Predictions of the glass transition temperature and viscosity of organic aerosols by volatility distributions

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Abstract:

Volatility and viscosity are important properties of organic aerosols (OA), affecting aerosol processes such as formation, evolution and partitioning of OA. Volatility distributions of ambient OA particles have often been measured, while viscosity measurements are scarce. We have previously developed a method to estimate the glass transition temperature \( T_g \) of an organic compound containing carbon, hydrogen, and oxygen. Based on analysis of over 2300 organic compounds including oxygenated organic compounds as well as nitrogen- and sulfur-containing organic compounds, we extend this method to include nitrogen- and sulfur-containing compounds based on elemental composition. In addition, parameterizations are developed to predict \( T_g \) as a function of volatility and the atomic oxygen-to-carbon ratio based on a negative correlation between \( T_g \) and volatility. The prediction method of \( T_g \) and viscosity is applied to ambient observations of volatility distributions at eleven field sites. The predicted \( T_g \) varies mainly from 290 K to 339 K and the predicted viscosities are consistent with the results of ambient particle phase state measurements in the southeastern US and the Amazonian rain forest. Reducing the uncertainties in measured volatility distributions would be helpful to improve predictions of viscosity especially at low relative humidity. We also predict the \( T_g \) of OA components identified via positive matrix factorization of aerosol mass spectrometer data. The predicted viscosity of oxidized OA is consistent with previously reported viscosity of SOA derived from \( \alpha \)-pinene, toluene, isoprene epoxydiol (IEPOX), and of diesel fuel. Comparison of the predicted viscosity based on the observed volatility distributions with the viscosity simulated by a chemical transport model implies that missing low volatility compounds in a global model can lead to underestimation of OA viscosity at some sites. The relation between volatility and viscosity can be applied in the molecular corridor or volatility basis set approaches to improve OA simulations in chemical transport models by consideration of effects of particle viscosity in OA formation and evolution.
1. Introduction

Organic aerosols (OA) contribute substantially to the mass loadings of atmospheric fine particulate matter (Hallquist et al., 2009; Jimenez et al., 2009). OA formed from various anthropogenic or biogenic precursors have complex physicochemical properties (Goldstein and Galbally, 2007; Nizkorodov et al., 2011; Ditto et al., 2018), which makes predictions of their role in air quality, climate and public health challenging (Kanakidou et al., 2005; Shrivastava et al., 2017). Volatility and viscosity are important properties of OA, both of which affect important aerosol processes such as gas–particle partitioning, new particle formation and evolution of size distribution, heterogeneous reactions, and cloud condensation and ice nucleation pathways of OA, as summarized in recent review articles (Krieger et al., 2012; Bilde et al., 2015; Poschl and Shiraiwa, 2015; Knopf et al., 2018; Reid et al., 2018).

Recent measurements show that OA can exist in liquid (low dynamic viscosity \(\eta; \eta < 10^2 \text{ Pa s}\)), semi-solid (\(10^2 \leq \eta \leq 10^{12} \text{ Pa s}\)), and amorphous solid (\(\eta > 10^{12} \text{ Pa s}\)) states (Reid et al., 2018). Even though there are several particle bounce measurements to infer ambient OA phase state, there are limited ambient measurements of particle phase state or viscosity depending on temperature (\(T\)), relative humidity (RH), and particle composition (Virtanen et al., 2010; O’Brien et al., 2014; Bateman et al., 2016; Pajunoja et al., 2016; Bateman et al., 2017; Liu et al., 2017; Ditto et al., 2019; Slade et al., 2019). Viscosity can be directly converted to bulk diffusivity in organic molecules using the Stokes–Einstein equation, which has been shown to work well for organic molecules diffusing through low viscous materials (Price et al., 2016; Chenyakin et al., 2017). This relation is inapplicable for predicting the bulk diffusivity of water and small molecules and it may also underestimate the diffusivity of organic molecules in a highly viscous matrix, which can be corrected using a fractional Stokes–Einstein equation (Price et al., 2016; Evoy et al., 2019).

Chemical composition of OA is complex and molecular specificity is often unavailable, which makes viscosity predictions of ambient OA challenging (Bosse,
Viscosity can be related to the glass transition temperature ($T_g$), at which phase transition between amorphous solid and semi-solid states occurs (Koop et al., 2011). Ambient temperature varies through 100 K throughout the troposphere, greatly influencing the viscosity of the mixture. When the ambient temperature is below $T_g$, an amorphous particle behaves as a solid, while a particle would be semi-solid or liquid when the ambient temperature is above $T_g$. OA particles contain a number of organic compounds and also a variable amount of liquid water depending on RH, which can act as a plasticizer to reduce $T_g$; these mixture effects can be estimated using the Gordon-Taylor relation (Mikhailov et al., 2009; Koop et al., 2011; Dette et al., 2014). In addition, ambient OA may often be internally mixed with inorganic species such as sulfate and nitrate, which would further lower $T_g$ and viscosity if they are well-mixed in one phase; when the phase separation occurs, the inorganic-rich and organic-rich phases may undergo glass transition at different temperatures (Dette and Koop, 2015).

We have recently developed a set of semi-empirical parameterizations using molar mass ($M$) and atomic O:C ratio (Shiraiwa and Li et al., 2017) or elemental composition (DeRieux and Li et al., 2018) to predict $T_g$ for compounds comprised of carbon, hydrogen, and oxygen (CHO compounds). These parameterizations have been applied to high-resolution mass spectrometry measurements to estimate viscosity of organic aerosols (DeRieux and Li et al., 2018; Schum et al., 2018; Ditto et al., 2019) and coupled into a thermodynamic model (Gervasi et al., 2019). Note that heteroatoms and the effects of molecular structure and functional groups on $T_g$ are not considered in parameterizations of Shiraiwa and Li et al. (2017) and DeRieux and Li et al. (2018).

