<td>159</td>
<td>129</td>
<td>18</td>
<td>37</td>
<td>18</td>
</tr>
</tbody>
</table>

Fig. 1. Sectoral emissions of major pollutants in the Nordic countries.
2.1.1. Tagging Method

Tagging method keeps track of contributions to the concentration field from a particular emission source or sector, as explained in detail in Brandt et al. (2013a). Tagging involves modelling the background concentrations and the $\delta$-concentrations (the contributions from a specific emission source or sector to the overall air pollution levels) in parallel, where special treatment is required for the non-linear process of atmospheric chemistry, since the $\delta$-concentrations are strongly influenced by the background concentrations in such processes. Although this treatment involves taking the difference of two concentration fields, it does not magnify the spurious oscillations (the Gibbs phenomenon), which are primarily generated in the advection step. The non-linear effects can be accounted for in the $\delta$-concentrations without losing track of the contributions arising from the specific emission source or sector.

2.1.2. Model evaluation

Surface concentrations modelled by the DEHM model were evaluated against data at selected urban background and regional or global monitoring stations in each Nordic country. The statistical comparisons included using correlation coefficient ($r$), mean bias ($MB$) and normalized mean error ($NME$) and root mean square error ($RMSE$), shown in the supplementary material. The station information is provided in Table S1.

The Danish Air Quality Monitoring Programme consists of an urban monitoring network that includes stations in the four largest Danish cities (Aalborg, Aarhus, Copenhagen and Odense) and two background stations in rural areas (Keldsnor and Risø). The design of the Danish air quality monitoring network is based on location of one or more pairs of stations in each of the four cities (Ellermann et al., 2015). In each city one of the stations is located at a pollution hot spot close (at the sidewalk) to a street lane with a high traffic density. The other station is located within a few hundred meters from the street station. It is placed so that it is not influenced by emissions from a single or a few streets or other nearby sources and hence is representative for the urban background pollution. In most cases the background stations are placed on rooftops. In addition, rural stations monitor the pollution outside city areas.

The measurement data for Finland represents regional and urban background levels. Data from the global and regional background stations are reported to European Monitoring and Evaluation Programme (EMEP) under the CLRTAP (Convention on Long-range Transboundary Air Pollution, http://www.unece.org/env/lrtap), and are available at http://ebas.nilu.no. The data for the urban background stations are reported at the ‘Air Quality in Finland’ web pages by the Finnish Meteorological Institute (https://en.ilmatieteenlaitos.fi/air-quality).

The measurement dataset from Norway is from the national monitoring program of air pollutants financed by the Norwegian Environment Agency (Aas et al, 2018), and also reported to European Monitoring and Evaluation Programme (EMEP) under the CLRTAP (Convention on Long-range Transboundary Air Pollution, http://www.unece.org/env/lrtap). The data is openly available at http://ebas.nilu.no. The data from the city background stations is reported to EEA (European Environmental Agency, http://www.eea.europa.eu) as required in the EU air quality directive (EU, 2008) and it is available at http://www.luftkvalitet.info.
The measurement dataset for Sweden were extracted from the openly available Shair data base (http://shair.smhi.se/portal/concentrations-in-air), which includes most national environmental data and is financed by the Swedish Environmental Agency. The observation sites used here were carefully selected to represent urban background at rooftop level, rural or regional background, and to have known good quality.

