



Global modeling of fungal spores with the EMAC chemistry-climate model: uncertainties in emission parametrizations and observations

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15 **Abstract.** Primary biological aerosol particles (PBAPs) may impact human health and aerosol-cloud-climate interactions. The role of PBAPs in the earth system is associated with large uncertainties, for example of source estimates and the atmospheric lifetime. We used a chemistry-climate model to simulate PBAPs in the atmosphere including bacteria and fungal spores. Three fungal spore emission parameterizations have been evaluated against an updated set of spore counts synthesized from observations reported in the literature. The comparison indicates an
20 optimal fit for the emission parameterization proposed by [Heald and Spracklen \(2009\)](#) and adapted by [Hoose et al. \(2010\)](#) for particle sizes of 5 μm or 3 μm , although the model still overpredicts PBAP concentrations in some locations. The correlations between the spore count observations and meteorological parameters simulated by the model show a strong dependence on the leaf area index in non-urban areas and the specific humidity in urban areas. Additional evaluation was performed by comparing our combined bacteria and fungal spore simulations to a global
25 dataset of fluorescent biological aerosol particle (FBAP) concentrations. The model predicts the total sum of measured PBAP concentrations relatively well, typically within a factor of two of FBAP. Further, the modeled fungal spore results deviate from the FBAP concentrations when used as a rough proxy for spores, depending on the particle size used in the parametrization. Uncertainties related to technical aspects of the FBAP and direct-



counting spore measurements challenge the ability to further refine quantitative comparison on this scale. Additional long-term data of better quality are needed to improve emission parameterizations.

1. Introduction

Primary biological aerosol particles (PBAPs) are diverse and include bacteria, fungal spores, viruses, pollen as well
5 as fragments of these and other organisms. The growing interest in PBAP is related to the effects they may have on
agricultural crops, human health and atmospheric chemistry ([Deguillaume et al., 2008](#)). Airborne bioparticles may
influence climate by acting as ice nuclei (IN) especially in mixed phase clouds. The abundance of atmospheric IN
can influence cloud development, impacting cloud radiative properties and the location and timing of rain
formation ([Bangert et al., 2012](#); [DeMott et al., 2010](#); [French et al., 2018](#); [Prenni et al., 2007](#)).

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Fungal spores are the most abundant and the most genetically diverse PBAPs in the atmosphere ([Lacey, 1981](#)).
They are also of critical importance because many species can induce considerable economic losses, acting as plant
pathogens or triggering respiratory diseases and allergenic processes in humans ([Reinmuth-Selzle et al., 2017](#)).
Inhalation of spores in sufficient quantities causes various respiratory diseases including various infectious
15 diseases, allergic rhinitis, asthma, and other allergic reactions ([Burge and Rogers, 2000](#); [Bush and Portnoy, 2001](#)).
More than 100 species of fungal spores have been shown to contribute to respiratory disorders ([Green et al., 2005](#)).
Fungal spores have been identified and characterized using a large range of methods, including traditional methods
such as microscopic analysis and cultivation methods, and modern methods using molecular techniques ([Caruana,
2011](#); [Després et al., 2012](#); [Griffiths and Decosemo, 1994](#)). Many studies have been used to estimate the
20 concentration of fungal spores in the atmosphere. However, a standardised procedure to quantify spore
concentrations is still lacking, which leads to very heterogeneous results and hinders direct comparisons of
observations across different studies.

Estimates of the total global emissions of fungal spores emitted into the atmosphere diverge greatly across the
literature, varying from 8 Tg yr⁻¹ ([Sesartic and Dallafior, 2011](#)) to 186 Tg yr⁻¹ ([Jacobson and Streets, 2009](#)). Fungal
25 spores contribute up to ~45% of the coarse particulate matter over the tropical rainforest ([Després et al., 2012](#)) and
their number and mass concentrations are typically about 10³ -10⁴ m⁻³ and ~1 µg m⁻³, respectively ([Fröhlich-
Nowoisky et al., 2016](#)). The number and composition of airborne fungal spores depend on complex interactions
between biological and environmental factors, including the climate and local ecological systems ([Grinn-Gofron
and Bosiacka, 2015](#)). Meteorological factors influence their production, release, and transport, which is contingent
30 on geographical areas, and vary seasonally. Interactions with PBAPs can also depend on the emitting species
involved ([Elbert et al., 2007](#); [Fröhlich-Nowoisky et al., 2009](#); [Hirst, 1953](#); [Levetin and Dorsey, 2006](#); [Li and
Kendrick, 1995](#); [Oliveira et al., 2009](#)).



Global and regional models have been used to evaluate fungal spore emissions, transport and their impact on the hydrological cycle by acting as cloud condensation nuclei (CCN) and IN (Heald and Spracklen, 2009; Hoose et al., 2010; Hummel, 2018; Hummel et al., 2015; Sesartic and Dall'afior, 2011; Sesartic et al., 2013; Spracklen and Heald, 2014). These models require the representation of the emissions and particle properties influencing transport and removal from the atmosphere through suitable parameterizations. Large uncertainties in the number concentrations remain, especially due to inherent uncertainties in the emission estimates of PBAPs. Past evaluations of these emission parameterizations have only used a few point measurements and have never been extended to a global long-term set of observations obtained using different measurement techniques.

The goal of this study is to evaluate three fungal spore emission parameterizations available in the literature and compare their number concentrations, simulated with a global model, to an updated new set of observations. We use a chemistry-climate model to simulate the PBAPs present in the atmosphere including bacteria and fungal spores. We compare our simulated fungal spore concentrations to a global dataset of spore counts synthesized from the literature and the sum of modeled fungal spore and bacteria concentrations to another global dataset of fluorescent biological aerosol particle (FBAP). These measurements have been performed with real-time techniques that detect the fluorescence signal through UV excitation of fluorophores commonly present biological materials (e.g., fungal spores, bacteria, and leaf fragments). To our knowledge, this is the first evaluation of fungal spore emission parameterizations against a global collection of these newly-available observations. It constitutes an important step towards understanding the many still-remaining uncertainties and questions related to the PBAP representation in global atmospheric models.