Viscosity of pure compounds has been found to be inversely correlated with vapor pressure (Thomas et al., 1979). The molecular corridor (Shiraiwa et al., 2014; Li et al., 2016) based analysis of hundreds of SOA components has shown that compounds with lower pure compound saturation mass concentration ($C^*$) have higher $T_g$ (Shiraiwa et al., 2017). Rothfuss & Petters (2017) found that there is a similar trend between the
sensitivity of viscosity to functional group addition and the sensitivity of vapor pressure to functional group addition. Measurements of the evaporation kinetics of maleic acid showed that decreasing particle viscosity leads to a suppression in the effective vapor pressure of maleic acid (Marshall et al., 2018). Champion et al. (2019) found secondary organic aerosols (SOA) with higher condensed-phase fractions of extremely low (ELVOC) and low volatile organic compounds (LVOC) showed an increased viscosity. Zhang et al. (2019) measured $T_g$ of isoprene SOA components including isoprene hydroxy hydroperoxide (ISOPOOH), isoprene-derived epoxydiols (IEPOX), 2-methyltetrols, and 2-methyltetrol sulfates (2-MT-OS), observing a tight correlation between $T_g$ and vapor pressure.

Based on the above evidence showing a close relation between volatility and viscosity, in this study we develop the parameterizations predicting $T_g$ as a function of $C^0$ based on data from over 2000 compounds. Functional group contribution approaches are often used to predict $C^0$ (Capouet and Müller, 2006; Pankow and Asher, 2008; Compernolle et al., 2011; O’Meara et al., 2014), thereby using volatility to predict $T_g$ would include the molecular structure effect indirectly. The developed parameterizations are applied to field observations of volatility distributions to predict viscosity of ambient OA.

### 2. Parameterizations

Figure 1a shows a dependence of $T_g$ on $C^0$ of 2325 organic compounds compiled in previous studies with available measured otherwise estimated $T_g$ and $C^0$ (Koop et al., 2011; Li et al., 2016; Rothfuss and Petters, 2017; Lessmeier et al., 2018). In this study “estimated $T_g$” represents the $T_g$ estimated from the melting temperature ($T_m$) applying the Boyer-Kauzmann rule of $T_g = g \cdot T_m$ with $g = 0.7$ as validated in a previous study (Koop et al., 2011). The 2325 compounds are classified into four classes based on chemical composition: CH, CHO, CHON, and CHOS, with the number of 433, 783, 274, and 835, respectively. Detailed information of this dataset is described
in the Supplement. Figure 1a shows compounds with lower $C^0$ have higher $T_g$ and the $T_g$ appears to level at around 420 K at $C^0 < 10^{-10}$ μg m$^{-3}$. The dependence of $T_g$ on the atomic O:C ratio is weaker (Fig. 1a), in agreement with previous studies (Koop et al., 2011; Shiraiwa et al., 2017). Based on the trend shown in Fig. 1a, we develop a parameterization (Eq. 1) to predict $T_g$ as a function of $C^0$ and the atomic O:C ratio, which are the parameters used in the two-dimensional VBS (2D-VBS) framework (Donahue et al., 2011):

$$T_g = 289.10 - 16.50 \times \log_{10}(C^0) - 0.29 \times [\log_{10}(C^0)]^2 + 3.23 \times \log_{10}(C^0) \text{ (O:C)}$$ (1)

The coefficients in Eq. (1) are obtained by fitting the $T_g$ of 2325 compounds in Fig. 1a with multi-linear least squares analysis with 68% prediction and confidence intervals. Figure 1b shows that the $T_g$ of those 2325 compounds are predicted well by Eq. (1) as indicated by a high correlation coefficient ($R$) of 0.92. The average absolute value of the relative error (AAVRE, Aiken et al., 2007) is 12%. $T_g$ of individual compounds can be predicted within ±31 K as indicated by the prediction band (dotted lines in Fig. 1b); however, this uncertainty may be much smaller for multicomponent SOA mixtures under ideal mixing conditions as indicated in the confidence band (dashed lines, almost overlapping with the 1:1 line) (Shiraiwa and Li et al., 2017; DeRieux and Li et al., 2018; Song et al., 2019).

Equation (1) is also evaluated with a different set of organic compounds as a test dataset, including 654 CHO compounds and 212 CHON compounds found in SOA oxidation products (Shiraiwa et al., 2014). The values of their $T_g$ are predicted by Eq. (1) with saturation mass concentrations estimated using the EVAPORATION model (Compernolle et al., 2011). The predicted $T_g$ by Eq. (1) are compared with the $T_g$ estimated from the $T_m$ applying the Boyer-Kauzmann rule. $T_m$ values are adopted from the estimations of the Estimation Programs Interface (EPI) Suite software (version 4.1) developed by the US Environmental Protection Agency (US EPA, 2015). As shown in
Fig. 1(c), Eq. (1) also presents a good performance for predicting the $T_g$ of these SOA components with $R = 0.96$ and AAVRE = 6%. Note that Eq. (1) was developed based on the training dataset containing a large number of compounds with measured $C^0$ (Table S1 in Supplement) and aimed to be applied in the 2D-VBS framework to predict the viscosity of SOA mixtures. For pure organic compounds with known molecular structure, viscosity can be predicted by group contribution approaches (Cao et al., 1993; Bosse, 2005; Song et al., 2016b; Gervasi et al., 2019; Rovelli et al., 2019).