2.2. EVA

The EVA system (Brandt et al., 2013a,b; Geels et al., 2015; Im et al., 2018) is based on the impact-pathway chain method (Friedrich and Bickel, 2001), and it calculates health impacts of ambient air pollution due to exposure to surface concentrations of \( O_3 \), \( CO \), \( SO_2 \) and \( PM_{2.5} \), and the associated external costs. The EVA system requires gridded concentrations along with gridded population data, exposure-response functions (ERFs) for health impacts, and economic valuation functions of the impacts from air pollution. The EVA system can estimate various health impacts, including different morbidity outcomes as well as acute and chronic mortality, related to short term (acute) exposure to \( O_3 \), \( CO \), and \( SO_2 \), and long term (chronic) exposure to \( PM_{2.5} \). EVA calculates and uses the annual mean concentrations of \( CO \), \( SO_2 \), and \( PM_{2.5} \), while for \( O_3 \) it uses the SOMO35 metric that is defined as the annual sum of the daily maximum of 8-hour running average over 35 ppb, following WHO (2013) and EEA (2017). In addition, EVA uses population densities over fixed age intervals, corresponding to babies, children, adults and elderlies.

Exposure response functions (ERF) for all-cause chronic mortality due to \( PM_{2.5} \) are based on Pope et al., 2002; Krewski et al., 2009; WHO, 2013). Following Pope et al. (2002), the relative risk (RR) is 1.062 (1.040-1.083) on 95% confidence interval. The counterfactual \( PM_{2.5} \) concentration is assumed to be 0 \( \mu g \) m\(^{-3}\) following the EEA methodology, meaning that the impacts have been estimated for the full range of modelled concentrations. Regarding short-term exposure to \( O_3 \), EVA uses the ERF recommended by the CAFE Programme (Hurley et al., 2005) and WHO (2013) that uses the daily maximum of 8-hour mean \( O_3 \) concentrations. The ERFs used in EVA to calculate mortality are presented in Table S2. For the valuation of the health impacts, a value of EUR 1.5 million was applied for preventing an acute death, following expert panel advice (EC, 2001), while for the valuation of a life year, a value of EUR 57,500 per year of life lost (YOLL) were applied (Alberini et al., 2006). More details can be found in Im et al. (2018a).

2.3. Scenarios (response and contribution)

We have applied a 30% reduction on land-based anthropogenic emissions from each of the continental Nordic countries, which include Denmark, Finland, Norway and Sweden. The perturbations are applied based on the SNAP sectors. Each simulation perturbed a SNAP sector from an individual Nordic country, which are listed in Table 2.

DEHM model has been run on “tagged” mode, explained in section 2.1., so each simulation included a “perturbed” and “non-perturbed” concentration, which we used to calculate the response to the 30% reduction in the particular country and sector. These responses are then converted to contributions by assuming a linear extrapolation to 100%. We have also simulated a 100% reduction scenario to all sectors per country (“All” in Table 2) to see the impact of a 100% reduction and how it compares to the scaled 30% response at each country.
Table 2. Source sectors used in the perturbation scenarios.

<table>
<thead>
<tr>
<th>Source Sectors</th>
<th>SNAP Code</th>
</tr>
</thead>
<tbody>
<tr>
<td>Combustion in energy and transformation industries</td>
<td>1</td>
</tr>
<tr>
<td>Non-industrial Combustion</td>
<td>2</td>
</tr>
<tr>
<td>Industry</td>
<td>3,4,5,6</td>
</tr>
<tr>
<td>Road transport</td>
<td>7</td>
</tr>
<tr>
<td>Other mobile sources and machinery</td>
<td>8</td>
</tr>
<tr>
<td>Others (waste and agriculture)</td>
<td>9,10</td>
</tr>
<tr>
<td>All</td>
<td>1,2,3,4,5,6,7,8,9,10</td>
</tr>
</tbody>
</table>

3. Results and Discussion

3.1. Evaluation

Surface ozone and PM$_{2.5}$ concentrations calculated by the DEHM model have been evaluated using surface observations from the Nordic countries, described in 2.1.2. The comparison of the mean of all observed concentrations in each country and the corresponding modelled concentrations are presented in Table 3 while Figs. 2 and 3 present Taylor diagrams for each station in each Nordic country. As seen in Table 3, O$_3$ levels are well reproduced by the DEHM model over all countries ($r > 0.7$), however with a slight overestimation of ~10% over Denmark, Finland and Sweden, and ~30% over Norway. The monthly variations of PM$_{2.5}$ levels, averaged over all stations in each Nordic country are well reproduced for Denmark and Norway ($r$~0.7), moderately over Sweden and poorly ($r$~0) over Finland (Table 3).