2. Materials and methods

2.1. Model description

The global ECHAM5/MESSy Atmospheric Chemistry – Climate (EMAC) model (ECHAM version 5.3.01, MESSy version 2.5.2 (Jöckel et al., 2005; Jöckel et al., 2016) was used to simulate the emissions and transport of biological particles. PBAPs have been included in the form of fungal spores and bacteria-containing particles by using the emission parameterizations described below. Removal processes of particles simulated by the model include sedimentation, dry deposition, impaction scavenging, and nucleation scavenging by liquid, mixed-phase, and ice clouds. PBAP dry and wet depositions are treated as described for other aerosol species (see (Burrows et al., 2009; Pozzer et al., 2012; Pringle et al., 2010) and references therein). All PBAP classes are treated as having a lognormal distribution with modal-scale parameter $\sigma = 1$ and density of 1 g cm^{-3} . We assumed that all particles can become activated as cloud condensation nuclei when calculating particle removal processes by wet deposition,



as described in (Burrows et al., 2009). Fungal spores and bacteria-containing particles of different size are transported as passive tracers, i.e., their concentrations are influenced by model processes such as dry deposition and scavenging by clouds and rain, but do not interact with radiation or change cloud microphysical properties. The sedimentation and dry deposition of the particles are treated as described in (Kerkweg et al., 2006). Wet deposition of the particles is described in (Tost et al., 2006). We use the global Leaf Area Index (LAI) distribution by (Deng et al., 2006) to calculate the fungal spores emission fluxes.

For the present study, we applied EMAC in the T63L31 resolution; with a spherical truncation of T63 (corresponding to a grid of approximately $1.9^\circ \times 1.9^\circ$ in latitude and longitude, or approximately 140×210 km at middle latitudes) and 31 vertical hybrid pressure levels up to 10hPa. The model was run for five consecutive years without meteorological nudging, from the year 2000 until 2004. Monthly sea surface temperatures and sea ice concentrations from AMIP-II climatologies were used to provide boundary conditions for the atmospheric circulation. Climatological averages for PBAP distribution for the last four years of the simulation were used after a 1-year spin-up period. The EMAC model, evaluated in (Jöckel et al., 2005; Jöckel et al., 2016) and used in similar configurations, has been shown to be capable of realistic simulations of aerosol transport and deposition for the transport of African dust to Europe (Glaser et al., 2012) and radioactive aerosol particles from the Chernobyl and Fukushima accidents (Christoudias and Lelieveld, 2013; Lelieveld et al., 2012). We emphasize that the simulation results represent a climatology rather than specific weather conditions under which some PBAB samples may have been collected, hence we expect mean concentrations and distributions to be represented by the model rather than distinct measurement data.

2.2. PBAP emissions

2.2.1. Fungal spores

We compare three fungal spore emission parameterizations previously used in global and regional modeling studies. Firstly, fungal spore emission fluxes have been derived by Heald and Spracklen (2009) for fine and coarse modes from an empirically optimized scheme where emissions are linear functions of LAI (in $\text{m}^2 \text{m}^{-2}$) and the specific humidity q (in kg kg^{-1}) at the surface. In order to match their emission estimates, Hoose et al. (2010) applied the following formulation to calculate the emission flux, assuming a mean spore diameter of $5 \mu\text{m}$ (HO5 hereafter):

$$F_{HO5} = 500 \text{m}^{-2} \text{s}^{-1} \times \frac{LAI}{5} \times \frac{q}{1.5 \times 10^{-2} \text{kg kg}^{-1}}$$



Another formulation of this parameterization has been rescaled in this work in order to match the emission estimates calculated by [Heald and Spracklen \(2009\)](#) assuming a spore diameter of 3 μm (HO3 hereafter):

$$F_{HO3} = 2315 \text{m}^{-2} \text{s}^{-1} \times \frac{LAI}{5} \times \frac{q}{1.5 \times 10^{-2} \text{kgkg}^{-1}}$$

The second parameterization we tested uses the emission number fluxes of fungal spores calculated by [Sesartic and Dallafior \(2011\)](#) (SD hereafter) for the five different ecosystems defined by [Olson et al. \(2001\)](#). We used the best-estimate number fluxes weighted by the area fraction of the respective MODIS ecosystems in the gridbox E_i . We lumped the categories defined in the MODIS classifications to match similar sets of lumped ecosystems used by [Sesartic and Dallafior \(2011\)](#) (i.e., derived from the Olson ecosystem types). The total emission flux for fungal spores is calculated assuming a mean diameter of 3 μm given as:

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$$F_{SD} = 194 \text{m}^{-2} \text{s}^{-1} \times E_{tropicalforest} + 214 \text{m}^{-2} \text{s}^{-1} \times E_{forest} + 1203 \text{m}^{-2} \text{s}^{-1} \times E_{shrub} + 165 \text{m}^{-2} \text{s}^{-1} \times E_{grassland} + 2509 \text{m}^{-2} \text{s}^{-1} \times E_{crop}$$

The third parameterization, derived by [Hummel et al. \(2015\)](#) (HU hereafter), is adapted to measurements of airborne fluorescent biological particles across northern Europe. Similar to the parameterization of [Heald and Spracklen \(2009\)](#), this recent parameterization depends on LAI and specific humidity, and is extended to include the surface temperature T (K):

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$$F_{HU} = b_1 \times (T - 275.82) + b_2 \times q \times LAI$$

where $b_1 = 20.426$ and $b_2 = 3.93 \cdot 10^4$.

For each parameterization, the mean diameter was assigned according to the recommendations made in the three studies: 5 μm for HO5, 3 μm for HU, and 3 μm for SD, and we used 3 μm for HO3.

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2.2.2. Bacteria

While the focus of this study is on the simulation of fungal spore concentrations, bacteria-containing particles are expected to contribute significantly to FBAP concentrations. In order to examine their possible contribution to FBAPs (discussed in more detail in the section 3.4), we also simulated bacteria-containing particles (see Supplementary Information), using an emission parameterization previously developed for EMAC on the basis of observed atmospheric concentrations of bacteria-containing particles in different ecosystems ([Burrows et al., 2009a](#); [Burrows et al., 2009b](#)). The uncertainties in this parameterization and the underlying observations have

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been discussed in detail in ([Burrows et al., 2009a](#); [Burrows et al., 2013](#)). Further evaluation of the simulated bacteria-containing particle concentrations is beyond the scope of this study; here we include these concentrations to enable a qualitative comparison with observed FBAP concentrations.

