

Response to Referee number 3

7 February 2019

The authors thank Referee #3 for his/her work. We have considered all comments thoroughly and profoundly. Unfortunately, many of them cannot be accepted at all or fully. Our specific responses are as follows, while the textual modifications are highlighted in red or by crossing out in the revised MS.

The paper summarizes a large body of data and tries to extract information on the underlying processes of new particle formation. This is rather difficult as the lowest particle size they measure is 6 nm and growth rates can only be determined around 10 nm.

1. The lower measurable particle diameter limit of DMPS/SMPS systems is important for identification of NPF and growth events and further data treatment. Evaluations of this type of atmospheric measurements are mostly based on particle diameter range <20 nm (e.g. Kulmala et al., *Nat. Protoc.*, 7, 1651–1667, 2012). In order to separate reliably the NPF and growth events from huge emission peaks which can occur in cities and which can temporarily influence the size intervals down to even smaller diameters, it is highly preferable to have the lower limit below 10 nm. Our limit value of 6 nm was proved to be already satisfactory since it allows to identify and separate different particle generation processes (see e.g. Fig. S1b of the present MS and Salma et al., *Atmos. Chem. Phys.* 16, 7837–7851, 2016). It is also worth mentioning that from 6 urban sites involved in a recent global analysis of NPF over long-term measurements (Nieminen et al., *Atmos. Chem. Phys.*, 18, 14737–14756, 2018), the lower diameter limit was 3 nm at 2 sites, it was 6 nm at 3 locations, while it was 11 nm at 1 of the sites, and both the J_{nuc} and GR were determined for the diameter interval of 10–25 nm. All these indicate that in atmospheric studies, our experimental systems and evaluation protocols seem completely adequate for the time being.

The authors do not provide much more insight than in the paper of Nieminen et al., where they are coauthors of, except that the results are now based on a larger data set.

2. The goals of the paper mentioned in the comment were largely different from our aims. We can list several important insights explicitly as examples which are part of the present MS and which were not dealt with in the referred paper. They primarily include 1) the evaluation

and discussion of monthly distributions of J_6 and GR_{10} together with their relationships with nucleation occurrence frequency and relevant atmospheric parameters, 2) timing properties of NPF and growth events, 3) refinements of J and GR calculations dedicated to urban environments, 4) statistical distributions of J_6 and GR_{10} , 5) occurrence and properties of extreme events and events with broad onset. These items represent a considerable piece of novelty and new knowledge. Furthermore, the results and conclusions are based on 247 quantifiable NPF and growth events in an urban environment, which means a rather strong background. Finally, we can quote from the Summary and conclusions section of the Nieminen et al. paper (p. 14750): “For future studies, it would be very valuable to make detailed investigations on the interdependencies among J_{nuc} , GR, and NPF event frequency, at both single measurement sites and among sites of seemingly similar environmental characteristics.” This is exactly what we did in our MS. In addition to our arguments, we can offer to all persons involved a recent and excellent review paper of Kerminen et al., *Environ. Res. Lett.* 13 (2018) 103003, 2018 dedicated to field observations, which also gives a scientific outlook and summarizes future research needs, and which can help putting our present results and conclusions more adequately into a scientific frame of international atmospheric NPF and particle growth studies.

Also the fact that the sulphuric acid proxy does not correlate with J_6 and GR_{10} has already been reported in an earlier paper. Although sulphuric acid does only contribute 12.3% to GR_{10} it does not mean that it is not relevant for NPF (line 608).

3. This conclusion was mentioned in the MS as a minor outcome of the study with the purpose of confirming earlier results (as explicitly stated in the line specified in the comment). The related sentence was modified now to emphasize the key role of H_2SO_4 in the nucleation process and early particle growth.

Many studies have shown a relation between NPF rates measured at small sizes and sulphuric acid, while the growth is dominated by organics. In Figure 4 the authors relate basically reciprocal (sulfuric acid proxy) versus reciprocal (sulfuric acid proxy) modulated by the GR. The linear relation is not surprising and does not lead to any conclusions. As the authors repeat several times in the paper NPF and growth is a complex process. Nevertheless, they test only relations of one single parameter with J6 or GR10. Why do the authors not make an attempt to combine parameters?

4. Figure 4 and the related discussion were removed from the MS to avoid any misunderstanding or incompleteness. The remaining part was also restructured, split into shorter pieces and clarified. Evaluation of the overall data set by multistatistical methods is indeed planned. This comprehensive evaluation is, however, to be accomplished after some markers or proxies for biogenic emission sources (such as e.g. photosynthetic activity) are also included. The extension of the present MS by this comprehensive statistical analysis would not fit among the present objectives and would not be advantageous or feasible considering both the length and timing of this MS as well. See also response no. 5.

It is known that low temperature stabilizes nucleating clusters and that organics promote growth and thus the survival probability. It might thus be worthwhile to look for a proxy representing condensing organics.

5. Chemical species including organics participating in the urban atmospheric NPF and growth were investigated in an intensive international measurement campaign in Budapest over March-May 2018 by deploying API TOF-MS with/without CI, PSM, AIS and DMPS systems. Some potential proxy values for condensing organics are under evaluation. This was mentioned in the Conclusions section, and it is further emphasized and explained in the revised version.

I also question if daily averages are the appropriate parameter to inquire NPF mechanisms. Although it is worth to report on this large data set, I find the paper does not provide much new information and I do not see what the authors' "consequences of dynamic and timing properties" are as announced in the title.

6. Daily averages were calculated for those variables which change slowly over a day (e.g. [SO₂]). For some other variables such as CS, we constrained the averaging for the time intervals from t_1 to t_2 , thus over the nucleation process itself. Some other variables, such as the gas-phase H₂SO₄ proxy, were characterized by their daily maximum. They are accurate

specified in Section 4.1 Ranges and averages. As far as the novelty of the MS is concerned, we must refer again to the list in response no. 2. The main conclusions drawn from the dynamic and timing properties are readily collected in the Abstract.

Line 151 and 494: What is the detection limit of the SO₂ detector? Are the low SO₂ concentrations measured significantly above DL?

7. The limit of determination (LOD) of the SO₂ analyzer system applied is approximately 0.2 µg m⁻³. More than 98% of the hourly-mean concentrations were above the LOD. The information is also included into the text now.

Line 318-319: I do not see a trend in particle concentrations.

8. The annual medians for the city centre in the measurement years 2008–2009, 2013–2014, 2014–2015, 2015–2016 and 2017–2018 are as follows: 11.5×10³, 9.7×10³, 9.3×10³, 7.5×10³ and 10.6×10³ cm⁻³, respectively. The first 4 data indicate unambiguously a decreasing tendency, while the last data point may look somewhat different. Rigorous statistical evaluation of the joint data set of particle number concentrations in various size fractions over a decennial time interval from 03–11–2008 to 02–11–2018 is in progress, and its preliminary results in the one hand, confirm the decreasing tendency, and in the other hand, reveal some fine structure to this dependency. This information was added to the revised MS.

Table 2: the authors use local time as time base. We know that photolysis is an important driver of sulfuric acid and oxidant production. Would it not be more appropriate to use time after sunrise for starting time?

9. The suggestion represents an option, which can be consider for specific studies. In the present MS, we selected the local time as the time base of most data on purpose and as a compromise because we had experienced in several earlier investigations (e.g. Salma et al., Atmos. Environ., 92, 154–161, 2014) that it is the daily activity time pattern of inhabitants that substantially influences or determines many atmospheric sources and important processes in Budapest. It was explained in lines 123–125 of the original MS, and a reference

for the statement was included as well. The timing parameters of the NPF were given in UTC+1.

Line 441: how can you conclude that NPF is not sensitive to temperature? Indeed the yearly average does not vary much, but is the yearly average really important? What matters more is the temperature during an event in combination with formation rates of nucleating and condensing vapors.

10. The sentence mentioned was replaced from its original location to section 4.2 Monthly distributions. It was largely corrected and extended to a discussion by involving the temperature profiles on nucleation and non-nucleation days, biogenic emissions, photochemistry and results from other international studies.

Line 498: What do you mean by “CO is less certain”?

11. The related sentence was modified to express our intention better that the variability of CO was without obvious tendentious temporal structure or feature.

Figure 2: Is the low value of H₂SO₄-proxy in May real or an artefact? What is the reason for that?

12. It is the monthly distribution of daily maximum gas-phase H₂SO₄ proxy that is shown in Fig. 2. The mean value for May represents 23 days. Its low value seems to be influenced by enhanced effect of multiplying relatively low GRad with relatively small [SO₂] for a few days particularly in 2015 (which was a strange year as far as the monthly distribution of nucleation frequency as well is concerned; see Fig. 1 of this response). The reliability of the monthly data is to be increased with the length of the overall data sets in the future. This additional information is added now in a synthesized manner to the text.

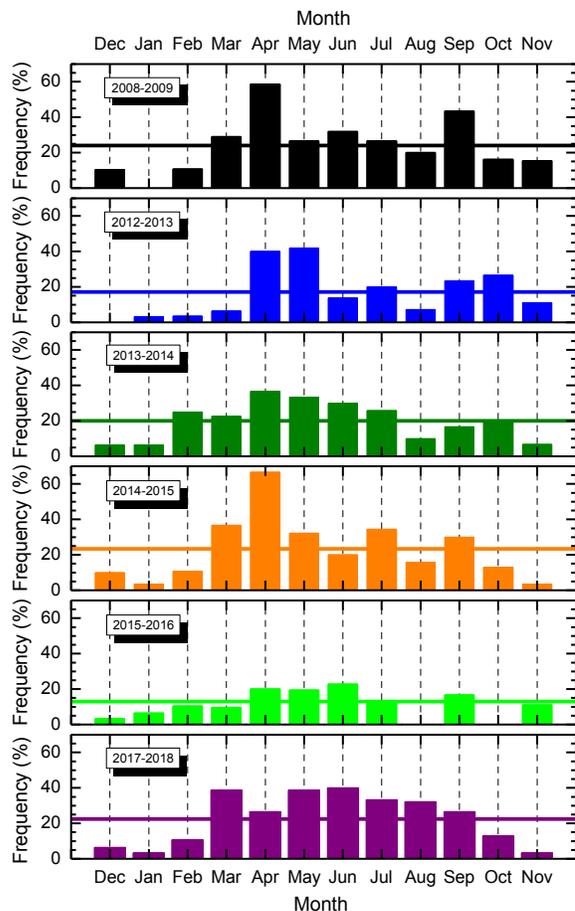


Figure 1. Monthly distribution of relative nucleation frequency in Budapest for measurement years of 2008–2009, 2012–2013, 2013–2014, 2014–2015, 2015–2016 and 2017–2018. The horizontal lines indicate the annual mean frequency. More information is given in the MS.

Line 545: This is not the line of equality. The units of each axis is different. There is also no discussion of this relation with respect to literature, e.g. Nieminen et al.

13. We used the expression “line of equality” in its broader sense, hence when the abscissa and ordinate are on the same scale even they do not have the same units. To the explicit request of Referee #3, however, we can change it to another expression, e.g. “line with a slope of 1”. We also amended the discussion of the relationship between J and GR at several places by considering the international results available in the literature.

Line 547: The difference between slopes for centre and near-city station is not very convincing. If the authors would also restrict the city centre plots to $GR < 10$ nm/h I expect a large scatter of the slopes. The near-city data do not seem to be different from the other data.

14. We were aware of this inherent limitation mainly caused by smaller dynamic properties (and partly by shorter measurement time interval) in the near-city background than in the city centre, and expressed it by ourselves in lines 554–556 of the original MS. Now, we reformulated the statement completely and turned it from a conclusion into a working hypothesis because a rigorous statistical treatment would indeed require stronger/larger variability in the near-city background data.

Line 559: It should say “that leads to $J_6 > 0$ ”. $J=0$ cannot be measured and is meaningless.

15. The suggestion was accepted and adopted.

Line 565: what do you mean by “weak phenomena”?

16. The related sentence was modified to express that we mean the NPF events with relatively small particle formation rate (weak events).

Line 611ff: This explanation is unclear. Surely, GR need to be faster in urban areas but that does not mean that there could be no correlation. Simply speaking higher CS should lead to lower GR. Apparently, a positive correlation is found, isn't it? This would be counterintuitive.

17. The GR of newly formed particles to larger sizes is primarily coupled to 1) CS, which is further linked to the entire aerosol particle population (including the newly formed particles, thus the NPF itself), 2) to the total concentration and some physicochemical properties of non-volatile gaseous compounds and 3) to their production rate in the gas phase from aerosol precursor compounds (e.g. Kerminen et al., *Environ. Res. Lett.* 13 (2018) 103003, 2018). Understanding these couplings is essential when analyzing atmospheric observations. It is not fully plausible to make intuitive expectations on simplified paired relationships, for instance between CS and GR, under such complexity. Therefore, we stuck to the experimental data and are to contribute to the phenomenological picture on the system of relationships in this part of the MS, which will be eventually leading to a comprehensive and

qualitative explanation of the connections in the future. We extended the sentence briefly with these additional arguments and explanation.

Section 4.4 needs much improvement.

18. We split the section into shorter parts and clarified it by clearer formulations.

Line 739: Where does this number of contribution of NPF to total particle concentration come from? How was the analysis done?

19. Typical number of particles generated by an NPF and growth event on a nucleation day was roughly estimated by considering the median J_6 and median duration of nucleation, Δt (their distribution function is lognormal; see Table 2) and mean relative coagulation loss, F_{coag} (see Table S1) as: $J_6 \times \Delta t \times (1 - F_{\text{coag}}) = 4.6 \times 180 \times 60 \times 0.83 = 41 \times 10^3 \text{ cm}^{-3} \approx 10^4 \text{ cm}^{-3}$. This concentration is in line with other results achieved by nucleation strength factor according to which the particle number concentration due to NPF and growth process on a general nucleation day is increased by a factor of approximately 2 (Salma et al., Atmos. Chem. Phys., 17, 15007–15017, 2017). A more detailed description of the estimation process and the mathematical expression utilized are added now together with the last reference mentioned.

In addition to the issues above, we also adopted some smaller changes and added a few recent papers as references to further improve the MS.