Table 3. Model evaluation for the daily mean concentrations of O$_3$ and PM$_{2.5}$ for all the selected stations in the Nordic countries.

<table>
<thead>
<tr>
<th>Country</th>
<th>O$_3$</th>
<th>PM$_{2.5}$</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>R</td>
<td>MB (µg m$^{-3}$)</td>
</tr>
<tr>
<td>Denmark</td>
<td>0.81</td>
<td>5.67</td>
</tr>
<tr>
<td>Finland</td>
<td>0.74</td>
<td>4.77</td>
</tr>
<tr>
<td>Norway</td>
<td>0.64</td>
<td>12.02</td>
</tr>
<tr>
<td>Sweden</td>
<td>0.74</td>
<td>7.00</td>
</tr>
</tbody>
</table>
Fig. 2. Taylor diagrams for daily mean O$_3$ for all stations in a) Denmark, b) Finland, c) Norway and d) Sweden.

Fig. 3. Taylor diagrams for daily mean PM$_{2.5}$ for all stations in a) Denmark, b) Finland, c) Norway and d) Sweden.

3.2. Sectoral contributions to surface concentrations
3.2.1. Nordic countries

In general, the long-term transport of air pollutants from one country to another is dependent on the global and regional atmospheric circulation and on the relative geographic positions of the countries. Nordic countries are influenced by substantial long-range transported contributions of air pollution especially from the central, western and central eastern parts of Europe. In the region containing the continental Nordic countries, the prevailing atmospheric flow directions near the ground surface are from the west, south-west and south. Caused by the atmospheric circulation patterns, it is therefore to be expected that, e.g., the emissions in Denmark will have a relatively larger influence on the pollution levels in the other Nordic countries than those in Finland.

Fig. 4 compares the contribution of the total contribution of each Nordic country on the surface concentrations over the country itself, with contributions from rest of the Scandinavian countries and rest of the world. The figure clearly shows that over 90% or more of PM2.5 surface levels are coming outside each country. Similar high contributions for other species including CO also shows that Scandinavian countries are exposed to airmasses coming from rest of the world while local pollution is low. The figure also shows that PM$_{2.5}$ levels are generally low in the Scandinavian countries, with annual means of 2-4 µg m$^{-3}$ (highest in Denmark and lowest in Finland). Similar to PM$_{2.5}$, annual mean surface O$_3$ levels are also low (~30 µg m$^{-3}$).

Fig.4. Contribution of national, Scandinavian and other sources on. The surface levels of major air pollutants over a) Denmark, b) Finland, c) Norway and d) Sweden. Note that CO concentrations are divided by 20 to scale with other pollutants.
Danish emissions contribute to only 0.4 µg m⁻³ (10%) of the surface PM₂.₅ concentrations over Denmark (3.8 µg m⁻³), while contributions to other Nordic countries are less than 1% (Fig. 5). Non-industrial combustion (SNAP2), which is dominated by residential combustion, is responsible for 0.24 µg m⁻³ (65%) of the Danish contribution to surface PM₂.₅ concentrations over Denmark. Residential combustion contributes to 0.16 µg m⁻³ (60%) of the Danish contribution to surface organic carbon (OC) concentrations over the country, suggesting the importance of residential wood burning for heating. Industry contributes to 0.7 g m⁻³ (31%) of the Danish contribution to the surface SO₂ concentrations over Denmark, while on-road and off-road transport contributes equally to the Danish share of the in surface NO₂ concentrations by 0.55 µg m⁻³ (~72% together).

Agriculture and waste handling are important sources for surface SO₄ levels over Denmark as well as over the other Nordic countries, via the formation of ammonium sulfate ((NH₄)₂SO₄) due to the large ammonia (NH₃) emissions from these sectors. Among the other Nordic countries, Danish emissions, in particular non-industrial combustion, have the largest contribution to the pollutant levels over Sweden.