2.3. Data description

5 2.3.1. Spores counts

We compare the fungal spore concentrations calculated by EMAC using the three emission parameterizations described above, to observations collected through literature review (see their geographical locations in Figure 1). [Sesartic and Dallafior \(2011\)](#) have reviewed more than 150 studies and found that only a relatively small number
10 (35) reported total fungal spore concentration measurements, excluding observations that employed cultivation of a subset of species (e.g., in a petri dish) and measurements that report only mass concentrations instead of spore counts. We updated this dataset with observations that meet the same criteria, mostly from studies published since 2011. Our updated review revealed that much of the relevant literature reports only concentrations of the genetic
15 diversity of fungal taxa and not their total concentrations, hence explaining the scarcity of data that can be used for model evaluations. The uncertainties related to these methods and discussed in detail in [Sesartic and Dallafior \(2011\)](#) include heterogeneous measurement methods, which differ in the devices and methods used for spore trapping, choice of impaction media, choice of airflow, period of measurements, placement of the sampling device, choice of the impaction media, choice of nutrient medium and incubation duration and temperature (for viable counts), and counting method. The observations used for comparison with model results are listed in Table S12.
20 Modeled concentrations have been sampled to match the period of observation for each location. Since we do not compute actual meteorology but rather climatological conditions, our model results do not represent instantaneous local processes, especially when they vary strongly on a small scale. However, we expect that time averaging limits such biases. We differentiate the data by ecosystem using the MODIS categories and the description provided by the reference. Most of the observations that met our criteria for inclusion have been taken in urban areas.

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2.3.2. FBAP observations

Over the last two decades real-time measurement techniques have provided opportunities to monitor airborne PBAP continuously at relatively high time-resolution. Techniques involving laser/light-induced fluorescence (LIF)
30 have been particularly effective in rapidly providing information about PBAP in real time e.g. ([Fennelly et al., 2018](#); [Huffman and Santarpia, 2017](#); [Kaye et al., 2005](#); [Pan et al., 2009](#); [Saari et al., 2014](#); [Sivaprakasam et al., 2009](#)). Among many available instruments, two commercially available LIF biosensors have been widely applied



to ambient bioparticle monitoring and have helped to reveal fine detail about atmospheric PBAP patterns not previously observed e.g. (Gabey et al., 2010; Huffman et al., 2013; Huffman et al., 2010; Perring et al., 2015; Schumacher et al., 2013). For example, the ultraviolet aerodynamic particle sizer (UV-APS; TSI, Inc.) and the wideband integrated bioaerosol sensor (WIBS; University of Hertfordshire or Droplet Measurement Technologies) both characterize biological particles in real-time based on the intensity of fluorescence emission observed from individual particles after pulsed excitation at wavelengths characteristic for common biofluorophores (Foot et al., 2008; Hairston et al., 1997; Pöhlker et al., 2012). Despite the uncertainties related to this type of measurement (Huffman et al., 2012; Pöhlker et al., 2012; Savage et al., 2017), the FBAPs detected by the UV-APS or WIBS have been successfully used in some cases as a lower-limit for the atmospheric abundance of PBAPs in the super-micron ($> 1 \mu\text{m}$) size range (Huffman et al., 2010).

In this context, however, it is important to mention a few caveats implicit with the assumption linking FBAPs to PBAPs. First, real-time LIF instruments can only detect the physical properties of particles (i.e., fluorescence and size) and cannot directly determine whether a particle is of biological origin. By applying certain analytical strategies, however, a given ensemble of particles may be assigned as PBAP with varying degrees of certainty. In some cases weakly fluorescing biological particles can escape the LIF detection e.g. (Huffman et al., 2012) and in other cases highly fluorescent particles of non-biological origin such as certain kinds of aged brown secondary organic aerosols, diesel soot particles, and some humic-like substances (HULIS) can interfere with the LIF detection leading to an overestimation of PBAPs e.g. (Gabey et al., 2013; Huffman et al., 2010; Saari et al., 2013; Savage et al., 2017). Bioparticle size also plays an important role in LIF detection. For example, viruses are generally too small to be detected by LIF instruments, and almost all species of pollen are too large to be detected without fragmentation or instrument modification (O'Connor et al., 2011). Additionally, technical differences in instrument design, the choice of detection channels, and operational parameters can have significant effects on the reported number concentration of FBAPs and the quality of their correlation with PBAP classes (Savage et al., 2017). Nevertheless, we use the FBAP numbers reported by UV-APS and WIBS instruments as a rough proxy for PBAP, comparing the observed FBAP numbers both with more direct PBAP measurements and with model outputs.

FBAP observations using UV-APS and the FL3 channel from the WIBS-3 and WIBS-4A instruments are listed in Table 3. The majority of FBAP data were extracted from published reports without additional analysis or as tabulated by previous reviews (Fennelly et al., 2018; Saari et al., 2015; Yu et al., 2016), and all original data sources are specified in Table 3. By limiting WIBS data to only the FL3 channel, the number concentrations are expected to be significantly lower than total FBAP numbers often reported, but are used here for a better correlation with the UV-APS, due to similarities in fluorescence excitation and emission bands (Foot et al., 2008;



Pöhlker et al., 2012; Savage et al., 2017). For all LIF data, mean FBAP number concentrations were integrated from either 0.8 μm to 15 μm or 1.0 μm to 20 μm and are reported for each of the 17 geographic locations (see Fig. 5) in Table 3 for a comparison with our simulated mean number concentrations of fungal spores and bacteria. The FBAP observations show a peak in the number distribution at about 1 – 4 μm , irrespective of location or instrument. It has been previously suggested for several geographic locations that the UV-APS and WIBS FL3 channel may yield lower limit proxies for fungal spores, due in part to the large number concentration of spores in this size range compared to other biological particles (Gosselin et al., 2016; Healy et al., 2014). The sizes of modeled bacteria-containing and fungal spores particles considered in this study are consistent with this observed size range of peak FBAP concentrations (1 – 4 μm), enabling a more accurate comparison.

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3. Results

3.1. Concentrations of modeled fungal spores

15 The total global emissions calculated here with HO5 and HO3, 17 Tg yr⁻¹ (see Table 1) corresponding to an average mass emission flux of 2.5 ng m⁻² s⁻¹, are within the range of uncertainties reported by Després et al. (2012) and Fröhlich-Nowoisky et al. (2016). Using the same HO5 parameterization, Heald and Spracklen (2009) and Hoose et al. (2010) calculated higher totals, 28 Tg yr⁻¹ and 31 Tg yr⁻¹, respectively. Nevertheless, calculations by Heald and Spracklen (2009) include not only the coarse mode (including a diameter of 5 μm , similar to HO5) but
20 also a fine mode for particles of a diameter lower than 2.5 μm . This demonstrates the model sensitivity to the leaf area index dataset used in each study and the specific humidity calculated by each model.

The total global emissions calculated using SD and HU are about 48 Tg yr⁻¹ and 77 Tg yr⁻¹, respectively. These stronger emissions lead to higher mean mass concentrations (~1.1 $\mu\text{g m}^{-3}$ for both the SD and HU parameterizations) as compared with HO3 and HO5 (see Table 1); and the characteristic magnitude reported in the literature.