We also develop a parameterization (Eq. 2) predicting $T_g$ as a function of $C^0$ solely, which could be applied to the information available with the one-dimensional VBS (1D-VBS) framework (Donahue et al., 2006), and can be used when the O:C ratio is not available in measurements.

\begin{equation}
T_g = 288.70 - 15.33 \times \log_{10}(C^0) - 0.33 \times [\log_{10}(C^0)]^2 \tag{2}
\end{equation}

The coefficients in Eq. (2) are obtained following the procedures developing Eq. (1) and the same training dataset is used. Figure S1 shows that Eq. (2) gives very similar predictions as Eq. (1) particularly for the compounds with low O:C ratio. As Eq. (1) and (2) are developed based on the compounds with their $C^0$ higher than $\sim 10^{-20} \mu g m^{-3}$, Eqs. (1–2) may not be applicable for compounds with $C^0 < \sim 10^{-20} \mu g m^{-3}$ (Fig. 1a).

We recently developed a parameterization (Eq. 3) predicting $T_g$ as a function of the number of carbon ($n_C$), hydrogen ($n_H$), and oxygen ($n_O$) atoms (DeRieux and Li et al., 2018), similar to the formulation used to predict $C^0$ (Donahue et al., 2011; Li et al., 2016).

\begin{equation}
T_g = (n_C^0 + \ln(n_C)) b_C + \ln(n_H) b_H + \ln(n_O) b_O + \ln(n_C) \ln(n_O) b_{CO} \tag{3}
\end{equation}

Values of the coefficients [$n_C^0, b_C, b_H, b_O, b_{CO}$] are [1.96, 61.99, -113.33, 28.74, 0, 0] for CH compounds and [12.13, 10.95, -41.82, 21.61, 118.96, -24.38] for CHO.
compounds. We broaden the parameterizations for CH and CHO compounds (Eq. 3) to the following equations applicable to CHON (Eq. 4) and CHOS compounds (Eq. 5):

\[ T_g = (n_C^0 + \ln(n_C)) \cdot b_C + \ln(n_O) \cdot b_O + \ln(n_S) \cdot b_S + \ln(n_C) \cdot \ln(n_O) \cdot b_{CO} + \ln(n_C) \cdot \ln(n_N) \cdot b_{CN} + \ln(n_O) \cdot \ln(n_S) \cdot b_{OS} \]  \hspace{1cm} (4) \\
\[ T_g = (n_C^0 + \ln(n_C)) \cdot b_C + \ln(n_O) \cdot b_O + \ln(n_S) \cdot b_S + \ln(n_C) \cdot \ln(n_O) \cdot b_{CO} + \ln(n_C) \cdot \ln(n_N) \cdot b_{CS} + \ln(n_O) \cdot \ln(n_S) \cdot b_{OS} \]  \hspace{1cm} (5)

Values of the coefficients \([n_C^0, b_C, b_O, b_S, b_{CO}, b_{CN} \text{ and } b_{ON}]\) in Eq. (4) are [5.34, 31.53, -7.06, 134.96, 6.54, -34.36, -15.35] and \([n_C^0, b_C, b_O, b_S, b_{CO}, b_{CS} \text{ and } b_{OS}]\) in Eq. (5) are [1.12, 68.41, 64.95, 35.77, -12.32, -9.85, 13.80], respectively. These values are obtained by fitting the \(T_g\) of CHON and CHOS compounds included in the training dataset (Fig. 1a, Table S1) with multi-linear least squares analysis. In the training dataset there are 35 CHON compounds with measured \(T_g\) available (Koop et al., 2011) and \(T_g\) of other 239 CHON compounds are estimated by the Boyer-Kauzmann rule using their measured \(T_m\) (Supplement). Figure 2(a) shows a fair agreement between the predicted \(T_g\) using Eq. (4) and the measured otherwise estimated \(T_g\) with \(R\) of 0.55 and relatively large AAVRE of 16 %. Figure 2 (b) shows a better prediction performance with \(R\) of 0.83 and AAVRE of 9 % for 212 CHON compounds included in the test dataset with estimated \(T_g\) by the Boyer-Kauzmann rule using the EPI-estimated \(T_m\). Measured \(T_g\) values for CHOS compounds are sparse (Zhang et al., 2019) and the \(T_g\) of the 835 CHOS compounds shown in Fig. 2c are estimated by the Boyer-Kauzmann rule using the EPI-estimated \(T_m\). Figure 2 (c) shows that Eq. 5 performs well for those CHOS compounds with \(R\) of 0.87 and AAVRE of 8 %.

Comparison of \(T_g\) predicted by Eqs. (4, 5) with \(T_g\) predicted by Eq. (1) shows that the agreement between the two sets of parameterizations for nitrogen- and sulfur-containing compounds is not as good as that for CHO compounds (Fig. S2), indicating that there are limitations of predicting \(T_g\) by the elemental composition for nitrogen-
and sulfur-containing compounds with complex elemental composition and molecular structures. As volatility depends significantly on functional groups contained in a molecule (Pankow and Asher, 2008; Compernolle et al., 2011), predicting $T_g$ by volatility (Eq. 1) indirectly incorporates the effects of molecular structure. Considering Eqs. (4, 5) are useful in high-resolution mass spectrometry measurements (DeRieux and Li et al., 2018; Schum et al., 2018), future experiments measuring more $T_g$ data for nitrogen- and sulfur-containing organics would help improve the $T_g$ parameterizations. Recently Zhang et al. (2019) developed a semi-empirical parameterization (Eq. 6) using vapor pressure ($p_0$ in atm) to predict $T_g$ based on measured $T_g$ of 11 SOA compounds:

$$T_g = 480.1 - \left(\frac{54395}{\log_{10}(p_0) - 1.7929}\right)^2 + 116.49$$  \hspace{2cm} (6)$$

$p_0$ can be converted to $C^0$ via $C^0 = (10^6 M p_0)/(RT)$, where $R$ is the ideal gas constant ($R = 8.2 \times 10^{-5}$ m$^3$ atm mol$^{-1}$ K$^{-1}$), $M$ is the molar mass (g mol$^{-1}$), and $T$ is the temperature (K). Figure 3 compares the measured $T_g$ (Koop et al., 2011; Rothfuss and Petters, 2017; Lessmeier et al., 2018; Zhang et al., 2019) to $T_g$ predicted by (a) $C^0$ and the atomic O:C (Eq. 1), (b) elemental composition (Eqs. 3-5), and (c) Eq. (6) by Zhang et al. (2019). While all three methods perform reasonably well, the predictions using elemental composition (Eqs. 3-5) show better performance (Fig. 3b) with $R$ of 0.93 and AAVRE of 11%, respectively.