Fig. 5. Contributions of sectoral Danish emissions on surface a) BC, b) SO₄, c) OC, d) PM₂.₅, e) SO₂ and f) NO₂ over the Nordic countries.

Contributions of the Norwegian emissions over the Nordic countries are presented in Fig. 6. Similar to the Danish emissions, Norwegian emissions contribute to 0.17 µg m⁻³ (7%) of the surface PM₂.₅ concentrations over Norway, while contributions to other Nordic countries are below 1%, except for NO₂, where on-road transport emissions from Norway contributes to almost 0.08 µg m⁻³ (15%) of the surface NO₂ levels over Finland. Non-industrial combustion is the main source of pollutant levels, in particular for OC, where Norwegian emissions are responsible for 0.07 µg m⁻³ (77%) of local contribution to the surface OC levels over Norway. Industry is a major source of surface SO₂ levels over Norway, contributing to 0.01 µg m⁻³ (76%) of the local contribution.
Fig. 6. Contributions of sectoral Norwegian emissions on surface a) BC, b) SO$_4$, c) OC, d) PM$_{2.5}$, e) SO$_2$ and f) NO$_2$ over the Nordic countries.

Fig. 7 shows the contributions of Finnish emissions on the pollutant levels over the Nordic countries. Similar to Denmark and Norway, non-industrial combustion, which is dominated by residential combustion, is the major source of pollution over Finland, although contributions are lower compared to Denmark and Norway (0.06 µg m$^{-3}$ (58%) of PM$_{2.5}$ and 0.04. µg m$^{-3}$ (66%) of OC). Another noticeable difference is that energy production is also an important contributor to surface SO$_2$ (0.01. µg m$^{-3}$: 43%) and SO$_4$ (0.02. µg m$^{-3}$; 43%) levels over Finland. Finnish emissions, in particular industrial combustion, contribute largest to the air pollution over Sweden.

Fig. 7. Contributions of sectoral Finnish emissions on surface a) BC, b) SO$_4$, c) OC, d) PM$_{2.5}$, e) SO$_2$ and f) NO$_2$ over the Nordic countries.

Contributions from the Swedish emission sources to surface pollutant levels over the Nordic countries are presented in Fig. 8. Unlike other Nordic countries, Swedish emissions have larger contributions to pollution levels over the other Nordic countries, in particular over Norway. The figure also shows that Sweden does not experience as dominant contribution from non-industrial...
combustion (28%) like the other Nordic countries show. Swedish emissions from SNAP2 are much lower than for the rest of the Nordic countries (official emissions reported to the CLRTAP), most probably due to lower emission factors. Residential combustion and industry contribute similarly to the surface PM$_{2.5}$ levels. Industry also has a dominant contribution to surface SO$_4$ levels (0.01 µg m$^{-3}$: 51%), as well to SO$_2$ (0.01. µg m$^{-3}$: 74%) and BC (0.004 µg m$^{-3}$: 31%).

Fig 8. Contributions of sectoral Swedish emissions on surface a) BC, b) SO$_4$, c) OC, d) PM$_{2.5}$, e) SO$_2$ and f) NO$_2$ over the Nordic countries.

3.2.2. Arctic

The contributions of the emission sources in the different Nordic countries on the surface aerosol concentrations over the Arctic region (defined as the area north of 67 °N latitude) are presented in Fig. 9. Results show that overall, Norway has the largest contribution to surface aerosol levels over the Arctic, while Denmark has the lowest contribution, although contributions are only a few percent. Norwegian emissions, in particular non-industrial combustion, contributes to about 2% of the surface BC levels over the Arctic. Non-industrial combustion in the Nordic countries is also the largest contributor to BC levels, except for Sweden, where industry plays a more important role. Non-industrial combustion is also the dominant contributor to OC levels over the Arctic.