25

Figures 2 and 3 show the spatial distribution of the annual mean fungal spore emission fluxes and number concentrations at the surface as simulated by all parameterizations. HO5 and HO3 show similar distributions but with higher number magnitude for HO3 because of the difference of the scaling mass factor due to the particle size. Whereas HO5/HO3 and HU result in highest values over the Amazon and tropical Africa, SD does not yield
30 maximum values in these regions and simulates much larger emission. Further, the comparison of fungal spore number fluxes calculated by Sesartic and Dallafior (2011) and by EMAC yields large discrepancies in magnitude and spatial distribution. This is most likely explained by large differences in the biome spatial distribution between MODIS and Olson data, leading to the higher emissions calculated by EMAC when using the SD parameterization. Interestingly, the spatial distributions of HU emission fluxes and concentrations are not restricted to land as for the



other parameterizations and have been extended to all ecosystems. This leads to the highest totals, as indicated in Table 1. Note that this parameterization has been optimized for a regional modeling study over northern Europe, and therefore is probably not suitable for use globally and for all ecosystems.

5 3.2. Comparison of the EMAC fungal spore concentrations with direct measurements

Our simulated, globally distributed, annual mean number concentrations of fungal spores obtained using the three emission parameterizations are compared to the set of spore counts at various locations in Figure 4. Observational data are differentiated by biome as defined by MODIS data, including the urban ecosystem. The concentrations are overestimated by all three parameterizations, but least by HO5. HO3 agrees better with observed spore counts with a correlation coefficient (R) of 0.24 and a median model-to-observation ratio of 6 for HO5, compared to 10 for HO3 and more than 100 for the other two parameterizations. Two additional metrics of the level of agreement between model results and observations are presented in Figure 4: the modified normalized mean bias (MNMB), a measure of bias that is symmetrical with respect to over-estimates and underestimates (ranging between -2 to 2 and equal to 0 for a “perfect” model), and the fractional gross error (FGE), a measure of the relative model error (ranging from 0 for a “perfect model” to a maximum value of 2), which behaves symmetrically with respect to under- and overestimation, without overemphasizing outliers (Huijnen and Eskes, 2012).

From the three parameterizations, HO5 performs best on all three scores. HO3 shows a similar correlation coefficient but higher MNMB and FGE than HO5. SD and HU show similar MNMB and FGE, but HU correlates slightly negatively with the observations. Although we compare local measurements limited in their representativeness in time and space to the relatively coarse grid size (approximately 140 km) of the climatological model data, only the comparison with HO5 can be rated as satisfactory. Surprisingly, the SD derived fungal concentrations are the least comparable to observations despite the fact that the formulation of the emission parameterization is based partly on the same observations. This might be due to the change of the global ecosystem distribution, as we used the MODIS classification on which the HO5 parameterization was based, instead of the Olson distribution on which the SD parameterization was based. The HU emission parameterization might not be suited for use in global modeling studies since it was optimized for a regional modeling study over northern Europe. Differences in model physics, including the parameterizations of turbulent transport, precipitation, wet and dry removal, can also produce different concentrations, given the same emissions, so these results cannot necessarily be extended to other atmospheric models.

Discrepancies between model results and observations may be explained by an overprediction of fungal spore sources via biases in the emission parameterization or long-range transport, or an underprediction of the rate of



removal by dry and wet deposition. Additionally, as outlined by [Sesartic and Dallafior \(2011\)](#) and references therein, the observational data quality is limited and should be considered with caution. The methods used to measure actual spore concentrations may involve biases, e.g. problems related to the identification of fungal spores. As mentioned in section 2.3.1, ([Sesartic and Dallafior, 2011](#)) showed that many direct-counting spore techniques
5 can significantly undercount spore number (i.e. by orders of magnitude). Additionally, any culture-based methods have significant biases in that only a very small fraction of spore species can be culturable in a given medium.

Since the HO simulations agree most closely with observations, we will show results only from this simulation hereafter.

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3.3. Correlation between direct measurements and meteorological parameters

Meteorological variables affect the initial release of fungal spores into the atmosphere and their dispersal once airborne. Temperature and humidity affect the size of the source and control the release of most actively released
15 fungal spores, where some species emit spores primarily during cool and wet conditions and other emit during warm and dry conditions ([Jones and Harrison, 2004](#)). The spore concentration and frequency of emission are equally dependent on geographical characteristics. Further, rainfall has been shown to correlate with fungal spore and other PBAP emission properties, leading to substantial local increases in short bursts during and following rain ([Joung et al., 2017](#); [Rathnayake et al., 2017](#)). Since the publications collected for this study do not always provide
20 information on the meteorological parameters of the observational site, we investigate the effects of physical parameters such as temperature, specific humidity, and leaf area index, as modeled by EMAC, on the observed particle concentrations, taking into account the differentiation between ecosystems as defined by MODIS data. Interestingly, we find a strong difference between non-urban and urban observations through their correlations with the three parameters. Table 2 shows a correlation coefficient of 0.41 between the observations and specific
25 humidity for the urban points and the same correlation between the non-urban observations and the leaf area index. This demonstrates the stronger effect of meteorological variables on urban sites, which could be taken into account in the formulation of a more advanced fungal spore emission parameterization. Additional observational evidence is needed to support this hypothesis.

30 3.4. Comparison of the EMAC simulated fungal spores and bacteria with FBAP observations

Figure 5 shows a comparison between the observed FBAP number concentrations and the modeled concentrations of bacteria and fungal spores using HO5 in Figure 5A and HO3 in Figure 5B. The comparison excludes data from Nanjing, China ($\sim 2 \times 10^6 \text{ m}^{-3}$), because the observed particle counts were approximately two orders of magnitude



higher than concentrations observed elsewhere, likely as a result of urban pollution influence. For the locations Saclay and Killaney, both UV-APS and WIBS observations are available, though, for the sake of clarity, only WIBS data are shown in Figure 5.

5 In general, the HO5 and HO3 model results compare relatively well with the observations of total fluorescent particles in almost all locations. The HO5 model concentrations are lower than FBAP concentrations in ca. 40% of the observations and are within a factor of two (high or low) in all cases. This model underprediction is largest in the summer season. The HO3 model concentrations, alternatively, are higher by a factor of two to three than FBAP concentrations in almost 95% of the locations. This is due to the higher number concentrations simulated using
10 HO3 than HO5, illustrated in Figure 5B. The model predicts higher bacteria loadings than fungal spores in almost all cases, with the only exceptions of Ucluelet (rural coastal) and the Amazon (near-pristine tropical forest) using HO5 and HO3. Additionally, the simulated bacteria number concentrations in Figure 5 dominate the total PBAP concentrations for the locations where observations are available in winter, when very low concentrations of fungal spores are expected. A possible explanation for this discrepancy is that bacteria emissions are assumed constant in
15 time, representing inferred “background” emissions, while they may exhibit seasonal variations not taken into account in our model. Consequently, the differences in the seasonal bacteria concentrations for each location are related to transport and deposition patterns reproduced in the model. Fungal spore emissions, in contrast, are assumed to include seasonal variability due to the seasonality of the leaf area index and the specific humidity, which are used as model inputs. Many reports support the observed seasonal cycle of fungal spore concentrations,
20 which are typically highest in summer and early fall, but depends on latitude and ecosystem ([Lacey, 1996](#); [Lang-Yona et al., 2012](#); [Manninen et al., 2014](#); [Schumacher et al., 2013](#)).