The prediction performance is influenced by the training dataset used developing parameterizations of $T_g$. The compounds shown in Fig. 3 contain mostly carboxylic acid and hydroxyl functional groups (Koop et al., 2011; Rothfuss and Petters, 2017) and are included in the training dataset used developing Eq. (1) and Eqs. (3-5). The training dataset used in Zhang et al. (2019) included 11 organic compounds, and their parameterization predicted $T_g$ of isoprene SOA very well (Zhang et al., 2019), while underpredicting some low-$T_g$ compounds (Fig. 3c). For compounds with their
measured $T_g$ higher than 200 K, predictions by Zhang et al. (2019) show good performance and are consistent with the predictions given by Eq. (1). Predicted $T_g$ of 2-MT-OS using the three methods are 297 K (Eq. 1), 275 K (Eq. 5) and 280 K (Eq. 6), comparable with the measured $T_g$ of 276 ± 15 K (Zhang et al., 2019).

Note that predictions using elemental composition (Eq. 3) overestimate the $T_g$ of phthalate compounds (the star markers in Fig. 3). For instance, the observed $T_g$ of dioctyl phthalate is 194 K (Zhang et al., 2018), while the prediction is higher than 300 K (Fig. 3b). The reason is that ester is not an effective functional group to increase viscosity compared to carboxylic acid and hydroxyl (Rothfuss and Petters, 2017). Parameterizations using volatility (Eqs. 1 and 6) improve the predicted $T_g$ of phthalate compounds (Fig. 3a, c). Figure A1 shows comparing to the predictions using Eq. (6) provided in Zhang et al. (2019), predictions by $C^0$ and the atomic O:C (Eq. 1) and elemental composition (Eq. 3) agree better with the $T_g$ estimated from the Boyer-Kauzmann rule. Future experiments measuring more $T_g$ of SOA components would help verify the $T_g$ predictions by different parameterizations.

3. Application in field observations

$T_g$ of ambient OA are estimated by applying the developed parameterizations to field observations of volatility distributions. The term volatility refers to the effective saturation mass concentration ($C^*$) and we assume ideal thermodynamic mixing in which case $C^*$ is equal to $C^0$ (Donahue et al., 2011). $T_g$ are then placed into the 2D-VBS framework of O:C ratio vs. volatility as shown in Fig. 4. The isolines in Fig. 4 show the predicted $T_g$ by Eq. (1), showing that $T_g$ would be below ~ 250 K for intermediate volatility organic compounds (IVOC; $300 < C^0 < 3 \times 10^6$ μg m$^{-3}$), from ~ 260 K to 290 K for semi-volatile organic compounds (SVOC; $0.3 < C^0 < 300$ μg m$^{-3}$), and higher than 300 K for low-volatile organic compounds (LVOC; $3 \times 10^{-4} < C^0 < 0.3$ μg m$^{-3}$) and extremely low-volatile organic compounds (ELVOC; $C^0 < 3 \times 10^{-4}$ μg m$^{-3}$). $T_g$ increases as the O:C ratio increases for SVOC and IVOC, consistent with previous studies (Koop
et al., 2011; Saukko et al., 2012; Berkemeier et al., 2014). $T_g$ slightly decreases as the O:C ratio increases for LVOC and ELVOC compounds, which might be due to the uncertainties in Eq. (1) which is derived from a dataset containing fewer LVOC and ELVOC compounds as shown in Fig. 1a, which exhibits lower $T_g$ with higher O:C.

3.1 Southern Oxidant and Aerosol Study (SOAS)

The volatility of ambient OA was measured during the SOAS campaign which took place in the southeastern United States (Centreville, Alabama) in summer 2013 (Carlton et al., 2018). $T_g$ in each volatility bin ($T_g,i$) is calculated by Eq. (1) under the assumption of ideal thermodynamic mixing and the measured O:C is used (Xu et al., 2015). The glass transition temperatures of organic aerosols under dry conditions ($T_{g,org}$) are calculated by the Gordon-Taylor equation assuming the Gordon-Taylor constant ($k_{GT}$) of 1: $T_{g,org} = \sum w_i T_{g,i}$ (Dette et al., 2014), where $w_i$ is the mass fraction in the particle phase for each volatility bin (Table S2). To place $T_{g,org}$ into the 2D-VBS framework, the average log$_{10}(C^*)$ is calculated by $\sum w_i \log_{10}(C^*_i)$ (Kostenidou et al., 2018).