Finally, we think that the comments of Referee #3 eventually helped us to formulate our thoughts and ideas better. We appreciate this. We wish, however, to emphasize that the major message of the MS lies in a considerable variety of contributions to the emerging research field of urban atmospheric NPF and growth, which have been becoming possible and increasingly recognized thank to gradually generating, several-year long, semi-continuous, critically evaluated, complex and coherent data sets. We further stressed this aspect of the MS now in the Conclusion section and added a new opening sentence to the Abstract as well.

Imre Salma

Response to Referee number 4

7 February 2019

The authors thank Referee #4 for his/her detailed, expertise and valuable comments to further improve and clarify the MS. We have considered all recommendations and made appropriate alterations. Our specific responses to the comments are as follows, while the detailed textual modifications are highlighted in red or by crossing out in the revised MS.

Page8, line 256: the mean new-to-old rate ratios of J_6 were 1.23 for city center and 1.20 for near-city background. I would expect that traffic emission causes overestimation of formation rates because it is a source of nanoparticles. Please specify why correcting traffic emission in formation rates calculation gives higher J_6 .

1. Several modifications were simultaneously adopted in the revisited and refined calculations protocol of the new set of J_6 for the measurement years of 2008–2009 (city centre) and 2012–2013 (near-city background). They include the subtraction of particle number concentrations emitted by road traffic from N_{6-25} , which usually leads to a decrease in the coagulation loss and loss due to growth out from the diameter range of 6–25 nm, and which can sensitively influence the slope of the concentration change in time (dN_{6-25}/dt) in a positive or negative manner depending on the actual time evolution of perturbing emission source. In addition to that, the time interval in which this slope is considered to be constant was determined within a completely new treatment. We would also like to mention the mean relative contribution of the concentration increment, coagulation loss and growth out from the diameter interval to J_6 have different weights of 71%, 17% and 12%, respectively (lines 246–249 of the original MS, and Table S1). Furthermore, J_6 itself also depends on GR_{10} , which makes the relationships even more complex. These explain why the overall effect of urban influence generally resulted in increased dynamic properties. The mean new-to-old ratio for J_6 was larger for the city centre (1.23) than for the near-city background (1.20). It should also be emphasized that the re-calculation mainly affected the individual dynamic properties with relatively small absolute values. The whole process is considered as a methodological improvement over the years of research. The MS was amended by a more detailed description of Equation (1) and by a brief explanation of the issues above.

Section 4.2: Discussion on NPF events frequency should include conditions of NPF days as well as non-NPF days. Properties discussed in the section are only based on events days. This could be misleading because non-events day conditions are not discussed.

2. Information on the average CS (calculated for whole days), gas-phase H₂SO₄ proxy, GRad, air *T* and concentration of some criteria pollutants on non-nucleation days were partly included now. Many properties are, however, biased by the seasonal cycle of solar electromagnetic radiation via the seasonal variation of new particle formation frequency, and therefore, their interpretation needs special attention. They are to be fully utilized and explained for investigating the changes in annual pattern of relative nucleation frequency over the years, and a more comprehensive evaluation and discussion is to be realized in a future study outlined in response no. 5.

Line 484 conclude gas-phase H₂SO₄ are unlikely to be the limiting factor of NPF occurrence in Carpathian Basin including Budapest from the misalignment between the monthly occurrence frequency and the other properties. To make this statement solid, H₂SO₄ proxy for events days and non-events days is needed.

3. Averages of several atmospheric properties involved in the H₂SO₄ proxy were derived separately for the event days and non-event days, their effects were briefly discussed, and as a result of it, the statement mentioned was removed.

Page18, line548: Direct compare the numbers of J and GR or saying something contribute equally to the formation of particle and to their growth don't make sense because they are different physical variables. Correlation between J and GR are expected but comparison of the regression line with $J_6 = GR_{10}$ doesn't give any useful information.

4. The sentence was modified to express clearly that we mean that the chemical species available in the air affect the formation rate and growth rate differently at the 2 urban sites. This could partially be caused by differences in chemical composition. We reformulated the whole statement completely and turned it from a conclusion into a working hypothesis because a rigorous statistical treatment would indeed require larger variability in the near-city background data.

Page20: Lacking correlation with single parameters to J/GR doesn't tell too much as NPF is controlled by multiple parameters. With the size of the data set, authors could perform analysis on subsets of the data with certain constrains like temperature or H2SO4 proxy.

5. Evaluation of the overall data set by multistatistical methods is indeed planned. This comprehensive evaluation is, however, to be accomplished after some markers or proxies for biogenic emission sources (such as e.g. photosynthetical activity) are also included. The extension of the present MS by this comprehensive statistical analysis would not fit among the present objectives and would not be advantageous or feasible considering both the length and timing of this MS as well. The present study can be considered as the first step in a larger statistical evaluation process and which supplied orienting ideas on the specific directions to proceed in. This perspective further study is very briefly mentioned in the Conclusions section now.

Page 20, line 625 to 636 and figure 4: $GR/H_2SO_4 \text{ proxy} = b * (1/H_2SO_4 \text{ proxy}) + a$ is equivalent to $a * H_2SO_4 \text{ proxy} + b = GR$. A negative 'a' means the higher H2SO4, the lower the GR. This is contradictory to the interpretation of increasing gas-phase H2SO4 related to larger contribution of other vapors to particle growth. Another concern would be special care should be taken when combine H2SO4 proxy at sub-urban site and urban site as the VOCs and NOx condition could be totally different but not taken into consideration.

6. Figure 4 and the related discussion were removed from the MS to avoid any misunderstanding or incompleteness. The remaining part of the section was restructured, split into shorter pieces and clarified.

Page 24, line 739: To make the full potential of the data set, more detailed studies on the contribution of NPF to regional particle concentration could be performed.

7. We fully agree on this item and will proceed in that direction in the future.

Spelling

Line 113: mean see level-> mean sea level

Line 751: cloud -> CLOUD

8. The typing errors were corrected.

In addition to the issues above, we also adopted some smaller changes and added a few recent papers as references to further improve the MS.

Imre Salma

Consequences of dynamic and timing properties of new aerosol particle formation and consecutive growth events

Imre Salma and Zoltán Németh

Institute of Chemistry, Eötvös University, H-1518 Budapest, P.O. Box 32, Hungary

Correspondence to: Imre Salma (salma@chem.elte.hu)

Abstract. A variety of contributions to the emerging research field of urban atmospheric new particle formation (NPF) and consecutive particle diameter growth based on gradually generating, several-year long, semi-continuous, critically evaluated, complex and coherent data sets are presented here. Dynamic properties, i.e. particle formation rate J_6 and particle diameter growth rate GR_{10} , and timing properties, i.e. starting time (t_1) and duration time interval (Δt) of 247 quantifiable atmospheric NPF and growth events identified in the city centre and near-city background of Budapest over 6 full measurement years together with related gas-phase H_2SO_4 proxy, condensation sink (CS) of vapours, basic meteorological data and concentrations of criteria pollutant gases were derived, evaluated, discussed and interpreted. In the city centre, nucleation ordinarily starts at 09:15 UTC+1, and it is maintained for approximately 3 h. The NPF and growth events produce 4.6 aerosol particles with a diameter of 6 nm in 1 cm³ of air in 1 s, and cause the particles with a diameter of 10 nm to grow with a typical rate of 7.3 nm h⁻¹. Nucleation starts approximately 1 h earlier in the near-city background, it shows substantially smaller J_6 (with a median of 2.0 cm⁻³ s⁻¹) and GR_{10} values (with a median of 5.0 nm h⁻¹), while the duration of nucleation is similar to that in the centre. Monthly distributions of the dynamic properties and daily maximum H_2SO_4 proxy do not follow the mean monthly pattern of the event occurrence frequency. The factors that control the event occurrence and that govern the intensity of particle formation and growth are not directly linked. ~~Condensing atmospheric chemical species and/or their processes in the city centre seem to contribute equally to new particle formation and particle growth. In the near city background, however, chemical compounds available and their processes power particle growth more than particle formation.~~ New particle formation and growth processes advance in a different manner in the city and its close environment. This could mainly be related to diversities in atmospheric composition, chemistry and physics. We showed that there is a minimum growth rate (1.8 nm h⁻¹ is our case) that is required for nucleated particles to reach the lower end of the diameter interval measured (in our case 6 nm). Monthly distributions and relationships among the properties mentioned

32 provided indirect evidence that chemical species other than H₂SO₄ largely influence the particle
33 growth and possibly atmospheric NPF process as well. The J_6 , GR₁₀ and Δt can be described
34 by log-normal distribution function. Most of the extreme dynamic properties could not be
35 explained by H₂SO₄ proxy, CS, meteorological data or pollutant gas concentrations.
36 Approximately 40% of the NPF and growth events exhibited broad beginning, which can be an
37 urban feature. For 9% of all quantifiable days, it was feasible to calculate 2 separate sets of
38 dynamic properties. The later onset frequently shows more intensive particle formation and
39 growth than the first onset by a typical factor of approximately 1.4. The first event is attributed
40 to regional type, while the second event, superimposed on the first, is often associated with sub-
41 regional, thus urban NPF and growth process.

42

43 **1 Introduction**

44

45 Molecules and molecular fragments in the air collide randomly and can form electrically neutral
46 or charged clusters. Most clusters decompose shortly. Chemical stabilising interactions among
47 certain components within a cluster can enhance its lifetime, during which it can grow further
48 by additional molecular collisions through some distinguishable size regimes (Kulmala et al.,
49 2014). If the diameter of these clusters reaches a critical value of 1.5 ± 0.3 nm (Kulmala et al.,
50 2013), they become thermodynamically stable, and their further growth turns into a spontaneous
51 process. Supersaturation is a necessary atmospheric condition for this principal transformation.
52 It is essentially a phase transition, which takes place in a dispersed manner in the atmosphere,
53 so it generates an aerosol system. The newly formed particles grow further by condensation to
54 larger sizes in most cases due to the existing supersaturation. Photochemical oxidation products
55 such as H₂SO₄ (Sipilä et al., 2010), extremely low-volatile organic compounds (ELVOCs, Ehn
56 et al., 2014; Jokinen et al., 2015) and highly oxygenated molecules (HOMs, Bianchi et al., 2016;
57 Kirkby et al., 2016; Tröstl et al., 2016) together with H₂O vapour, NH₃ (Kirkby et al., 2011),
58 amines (Almeida et al., 2013), other oxidation products of volatile organic compounds (VOCs;
59 Metzger et al., 2010; Schobesberger et al., 2013; Riccobono et al., 2014) and NO₂ can play an
60 important role in both the particle formation and growth. The VOCs include compounds of both
61 anthropogenic and biogenic origin, mainly isoprenoids such as α -pinene (Kirkby et al., 2016).
62 In some specific coastal regions, iodine oxides produced from marine biota are involved
63 (O'Dowd et al., 2002). Atmospheric concentration of these key compounds at a level that is
64 smaller by 12–14 orders of magnitude than the concentration of air molecules is already
65 sufficient for the phenomenon (Kulmala et al., 2014). Relative importance of the organics

66 increases with particle size (Riipinen et al., 2011; Ehn et al., 2014), and their supersaturation is
67 maintained by fast gas-phase autooxidation reactions of VOCs (Crouse et al., 2013). The
68 overall phenomenon is ordinarily confined in time for 1 day or so, and, therefore, it can be
69 regarded as an event in time, and is referred as new aerosol particle formation (NPF) and
70 consecutive particle growth event.

71
72 Such events appear to take place almost everywhere in the world and anytime (Kulmala et al.,
73 2004; Kerminen et al., 2018; Nieminen et al., 2018). Their occurrence frequency and, more
74 importantly, their contribution to particle number concentrations were found to be substantial
75 or determinant in the global troposphere (Spracklen et al., 2006; Kulmala et al., 2014).
76 Moreover, their contribution to the number of cloud condensation nuclei (CCN) can be 50% or
77 even more (Makkonen et al., 2009; Merikanto et al., 2009; Sihto et al., 2011), which links the
78 events to climate system, and emphasizes their global relevance (Kerminen et al., 2012;
79 Makkonen et al., 2012; Carslaw et al., 2013; Gordon et al., 2016). New particle formation and
80 growth events were proved to be common in polluted air of large cities as well with a typical
81 relative occurrence frequency between 10% and 30% (Woo et al., 2001; Baltensperger et al.,
82 2002; Alam, et al., 2003; Wehner et al., 2004; Salma et al., 2011; Dall'Osto et al., 2013; Xiao
83 et al., 2015; Zhang et al., 2015; Kulmala et al., 2017, Nieminen et al., 2018). The coupling and
84 relationships between regional and urban (sub-regional) NPF were demonstrated at least under
85 favourable orographic conditions (Salma et al., 2016b). New particle formation can increase
86 the existing particle number concentrations in city centres by a factor of approximately 2 on
87 nucleation days, while it can produce approximately 28% of ultrafine (UF) particles on a longer
88 (e.g. annual) time scale (Salma et al., 2017). Particle concentrations from NPF are also
89 important when compared to the (primary) particles emitted by their dominant source in cities,
90 namely by road vehicles with internal combustion engines (Paasonen et al., 2016). These results
91 jointly suggest that particles from NPF and growth events in cities can influence not only the
92 urban climate but can contribute to the public's excess health risk from particle number
93 exposures (Oberdörster et al., 2005; Braakhuis et al., 2014; Salma et al., 2015), and,
94 furthermore, could be linked to the role of human actions in all these effects.