Sulfate levels are largely influenced by the contributed from the agriculture and waste treatment facilities over the Nordic countries. Contributions to Arctic PM$_{2.5}$ levels are similar to the contributions to the BC levels.
3.2.3. Spatial distributions of contributions

The geographical distributions of total anthropogenic emissions from each Nordic country to surface PM$_{2.5}$ and O$_3$ levels are calculated to investigate the extent of contributions from each Nordic country to its neighbours and to the Arctic. Fig. 10 shows the annual-mean relative contributions (%) of total land-based anthropogenic emissions to surface O$_3$ levels in the Nordic region from each country. The annual-mean contributions are very low, generally lower than 5% and are mainly calculated in the source country itself. Danish anthropogenic emissions (Fig. 10a) leads to a titration of around 4-5%, particularly over the Zealand region over the country where it leads to a very small O$_3$ increase (>1%) in the downwind towards south. The largest impact of Finnish emissions is around the Helsinki area, responsible for up to 5% of surface O$_3$ destruction over the area (Fig. 10b). Similar to Denmark, Finnish emissions also lead to an increase of surface O$_3$ levels by less than 1% over the downwind regions to the southeast and northwest. Impact of Norwegian emissions to surface O$_3$ levels (Fig. 10c) are largest (2%) over the Oslo area and the impact extents over the northern part of Oslo with a slightly larger contribution to O$_3$ levels compared to Denmark and Finland. The Swedish emissions have a larger geographical impact on the surface O$_3$ levels (Fig. 10d) over the country itself compared to the other Nordic countries but the magnitude is similar to the impact from the Norwegian emissions.
Fig. 10. Spatial distributions of annual-mean relative contributions (%) of total emissions from a) Denmark, b) Finland, c) Norway, and d) Sweden to surface $O_3$ levels in the Nordic region.

Fig. 11 shows the annual-mean relative contributions of each Nordic country on the surface PM$_{2.5}$ levels in the entire model domain. Danish anthropogenic emissions are responsible for up to 20% of surface PM$_{2.5}$ levels over Denmark, with largest contributions over the Zealand region (Fig. 11a). Danish land emissions also impact the surface PM$_{2.5}$ levels over the southern part of Sweden and Norway, by around 4% and 2%, respectively. The Finnish anthropogenic emissions have the largest impact on surface PM$_{2.5}$ levels over the southern part of the country, around the capital region by up to 30% (Fig. 11b). Finnish emissions also have a small impact, lower than 3%, on the central part of Sweden and northern parts of Norway. Norwegian anthropogenic emissions have largest contributions to surface PM$_{2.5}$ level around the capital region by up to 30%, while there is also a significant impact on surface PM$_{2.5}$ levels over Sweden by around 7% (Fig. 11c). Finally, Swedish anthropogenic emissions have large contribution to surface PM$_{2.5}$ levels over the Stockholm area by around 15% and also contributes to PM$_{2.5}$ levels over Finland, in particular over the southwestern parts of Finland, by up to 5% (Fig. 11d).

Fig. 11 also shows the impact of anthropogenic emissions from each Nordic country to the surface PM$_{2.5}$ over the Arctic. Overall, the impacts are very small, around a few per cent, as seen in the figure. The Danish emissions (Fig. 11a) have a more local contribution compared to other Nordic countries and the impact does not reach above roughly 70 °N. The outflow from Finland, Norway...
and Sweden can reach to the central Arctic ocean over to the northern parts of Greenland, however contributions are around 1-2% (Figs. 1b-d).

Fig. 1. Spatial distributions of annual-mean relative contributions (%) of total emissions from a) Denmark, b) Finland, c) Norway, and d) Sweden to surface PM$_{2.5}$ levels over the Nordic and the Arctic regions (north of 67°N).