Figure 4 shows that $T_{g,org}$ of total OA (TOA) ranges from 232 K to 334 K, depending on volatility distributions measured by different methods. Stark et al. (2017) used three methods to derive volatility distributions applying the measurements of organic acids (which comprised half of the total OA; Yatavelli et al., 2015) from a high-resolution chemical ionization time-of-flight mass spectrometer (HRTof-CIMS) equipped with a filter inlet for gases and aerosols (FIGAERO, Lopez-Hilfiker et al., 2014; Thomson et al., 2017). The “Formulas” method used the SIMPOL group contribution method (Pankow and Asher, 2008) to calculate vapor pressures from the composition of the identified ions. Many of the detected species can be thermal decomposition products rather than actual SOA molecules (Stark et al., 2015; Stark et al., 2017), which can lead to overestimations of volatilities mainly in the IVOC range, resulting in the unlikely low $T_{g,org}$ of 232 K. The “Partitioning” method used the
measured particle-phase mass fractions of each species to estimate $C^*$ based on the partitioning theory (Pankow, 1994). The estimated $C^*$ is distributed mainly in the SVOC range (Stark et al., 2017), leading to a $T_{g,org}$ of 279 K (Fig. 4). This value is very close to the $T_{g,org}$ (281 K) simulated by a global chemical transport model EMAC-ORACLE (Shiraiwa et al., 2017). However, Stark et al. (2017) note that the partitioning-based volatility distribution is likely too high due to an artifact of signal-to-noise limitations, confining the $C^*$ characterizable by the partitioning method to a relatively narrow range centered around the ambient OA concentration (by definition, the semi-volatile range). In the thermogram method, $C^*$ at 298 K is estimated from the desorption temperature after calibration with known species (Faulhaber et al., 2009). This method results in 93% of OA mass distributed in the LVOC and ELVOC (Stark et al., 2017), and a high $T_{g,org}$ of 330 K is predicted (Fig. 4). These analyses indicate that the volatility distributions derived from different methods, even when based on the same measurements, significantly affect the predicted $T_{g,org}$.

Among these three methods using CIMS measurements of organic acids, Stark et al. (2017) note that the result from the thermogram method is more consistent with those measured by an aerosol mass spectrometer (AMS) with a thermo denuder (TD), which also applied the thermogram method to estimate the $C^*$ distributions (Hu et al., 2016). Saha et al. (2017) applied an evaporation kinetic model (Lee et al., 2011) based on the VBS approach to extract the $C^*$ distributions, and the effects of enthalpy of vaporization ($\Delta H_{vap}$) and accommodation coefficient ($\alpha$) are considered, resulting in the estimated $T_{g,org}$ of 313 K. This study retrieved $\alpha \sim 0.5$, which is consistent with recent observations (Krechmer et al., 2017; Liu et al., 2019). In summary, $T_{g,org}$ during the SOAS campaign span the range of 313 – 330 K.

Figure 4 also shows $T_{g,org}$ of OA components identified via positive matrix factorization (PMF) of the aerosol mass spectrometer data. Isoprene-epoxydiols-derived SOA (IEPOX-SOA) has $T_{g,org}$ of 345 K with very low volatility with the average $C^*$ lower than $10^{-4} \, \mu g \, m^{-3}$ (Hu et al., 2016; Lopez-Hilfiker et al., 2016; D’Ambro...
et al., 2019), which may be due to substantial formation of organosulfates and other oligomers (Lin et al., 2012; Hu et al., 2015; Riva et al., 2019). The predicted $T_{g,\text{org}}$ of IEPOX-SOA is higher than previously reported $T_{g,\text{org}}$ of 263 - 293 K for monoterpene-derived ($\alpha$-pinene, $\Delta^3$-carene, myrcene, limonene and ocimene) SOA (Petters et al., 2019).

The phase state of aerosol particles strongly depends on their water content (Mikhailov et al., 2009; Koop et al., 2011). Under humid conditions, $T_s$ of organic-water mixtures at given RH can be estimated considering hygroscopic growth combined with the Gordon-Taylor equation. The effective hygroscopicity parameter ($\kappa$) (Petters and Kreidenweis, 2007) of TOA during the SOAS is measured as 0.14 (Cerully et al., 2015). The Gordon-Taylor constant for organic-water mixtures is suggested to be 2.5 (Zobrist et al., 2008; Koop et al., 2011). Using the $T_s$ of organic-water mixtures, the temperature-dependence of viscosity can be calculated applying the Vogel-Tammann-Fulcher (VTF) equation (Angell, 1991; Rothfuss and Petters, 2017; DeRieux and Li et al., 2018). Figure 5(a) shows the predicted viscosity of total OA at different RH. $T$ is adopted as 298 K, the average value during the SOAS campaign (Hu et al., 2016). The characteristic timescale of mass transport and mixing by molecular diffusion ($\tau_{\text{mix}}$) is also calculated: $\tau_{\text{mix}} = d_p^2 / (4\pi D_b)$ (Seinfeld and Pandis, 2006), where $d_p$ is the particle diameter and the bulk diffusion coefficient $D_b$ is calculated from the predicted viscosity by the fractional Stokes–Einstein relation (Evoy et al., 2019). We assume the radius of the diffusing molecule of $10^{-10}$ m and the particle diameter of 200 nm (Shiraiwa et al., 2011). Note that these estimated timescales represent rough estimations, as molecular interactions in complex mixtures are not considered.

The viscosity of TOA at RH of 83% (average RH during SOAS) is predicted to be less than $10^2$ Pa s with $\tau_{\text{mix}}$ less than 1 s, which is consistent with the particle bounce measurements which suggested organic-dominated particles were mostly liquid during the SOAS campaign (Pajunoja et al., 2016). When RH was below ~50% in the sampling inlet, the particles were found to adopt a semi-solid state (Pajunoja et al.,
which agrees with the predicted viscosity of $10^7 - 10^{11}$ Pa s and $\tau_{\text{mix}}$ can be higher than 1 hour at RH of 50% (Fig. 5a).