95
96 Despite these potentials, conclusive interpretation of the data obtained, and results derived
97 specifically for cities remained hindered so far. Several-year long, semi-continuous, critically
98 evaluated, complex and coherent data sets are required for this purpose, which have been
99 generating gradually. As part of this international progress, investigations dedicated to urban

100 NPF and growth events in Budapest have been going on since November 2008. Measurements
101 for 5 full years were realised in the city centre at a fixed location, 1 full year was devoted to
102 measurements in a near-city background environment, and some other measurements were
103 accomplished in different urban microenvironments for time intervals of a few months. The
104 main objectives of this study are to determine, present and analyse the dynamic properties, i.e.
105 particle formation rate and particle diameter growth rate, timing properties, i.e. starting time
106 and duration time interval of nucleation process of NPF and growth events together with the
107 major sources and sink of condensing vapours, basic meteorological data and criteria pollutant
108 gases for 6 years, to investigate and interpret their relationships, to discuss their monthly
109 distributions, to evaluate and detect some of their features specific for urban atmospheric
110 environments, and to demonstrate some specific urban influence on the calculation of the
111 properties. These quantities and relationships are of basic importance in many atmospheric
112 processes for several reasons. **Our goals are in line with the research needs for global
113 atmospheric nucleation studies (Kerminen et al., 2018; Nieminen et al., 2018).**

114

115 **2 Experimental methods**

116

117 The measurements took place at two urban locations in Budapest, Hungary. Most measurements
118 were realised at the Budapest platform for Aerosol Research and Training (BpART) facility (N
119 47° 28' 29.9", E 19° 3' 44.6", 115 m above mean sea level (a.s.l.; Salma et al., 2016a). This site
120 represents a well-mixed, average atmospheric environment for the city centre. The other
121 location was situated at the NW border of Budapest in a wooded area of the Konkoly
122 Astronomical Observatory of the Hungarian Academy of Sciences (N 47° 30' 00.0", E 18° 57'
123 46.8", 478 m a.s.l.). This site characterises the air masses entering the city since the prevailing
124 wind direction in the area is NW. The measurements were accomplished for 6 full-year long
125 time intervals, i.e. from 03–11–2008 to 02–11–2009, from 19–01–2012 to 18–01–2013, from
126 13–11–2013 to 12–11–2014, from 13–11–2014 to 12–11–2015, from 13–11–2015 to 12–11–
127 2016 and from 28–01–2017 to 27–01–2018. In the measurement year 2012–2013, the
128 instruments were set up in the near-city background, while in all other years, they were installed
129 in the city centre. Local time (LT=UTC+1 or daylight-saving time, UTC+2) was chosen as the
130 time base of the data unless otherwise indicated because **it had been observed in earlier
131 investigations that** the daily activity time pattern of inhabitants substantially influences the
132 atmospheric processes in cities (Salma et al., 2014).

133

134 The main measuring system was a flow-switching type differential mobility particle sizer
135 (DMPS). It consists of a radioactive (^{60}Ni) bipolar charger, a Nafion semi-permeable membrane
136 dryer, a 28-cm long Vienna-type differential mobility analyser and a butanol-based
137 condensation particle counter (TSI, model CPC3775). The sample flow was 2.0 L min^{-1} in the
138 high-flow mode, and 0.31 L min^{-1} in the low-flow mode with sheath air flow rates 10 times
139 larger than for the sample flows. The DMPS measures particle number concentrations in an
140 electrical mobility diameter range from 6 to 1000 nm in the dry state of particles (with a relative
141 humidity of $\text{RH} < 30\%$) in 30 channels, which finally yields 27 channels after averaging 3
142 overlapping channels when joining the data for the 2 flow modes. The time resolution of the
143 measurements was approximately 10 min till 18-01-2013, and 8 min from 13-11-2013 (after
144 a planned update of the DMPS system). There was no upper size cut-off inlet applied to the
145 sampling line, and a weather shield and insect net were only attached. The sampling inlet was
146 installed at a height of 12.5 m above the street level in the city centre, and of approximately 1.7
147 m in the near-city background. The measurements were performed according to the
148 international technical standard (Wiedensohler et al., 2012). The availability of the DMPS data
149 over 1-year long time intervals are summarised in Table 1. Synoptic meteorological data for air
150 temperature (T), RH, wind speed (WS) and wind direction (WD) were obtained from a
151 measurement station of the Hungarian Meteorological Service (HMS, no. 12843) by
152 standardised methods with a time resolution of 1 h. Global solar radiation (GRad) data were
153 measured by the HMS at a distance of 10 km in E direction with a time resolution of 1 h.
154 Meteorological data were available in >90% of the possible cases in each year. Concentrations
155 of SO_2 , O_3 , NO_x and CO were obtained from measurement stations of the National Air Quality
156 Network in Budapest (in a distance of 4.5 km from the urban site, and of 6.9 km from the near-
157 city background site) located in the upwind prevailing direction from the measurement sites.
158 They are measured by UV fluorescence (Ysselbach 43C), UV absorption (Ysselbach 49C),
159 chemiluminescence (Thermo 42C) and IR absorption methods (Thermo 48i), respectively with
160 a time resolution of 1 h. The concentration data were available in >85% of the yearly time
161 intervals, and >98% of them were above the limit of determinations. It is worth mentioning that
162 concentration of SO_2 in the Budapest area is ordinarily distributed without larger spatial (and
163 temporal) gradients (Salma et al., 2011). For the present study, this was proved by evaluating
164 the concentration ratios from 2 different municipal stations which are in the closest distance
165 from the BpART in 2 different directions with an angle of 60° between them. The mean SO_2
166 concentration ratio and standard deviation (SD) for the 2 stations were $81 \pm 20\%$ over the 5-year

167 long measurement time interval. The assumption can also be justified indirectly by a conclusion
168 on the monthly distribution of SO₂ concentration in Sect. 4.2.

169

170 **3 Data treatment**

171

172 The measured DMPS data were evaluated according to the procedure protocol recommended
173 by Kulmala et al. (2012) with some refinements that are related to urban features (see Sect. 3.1).
174 Particle number concentrations in the diameter ranges from 6 to 1000 nm (N), from 6 to 25 nm
175 (N_{6-25}) and from 6 to 100 nm (N_{6-100} or UF particles) were calculated from the measured and
176 inverted DMPS concentrations. Particle number size distribution surface plots showing jointly
177 the variation in particle diameter and particle number concentration density in time were also
178 derived. Identification and classification of NPF and growth events was accomplished on these
179 surface plots (Dal Maso et al., 2005; Németh et al., 2018) on a daily basis into the following
180 main classes: NPF event days, non-event days, days with undefined character, and days with
181 missing data (for more than 4 h during the midday). Relative occurrence frequency of events
182 was determined for each month and year as the ratio of the number of event days to the total
183 number of relevant (i.e. all–missing) days. A subset of NPF events with uninterrupted evolution
184 in time, which are called quantifiable (class 1A) events, were further separated because the time
185 evolution of their size distribution functions was utilised to determine the dynamic and timing
186 properties with good accuracy and reliability.

187

188 **3.1 Dynamic and timing properties**

189

190 Growth rate (GR) of nucleation-mode particles was calculated by mode-fitting method
191 (Kulmala et al., 2012). Particle number median mobility diameter (NMMD) of the nucleation
192 mode were obtained from fitting the individual size distributions by DoFit algorithm (Hussein
193 et al., 2004). The growth rate was determined as the slope of the linear line fitted to the time
194 series of the NMMD data within a time interval around a diameter d , where the dependency
195 could be satisfactorily approximated by linear fit. Since the nucleation mode was mostly
196 estimated by N_{6-25} in the calculations of the formation rate (see below), and since the median
197 of the related diameter interval (from 6 to 25 nm) is close to $d=10$ nm, GRs for particles with a
198 diameter of 10 nm were determined (GR₁₀). This type of GR can be interpreted as an average
199 GR as far as the given particle diameter range is concerned, but it actually expresses the

200 beginning of the growth process only, which may have considerable effects on the formation
201 rate calculations in specific cases (see later).

202

203 Time evolution of an aerosol population is described by the general dynamic equation which
204 was rearranged, simplified and approximated by several quantities (Kulmala et al., 2001; Dal
205 Maso et al., 2002; Kulmala et al., 2012; Cai and Jiang, 2017) to express the formation rate J_6
206 of particles with the smallest detected diameter of $d_{\min}=6$ nm in a form utilised in the present
207 evaluation as

208

$$209 \quad J_6 = \frac{dN_{6-25}}{dt} + \text{CoagS}_{10} N_{6-25} + \frac{\text{GR}_{10}}{(25-6)} N_{6-25} - \frac{dN_{\text{Ai}, <25}}{dt}. \quad (1)$$

210

211 The first term on the right side of Eq. 1 expresses the concentration increment. The particle
212 number concentration in the size range from 6 to 25 nm (i.e. N_{6-25}) was selected to approximate
213 the nucleation-mode particles $N_{\text{nuc}} \approx N_{6-25}$. This is a usual and reasonable choice because it was
214 proved to be advantageous and effective way in handling fluctuating data sets since N_{6-25} often
215 exhibits less sensitivity and smaller scatter in time than the fitted area of the nucleation mode.
216 It is implicitly assumed that the intensity of the NPF is constant for a certain time interval, and,
217 therefore, dN_{6-25}/dt can be determined as the slope of the linear function of N_{6-25} versus time t
218 within an interval where the dependence could be satisfactorily approximated by linear fit. A
219 limitation of the relatively wide size range (6–25 nm) selected can be manifested by
220 disturbances from primary particles particularly in urban environments. This is taken into
221 account by the last term of Eq. 1 and is discussed later.

222

223 The second term on the right side of Eq. 1 represents the loss of particles due to coagulation
224 scavenging (with pre-existing particles). The coagulation scavenging efficiency for particles
225 with a diameter of 10 nm (CoagS_{10}) was selected to approximate the mean coagulation
226 efficiency of nucleation-mode particles ($\text{CoagS}_{\text{nuc}}$). This diameter was chosen by considering
227 the median of the related diameter range, which was discussed above for GR. The coagulation
228 efficiency was calculated from classical aerosol mechanics with adopting a mass
229 accommodation coefficient of 1 and utilizing the Fuchs' transition-regime correction factor
230 (Kulmala et al., 2001; Dal Maso et al., 2005; Kulmala et al., 2013) by using computation scripts
231 developed at the University of Helsinki. Self-coagulation within the nucleation mode was

232 neglected due to its limited concentration. Hygroscopic growth of particles was not considered
233 since this depends on chemical composition of particles, which is unknown.

234

235 The third term on the right side of Eq. 1 expresses the particle growth out of the considered size
236 range by condensation of vapours. The GR_{10} was selected to approximate a representative value
237 at the median of the particle diameter range considered (Vuollekoski et al., 2012). It is implicitly
238 assumed that GR_{10} can be regarded to be constant over the time interval under consideration.
239 Nevertheless, the growth of nucleation-mode particles in time is occasionally limited (Fig. S1a).
240 In these specific cases, the mean relative area of the nucleation mode below 25 nm was
241 determined by fitting the individual size distributions around the time of the maximum
242 nucleation-mode NMMD, and the ratios were averaged. A correction in form of the mean
243 relative area was adopted as a multiplication factor for the growth out term in Eq. 1. On very
244 few days, the growth of newly formed particles was followed by a decrease in nucleation-mode
245 NMMD (Salma et al., 2016a). In these cases, the shrinkage rate (with a formal $GR_{10}<0$) was
246 derived and adopted in Eq. 1.

247

248 The fourth term on the right side of Eq. 1 expresses the contribution of high-temperature
249 emission sources, usually of vehicular road traffic (Paasonen et al., 2016; Salma et al., 2017) to
250 N_{6-25} , which can provisionally disturb the assumption of $N_{nuc}\approx N_{6-25}$. A typical example of such
251 a situation is shown in Fig. S1b from 10:09 to 12:23 LT. In these specific cases, the contribution
252 of primary emissions was estimated from the slope of the time series of the fitted peak area of
253 the Aitken mode below $d<25$ nm ($N_{Ai,<25}$) in the time region under consideration. Reliable
254 separation of the nucleation and Aitken modes from each other was hindered or was not possible
255 for a few individual size distributions due to overlapping modes and the scatter in the
256 concentration data, and these individual Aitken-mode areas were excluded from or skipped in
257 the time series. Relative contributions of the concentration increment coagulation loss and
258 growth out from the diameter interval to J_6 are decreasing in this order with mean values of
259 71%, 17% and 12%, respectively (Table S1).

260

261 The formation and growth rates for the measurement years of 2008–2009 and 2012–2013 were
262 calculated earlier by a slightly different way and neglecting the urban features discussed above
263 (Salma et al., 2011, 2016b). To obtain consistent data sets, the dynamic properties for these 2
264 years were re-evaluated by adopting the present improved protocol and implementing the
265 experience gained over the years. The mean new-to-old rate ratios with SDs for the GR_{10} and

266 J_6 were 1.06 ± 0.32 and 1.23 ± 0.37 , respectively in the city centre (2008–2009) and 1.04 ± 0.21
267 and 1.20 ± 0.35 , respectively in the near-city background (2012–2013). It was the smaller rates
268 that were primarily and sometimes substantially impacted. The modifications were
269 simultaneously adopted. The subtraction of particle number concentrations emitted by road
270 traffic from N_{6-25} , which usually leads to a decrease in the coagulation loss and loss due to
271 growth out from the diameter range of 6–25 nm, and which can influence the slope of the
272 concentration change in time (dN_{6-25}/dt) in a positive or negative manner depending on the
273 actual time evolution of perturbing emission source. In addition to that, the time interval in
274 which this slope is considered to be constant was set in a new treatment. It is mentioned that
275 the relative contributions of the concentration increment, coagulation loss and growth out from
276 the diameter interval to J_6 have different weights in propagating their effects. Furthermore, J_6
277 itself also depends on the GR_{10} , which makes the relationships even more complex. These
278 explain why the changes resulted in overall increments. The re-calculation is considered as a
279 methodological improvement over the years of research.

280

281 The assumptions and estimations above usually represent a reasonable approximation to reality.
282 The N_{6-25} is derived from the experimental data in a straightforward way, the GR_{10} and the
283 corrections for primary particles and limited particle growth depend on the quality of the size
284 distribution fitting as well, while the $CoagS_{10}$ is determined by using a theoretical model. The
285 resulting accuracies of the dynamic properties, in particular of J_6 , look rather complicated. They
286 also depend on the spatial heterogeneity in the air masses measured particularly for the
287 observations performed at the fixed site, size and time resolution of the concentrations
288 measured, diameter range of the size distributions, fluctuations in the experimental data,
289 selection of the particle diameter interval, choice of the time interval of interest (for linear fits),
290 sensitivity of the models to the input uncertainties (Vuollekoski et al., 2012), and also on the
291 extent of the validity of the assumptions applied under highly polluted conditions (Cai and
292 Jiang, 2017). The situation is further complicated with the fact that the dynamic (and also the
293 timing) properties are connected to each other. Finally, it is important to recognise that some
294 NPF and growth curves on the surface plots have rather broad starting time interval (Fig. S1a
295 and S1c). They occur in a considerable abundance in cities, e.g. in 40% of all quantifiable events
296 in Budapest (Sect. 4.4). This may yield badly defined or composite dynamic properties, whose
297 uncertainty can have principle limitations which can prevail on the experimental and model
298 uncertainties.