3.3. Contribution to premature mortality and costs

The number of acute and chronic premature mortality in the four selected Nordic countries and the Arctic region (north of 67°N), along with the associated costs are presented in Table 4. As seen in the Table, chronic mortality due to PM$_{2.5}$ is the major source for premature mortality, as EVA calculates chronic mortality only due to exposure to PM$_{2.5}$ (see Table S2). The highest number of cases is calculated for Sweden (~4 200 cases), followed by Denmark (~3 500 cases), Finland (~1 800) and Norway (~1 700). These numbers lead to an associated cost of more than 2 billion Euros in Sweden and Denmark and ~ 1 billion Euros in Finland and Norway. The number of premature death cases are comparable with existing literature (e.g. Brandt et al., 2013a for
Denmark; Solazzo et al., 2018 for all four Nordic countries; EEA, 2017 for all four Nordic countries. In the Arctic region, the total number of premature mortality cases is calculated to be 94, 93 of which are due to exposure to PM$_{2.5}$ (chronic), leading to a cost of 58 million Euros.

Table 4. Acute and chronic premature death cases in the Nordic countries and the Arctic region (north of 67°N) in 2015 and the associated costs.

<table>
<thead>
<tr>
<th></th>
<th>Denmark</th>
<th>Finland</th>
<th>Norway</th>
<th>Sweden</th>
<th>Arctic</th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Premature Mortality (number of cases)</strong></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Acute</td>
<td>19</td>
<td>18</td>
<td>6</td>
<td>25</td>
<td>1</td>
</tr>
<tr>
<td>Chronic</td>
<td>3 332</td>
<td>1 707</td>
<td>1 596</td>
<td>4 091</td>
<td>93</td>
</tr>
<tr>
<td><strong>Total</strong></td>
<td>3 351</td>
<td>1 725</td>
<td>1 602</td>
<td>4 116</td>
<td>94</td>
</tr>
<tr>
<td><strong>Cost (million Euros)</strong></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Acute</td>
<td>30</td>
<td>28</td>
<td>9</td>
<td>38</td>
<td>1</td>
</tr>
<tr>
<td>Chronic</td>
<td>2 031</td>
<td>1 040</td>
<td>973</td>
<td>2 494</td>
<td>57</td>
</tr>
<tr>
<td><strong>Total</strong></td>
<td>2 061</td>
<td>1 068</td>
<td>982</td>
<td>2 532</td>
<td>58</td>
</tr>
</tbody>
</table>

The EVA model has been used to calculate the contributions of Nordic emissions to the total premature mortality (acute + chronic) in the Nordic countries for the year 2015. Table 5 presents a source/receptor matrix of the contributions to premature mortality on the Nordic countries. Danish emissions contribute to ~400 premature deaths in Denmark, dominated by agriculture (33%), non-industrial combustion (31%) and traffic (18%). In Norway, the dominating sector contributing is non-industrial combustion, responsible for 48% of the ~200 premature deaths in Norway. In Finland, the total number of premature deaths in 2015 is calculated to be ~270, where non-industrial combustion and traffic are responsible for more than half. Finally, in Sweden, traffic and waste management/agriculture are responsible for 50% of the total premature death in Sweden (~330).

Table 5. Source/Receptor relationships of the contributions of anthropogenic emissions from the Nordic countries to the premature mortality in the Nordic area.

<table>
<thead>
<tr>
<th>Source/Receptor</th>
<th>Denmark</th>
<th>Finland</th>
<th>Norway</th>
<th>Sweden</th>
</tr>
</thead>
<tbody>
<tr>
<td>Denmark</td>
<td>422</td>
<td>24</td>
<td>29</td>
<td>198</td>
</tr>
<tr>
<td>Finland</td>
<td>8</td>
<td>274</td>
<td>9</td>
<td>42</td>
</tr>
<tr>
<td>Norway</td>
<td>33</td>
<td>26</td>
<td>203</td>
<td>86</td>
</tr>
<tr>
<td>Sweden</td>
<td>57</td>
<td>64</td>
<td>27</td>
<td>340</td>
</tr>
</tbody>
</table>