Figure 5(b) shows diurnal variations of predicted viscosity of total OA using measured $T$ and RH during the SOAS campaign (Hu et al., 2016). During 10:00 – 20:00 when RH < 70 % and $T$ > 298 K, three simulations using different $T_{g,\text{org}}$ values predict that total OA occur as semi-solid with the predicted viscosity of $10^7 - 10^9$ Pa s and the mixing times less than 1 hour. Particles are predicted to have low viscosity of < 1 Pa s adopting a liquid phase during nighttime. The lowest viscosity occurs around 5:00 – 6:00 with RH > 95 %. Here we did not consider the effects of the diurnal variations of volatility distributions, as they did not vary dramatically over the campaign period (Saha et al., 2017). Besides $T$ and RH, diurnal variation of ambient aerosol phase state also depends on particle chemical composition and mixing states. Organic particles in Amazon were found to be more viscous at night than the daytime due to the influence of biomass burning that may form non-liquid particles (Bateman et al., 2017). Particles in a mixed forest in northern Michigan were also found more viscous at night despite higher RH than the daytime, due to the formation of high molar mass organic compounds and smaller inorganic sulfate mass fractions (Slade et al., 2019). Phase state measurements during daytime and nighttime at Atlanta suggested that the ambient particle phase state was influenced by OA composition, the presence of inorganic ions, aerosol liquid water and particle mixing state (Ditto et al., 2019).

### 3.2 $T_{g,\text{org}}$ at 11 global sites

Figure 6 summarizes $T_{g,\text{org}}$ at 11 sites where the measured volatility distributions with volatility bins of four or more are available (Table S2). We did not include the data with narrower volatility ranges which may not correctly characterize the properties of ambient SOA (Bilde et al., 2015), and thus may not be appropriate for estimating fundamental particle properties and volatility distributions without considering realistically low $C^*$ bins would result in too low $T_g$. Note that a narrow VBS
may still be useful for efficiency in 3-dimentional chemical transport models for SOA evaporation and condensation under the narrow range of ambient temperature variations (Kostenidou et al., 2018). $T_{g,\text{org}}$ is color-coded in the 2D-VBS framework of O:C vs. $\log_{10}C^*$ in the panel (a), whereas the panel (b) shows $T_{g,\text{org}}$ vs. $\log_{10}C^*$ with markers color-coded with the O:C ratio. $T_{g,\text{org}}$ of total OA (TOA) varies from 290 K to 339 K. The lower $T_{g,\text{org}}$ occurs at Beijing, China in June 2018 (Xu et al., 2019). OA in Beijing was found to be overall more volatile with the particle-phase semi-volatile fraction of 63%, leading to a lower $T_{g,\text{org}}$. The predicted $T_{g,\text{org}}$ of total OA at numerous other sites range between 300 K and 320 K, including Paris (Paciga et al., 2016), Mexico city (Cappa and Jimenez, 2010), Centreville (Hu et al., 2016; Saha et al., 2017; Stark et al., 2017), Raleigh (Saha et al., 2017), and Durham (Saha et al., 2018) in southeastern US. The $T_{g,\text{org}}$ (316 K) is higher at 220 m from a highway in Durham than the $T_{g,\text{org}}$ (309 K) at 10 m downwind distance due to the dilution and mixing of traffic-sourced particles with background air and evaporation of semi-volatile species during downwind transport (Saha et al., 2018). $T_{g,\text{org}}$ is predicted to be higher (>320 K) at the sites in Athens, Pasadena, Colorado Rocky Mountain and Amazon (Fig. 6).

At the same site in Mexico city or Athens, Figure 6 shows that the $T_{g,\text{org}}$ is lowest for the BBOA factor (287 K – 305 K) followed by HOA (290 K – 331 K) and OOA (> 350 K for the MO-OOA in Mexico city and Paris), which follows the trend of the decrease in the average volatility of these OA factors. Cooking OA (COA) has comparable $T_{g,\text{org}}$ (290 K – 324 K) with HOA. Marine OA (MOA) was characterized with high volatility containing around 60% SVOC, leading to the $T_{g,\text{org}}$ of 295 K in summer in Paris.

Figure 7 shows the OA viscosity variation of OA components against RH. The hygroscopic growth is considered based on hygroscopicity ($\kappa$), which is estimated as a function of the O:C ratio (Lambe et al., 2011) when $\kappa$ was not measured (Table S2). Figure 7a shows OA factors with low O:C ratio, i.e., HOA, COA and BBOA, occur as liquid only when RH is very high (> ~80 %), as their $\kappa$ is estimated to be low (<
HOA, COA and BBOA in Athens and COA in Paris can undergo glass transition (i.e., viscosity reaches $10^{12}$ Pa s) at RH between 25% and 68%. OA factors with higher O:C ratio and hygroscopicity including LO-OOA, MO-OOA, IEPOX SOA and isoprene-OA tend to become liquid (viscosity < $10^2$ Pa s) at intermediate RH (Fig. 7b).

The predicted ambient viscosity is compared with the experimental observed viscosity of SOA formed from isoprene (Song et al., 2015), α-pinene (Abramson et al., 2013; Renbaum-Wolff et al., 2013; Kidd et al., 2014; Pajunoja et al., 2014; Bateman et al., 2015; Zhang et al., 2015; Grayson et al., 2016; Petters et al., 2019), toluene (Song et al., 2016a) and diesel fuel (Song et al., 2019). The majority of experimental values are well bounded by the predicted viscosity of OOA, represented by the pink shaded area. One exception is the measured viscosity of isoprene SOA is lower than the predicted viscosity of IEPOX SOA at low RH (<30%). One possible reason is that the isoprene SOA in experiments was formed with high oxidant concentrations with short reaction time in an oxidation flow reactor in the absence of inorganic seed particles (OFR) (Song et al., 2015). In ambient environments heterogeneous reactions with acidic sulfate particles forming oligomers are suggested to be an important pathway (Surratt et al., 2010; Lin et al., 2013; Hu et al., 2015, 2016). These particle-phase organosulfates may contribute to a higher viscosity, as indicated by the predicted viscosity of IEPOX-derived organosulfate mixtures with their $T_{g,\text{org}}$ estimated to be 313 K (Riva et al., 2019). Another reason could be the mass concentrations (100 ~ 1000 μg m$^{-3}$) of isoprene SOA generated in the OFR (Song et al., 2015) are much higher than the ambient OA concentrations (5 μg m$^{-3}$, Stark et al., 2017) observed during the SOAS campaign. Higher mass concentrations can lead to lower viscosity, as more semi-volatile compounds can partition into the particle phase (Pankow, 1994; Donahue et al., 2006; Grayson et al., 2016; Jain et al., 2018; Champion et al., 2019).