299

300 Timing properties of NPF and growth events are increasingly recognised, and they can provide
301 valuable information even if they are estimated indirectly from the observed diameter interval
302 >1.5 nm (Sect. 1). The earliest estimated time of the beginning of a nucleation (t_1) and the latest
303 estimated time of the beginning of a nucleation (t_2) were derived by a comparative method
304 (Németh and Salma, 2014) based on the variation in the content of the first size channel of the
305 DMPS system. Both time parameters include a time shift that accounts for the particle growth
306 from the stable neutral cluster mode at approximately 2 nm to the smallest detectable diameter
307 limit of the DMPS systems (6 nm in our case) by adopting the GR value in the size window
308 nearest to it in size space. The difference $\Delta t=t_2-t_1$ is considered as the duration time interval of
309 the nucleation process. It represents the time interval during which new aerosol particles are
310 generated in the air. **The timing properties are expressed in UTC+1**, and their uncertainty is
311 regarded to be ca. 30 min under ordinary NPF and growth situations.

312

313 **3.2 Sources and sink**

314

315 Relative effects and role of gas-phase H_2SO_4 were estimated by its proximity measure (proxy
316 value) containing both its major source and sink terms under steady-state conditions according
317 to Petäjä et al. (2009). It was calculated for $\text{GRad} > 10 \text{ W m}^{-2}$. Formally, it is possible to convert
318 the H_2SO_4 proxy values to H_2SO_4 concentrations by an empirical scaling factor of $k=1.4 \times 10^{-7}$
319 $\times \text{GRad}^{-0.70}$, where GRad is expressed in a unit of W m^{-2} (Petäjä et al., 2009). The factor was,
320 however, derived for a remote boreal site, and, therefore, we prefer not to perform the
321 conversion since urban areas are expected to differ from the boreal regions, and adopting the
322 factor could distort the dynamic relationships or time trends investigated. The conversion was
323 applied only to estimate the order of average H_2SO_4 atmospheric concentration levels. The
324 results derived by utilising the proxy are subject to larger uncertainties than for the other
325 properties because of these limitations, but they may indicate well gross tendencies.

326

327 Condensation sink for vapour molecules onto the surface of existing aerosol particles was
328 computed for discrete size distributions as described in earlier papers (Kulmala et al., 2001; Dal
329 Maso et al., 2002, 2005) and summarised by Kulmala et al. (2013). The equilibrium vapour
330 pressure of the condensing species was assumed to be negligible at the surface of the particles,
331 so similar to sulfuric acid. Dry particle diameters were considered in the calculations.

332

333 **4 Results and discussion**

334

335 Annual median particle number concentrations based on the individual data in the near-city
336 background in 2012–2013, and in the city centre for the separate measurement years of 2008–
337 2009, 2013–2014, 2014–2015, 2015–2016 and 2017–2018 were 3.4×10^3 , and 11.5×10^3 ,
338 9.7×10^3 , 9.3×10^3 , 7.5×10^3 and $10.6 \times 10^3 \text{ cm}^{-3}$, respectively. The data for the city centre indicate
339 a decreasing trend. **The first 4 values unambiguously show a decrease, while the last data point**
340 **may look somewhat differently. Rigorous statistical evaluation of the joint data set of particle**
341 **number concentrations in various size fractions over a decennial time interval from November**
342 **2008 to November 2018 is in progress, and its preliminary results in the one hand, confirm the**
343 **decreasing tendency, and in the other hand, reveal some fine structure to this dependency.** The
344 mean UF/N ratio with SD for the same measurement time intervals were $67 \pm 14\%$, and $79 \pm 6\%$,
345 $75 \pm 10\%$, $75 \pm 11\%$, $76 \pm 11\%$ and $80 \pm 10\%$, respectively. The values correspond to ordinary
346 urban atmospheric environments in Europe (Putaud et al., 2010).

347

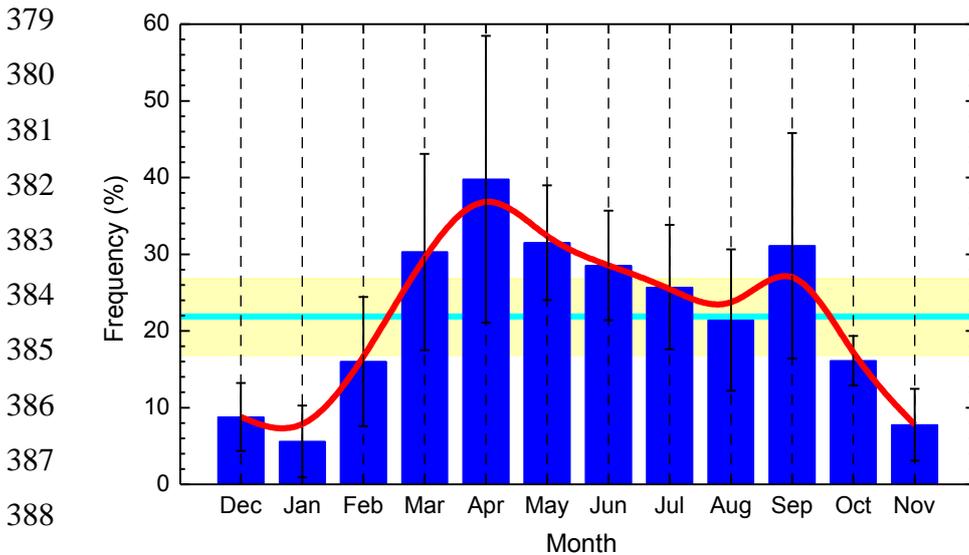
348 An overview on the number of classified days separately for the 1-year long measurement time
349 intervals is given in Table 1. The availability of the daily size distribution surface plots with
350 respect to all days ensures that the data are representative on yearly and monthly time scales,
351 except for the months August and September 2015, when there were missing days in larger
352 ratios. The number of quantifiable event days (248 cases) is also considerable, which establishes
353 to arrive at firm conclusion on the NPF and growth events as well.

354

355 **Table 1.** Number of days with new aerosol particle formation and growth event, quantifiable (class 1A)
 356 event days, non-event days, undefined days, missing days and the coverage of relevant days with respect
 357 to all days in the near-city background and city centre separately for the 1-year long measurement time
 358 intervals.
 359

Environment	Background		Centre				
	Time interval	2012–2013	2008–2009	2013–2014	2014–2015	2015–2016	2017–2018
Event days		96	83	72	81	35	83
Quantifiable days		43	31	48	56	18	52
Undefined days		19	34	24	25	8	23
Non-event days		231	229	267	240	226	257
Coverage (%)		95	95	99	95	73	99
Missing days		20	19	2	19	97	2

360
 361
 362 It was previously shown that the NPF and growth events observed in the city centre of Budapest
 363 and its background ordinarily happen above a larger territory or region in the Carpathian Basin
 364 (Salma et al., 2011; Németh and Salma, 2014), and they are linked to each other (Salma et al.,
 365 2016b). From the point of the occurrence frequency distribution, they can, therefore, be
 366 evaluated jointly in the first approximation. An overall monthly mean relative occurrence
 367 frequency of nucleation days derived for all 6 measurement years is shown in Fig. 1. The annual
 368 mean frequency with SD was $22\pm 5\%$, which is considerable and is in line with other urban sites
 369 (Sect. 1). The monthly mean frequency has a temporal variation, which can be characterised by
 370 a noteworthy pattern. The mean monthly dependency exhibits an absolute and a local minimum
 371 in January (5.6%) and August (21%), respectively, and an absolute and a local maximum in
 372 April (40%), and September (31%), respectively. Nevertheless, the SDs of the monthly (and
 373 annual) means indicate prominent variability from year to year. The pattern can be related to
 374 multivariate relationships and complex interplay among the influencing factors, which include
 375 the air temperature (January is the coldest month, while August is the warmest month in the
 376 Carpathian Basin) and enhanced emission of biogenic VOCs in springtime (March–April) and
 377 early autumn (September) as well (Salma et al., 2016b).
 378



379
380
381
382
383
384
385
386
387
388
389
390 **Figure 1.** Monthly mean relative occurrence frequency of new aerosol particle formation and
391 consecutive particle diameter growth events with respect to the number of relevant days for the joint 6-
392 year long data set. The error bars show ± 1 standard deviation, the horizontal line in cyan indicates the
393 overall annual mean frequency, the yellow bands represent ± 1 standard deviation of the annual mean,
394 and the smooth curve in red serves to guide the eye.

395
396 The properties and variables studied were derived in full time resolution. They were averaged
397 in several ways for different conditions and for various purposes to obtain typical average
398 descriptive characteristics. In 1 case (31-08-2016), the NPF and growth event could reliable
399 be identified, while the measured absolute particle number concentrations could not be
400 validated due to some experimental troubles, and, therefore, it was left out from the further
401 calculations. Similarly, there were 1 and 4 events with unusually/extraordinarily large dynamic
402 properties in the measurement years 2014-2015 and 2017-2018, respectively. More
403 specifically, 5 individual J_6 data when expressed in a unit of $\text{cm}^{-3} \text{s}^{-1}$ and 1 individual GR_{10}
404 data when given in nm h^{-1} were >20 (Table 3). These extremes were left out from the overview
405 statistics to maintain the representativity (they could be influenced by some unknown extra or
406 very local sources) and to fulfil better the basic requirements of correlation analysis. If an event
407 showed a double beginning then the dynamic properties for the first onset were considered in
408 the basic overview since this onset is of regional relevance (Salma et al., 2016b). The extreme
409 NPF and growth events and the characteristics for the second onsets were, however, evaluated
410 separately and are discussed in detail and interpreted in Sect. 4.4.

411

412 **4.1 Ranges and averages**

413

414 Ranges and averages with SDs of formation rate J_6 , growth rate GR_{10} , starting time of
415 nucleation (t_1) and duration time interval of nucleation (Δt) are summarised in Table 2 for
416 separate measurement years and for the joint 5-year long city centre data set. In the city centre,
417 nucleation generally starts at 09:15 UTC+1, and it is typically maintained for approximately 3
418 h. The NPF and growth events ordinarily produce 5.6 new aerosol particles with a diameter of
419 6 nm in 1 cm³ of air in 1 s, and cause the particles with a diameter of 10 nm to grow with a
420 typical rate of 7.6 nm h⁻¹. The statistics for J_6 and GR_{10} are based on 199 and 203 events,
421 respectively. The corresponding data for the separate years show considerable variability
422 without obvious trends or tendencies. The differences between the years can likely be related
423 to changes in actual atmospheric chemical and physical situations and conditions, and to the
424 resulting modifications in the sensitive balance and delicate coupling among them from year to
425 year. Spread of the individual data for GR_{10} is smaller than for J_6 ; the relative SDs for the joint
426 5-year long city centre data set were 38% and 68%, respectively, while the (external) relative
427 SDs calculated from the annual mean values were 4.2% and 14.0%, respectively.

428

429 The dynamic properties and t_1 data tend to be smaller in the near-city background than in the
430 city centre. In general, nucleation starts 1 h earlier in the background, and the events typically
431 show significantly smaller J_6 (with a median of 2.0 cm⁻³ s⁻¹) and GR_{10} (with a median of 5.0
432 nm h⁻¹). Duration of the nucleation is very similar to that in the city centre. All starting times
433 of nucleation were larger than (in a few cases, very close to) the time of the sunrise. This implies
434 that no nocturnal NPF and growth event has been identified in Budapest so far. The particle
435 growth process (the so-called banana curve) could be traced usually for a longer time interval
436 (up to 1.5 d) in the background than in the centre.

437

438 These results are in line with the ideas on atmospheric nucleation and consecutive particle
439 growth process (e.g. Kulmala et al., 2014; Zhang et al., 2015; Kerminen et al., 2018). It was
440 observed in a recent overview study (Nieminen et al., 2018) that the formation rate of 10–25
441 nm particles increased with the extent of anthropogenic influence, and in general, it was 1–2
442 orders of magnitude larger in cities than at sites in remote and clean environments. This
443 highlights the importance of some anthropogenic vapours such as SO₂, NH₃ and amines to NPF
444 and growth. The data also confirm our earlier findings with respect to Budapest and its regional

445 background within the Carpathian Basin achieved with shorter, 2-year long data sets (Salma et
 446 al., 2016b)

447
 448 **Table 2.** Ranges, averages and standard deviations of aerosol particle formation rate J_6 , particle diameter
 449 growth rate GR_{10} , starting time (t_1) and duration time interval ($\Delta t=t_2-t_1$) of nucleation process of
 450 quantifiable (class 1A) new particle formation and growth events in the near-city background and city
 451 centre separately for the 1-year long measurement time intervals and for the joint 5-year long city centre
 452 data set.