Laboratory-based observations have shown that the WIBS FL3 channel utilized here is less efficient at detecting bacteria ([Hernandez et al., 2016](#); [Savage et al., 2017](#)) and thus is likely to detect fungal spores or pollen with
25 relatively higher efficiency. The large size of pollen grains limits their ability to be detected by the WIBS, however, and so fungal spores are assumed to represent the largest fraction of biological particles detected by the WIBS FL3 channel. Moreover, evidence in certain campaigns suggests that these fluorescence channels correlate well with fungal spores (e.g., ([Fernandez-Rodriguez et al., 2018](#); [Healy et al., 2014](#); [Huffman et al., 2012](#))). Other important technical considerations were discussed in Section 2.3.2. In particular, note that our comparison here including the
30 bacteria-containing particle concentrations is only qualitative, as their contribution to the total FBAP observations as utilized here is unclear. We include bacteria-containing particles in our comparison in order to emphasize the uncertainty they contribute to any direct comparison between FBAP and fungal spores.



Therefore, the differentiation between HO5 and HO3 fungal spore concentrations is relevant in the comparison. If we consider that the FBAPs observed in these locations mainly represent fungal spores, HO5 shows an underestimation compared to FBAP at all stations, while HO3 shows an underestimation at only 50% of the stations. In both cases, the simulated fungal spores occur at lowest concentrations during the winter season, and in the case of HO3 concentrations are also low in spring and fall. For HO3, the concentrations are overestimated, especially in the summer when they are a factor of two higher than observations. This contradicts the results presented in Section 3.2, which shows a modeled overestimation (by a factor of 6 and 10) with respect to spore counts measured by optical microscopy (Figure 4). As mentioned, the comparison of modeled fungal spores with direct counts of spores may also be biased because of model inputs and the frequently observed undercounting of spore by some collection and detection methods. FBAP and spore counts measured via optical microscopy were compared for two sets of new measurements shown here (Saclay and Cyprus). The observed values from each method are reported here with the relative factor that the FBAP concentration overcounts the spore counts from the optical technique shown in parentheses: Saclay - $0.048 \times 10^6 \text{ m}^{-3}$ (spore count), $0.088 \times 10^6 \text{ m}^{-3}$ (WIBS-4A; x1.8), and $0.027 \times 10^6 \text{ m}^{-3}$ (UV-APS; x0.6); Cyprus - $0.0015 \times 10^6 \text{ m}^{-3}$ (spore count), and $0.0433 \times 10^6 \text{ m}^{-3}$ (WIBS-4A; x29). In several other cases, collocated measurements of FBAP and spore counts also show spore count to be lower than FBAP by a factor of ~ 2 (Fernandez-Rodriguez et al., 2018) to ~ 10 (Healy et al., 2014; Huffman et al., 2012), again depending partially on instrumental parameters and differences in observed bioaerosols. Quantitative differences are a combination of undercounting of spores by direct measurements and counting of non-spore particles (bacteria and some non-biological particles) by FBAP measurements. Additional direct comparison of FBAP concentrations with spore count concentrations summarized in this study (Table SII and Fig. 1) is not possible because of differences in locations and seasons. Therefore, the qualitative relationship between UV-LIF and other spore counting techniques has been demonstrated, but quantitative comparisons often show significant differences.

Significant uncertainties also remain in the interpretation of the UV-LIF measurements. For example, UV-LIF measurements do not detect all spore types equally. (Hernandez et al., 2016; Savage et al., 2017) show differences in fluorescence profiles of a number of spore types and discuss that different instrument units detect particles with different detection efficiencies. This implies that both biological and instrumental factors can lead to differences in observed FBAP concentrations. (Healy et al., 2014; Huffman et al., 2012) both discuss how certain types of spores may escape detection. In particular, (Fernandez-Rodriguez et al., 2018; Healy et al., 2014) discuss how the genus *Cladosporium* (among the most commonly observed spore type in many environments) correlated very poorly with fluorescent measurements, suggesting that dark-walled cell walls present in this type of spores may inhibit some types of real-time fluorescence detection. This spore type is a dominant spore type during dry weather, therefore it might be undercounted during the day and in certain locations where such spore types are a high fraction of the



spore number. In some cases the FL3 signal is also influenced by non-biological particles (Savage et al., 2017), and so FBAP number concentrations from WIBS and UV-APS presented here should be used only for a rough comparison.

5 4. Discussion and Conclusions

We used the chemistry-climate model EMAC to simulate the total PBAPs present in the atmosphere including bacteria, but with a focus on fungal spores. We compared the simulated fungal spore number concentrations, simulated using three emission parameterizations, to an updated data set of spore counts synthesized from the literature. Additional evaluation of the modeled simulated PBAP concentrations was performed using a comparison with a global data set of FBAP concentrations.

Among the four fungal spore emission schemes examined here, the HO5 parameterization using a fixed spore diameter of 5 μm showed the best fit to observations, although the model overpredicts the concentrations by up to a factor of 6 in some locations. Using the same parameterization (after rescaling), HO3 with a diameter of 3 μm showed a larger overprediction. Spore count observations are limited in time and space and are subject to several methodological issues. Therefore, these direct-count measurements cannot provide a rigorous evaluation of the model results. Bacteria and fungal spore concentrations predicted by the model agree well with FBAP observations from real-time measurement techniques. The overprediction/underprediction is estimated to be of a factor of 2 or less for all measurement locations. Modeled bacteria concentrations exceed fungal spores in most locations, particularly in winter, while simulated fungal spores might be underestimated/overestimated in near-pristine environments, depending on the diameter choice of the parametrization. The FBAP concentrations used here (from UV-APS data and WIBS-FL3 channel) are likely to underestimate fungal spores somewhat, and dramatically underrate the abundance of bacteria in most locations. Most bacteria are not strongly fluorescent in the applied wavelength channels and are therefore underestimated. This might explain the differences when we expect high bacterial counts, e.g. (Savage et al., 2017).