4. Implications on SOA simulations in chemical transport models
Shiraiwa et al. (2017) simulated the global distribution of annual averages of SOA phase state using the chemical transport model EMAC (Jöckel et al., 2006) coupled with the organic aerosol module ORACLE (Tsimpidi et al., 2014). ORACLE uses the 1D-VBS framework with four $C^*$ bins ($1$, $10$, $10^2$, and $10^3$ μg m$^{-3}$). To estimate $T_g$ the values of molar mass and O:C ratio were assigned for each volatility bin based on molecular corridors (Shiraiwa et al., 2014). Note that the molar mass assigned for the volatility bin of $1$ μg m$^{-3}$ was assumed to have relatively high molar mass to partially compensate for the fact that ORACLE does not consider lower volatility bins with higher molar mass. Global distributions of $T_g/T$ presented in Shiraiwa et al. (2017) is converted to viscosity using the VTF equation, as shown in Fig. 8. Figure 8 also includes the viscosity of total OA at 11 sites by applying the model simulated 5 years’ average $T$ and RH with $κ$ assumed to be 0.1 (Pringle et al., 2010). It shows the predicted viscosity at 11 sites (markers) generally agree well with the global simulations: the amorphous solid or semi-solid phase occurs over relatively dry lands, including western US, Mexico city, and Beijing; the lower viscosity occurs in southeastern US and Paris.

The global simulations show that the particles are liquid in the Amazon, while they occur as semi-solid in our predictions based on measured volatility distributions. The reason of this disagreement may be mainly due to the substantial fraction of low volatility compounds observed in ambient measurements largely missing from global simulations. Hu et al. (2016) observed that 90% of OA have volatilities lower than $1$ μg m$^{-3}$, which is the lowest $C^*$ bin in the global simulations. The ambient phase state measurements show that for background conditions of the Amazonian tropical forest, particles are mostly liquid, while with the anthropogenic influence of biomass burning, they occur as a nonliquid phase (Bateman et al., 2016; Bateman et al., 2017). The volatility distributions were measured in the dry season heavily influenced by biomass burning (Hu et al., 2016), which can lead to the higher predicted viscosity. A similar case is observed in Athens, that our predictions indicate
the glassy phase state while the global model predicts the occurrence of a semi-solid phase.

Most of the current chemical transport models treat particles as liquid or homogeneously well-mixed without considering particle-phase diffusion limitations (Pankow, 1994), which can lead to bias in simulations of SOA mass concentrations and evolution of size distributions (Shiraiwa and Seinfeld, 2012; Zaveri et al., 2018). The SOA simulations applying the VBS framework have not included the effects of viscosity on SOA formation and evolution. When the gas-particle partitioning is in the bulk diffusion-limited regime, kinetic treatments of SOA partitioning may need to be applied (Perraud et al., 2012; Liu et al., 2016; Yli-Juuti et al., 2017; Li and Shiraiwa, 2019). Some chamber experiments probing the mixing timescales of SOA particles did not observe significant kinetic limitations at moderate and high RH under room temperature (Loza et al., 2013; Ye et al., 2018), warranting further investigations on the degree of kinetic limitations in ambient tropospheric conditions. The method developed in this study can be applied in the VBS or the molecular corridor-based approach to improve OA simulations in chemical transport models (Pye et al., 2017; Schmedding et al., 2019). Several important aspects should be further explored in future studies. The interplay of diffusion limitations and phase separation impacts heterogeneous and multiphase chemistry (Vander Wall et al., 2018; DeRieux et al., 2019; Zhou et al., 2019) and gas-particle partitioning (Zuend and Seinfeld, 2012; Freedman, 2017; Pye et al., 2017; Gorkowski et al., 2019a). The particle morphology and the degree of non-ideal mixing and liquid-liquid phase separation can evolve upon atmospheric aging (Gorkowski et al., 2019b). These aspects may also need to be considered for better representation of organic aerosols in future studies.

Author contribution.

YL, JLJ and MS designed the research. YL developed the parameterizations. DAD, HS and JLJ provided measured volatility distributions for the SOAS campaign. YL and MS
wrote the manuscript. All authors discussed the results and contributed to manuscript editing.

Data availability. The data used in this study is available in the supplement.

Competing interests. The authors declare that they have no conflict of interest.

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Figure 1. (a) \(T_g\) of organic compounds as measured and estimated by the melting temperature \((T_m)\) with the Boyer-Kauzmann rule plotted against pure compound saturation mass concentration \((C^0)\) as measured and estimated from the EPI suite (see supplement for details of the dataset). (b) Predicted \(T_g\) for compounds shown in (a) using a parameterization (Eq. 1) developed in this study compared to measured and estimated \(T_g\) from \(T_m\) with the Boyer-Kauzmann rule. (c) Predicted \(T_g\) for SOA components (Shiraiwa et al., 2014) using Eq. (1) plotted against estimated \(T_g\) from \(T_m\) with the Boyer-Kauzmann rule. The markers are color-coded by the atomic O:C ratio. Organic compounds include CH compounds (squares), CHO compounds (circles), CHON compounds (triangles), and CHOS compounds (inverted triangles). The correlation coefficient \((R)\) and the average absolute value of the relative error \((AAVRE)\) are shown. The solid line shows the 1:1 line and the dashed and dotted lines show 68% confidence and prediction bands, respectively.