453

Environment	Background		Centre				
	2012– 2013	2008– 2009	2013– 2014	2014– 2015	2015– 2016	2017– 2018	All 5 years
Formation rate J_6 ($\text{cm}^{-3} \text{s}^{-1}$)							
Minimum	0.48	1.47	1.13	0.81	1.19	1.60	0.81
Median	2.0	4.2	3.5	4.4	4.6	6.3	4.6
Maximum	5.6	15.9	17.8	18.0	15.3	17.3	18.0
Mean	2.2	4.7	5.2	5.6	5.0	6.6	5.6
St. deviation	1.3	2.6	3.7	4.2	3.7	3.3	3.8
Growth rate GR_{10} (nm h^{-1})							
Minimum	3.0	3.7	3.1	2.8	3.2	3.3	2.8
Median	5.0	7.6	6.6	6.5	8.0	7.5	7.3
Maximum	9.8	17.4	19.0	18.0	15.5	19.8	19.8
Mean	5.2	7.8	7.2	7.3	7.7	8.0	7.6
St. deviation	1.4	2.6	2.8	3.2	3.0	2.8	2.9
Starting time, t_1 (HH:mm UTC+1)							
Minimum	05:51	07:14	06:44	05:48	07:31	05:57	05:48
Median	08:19	09:26	09:22	08:48	09:45	09:18	09:15
Maximum	11:09	11:38	12:21	11:23	12:45	12:15	12:45
Mean	08:17	09:27	09:25	08:49	10:02	09:24	09:19
St. deviation	01:11	01:05	01:26	01:22	01:23	01:36	01:26
Duration time, Δt (HH:mm)							
Minimum	01:23	00:52	00:42	00:31	01:03	01:26	00:31
Median	03:16	02:36	02:04	03:53	02:31	03:49	02:57
Maximum	06:44	06:04	05:34	07:46	06:05	07:55	07:55
Mean	03:30	02:44	02:14	03:52	02:58	03:57	03:18
St. deviation	01:40	01:11	01:01	01:40	01:47	01:39	01:40

454
 455 Ranges and averages with SDs of some related atmospheric properties, namely of mean CS
 456 averaged for the time interval from t_1 to t_2 , daily maximum gas-phase H_2SO_4 proxy, daily mean
 457 T and RH (Table S2), and of daily median concentrations of SO_2 (as the major precursor of gas-

458 phase H₂SO₄), O₃ (as an indicator of photochemical activity), NO_x and CO gases (as indicators
459 of anthropogenic combustion activities and road vehicle emissions) (Table S3) were also
460 derived for quantifiable NPF and growth event days, and are further evaluated. The annual
461 mean CS values exhibited decreasing tendency in the city centre over the years (as can be
462 expected from the particle number concentrations as well). The individual values remained
463 below approximately $20 \times 10^{-3} \text{ s}^{-1}$, which agrees well with the results of our earlier study (Salma
464 et al., 2016b) according to which the CS suppresses NPF above this level in the Carpathian
465 Basin. Maximum H₂SO₄ proxy values reached substantially higher levels (by a factor of
466 approximately 2) in the near-city background than in the city centre due mainly to the
467 differences in the CS and [SO₂]. The differences between the 2 sites are particularly evident
468 when considering their smallest values. The largest variability in the annual average values
469 were observed for the proxy. Median concentration of H₂SO₄ molecules was roughly estimated
470 to be approximately $5 \times 10^5 \text{ cm}^{-3}$ by adopting the scaling factor (Sect. 3.2). ~~The dynamic~~
471 ~~properties seem to be not very sensitive to air T~~ The air *T* displayed quite similar and comparable
472 values over the years at both sites. **The discussion of its overall effect on the dynamic properties**
473 **is accomplished in Sec. 4.2, where the monthly distributions are presented.** Some events
474 happened at daily mean temperatures below zero. The daily mean RH and its SD for the city
475 centre and near-city background were $54 \pm 11\%$ and $64 \pm 12\%$, respectively. There were events
476 that occurred at RHs as high as 90%. Relationships of the dynamic properties with *T* and RH
477 are also obscured with strong seasonal cycle of these meteorological data and with the fact that
478 air masses arriving to the receptor site in different trajectories are often characterised by distinct
479 levels of meteorological data.

480

481 As far as the pollutant gases are concerned (Table S3), SO₂ showed somewhat smaller daily
482 median values, and O₃ exhibited substantially smaller levels on event days in the city centre
483 than in the near-city background, while concentrations of NO_x and CO were obviously larger
484 in the city than in its close background. The differences can primarily be explained by the
485 intensity and spatial distribution of their major sources and atmospheric chemical reactions, and
486 the joined concentration data resembles typical situations without photochemical smog
487 episodes in cities. There was no obvious decrease in SO₂ concentration during these years in
488 contrast with an earlier decreasing trend from mid-1980s till about 2000. ~~No evident or sensitive~~
489 ~~effect of atmospheric gases on the dynamic or timing properties could be deduced from the~~
490 ~~averaged data. This can probably be explained by a dedicated balance between the intensifying~~

491 ~~and suppressing effects, which were averaged out on a yearly time scale. Relationships on~~
492 ~~shorter scales are further investigated and discussed in more detail in the following sections.~~

493

494 **4.2 Monthly distributions**

495

496 Distributions of the monthly mean J_6 , GR₁₀, daily maximum gas-phase H₂SO₄ proxy, mean CS,
497 daily mean air T and RH, and daily median SO₂, O₃, NO_x and CO concentrations for quantifiable
498 NPF and growth events for the joint 5-year long city centre data sets are shown in Fig. 2. The
499 distributions – eminently for J_6 , GR₁₀, H₂SO₄ proxy and SO₂ – do not follow the monthly pattern
500 of the event occurrence frequency at all (cf. Fig. 1). Instead, the J_6 , GR₁₀ and H₂SO₄ proxy tend
501 to exhibit larger values in summer months, and they temporal changes over the other months
502 are smooth and do not show distinctive features. The elevations are substantial; the estimated
503 maximum level was larger than the baseline by a factor of 2.1 for the J_6 , and by a factor of
504 approximately 1.4 for the GR₁₀ and H₂SO₄ proxy. The intensity of solar radiation at the surface,
505 its seasonal cycling, concentration of atmospheric precursors in different months, biogenic
506 processes, anthropogenic activities and the fact that rate coefficients of many thermal
507 chemical/physicochemical processes in the nature (including GR, Paasonen et al., 2018)
508 increase with T could play an important role in explained the distributions. A more
509 comprehensive study involving chemicals and their photochemistry is required for more
510 detailed explanation.

511

512 **It is worth mentioning that [SO₂] did not change substantially for the NPF event and non-event**
513 **days, while GRad was typically larger by a factor of ca. 2 and CS was smaller by approximately**
514 **30% on event days than on non-event days. The differences in the GRad (and some other**
515 **properties) are, however, biased by the seasonal cycle of solar electromagnetic radiation via the**
516 **seasonal variation of NPF occurrence frequency, while the CS indicated a modest seasonal**
517 **dependency. Interpretation of their joint effect should be approached by care, requires further**
518 **evaluations and is to be realised fully in a further study.** Nevertheless, the misalignment among
519 the monthly distributions of NPF and growth event occurrence frequency and all the other
520 properties indicates that the occurrence or its basic causes are not linked with the dynamic
521 properties in a straightforward or linear manner, ~~and 2) gas-phase H₂SO₄ does not seem to be~~
522 ~~the controlling factor of NPF occurrence in the Carpathian Basin including Budapest.~~

523

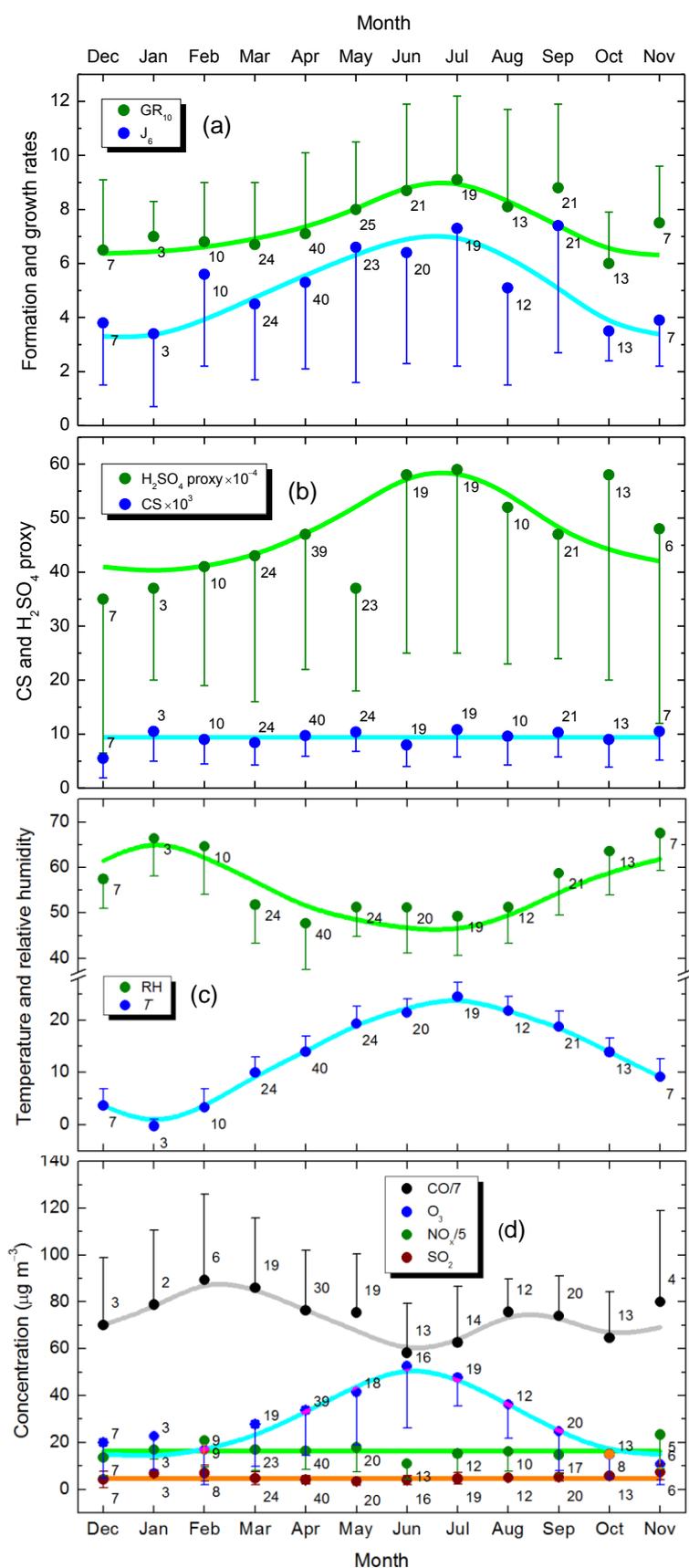
524 Some of our results are in line with other observations according to which GR exhibited almost
525 exclusively a summer maximum, while some other finding are different in the sense that the
526 seasonal variability in particle formation rate was quite modest and could not be established
527 earlier (Nieminen et al., 2018). There is one more aspect which may be worth realising in this
528 respect. A large fraction of compounds contributing to NPF and growth in cities can originate
529 from anthropogenic precursors (Vakkari et al., 2015). Their emissions may peak any time of
530 year depending on human habits and requirements (Nieminen et al., 2018). Nevertheless, the
531 fact that our monthly distributions of the dynamic properties in urban environments follow the
532 universal summer maximum behaviour may indicate the overall prevailing role of atmospheric
533 photochemistry coupled with biogenic emissions of aerosol precursor vapours.

534

535 The monthly mean J_6 , GR_{10} and H_2SO_4 proxy data still have considerable uncertainty, which
536 makes their interpretation not yet completely conclusive. The uncertainties are influenced by
537 inherent fluctuations in the primary data sets, enhancing effects caused by combining some
538 individual primary data into compound variables (such as H_2SO_4 proxy), number of data items
539 available for different properties and months, variations in other or unknown relevant
540 environmental conditions, and by the variability in relative nucleation occurrence frequency
541 from year to year. The resulting uncertainties are expected to decrease with the length of the
542 available data sets, which emphasized the need to continue the measurements.

543

544 The monthly distributions of CS, and SO_2 and NO_x concentrations could be represented by
545 constant values of the overall means and SDs of $(9.4 \pm 4.3) \times 10^{-3} \text{ s}^{-1}$, $4.7 \pm 2.1 \mu\text{g m}^{-3}$ and 81 ± 38
546 $\mu\text{g m}^{-3}$, respectively with an acceptable accuracy. This suggests that CS, SO_2 and NO_x in
547 Budapest do not critically or substantially affect either the dynamic properties (or the event
548 occurrence). Monthly distributions of air T and O_3 concentration showed a maximum over
549 summer months, while RH reflected the T tendency. The monthly variation of T on event days
550 and on non-event days were similar. More importantly, higher biogenic emissions and typically
551 stronger photochemistry are expected during the summer, which both enhance the production
552 rate of nucleating and condensing vapours, while there is practically nothing extra that would
553 suppress the dynamical properties (Kerminen et al., 2018). As a result of these, the dynamic
554 rates show a summer maximum. This is completely consistent with the results from other urban
555 and non-urban studies (Nieminen et al., 2018). Distribution of CO was more changing and
556 without obvious tendentious temporal structure or feature than for the other gases, and,
557 therefore, its interpretation is encumbered so far. However, it doesn't seem to substantially
558 affect the dynamic properties.



560 **Figure 2.** Distribution of
 561 monthly mean aerosol
 562 particle formation rate J_6 in
 563 a unit of $\text{cm}^{-3} \text{s}^{-1}$ and
 564 particle diameter growth
 565 rate GR_{10} in a unit of nm h^{-1}
 566 (a), mean condensation sink
 567 for vapours (CS) in a unit of
 568 s^{-1} averaged over the
 569 nucleation time interval (t_1 ,
 570 t_2) and daily maximum gas-
 571 phase H_2SO_4 proxy in a unit
 572 of $\mu\text{g m}^{-5} \text{W s}$ (b), daily
 573 mean air temperature (T) in
 574 a unit of $^\circ\text{C}$ and daily mean
 575 relative humidity (RH) in %
 576 (c), and daily median
 577 concentrations of SO_2 , O_3 ,
 578 NO_x and CO for
 579 quantifiable (class 1A) new
 580 particle formation and
 581 growth events in the city
 582 centre for the joint 5-year
 583 long time interval. The error
 584 bars are shown for one side
 585 for clarity and indicate 1
 586 standard deviation. Number
 587 of the individual data
 588 averaged in each month is
 589 displayed next to the
 590 symbols. The horizontal
 591 lines indicate the overall
 592 mean. The nonlinear curves
 593 assist to guide the eye.