Fig. 12 shows the contributions of sectoral emissions from each Nordic country to the total premature death cases in 2015 in the different Nordic countries. Overall, Nordic countries contribute to low premature death cases in their Nordic neighbours (≤50). The largest transboundary contribution is calculated for the Danish emissions, dominated by agriculture, non-industrial combustion and traffic, contributing to ~200 premature death cases in Sweden.
Table 6 shows the cost of air pollution on human health in each of the Nordic countries in the source country and the neighbouring Nordic countries. Among the four Nordic countries, Denmark has the largest external costs due to air pollution, followed by Sweden, Finland and Norway, respectively. Following the mortality rates, Denmark, Finland and Norway have the largest cost contribution to Sweden, while Sweden contributes largest to Denmark.

Table 6. Contribution of costs (million €) of air pollution impacts on human health in the Nordic countries.

<table>
<thead>
<tr>
<th>Source</th>
<th>Receptors</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>Denmark</td>
</tr>
<tr>
<td>Denmark</td>
<td>260</td>
</tr>
<tr>
<td>Finland</td>
<td>5</td>
</tr>
<tr>
<td>Norway</td>
<td>20</td>
</tr>
<tr>
<td>Sweden</td>
<td>35</td>
</tr>
</tbody>
</table>

Regarding the costs attributed to each of the source sectors, Fig.S1 summarizes the contributions per country. For Denmark, results suggest that non-industrial combustion and agriculture/waste management are the main sectors to be targeted to reduce the negative impacts of air pollution. In Norway, reduction of non-industrial combustion emissions alone can substantially reduce the costs of air pollution. In Finland, similar to Denmark and Norway, non-industrial combustion should be...
targeted for developing emission reduction strategies, along with the traffic emissions, which
contribute as large as the residential combustion. Finally, in Sweden, traffic and agriculture/waste
management sectors should be targeted to reduce the adverse impacts of air pollution and their
associated costs.

4. Conclusions

The sectoral contributions of land-based anthropogenic emission sources in the four Nordic
countries: Denmark, Finland, Norway and Sweden, on air pollution levels and premature mortality
in these countries and over the Arctic have been estimated using the DEHM/EVA impact
assessment system for the year 2015. The chemistry and transport model, DEHM, was run with
tagging mode in order to calculate the sectoral contributions based on 30% reductions of each
sector separately. Using the modelled surface concentrations of O₃, SO₂ and PM₂.₅, the EVA model
calculated the acute (O₃ and SO₂) and chronic (PM₂.₅) premature mortality due to exposure to these
pollutants.

Results show that the Nordic countries are responsible for 5-10% of the regional background
surface PM₂.₅ concentrations in the countries itself. The non-industrial combustion (SNAP2), which
is dominated by the residential wood combustion, is responsible for 50% to 80% of the contribution
to surface PM₂.₅ in the Nordic countries. In Denmark, Finland and Norway, non-industrial
combustion contributes largely to surface OC (by 60% - 80%). In Sweden, SNAP2 is responsible
for 43% of the contribution to surface OC, while 43% comes from industrial activities. Similar to
OC, BC is also dominated by non-industrial combustion (by 50%-65%), except for Sweden, where
25% originates from non-industrial combustion and 31% from industrial activities. The dominant
source for surface SO₂ and SO₂ in all four Nordic countries is calculated to be industrial activities.
In Norway and Sweden, around 70% of SO₂ are coming from industrial activities, while in
Denmark and Finland, industrial activities are responsible for around 30% of SO₂. Off-road traffic
is responsible for 21% of SO₂, while energy production is responsible for 50% of SO₂ in Finland.
Industrial activities are also responsible for 60% of SO₄ in Norway and Sweden and 30% in
Denmark and Finland. The dominant source for NO₂ is calculated to mobile sources, and the share
between on-road and off-road traffic varies depending on the country. Almost 35% of NO₂ comes
from on-road traffic in all four Nordic countries while off-road traffic contributes by 25% to 35%.