In general, the HO3 and HO5 parameterizations perform better than the HU and SD parameterizations, confirming that process-based parametrizations accomplish the best results. The differences between HO5 and HO3 emphasize the importance of the choice of the particle sizes in comparisons with observations, expressed in number concentrations and not in mass concentrations. This should be taken into account in the further development of emission parameterizations for fungal spores. New parametrizations should consider a range of particle sizes, as the size and mass are usually not measured in observations of atmospheric fungal spores, and remain highly uncertain.



Our calculated correlations between observed spore counts and meteorological parameters confirm the seasonal leaf area index as the main factor driving fungal spore emissions in non-urban areas, while the specific humidity controls the emissions in urban areas. Additional measurements of the meteorological parameters during periods of observations are needed to confirm this hypothesis. This might also contribute to a differentiation between urban
5 and non-urban areas in the emission fluxes, to be considered by models in the future.

In conclusion, fungal spore concentrations are underestimated by the EMAC model in most locations with respect to the FBAP data used, but overestimated with respect to direct count observations of spores. The bacteria emission scheme we applied does not represent seasonal nor diurnal variability, a weakness that needs to be addressed in
10 future work. Further comparisons of modeled results with long-term FBAP measurements, taking into account daily and seasonal variability, might offer new opportunities to better constraint our model emission parameterizations. Improvements in the way FBAP data are analyzed will also allow for better separation of bacteria and fungal spores and will thus allow for better comparison with improved models in the future ([Ruske et al., 2017](#)).

15



Author Contribution

MT and SMB designed the research study and AP, JAH, SB, MT carried it out. MT and SB did the model simulations. AP contributed to the model data processing. KMP, NJB and RSE and JAH provided the FBAP data.
5 JL discussed the results and contributed to the final manuscript.

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	Emission (Tg/yr)	Burden (Gg)	Lifetime (days)	Number concentrations (m ⁻³)	Mass concentrations (µg/m ⁻³)
Fungal spores HO (D=5µm)	17.58	85	1.2	2620	0.17
Fungal spores HO (D=3µm)	17.58	82	1.7	15824	0.22
Fungal spores HU (D=3µm)	77.47	423	2	81755	1.15
Fungal spores SD (D=3µm)	48.09	223	1.6	77994	1.1
Literature estimates	8-186 ^a	94 ^b , 186 ^c	1.1 ^b , 2.3 ^c	~10 ³ - 10 ⁴ ^a	~0.1 - 1 ^a

Literature values are from ^aDesprés et al. (2012), ^bHoose et al. (2010), ^cHeald and Spracklen (2009).

5 Table 1: Fungal spore emissions, burdens, lifetimes, and number and mass concentrations in near surface air.

	Specific humidity	Temperature	Leaf Area Index
Urban Observations	0.41	0.26	-0.07
Non-Urban Observations	0.02	0.15	0.41

Table 2: Correlation between observations and the modeled specific humidity, temperature, and leaf area index. The values in bold are statistically significant ($p < 0.05$).



Location	LON	LAT	ALT	Instrument	Start	End	Season	Site category	Publication	Size Range (μm)	Mean (10^6m^{-3})
Amazon, Brazil	-60.2	-2.58	110	UV-APS	03.02.2008	15.03.2008	Winter	rainforest	Huffman et al., 2012)	>1	0.093
Karlsruhe, Germany	8.42	49.09	117	WIBS-4A	01.04.2010	01.04.2011	Yearly	semi-rural	(Toprak and Schnaiter, 2013)	0.8 -16	0.031
Karlsruhe, Germany	8.42	49.09	117	WIBS-4A	01.04.2010	01.07.2010	Spring	semi-rural	(Toprak and Schnaiter, 2013)	0.8 -16	0.029
Karlsruhe, Germany	8.42	49.09	117	WIBS-4A	01.07.2010	01.10.2010	Summer	semi-rural	(Toprak and Schnaiter, 2013)	0.8 -16	0.046
Karlsruhe, Germany	8.42	49.09	117	WIBS-4A	01.10.2010	01.01.2011	Autumn	semi-rural	(Toprak and Schnaiter, 2013)	0.8 -16	0.029
Karlsruhe, Germany	8.42	49.09	117	WIBS-4A	01.01.2011	01.04.2011	Winter	semi-rural	(Toprak and Schnaiter, 2013)	0.8 -16	0.019
Colorado, USA	-105.15	39.16	2290	UV-APS	20.07.2011	22.08.2012	Yearly	rural	Schumacher et al., 2013)	>1	
Colorado, USA	-105.15	39.16	2290	UV-APS	01.03	31.05	Spring	rural	(Schumacher et al., 2013)	>1	0.015
Colorado, USA	-105.15	39.16	2290	UV-APS	01.06	31.08	Summer	rural	(Schumacher et al., 2013)	>1	0.03
Colorado, USA	-105.15	39.16	2290	UV-APS	01.09	30.11	Autumn	rural	(Schumacher et al., 2013)	>1	0.017
Colorado, USA	-105.15	39.16	2290	UV-APS	01.12	29.02	Winter	rural	(Schumacher et al., 2013)	>1	0.0053
Hyttälä, Finland	24.17	61.85	181	UV-APS	27.08.2009	17.04.2011	Yearly	boreal forest	(Schumacher et al., 2013))	>1	
Hyttälä, Finland	24.17	61.85	181	UV-APS	01.03	31.05	Spring	boreal forest	(Schumacher et al., 2013)	>1	0.015
Hyttälä, Finland	24.17	61.85	181	UV-APS	01.06	31.08	Summer	boreal forest	(Schumacher et al., 2013)	>1	0.046
Hyttälä, Finland	24.17	61.85	181	UV-APS	01.09	30.11	Autumn	boreal forest	(Schumacher et al., 2013)	>1	0.027
Hyttälä, Finland	24.17	61.85	181	UV-APS	01.12	29.02	Winter	boreal forest	(Schumacher et al., 2013)	>1	0.004
Killaney, Ireland	-9.5	52.05	34	WIBS-4A	02.08.2010	02.09.2010	Summer	rural	(Healy et al., 2014)	>1	0.035
Killaney, Ireland	-9.5	52.05	34	UV-APS	02.08.2010	02.09.2010	Summer	rural	(Healy et al., 2014)	>1	0.015
Mainz, Germany	8.23	49.98	100	UV-APS	03.08.2006	04.12.2006	Autumn	semi-urban	Huffman et al., 2010)	>1	0.027
Borneo, Indonesia	117.84	4.98		WIBS-3	01.06.2008	31.07.2008	Summer	forest	(Gabey et al., 2010)	0.8-20	0.2
Nanjing, China	118.95	32.12		WIBS-4A	29.10.2013	15.11.2013	Autumn	urban	Yu et al., 2013)	>1	2.09
Puy de dôme, France	2.96	45.43	1465	WIBS-3	22.06.2010	03.07.2010	Summer	mountain	(Gabey et al., 2013)	>1	0.095
Manchester, UK	-2.25	53.48		WIBS-3	04.12.2009	21.12.2009	Winter	urban	(Gabey et al., 2011)	0.8-20	0.11
Helsinki, Finland	24.65	60.2		UV-APS	02.02.2012	25.02.2012	Winter	suburban	(Saari et al., 2015)	<1	NA
Helsinki, Finland	24.65	60.2		UV-APS	16.06.2012	22.08.2012	Summer	urban	(Saari et al., 2015)	>1	0.013
Helsinki, Finland	24.65	60.2		BIO-SCOUT	02.02.2012	25.02.2012	Winter	suburban	(Saari et al., 2015)	>1	0.01
Helsinki, Finland	24.65	60.2		BIO-SCOUT	16.06.2012	22.08.2012	Summer	urban	(Saari et al., 2015)	>1	0.028
Munnar, India	77.06	10.09	1605	UV-APS	01.06.2014	20.08.2014	Summer	rural	(Valsan et al.,	>1	0.017