595 4.3 Relationships

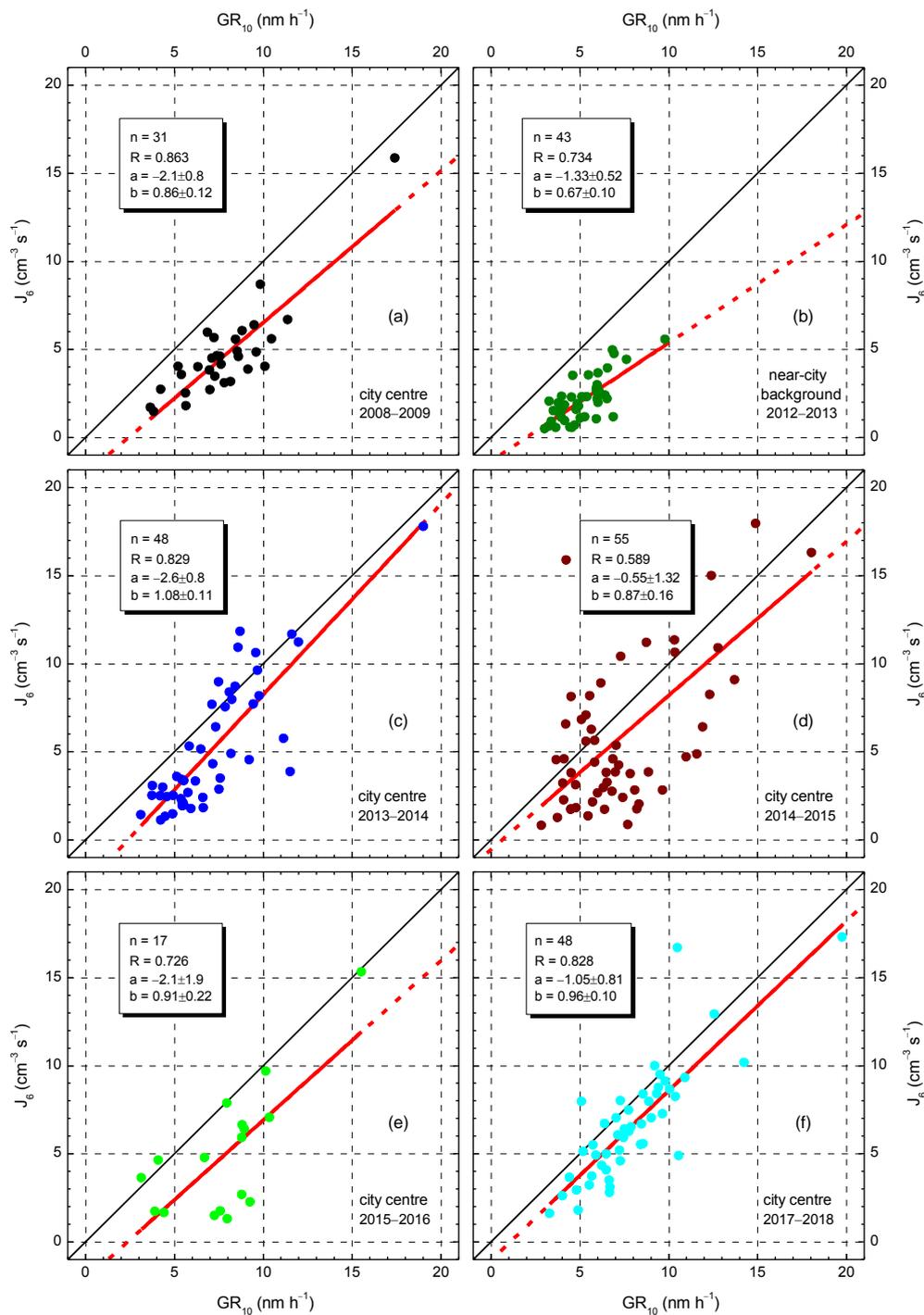
596

597 Pearson's coefficients of correlation (R) between J_6 and GR_{10} revealed significant linear
598 relationship between them for all annual data sets (the mean R and SD were 0.768 ± 0.099 ,
599 number of data pairs $n=243$). This confirms that formation of new aerosol particle and their
600 growth to larger sizes in the atmosphere are tightly and positively linked together. It should be
601 noted that J_6 and GR_{10} are not completely independent variables (see Eq. 1 and Table S1). The
602 linear relationship between the dynamic properties was observed under different atmospheric
603 conditions in many environments (Nieminen et al., 2018). At some sites, this relationship could
604 not be proved due to the weak variability in the variables.

605

606 The dynamic properties can also be coupled to the concentrations of aerosol precursor
607 compounds and properties of a pre-existing particle population, thus to atmospheric
608 environment (Kerminen et al., 2018). It is, therefore, sensible to investigate the city centre and
609 near-city background data separately. Scatter plots between J_6 and GR_{10} for the 1-year long
610 measurement time intervals are shown in Fig. 3. For the city centre, the regression lines follow
611 the line with a slope of 1 in all 5 years. The mean slope (b) with SD for the joint 5-year long
612 city centre data set was $b=0.94 \pm 0.07$ expressed formally in a unit of $\text{cm}^{-3} \text{s}^{-1} \text{nm}^{-1} \text{h}$. This
613 ~~implies that the relevant chemical species and/or their processes in the air of the city centre~~
614 ~~contribute equally to the formation of 6 nm particles and to their growth process.~~ At the same
615 time, the regression line for the near-city background deviated significantly with a $b=0.67 \pm 0.10$
616 $\text{cm}^{-3} \text{s}^{-1} \text{nm}^{-1} \text{h}$ from the J_6 vs. GR_{10} dependency for the city centre. This can imply that NPF
617 and growth processes advance in a different manner in these 2 environments, ~~and that the~~
618 ~~chemical compounds available and their processes power particle growth more than new~~
619 ~~particle formation in the near-city background.~~ This can be related to the differences between
620 the city and its close environment as far as the atmospheric composition (for instance, the VOC
621 and NO_x concentrations), chemistry and physics, and other delicate conditions are concerned
622 (Paasonen et al., 2018). The narrower range and smaller number of individual dynamic
623 properties available for the near-city background relative to those in the city centre represent
624 some inherent limitation or weakness in the explanation, and, therefore, it can strictly be
625 regarded as a working hypothesis because a rigorous statistical treatment would require larger
626 variability in the near-city background data.

627
 628
 629
 630
 631
 632
 633
 634
 635
 636
 637
 638
 639
 640
 641
 642
 643
 644
 645
 646
 647
 648
 649
 650
 651
 652
 653



654 **Figure 3.** Scatter plots for aerosol particle formation rate J_6 and consecutive particle diameter growth
 655 rate GR_{10} in city centre (a and c–f) and near-city background (b) separately for the 1-year long
 656 measurement time intervals. Number of data point (n), their coefficient of correlation (R) and the
 657 intercept (a) and slope (b) of the regression line with standard deviations are also indicated. The lines in
 658 black represent the line with a slope of 1, the solid lines in red show the regression lines, while the
 659 dashed parts in red are extrapolated from the regression line.

660

661 The intercepts (a) of the regression lines were identical for all data sets within their uncertainty
662 interval. The mean intercept and SD were estimated to be $-1.7 \pm 0.8 \text{ cm}^{-3} \text{ s}^{-1}$. This finding is
663 interpreted as the existence of a minimum GR or more exactly of a minimally required GR that
664 leads to $J_6 > 0$. Particles that exhibit at least this level of GR can escape coagulation mainly with
665 larger particles and reach the detectable diameter (6 nm in our case) by condensational growth.
666 The minimal GR was derived as $\text{GR}_{\min} = -a/b$, and its mean and SD are $1.8 \pm 1.0 \text{ nm h}^{-1}$ for the
667 conditions ordinarily present in the Budapest air. Nucleation processes which are initiated under
668 circumstances that cause the newly formed particle with a diameter of 10 nm to grow with a
669 rate $< \text{GR}_{\min}$ are normally not observed. Anyway, these are expected to be events with relatively
670 small J_6 (weak phenomena) due to the relationship between GR_{10} and J_6 . The events with GR
671 larger but close to this limit could be still masked by fluctuating experimental data. Their
672 identification and evaluation can be made feasible by decreasing the lower measurement
673 diameter limit of DMPS systems down to 3 nm, or by different instruments such as particle size
674 magnifier or neutral cluster and air ions spectrometer.

675

676 Correlations between individual H_2SO_4 proxy values on one side and J_6 or GR_{10} on the other
677 side were not significant. This is consistent with the corresponding conclusion of Sect. 4.2 and
678 with the earlier results according to which the mean contribution of H_2SO_4 condensation to the
679 particle GR_{10} was only 12.3% in Budapest (Salma et al., 2016b). The lack of correlation and
680 the average concentrations of SO_2 derived separately for the NPF and growth event and non-
681 event days suggest that this precursor gas is ordinarily available in excess and, therefore, the
682 formation of H_2SO_4 is likely governed by photochemical conditions, and that other chemical
683 species than H_2SO_4 can have larger influence on the particle growth. The role of H_2SO_4 in the
684 nucleation process and early particle growth could be still relevant or determinant.

685

686 Coefficients of correlation between CS on one side and J_6 or GR_{10} on the other side for the joint
687 city centre data sets were modest ($R=0.41$ and 0.32 , respectively with $n=194$ and 197 ,
688 respectively). This is simply related to the fact that larger GR values are typical for polluted
689 urban air (Kulmala et al., 2017) since particles capable of escaping coagulation scavenging need
690 to grow faster in comparison to cleaner environments, and the enhanced requirements for the
691 growth are linked to increased formation rates as well. It should be noted here that the GR of
692 newly formed particles to larger sizes is primarily coupled to 1) CS, which is further linked to
693 the entire aerosol particle population (including the newly formed particles, thus the NPF itself),

694 2) to the total concentration and some physicochemical properties of non-volatile gaseous
695 compounds and 3) to their production rate in the gas phase from aerosol precursor compounds
696 (e.g. Kerminen et al., 2018). These couplings could result in rather complex behaviour, and
697 their understanding is essential when analysing atmospheric observations.

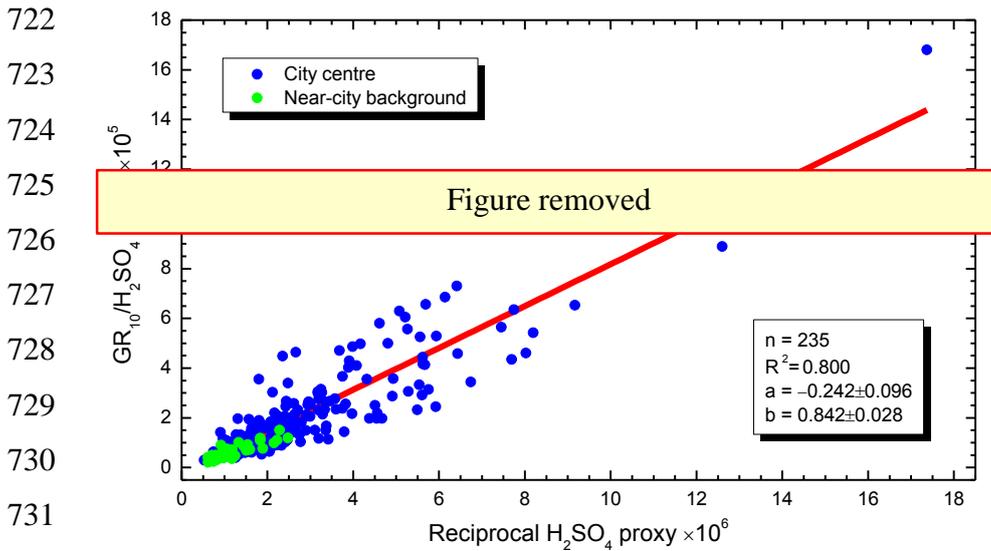
698

699 As far as the pollutant gases are concerned, no correlation could be identified between J_6 or
700 GR_{10} on one side and the gas concentrations on the other side. The coefficients of correlation
701 between CS and NO_x or CO were modest ($R=0.37$ and 0.42 , respectively with $n=164$ and 152 ,
702 respectively), while correlation of NO_x and CO on one side with WS was also modest but
703 negative ($R= -0.32$ and -0.42 , respectively with $n=167$ and 155 , respectively). The former
704 relationships can be explained by the fact that vehicular road traffic in cities is a considerable
705 and common source of NO_x , CO and primary particles (Paasonen et al., 2016), and the emitted
706 particles largely contribute to CS levels. The latter relationships are linked to the effect of large-
707 scale air mass transport (often connected to high WSs) on urban air pollution or air quality.

708

709 ~~Importance and contribution of condensing vapours other than H_2SO_4 are further demonstrated~~
710 ~~in Fig. 4. The data for the city centre and near-city background were presented in a linearized~~
711 ~~form and separately for the 2 sites. Nevertheless, the fitting of the correlation line was~~
712 ~~accomplished for the joint 6 year long data set. It can be demonstrated in particular on non-~~
713 ~~linearized plot of the GR_{10}/H_2SO_4 proxy as function of H_2SO_4 proxy (not shown here) that the~~
714 ~~2 data sets merge into each other without any relevant structure, and, therefore, that they can be~~
715 ~~regarded to be coherent. This approach seems sensible when considering also the limited~~
716 ~~accuracy of the values. The relationship between the 2 composite variables in Fig. 4 was~~
717 ~~significant ($R^2=0.800$, $p<0.05$). It can be interpreted as the increasing atmospheric concentration~~
718 ~~of gas phase H_2SO_4 can be related to larger contributions of other vapours than H_2SO_4 to~~
719 ~~particle growth. The other or competing compounds may include oxidation products and their~~
720 ~~dimers from photooxidation of VOC precursors from both biogenic and anthropogenic sources.~~

721



732 **Figure 4.** Dependence of the growth rate GR_{10} (in a unit of nm h^{-1}) of new particle formation and growth
 733 events normalised to the daily maximum gas phase H_2SO_4 proxy ($\mu\text{g m}^{-5} \text{W s}$) on the reciprocal proxy
 734 value in the city centre and near city background. The linear line in red represents the line fitted to the
 735 joint data set. Number of individual data considered (n), their coefficient of determination (R^2) and the
 736 intercept (a) and slope (b) of the fitted regression line with standard deviations are also shown.

738 4.4 Extreme and multiple events

739
 740 The joint 6-year long data sets of J_6 , GR_{10} and Δt containing all, 247 individual data each could
 741 be characterised by lognormal distribution function. This is demonstrated by log-probability
 742 graph for J_6 in Fig. S2 as example. The coefficient of determination, median and geometric
 743 standard deviation for J_6 , GR_{10} and Δt data sets were 0.990, 4.0 cm^{-3} and 2.3; 0.993, 6.8 nm h^{-1}
 744 and 1.46; and 0.998, 02:57 (0.123 d) and 1.74, respectively.