Figure 2. (a) Predicted \(T_g\) for CHON compounds using a parameterization (Eq. 4) developed in this study compared to measured or estimated \(T_g\) by the Boyer-Kauzmann rule using measured \(T_m\). (b) Predicted \(T_g\) for CHON compounds using Eq. (4) compared to estimated \(T_g\) by the Boyer-Kauzmann rule with \(T_m\) estimated by the EPI suite. (c) Predicted \(T_g\) for CHOS compounds using Eq. (5) compared to estimated \(T_g\) by the Boyer-Kauzmann rule with \(T_m\) estimated by the EPI suite. The markers are color-coded by the molar mass. The solid line shows the 1:1 line and the dashed and dotted lines show 68% confidence and prediction bands, respectively. The correlation coefficient \((R)\) and the average absolute value of the relative error \((AAVRE)\) are included in each figure legend.
Figure 3. Comparison between measured $T_g$ and $T_g$ predicted by (a) $C^0$ and O:C (Eq. 1), (b) elemental composition (Eqs. 3-5), and (c) the parameterization (Eq. 6) in Zhang et al. (2019). The markers are color-coded by the atomic O:C ratio. The solid line shows the 1:1 line. The compounds include CH compounds (squares), CHO compounds (circles), phthalates (stars), CHON compounds (triangles), and 2-methyltetrol sulfates (2-MT-OS, inverted triangle). The correlation coefficient (R) and the average absolute value of the relative error (AAVRE) are included in each figure legend.
Figure 4. Predicted glass transition temperatures of organic aerosols under dry conditions ($T_{g,\text{org}}$) during the SOAS campaign, which are placed into the two-dimensional VBS framework. The isopleths correspond to the $T_{g,\text{org}}$ calculated using Eq (1) with the effective saturation mass concentration ($C^\ast$) and the O:C ratio defined in the 2D-VBS. The markers represent the $T_{g,\text{org}}$ of total OA (TOA) and IEPOX SOA calculated from the volatility distributions simulated by a global chemical transport model EMAC-ORACLE (Shiraiwa et al., 2017) or measured during the SOAS campaign (Hu et al., 2016; Saha et al., 2017; Stark et al., 2017). Three methods ("Formulas", ‘Partitioning’, and ‘Thermograms’) are applied in Stark et al. (2017) to derive the $C^\ast$ distributions, while the “Thermograms” method provides the most credible volatility distributions compared to ‘Formulas’ and ‘Partitioning’ (edge lines are in gray).
Figure 5. (a) Predicted viscosity of total OA measured during the SOAS campaign as a function of RH. (b) Diurnal variations of viscosity of total OA predicted employing the measured RH and $T$ (Hu et al., 2016) during the SOAS campaign. $T_{g,org}$ are calculated using the volatility distributions measured in Hu et al., (2016), Saha et al. (2017), and the “Thermograms” method in Stark et al. (2017). Characteristic mixing timescales of organic molecules with the radius of $10^{-10}$ m within 200 nm particles are also shown in the right axis.
Figure 6. Predicted glass transition temperatures of organic aerosols under dry conditions ($T_{g,\text{org}}$) at 11 sites. The isopleths in (a) correspond to the $T_g$ varied with the effective saturation mass concentration ($C^*$) and the O:C ratio in the two-dimensional VBS (2D-VBS) framework. The markers represent the $T_{g,\text{org}}$ of total OA (TOA) and its components calculated from measured volatility distributions (Table S2) and color-coded with $T_{g,\text{org}}$ in (a) and the O:C ratio in (b). The edge color of the markers indicates the OA components identified via positive matrix factorization (PMF) of aerosol mass spectrometer data.
Figure 7. Predicted viscosity of (a) HOA, COA and BBOA and (b) LO-OOA, MO-OOA, IEPOX SOA and Isoprene OA in different locations at 298 K as a function of RH. Experimentally measured viscosity of laboratory-generated SOA formed from isoprene (Song et al., 2015), α-pinene (Abramson et al., 2013; Renbaum-Wolff et al., 2013; Kidd et al., 2014; Pajunoja et al., 2014; Bateman et al., 2015; Zhang et al., 2015; Grayson et al., 2016; Petters et al., 2019), toluene (Song et al., 2016), and diesel fuel (Song et al., 2019) are also shown. Predicted viscosity of IEPOX-derived OS mixtures (solid blue line) is from Riva et al. (2019). Note that in case these PMF factors are internally mixed with other components, the predicted viscosity would not represent real ambient complex organic mixtures.

Figure 8. Comparison of global distributions of annual averages of SOA viscosity at the surface simulated by a chemical transport model (Shiraiwa et al., 2017) with the viscosity predicted by measured volatility distributions at 11 global sites as indicated with markers (triangle, circle and square represent remote, urban and forested sites, respectively, Table S2).
Figure A1. Predicted $T_g$ for CHO compounds by (a) $C^0$ and O:C (Eq. 1), (b) elemental composition (Eq. 3), and (c) the parameterization (Eq. 6) in Zhang et al. (2019) plotted against estimated $T_m$. $T_m$ and $C^0$ in (a) – (c) are measured; $T_m$ and $C^0$ in (d) – (f) are estimated by the EPI Suite and the EVAPORATION model, respectively (see Supplement). Estimated $T_g$ is then calculated from $T_m$ applying the Boyer-Kauzmann rule. The markers are color-coded by the atomic O:C ratio. The solid line shows the 1:1 line. The correlation coefficient (R) and the average absolute value of the relative error (AAVRE) are shown.