									2016)		
Ucluelet, Canada	-125.54	48.92	5	WIBS-4A	06.08.2013	28.08.2013	Summer	rural/marine	unpublished	>1	0.059
Paris, France	2.35	48.85		UV-APS	14.01.2010	15.02.2010	Winter	urban	unpublished	>0.8	0.0276
Saclay, France	2.17	48.71		UV-APS	16.06.2014	05.08.2014	Summer	semi-urban	unpublished	>0.8	0.027
Saclay, France	2.17	48.71		WIBS-4A	16.06.2014	05.08.2014	Summer	semi-urban	unpublished	>0.8	0.088
Cyprus	33.06	35.03	550	WIBS-4A	01.04.2016	26.04.2016	Spring	rural/mountain	unpublished	>0.8	0.0433
Barbados	-59.43	13.16	5	WIBS-4A	16.07.2016	04.09.2016	Summer	rural/marine	unpublished	>0.8	0.0951

Table 3: List of the FBAP observations.

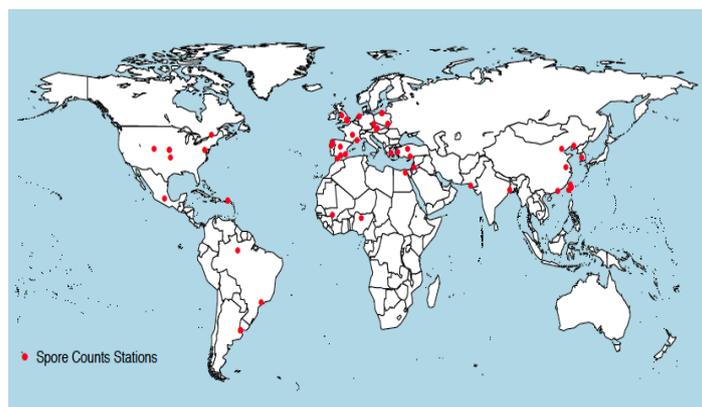


Figure 1: Geographical locations of the fungal spore counts listed in Table SI2 and used in Fig. 4

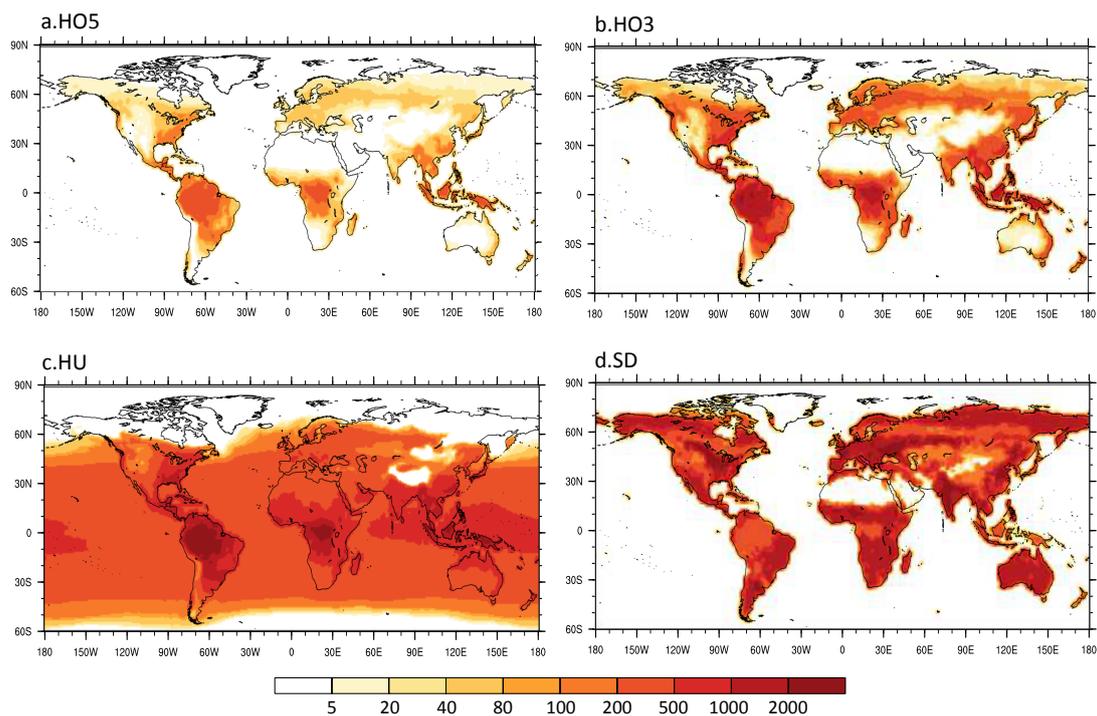


Figure 2: Modeled annual mean emission fluxes of fungal spores (in $\text{m}^{-2} \text{s}^{-1}$).