745
 746 One of the major properties of this distribution type is that it contains relatively large individual
 747 data with considerably high abundances. There were 5 individual J_6 and 5 individual GR_{10} data
 748 above the 98% percentile of the data sets, which belonged to 9 separate NPF and growth events
 749 (days). Their specifications, properties and parameters are summarised in Table 3. All these
 750 events occurred in the city centre from April to September. ~~Their number in the separate~~
 751 ~~consecutive measurement years (Sect. 2) were 1, 0, 1, 2, 0 and 5, respectively.~~ The medians of
 752 J_6 , GR_{10} , CS and air T for the subsets of these 9 extreme event days were larger by factors of
 753 5.2, 2.4, 1.5 and 1.4, respectively than for the city centre data. **At the same time**, the medians
 754 of the other atmospheric properties and concentrations in these 2 respective data sets agreed
 755 within approximately 10%. There was a single event associated with an extreme H_2SO_4 proxy

756 (of $23 \times 10^5 \mu\text{g m}^{-5} \text{ W s}$) and relatively low NO_x concentration ($44 \mu\text{g m}^{-3}$), which indicate
757 exceptionally favourable conditions for NPF and growth. In addition to this case, there were
758 only a few days that were characterised by an unusually large CS ($23 \times 10^{-3} \text{ s}^{-1}$) – which could
759 in turn be linked to higher dynamic rates (Sect. 4.3) – or by somewhat larger SO_2 ($8.1 \mu\text{g m}^{-3}$)
760 or lower NO_x concentration ($34 \mu\text{g m}^{-3}$). For all the other events, however, no simple or
761 compound property of the investigated variables could explain the extreme rates. Instead, they
762 may be related to some other chemical species and/or atmospheric processes, which were not
763 including in the present study. ~~Since the extreme NPF and growth events usually resembled the~~
764 ~~time evolution for regional events (well-developed banana curves) – sometimes with multiple~~
765 ~~onsets –, the missing atmospheric players in increased concentrations or their relevant processes~~
766 ~~are expected to appear on a larger horizontal spatial scale.~~

767

768 **Table 3.** Date (in a format of dd–MM–yyyy), new particle formation rate J_6 (in a unit of $\text{cm}^{-3} \text{ s}^{-1}$),
769 particle diameter growth rate GR_{10} (nm h^{-1}), starting time t_1 of nucleation (HH:mm UTC+1), duration
770 time interval $\Delta t = t_2 - t_1$ of nucleation (HH:mm), mean condensation sink CS during the nucleation process
771 (10^{-3} s^{-1}), daily maximum gas-phase H_2SO_4 proxy ($10^4 \mu\text{g m}^{-5} \text{ W s}$), daily mean air temperature T ($^\circ\text{C}$),
772 daily mean relative humidity RH (%), daily median concentrations of SO_2 , O_3 , NO_x ($\mu\text{g m}^{-3}$) and CO
773 (mg m^{-3}) gases, and the type of the onset for extreme quantifiable (class 1A) new particle formation and
774 growth events. The cells in yellow indicate the values which are above the 98% percentile of the
775 corresponding data sets. N.a.: not available.

776

Date/ property	15– 09– 2009	20– 04– 2014	19– 05– 2015	04– 07– 2015	28– 05– 2017	25– 06– 2017	02– 08– 2017	31– 08– 2017	09– 09– 2017
J_6	15.9	17.8	24	16.3	27	33	30	47	17.3
GR_{10}	17.4	19.0	12.2	18.0	9.2	17.0	11.8	21	19.8
t_1	10:20	08:52	08:52	09:38	06:34	10:18	07:39	10:06	11:38
Δt	01:23	01:42	03:57	02:06	07:15	02:46	06:58	06:19	02:06
Proxy	38	42	25	16	229	41	69	92	45
CS	13.4	8.9	13.7	11.9	6.9	10.5	23	18.2	15.5
T	20	13.0	22	26	20	24	29	23	19.1
RH	60	62	48	40	40	68	49	47	58
SO_2	6.1	2.5	4.4	2.3	3.4	3.1	5.6	8.1	6.6
O_3	16.3	43	n.a.	33	61	56	34	24	12.9
NO_x	69	34	174	70	44	66	n.a.	109	112
CO	0.42	n.a.	0.71	0.33	0.31	0.50	0.97	0.62	0.71
Onset	ordinary	double	broad	ordinary	broad	broad	broad	broad	ordinary

777

778

779 Each quantifiable NPF and growth event was labelled as ordinary or broad by visual inspection
780 of its beginning part. If the width of the beginning was smaller than approximately 2 h or there
781 was a determinant single growth curve (rib) on the size distribution surface plot then the onset
782 was labelled as ordinary, otherwise as broad (Fig. S1a and S1c for broad onsets). Broad onsets
783 can be generated by 1) long-lasting nucleation process, 2) disrupted and started over nucleation
784 due to changing atmospheric and meteorological conditions or 3) multiple nucleation processes
785 close to each other in time (Salma et al., 2016b). The broad onsets were specified as doublets
786 if the nucleation mode could be separated into 2 submodes by size distribution fitting.
787 Approximately 40% of all quantifiable events had a broad onset. This indicates that NPF and
788 growth events with broad/multiple onsets are abundant in the urban environment, which could
789 be an important difference from remote or clean atmospheres.

790

791 For 9% of all quantifiable days, it was feasible to calculate 2 sets of dynamic properties for
792 onsets 1 and 2 with a reasonable accuracy. In the near-city background, the medians of J_6 and
793 GR_{10} for the onset 1 were similar or somewhat smaller than the corresponding medians for the
794 whole near-city background data set, while for the onset 2, they were substantially larger,
795 namely $4.1 \text{ cm}^{-3} \text{ s}^{-1}$ and 10.0 nm h^{-1} , respectively (cf. Table 2). Actually, the latter values were
796 closer to the medians of the city centre than for the near-city background. ~~The dynamic~~
797 ~~properties for the city centre for both the onset 1 and onset 2 were somewhat larger than for the~~
798 ~~whole the city centre data set.~~ Approximately 75% of the doublets resulted in individual
799 onset2/onset1 ratios larger than unity. Their overall median ratios for J_6 and GR_{10} were similar
800 and approximately 1.4, while for the near-city background, they were about 2. The results are
801 in line with ~~and confirm~~ our earlier conclusion according to which the second onsets (if it is a
802 new formation process and not just a started over event) ~~often generate new particles more~~
803 ~~intensively~~ **are more intensive** than the first onsets (Salma et al., 2016b). These particles also
804 grow faster. This can be explained by the fact that the first event is ~~often or likely~~ of regional
805 scale since its dynamic properties resemble those of the regional background ~~process~~ (Yli-Juuti
806 et al., 2009), while the later event can be characterised by values typical for the city centre
807 (Salma et al., 2016b). The later event (or events) are mainly caused and governed by sub-
808 regional processes. These findings are also coherent with a previous observation of NPF and
809 growth events with multiple onsets in semi-clean savannah and industrial environments
810 (Hirsikko et al., 2013), and they also fit well into the existing ideas on mixing regional and

811 urban air parcels that exhibit different properties such as precursor concentrations, T and RH
812 (Kulmala et al., 2017).

813

814 **5 Conclusions**

815

816 Magnitude of the particle number concentration level produced solely by NPF and growth
817 (strength of the events) can roughly be estimated by considering the median J_6 , median duration
818 of nucleation Δt (their distribution function is lognormal; Table 2) and the mean coagulation
819 loss of these particles F_{coag} (0.17; Sect. 3.1 and Table S1) as: $J_6 \times \Delta t \times (1 - F_{\text{coag}})$. In central
820 Budapest, it yields a concentration of 10^4 cm^{-3} . This is in line with another result achieved by
821 nucleation strength factor (Salma et al., 2017). More importantly, the estimated concentration
822 from NPF and growth process is comparable to the annual median atmospheric concentrations
823 (Sect. 4). This simple example indicates that the phenomenon is not only relevant for aerosol
824 load and climate issues on regional or global spatial scales – which were first recognised – but
825 it can affect the urban climate and the health risk of inhabitants of cities as well.

826

827 Similar recognitions have led to the emerge of urban atmospheric nucleation studies. As part of
828 this international progress, we presented here a considerable variety of contributions, which
829 became feasible thank to gradually generating, multi-year long, critically evaluated, complex
830 and coherent data sets. Dynamic and timing properties of 247 NPF and growth events were
831 studied together with supporting aerosol properties, meteorological data and pollutant gas
832 concentrations in near-city background and city centre of Budapest for 6 years. The results and
833 conclusions derived form in important component that is based on atmospheric observations.
834 The present study can also be considered as the first step toward a larger statistical evaluation
835 process. The results are to be combine with results from laboratory experiments and finally,
836 with theoretical models to further improve our understanding on the atmospheric processes in
837 cities. Specialities and features of the urban atmospheric NPF and growth phenomena are finally
838 to be also considered when assessing their potentials to increase UF and CCN concentrations
839 or their health implications.

840

841 The present research based on ambient atmospheric measurements provided evidence that some
842 important chemical players in the NPF and growth events are still missing. Considering the
843 results and conclusions of cloud chamber experiments, these factors are expected to be related
844 mainly to oxidation products of VOCs and/or their processes. Further dedicated research

845 including sophisticated measurements, data evaluations and modelling studies is required to
846 find and identify these species and their processes, and to account their multifactorial role in
847 more detail. Such measurement campaign focusing on chemical composition of molecular
848 clusters, precursors and nucleating vapours by applying recent expedient instruments in
849 Budapest over the months of the highest expected event occurrence has been just realised within
850 a frame of an international cooperation. Its perspective results can hopefully provide additional
851 valuable information for some of the conclusion base on indirect evidence for the time being
852 and can further clarify the overall picture on urban multicomponent nucleation and growth
853 phenomenon.

854

855 **Data availability.** The observational data used in this paper are available on request from the
856 corresponding author or at the website of the Budapest platform for Aerosol Research and Training
857 (<http://salma.web.elte.hu/BpART>).

858

859 **Competing interest.** The authors declare that they have no conflict of interest.

860

861 **Acknowledgements.** The authors thank Markku Kulmala and his research team at the University of
862 Helsinki for their cooperation. Financial support by the National Research, Development and Innovation
863 Office, Hungary (contracts K116788 and PD124283); by the European Regional Development Fund
864 and the Hungarian Government (GINOP-2.3.2-15-2016-00028) is gratefully acknowledged.

865

866 **References**

- 867 Alam, A., Shi, J. P., and Harrison, R. M.: Observations of new particle formation in urban air, J.
868 Geophys. Res., 108 (D3), 4093, doi:10.1029/2001JD001417, 2003.
- 869 Almeida, J., Schobesberger, S., Kurten, A., Ortega, I. K., Kupiainen-Maatta, O., Praplan, A. P.,
870 Adamov, A., Amorim, A., Bianchi, F., Breitenlechner, M., David, A., Dommen, J., Donahue, N.
871 M., Downard, A., Dunne, E., Duplissy, J., Ehrhart, S., Flagan, R. C., Franchin, A., Guida, R.,
872 Hakala, J., Hansel, A., Heinritzi, M., Henschel, H., Jokinen, T., Junninen, H., Kajos, M.,
873 Kangasluoma, J., Keskinen, H., Kupc, A., Kurten, T., Kvashin, A. N., Laaksonen, A., Lehtipalo,
874 K., Leiminger, M., Leppa, J., Loukonen, V., Makhmutov, V., Mathot, S., McGrath, M. J.,
875 Nieminen, T., Olenius, T., Onnela, A., Petäjä, T., Riccobono, F., Riipinen, I., Rissanen, M., Rondo,
876 L., Ruuskanen, T., Santos, F. D., Sarnela, N., Schallhart, S., Schnitzhofer, R., Seinfeld, J. H.,
877 Simon, M., Sipilä, M., Stozhkov, Y., Stratmann, F., Tome, A., Tröstl, J., Tsagkogeorgas, G.,
878 Vaattovaara, P., Viisanen, Y., Virtanen, A., Vrtala, A., Wagner, P. E., Weingartner, E., Wex, H.,
879 Williamson, C., Wimmer, D., Ye, P. L., Yli-Juuti, T., Carslaw, K. S., Kulmala, M., Curtius, J.,

880 Baltensperger, U., Worsnop, D. R., Vehkamäki, H., and Kirkby, J.: Molecular understanding of
881 sulphuric acid–amine particle nucleation in the atmosphere, *Nature*, 502, 359–363, 2013.

882 Baltensperger, U., Streit, N., Weingartner, E., Nyeki, S., Prévôt, A. S. H., Van Dingenen, R., Virkkula,
883 A., Putaud, J. P., Even, A., Brink, H., Blatter, A., Neftel, A., and Gaggeler, H. W.: Urban and rural
884 aerosol characterization of summer smog events during the PIPAPO field campaign in Milan, Italy,
885 *J. Geophys. Res.*, 107(D22), 8193, doi:10.1029/2001JD001292, 2002.

886 Bianchi, F., Tröstl, J., Junninen, H., Frege, C., Henne, S., Hoyle, C. R., Molteni, U., Herrmann, E.,
887 Adamov, A., Bukowiecki, N., Chen, X., Duplissy, J., Gysel, M., Hutterli, M., Kangasluoma, J.,
888 Kontkanen, J., Kürten, A. Manninen, H. E., Münch, S., Peräkylä, O., Petäjä, T., Rondo, L.,
889 Williamson, C., Weingartner, E. Curtius, J., Worsnop, D. R., Kulmala, M., Dommen, J., and
890 Baltensperger, U.: New particle formation in the free troposphere: A question of chemistry and
891 timing, *Science*, 352, 1109–1112, <https://doi.org/10.1126/science.aad5456>, 2016.

892 Braakhuis, H. M., Park, M. V., Gosens, I., De Jong, W. H., and Cassee, F. R.: Physicochemical
893 characteristics of nanomaterials that affect pulmonary inflammation, *Part. Fibre Toxicol.*, 11:18,
894 doi: 10.1186/1743-8977-11-18, 2014.

895 Cai, R. and Jiang, J.: A new balance formula to estimate new particle formation rate: reevaluating the
896 effect of coagulation scavenging, *Atmos. Chem. Phys.*, 17, 12659–12675, 2017.