Norway has the largest contribution to aerosol levels over the Arctic, while Denmark has the lowest
contribution, although contributions are only a few percent. Non-industrial combustion in the
Nordic countries is also the largest contributor to Arctic OC and BC levels, except for Sweden,
where industry plays a more important role in relation to the Arctic levels. Agriculture and waste
treatment facilities over the Nordic countries are responsible contribute to the sulfate levels over the
Arctic.

Anthropogenic emissions lead to a titration of around 4-5%, particularly over the source countries
and lead to a very small surface O₃ increase (>1%) in the downwind regions. The largest impacts
are calculated to be around the capital regions. Danish emissions also impact the surface PM₂.₅
levels over the southern part of Sweden and Norway, by around 3%. Finnish emissions also have a
small impact, lower than 3%, on the central part of Sweden and northern parts of Norway.
Norwegian anthropogenic emissions impacts PM₂.₅ levels over Sweden by around 7% while
Swedish anthropogenic emissions contribute to PM₂.₅ levels over the southwestern parts of Finland,
by up to 5%. It should be noted that these results are calculated for a specific year, 2015, therefore
transport from one country to others can significantly vary in different years due to meteorology, in particular wind speed and direction.

The total number of premature mortality cases due to air pollution are calculated to be ~4 000 in Denmark and Sweden and ~2 000 in Finland and Norway, leading to a total cost of 7 billion Euros in the selected Nordic countries. The contributions of emission sectors to premature mortality in each of the Nordic countries vary. Danish agriculture and industrial emissions contribute similarly (by 33%) to ~400 premature mortality cases in Denmark. In Norway, non-industrial combustion, dominated by residential wood combustion, is responsible for 48% of the ~200 premature deaths in Norway. In Finland, non-industrial combustion and traffic are responsible for more than half of the ~270 premature deaths in 2015. Finally, in Sweden, traffic and waste management/agriculture are responsible for 50% of the total premature death in Sweden (~330). In Denmark, Finland and Norway, non-industrial combustion is the main sectors to be targeted to reduce the negative impacts of air pollution, while in Sweden, traffic and agriculture/waste management sectors should be targeted to reduce the adverse impacts of air pollution and their associated costs. Overall, Nordic countries contribute to low premature death cases in their Nordic neighbours (≤50). Among the four Nordic countries, Denmark has the largest external costs due to air pollution, followed by Sweden, Finland and Norway, respectively. Following the mortality rates, Denmark, Finland and Norway have the largest cost contribution to Sweden, while Sweden contributes largest to Denmark.

Overall, results from the estimates of pollution export, premature mortality and associated costs suggest that in the Nordic countries, non-industrial combustion, which is dominated by residential wood combustion, together with industry and traffic are the main sectors to be targeted for emission mitigation strategies. The contributions of emissions from Nordic countries to each other are small (≤10%), and to the Arctic (up to 2%), meaning that large reductions can be achieved only by coordinated efforts to decrease emissions in the upwind countries.

**Author Contribution**

UI and JHC conducted the model simulations. JHC and OKN worked with the emissions input. MS and RM contributed to the experimental design of the model simulations. UI, JK, CA and SL-A extracted measurement data from Denmark, Finland, Sweden and Norway, respectively. CG and JB contributed to premature mortality and cost calculations. All co-authors contributed to the manuscript.

**Acknowledgements**

This study has been conducted under the FREYA project, funded by the Nordic Council of Ministers, Climate and Air Pollution Group (grant agreement no. MST-227-00036). AU gratefully acknowledges the NordicWelfAir project funded by the NordForsk’s Nordic Programme on Health and Welfare (grant agreement no. 75007). The work has also been funded by the Academy of Finland within the project GLOROIA and by the Research Council of Norway under the project BlackArc (contract no 240921).

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