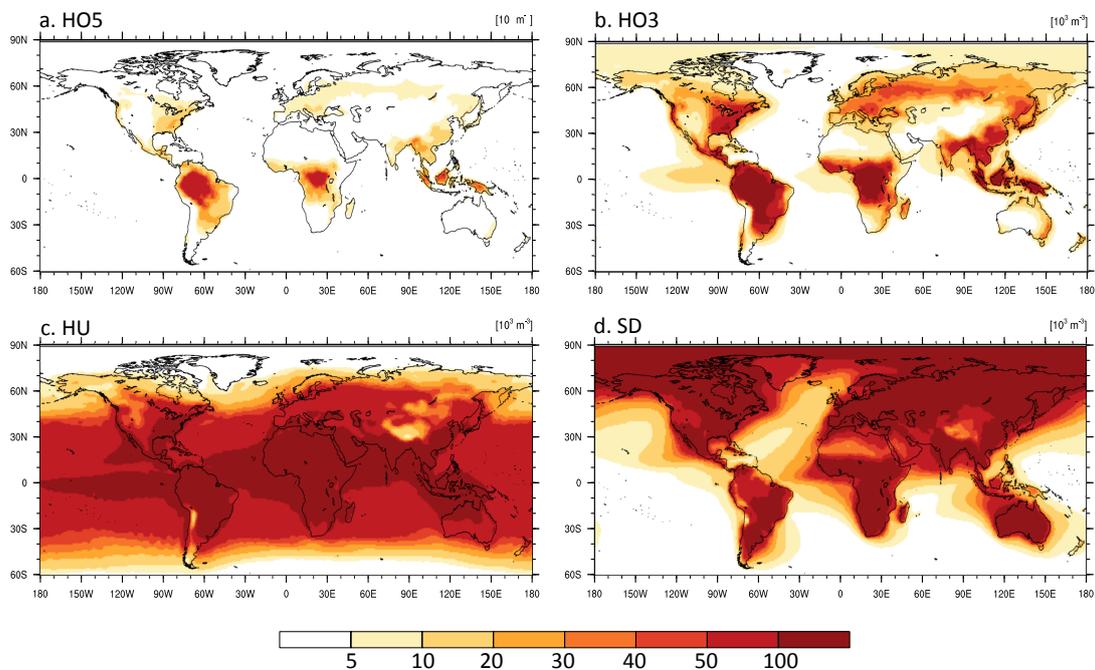


Figure 3: Modeled near-surface annual mean number concentration of fungal spores (in 10^3 m^{-3}).

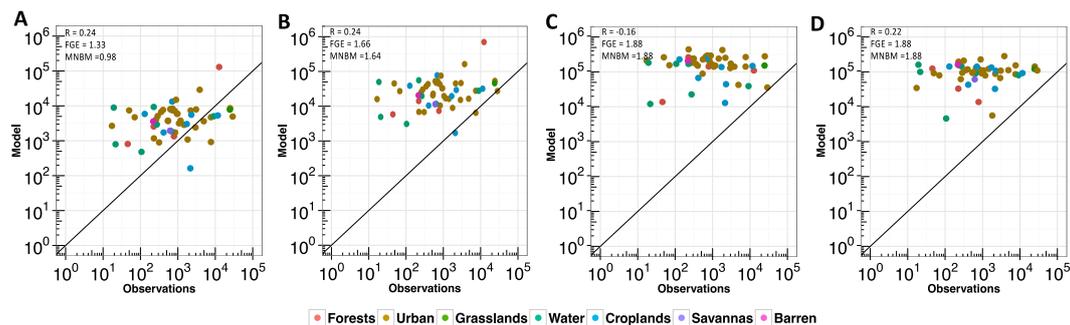
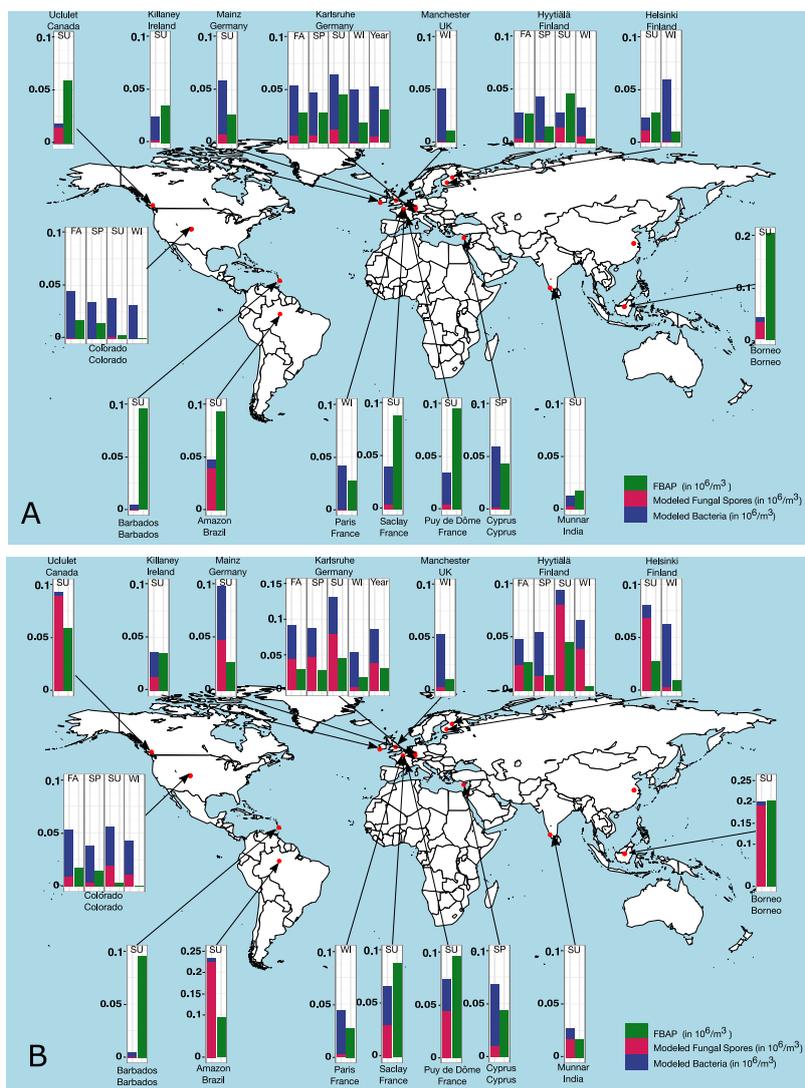


Figure 4: Comparisons between fungal spore number concentrations (in m^{-3}) observed and simulated by EMAC using the three emission parameterizations (including the two versions of (Hoose et al., 2010) HO5 and HO3). A. HO5, B. HO3, C. HU and D. SD. Colored points depict the ecosystems of the observational stations as defined by MODIS.

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5 Figure 5: Comparisons between observed FBAP and modeled bacteria and fungal spores sampled for the campaigns described in Table 3. Units: $10^6 m^{-3}$. Observations are seasonal, Fall (FA), Spring (SP), and Winter (WI), and yearly (Year)



A. Fungal spores are calculated using HO5. Data from Borneo are plotted on a different scale because of the high particle count observations and data from Nanjing, China were removed due to large urban influence. B. Fungal spores are calculated using HO3. Data from the Amazon, Borneo, and Karlsruhe are plotted on a different scale because of the high particle count observations and data from Nanjing, China were removed due to large urban influence.

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