897 Carslaw, K. S., Lee, L. A., Reddington, C. L., Pringle, K. J., Rap, A., Forster, P. M., Mann, G. W.,
898 Spracklen, D. V., Woodhouse, M. T., Regayre, L. A., and Pierce, J. R: Large contribution of
899 natural aerosols to uncertainty in indirect forcing, *Nature*, 503, 67–71, 2013.

900 Crouse, J. D., Nielsen, L. B., Jørgensen, S., Kjaergaard, H. G., and Wennberg, P. O.: Autoxidation of
901 organic compounds in the atmosphere, *J. Phys. Chem. Lett.*, 4, 20, 3513–3520, 2013.

902 Dal Maso, M., Kulmala, M., Lehtinen, K. E. J., Mäkelä, J. M., Aalto, P. P., and O’Dowd, C.:
903 Condensation and coagulation sinks and formation of nucleation mode particles in coastal and
904 boreal forest boundary layers, *J. Geophys. Res.*, 107(19D), 8097, 10.1029/2001jd001053, 2002.

905 Dal Maso, M., Kulmala, M., Riipinen, I., Wagner, R., Hussein, T., Aalto, P. P., and Lehtinen, K. E. J.:
906 Formation and growth of fresh atmospheric aerosols: eight years of aerosol size distribution data
907 from SMEAR II, Hyytiälä, Finland, *Boreal Environ. Res.*, 10, 323–336, 2005.

908 Dall’Osto, M., Querol, X., Alastuey, A., O’Dowd, C., Harrison, R. M., Wenger, J., and Gómez-
909 Moreno, F. J.: On the spatial distribution and evolution of ultrafine particles in Barcelona, *Atmos.*
910 *Chem. Phys.*, 13, 741–759, 2013.

911 Ehn, M., Thornton, J. A., Kleist, E., Sipilä, M., Junninen, H., Pullinen, I., Springer, M., Rubach, F.,
912 Tillmann, R., Lee, B., Lopez-Hilfiker, F., Andres, S., Acir, I. H., Rissanen, M., Jokinen, T.,
913 Schobesberger, S., Kangasluoma, J., Kontkanen, J., Nieminen, T., Kurten, T., Nielsen, L. B.,
914 Jorgensen, S., Kjaergaard, H. G., Canagaratna, M., Dal Maso, M., Berndt, T., Petäjä, T., Wahner,
915 A., Kerminen, V. M., Kulmala, M., Worsnop, D. R., Wildt, J., and Mentel, T. F.: A large source of
916 low-volatility secondary organic aerosol, *Nature*, 506, 476–479, 2014.

917 Gordon, H., Sengupta, K., Rap, A., Duplissy, J., Frege, C., Williamson, C., Heinritzi, M., Simon, M.,
918 Yan, C., Almeida, J., Tröstl, J., Nieminen, T., Ortega, I. K., Wagner, R., Dunne, E. M., Adamov,
919 A., Amorim, A., Bernhammer, A. K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven,
920 J. S., Dias, A., Ehrhart, S., Fischer, L., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Hakala, J.,
921 Hoyle, C. R., Jokinen, T., Junninen, H., Kangasluoma, J., Kim, J., Kirkby, J., Krapf, M., Kürten,
922 A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Molteni, U., Monks, S. A., Onnela,
923 A., Peräkylä, O., Piel, F., Petäjä, T., Praplan, A. P., Pringle, K. J., Richards, N. A. D., Rissanen, M.
924 P., Rondo, L., Sarnela, N., Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sharma, S., Sipilä, M.,
925 Steiner, G., Stozhkov, Y., Stratmann, F., Tomé, A., Virtanen, A., Vogel, A. L., Wagner, A. C.,
926 Wagner, P. E., Weingartner, E., Wimmer, D., Winkler, P. M., Ye, P., Zhang, X., Hansel, A.,
927 Dommen, J., Donahue, N. M., Worsnop, D. R., Baltensperger, U., Kulmala, M., Curtius, J., and
928 Carslaw, K. S.: Reduced anthropogenic aerosol radiative forcing caused by biogenic new particle
929 formation, *Proc. Natl. Acad. Sci. U.S.A.*, 113, 12053–12058,
930 <https://doi.org/10.1073/pnas.1602360113>, 2016.

931 Hirsikko, A., Vakkari, V., Tiitta, P., Hatakka, J., Kerminen, V.-M., Sundström, A.-M., Beukes, J. P.,
932 Manninen, H. E., Kulmala, M., and Laakso, L.: Multiple daytime nucleation events in semi-clean
933 savannah and industrial environments in South Africa: analysis based on observations, *Atmos.*
934 *Chem. Phys.*, 13, 5523–5532, 2013.

935 Hussein, T., Puustinen, A., Aalto, P. P., Mäkelä, J. M., Hämeri, K., and Kulmala, M.: Urban aerosol
936 number size distributions, *Atmos. Chem. Phys.*, 4, 391–411, 2004.

937 Hussein, T., Martikainen, J., Junninen, H., Sogacheva, L., Wagner, R., Dal Maso, M., Riipinen, I.,
938 Aalto, P. P., and Kulmala, M.: Observation of regional new particle formation in the urban
939 atmosphere, *Tellus 60B*, 509–521, 2008.

940 Jokinen, T., Berndt, T., Makkonen, R., Kerminen, V.-M., Junninen, H., Paasonen, P., Stratmann, F.,
941 Herrmann, H., Guenther, A. B., Worsnop, D. R., Kulmala, M., Ehn, M. and Sipilä, M.: Production
942 of extremely low volatile organic compounds from biogenic emissions: Measured yields and
943 atmospheric implications, *Proc. Natl. Acad. Sci. U.S.A.*, 112, 7123–7128, 2015.

944 Kerminen, V.-M., Paramonov, M., Anttila, T., Riipinen, I., Fountoukis, C., Korhonen, H., Asmi, E.,
945 Laakso, L., Lihavainen, H., Swietlicki, E., Svenningsson, B., Asmi, A., Pandis, S. N., Kulmala, M.,
946 and Petäjä, T.: Cloud condensation nuclei production associated with atmospheric nucleation: a
947 synthesis based on existing literature and new results, *Atmos. Chem. Phys.*, 12, 12037–12059,
948 2012.

949 Kerminen, V.-M., Chen, X., Vakkari, V., Petäjä, T., Kulmala, M., and Bianchi, F.: Atmospheric new
950 particle formation and growth: review of field observations, *Environ. Res. Lett.*, 13 (2018) 103003,
951 2018.

952 Kirkby, J., Curtius, J., Almeida, J., Dunne, E., Duplissy, J., Ehrhart, S., Franchin, A., Gagné, S., Ickes,
953 L., Kürten, A., Kupc, A., Metzger, A., Riccobono, F., Rondo, L., Schobesberger, S.,
954 Tsagkogeorgas, G., Wimmer, D., Amorim, A., Bianchi, F., Breitenlechner, M., David, A.,

955 Dommen, J., Downard, A., Ehn, M., Flagan, R. C., Haider, S., Hansel, A., Hauser, D., Jud, W.,
956 Junninen, H., Kreissl, F., Kvashin, A., Laaksonen, A., Lehtipalo, K., Lima, J., Lovejoy, E. R.,
957 Makhutov, V., Mathot, S., Mikkilä, J., Minginette, P., Mogo, S., Nieminen, T., Onnela, A., Pereira,
958 A., Petäjä, T., Schnitzhofer, R., Seinfeld, J. H., Sipilä, M., Stozhkov, Y., Stratmann, F., Tome, A.,
959 Vanhanen, J., Viisanen Y., Vrtala, A., Wagner, P.E., Walther, H., Weingartner, E., Wex, H.,
960 Winkler, P.M., Carslaw, K. S., Worsnop, D. R., Baltensperger, U., and Kulmala, M.: The role of
961 sulfuric acid, ammonia and galactic cosmic rays in atmospheric aerosol nucleation, *Nature*, 476,
962 429–433, 2011.

963 Kirkby, J., Duplissy, J., Sengupta, K., Frege, C., Gordon, H., Williamson, C., Heinritzi, M., Simon,
964 M., Yan, C., Almeida, J., Tröstl, J., Nieminen, T., Ortega, I. K., Wagner, R., Adamov, A., Amorim,
965 A., Bernhammer, A.-K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven, J., Dias, A.,
966 Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Hakala, J., Hoyle, C. R., Jokinen, T.,
967 Junninen, H., Kangasluoma, J., Kim, J., Krapf, M., Kürten, A., Laaksonen, A., Lehtipalo, K.,
968 Makhmutov, V., Mathot, S., Molteni, U., Onnela, A., Peräkylä, O., Piel, F., Petäjä, T., Praplan, A.
969 P., Pringle, K., Rap, A., Richards, N. A. D., Riipinen, I., Rissanen, M. P., Rondo, L., Sarnela, N.,
970 Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sipilä, M., Steiner, G., Stozhkov, Y., Stratmann, F.,
971 Tomé, A., Virtanen, A., Vogel, A. L., Wagner, A., Wagner, P. E., Weingartner, E., Wimmer, D.,
972 Winkler, P. M., Ye, P., Zhang, X., Hansel, A., Dommen, J., Donahue, N. M., Worsnop, D. R.,
973 Baltensperger, U., Kulmala, M., Carslaw, K. S., and Curtius, J.: Ion-induced nucleation of pure
974 biogenic particles, *Nature*, 533, 521–526, <https://doi.org/10.1038/nature17953>, 2016.

975 Kulmala, M., Dal Maso, M., Mäkelä, J. M., Pirjola, L., Väkevä, M., Aalto, P., Miikkulainen, P.,
976 Hämeri, K., and O'Dowd, C. D.: On the formation, growth and composition of nucleation mode
977 particles, *Tellus B*53, 479–490, 2001.

978 Kulmala, M., Vehkamäki, H., Petäjä, T., Dal Maso, M., Lauri, A., Kerminen, V.-M., Birmili, W., and
979 McMurry, P.: Formation and growth rates of ultrafine atmospheric particles: a review of
980 observations, *J. Aerosol Sci.*, 35, 143–176, 2004.

981 Kulmala, M., Petäjä, T., Nieminen, T., Sipilä, M., Manninen, H. E., Lehtipalo, K., Dal Maso, M.,
982 Aalto, P. P., Junninen, H., Paasonen, P., Riipinen, I., Lehtinen, K. E. J., Laaksonen, A., and
983 Kerminen, V.-M.: Measurement of the nucleation of atmospheric aerosol particles, *Nat. Protoc.*, 7,
984 1651–1667, doi:10.1038/nprot.2012.091, 2012.

985 Kulmala, M., Kontkanen, J., Junninen, H., Lehtipalo, K., Manninen, H. E., Nieminen, T., Petäjä, T.,
986 Sipilä, M., Schobesberger, S., Rantala, P., Franchin, A., Jokinen, T., Järvinen, E., Äijälä, M.,
987 Kangasluoma, J., Hakala, J., Aalto, P.P., Paasonen, P., Mikkilä, J., Vanhanen, J., Aalto, J., Hakola,
988 H., Makkonen, U., Ruuskanen, T., Mauldin, R. L. III, Duplissy, J., Vehkamäki, H., Bäck, J.,
989 Kortelainen, A., Riipinen, I., Kurtén, T., Johnston, M. V., Smith, J. N., Ehn, M., Mentel, T. F.,
990 Lehtinen, K. E. J., Laaksonen, A., Kerminen, V.-M., and Worsnop, D. R.: Direct observations of
991 atmospheric aerosol nucleation, *Science*, 339, 943–946, 2013.

992 Kulmala, M., Petäjä, T., Ehn, M., Thornton, J., Sipilä, M., Worsnop, D. R., and Kerminen, V.-M.:
993 Chemistry of atmospheric nucleation: On the recent advances on precursor characterization and
994 atmospheric cluster composition in connection with atmospheric new particle formation, *Annu.*
995 *Rev. Phys. Chem.*, 65, 21–37, 2014.

996 Kulmala, M., Kerminen, V. M., Petäjä, T., Ding, A. J., and Wang, L.: Atmospheric gas-to-particle
997 conversion: why NPF events are observed in megacities, *Faraday Discuss.*,
998 doi:10.1039/C6FD00257A, 2017.

999 Makkonen, R., Asmi, A., Korhonen, H., Kokkola, H., Järvenoja, S., Räisänen, P., Lehtinen, K. E. J.,
1000 Laaksonen, A., Kerminen, V.-M., Järvinen, H., Lohmann, U., Bennartz, R., Feichter, J., and
1001 Kulmala, M.: Sensitivity of aerosol concentrations and cloud properties to nucleation and
1002 secondary organic distribution in ECHAM5-HAM global circulation model, *Atmos. Chem. Phys.*,
1003 9, 1747–1766, 2009.

1004 Makkonen, R., Asmi, A., Kerminen, V.-M., Boy, M., Arneth, A., Hari, P., and Kulmala, M.: Air
1005 pollution control and decreasing new particle formation lead to strong climate warming, *Atmos.*
1006 *Chem. Phys.*, 12, 1515–1524, 2012.

1007 Merikanto, J., Spracklen, D. V., Mann, G. W., Pickering, S. J., and Carslaw, K. S.: Impact of
1008 nucleation on global CCN, *Atmos. Chem. Phys.*, 9, 8601–8616, 2009.

1009 Metzger, A., Verheggen, B., Dommen, J., Duplissy, J., Prévôt, A. S. H., Weingartner, E., Riipinen, I.,
1010 Kulmala, M., Spracklen, D. V., Carslaw, K. S., and Baltensperger, U.: Evidence for the role of
1011 organics in aerosol particle formation under atmospheric conditions, *Proc. Natl. Acad. Sci. U. S.*
1012 *A.*, 107, 6646–6651, 2010.

1013 Németh, Z. and Salma, I.: Spatial extension of nucleating air masses in the Carpathian Basin, *Atmos.*
1014 *Chem. Phys.*, 14, 8841–8848, 2014.

1015 Németh, Z., Rosati, B., Ziková, N., Salma, I., Bozó, L., Dameto de España, C., Schwarz, J., Ždímal,
1016 V., and Wonaschütz, A.: Comparison of atmospheric new particle formation and growth events in
1017 three Central European cities, *Atmos. Environ.*, 178, 191–197, 2018.

1018 Nieminen, T., Kerminen, V.-M., Petäjä, T., Aalto, P. P., Arshinov, M., Asmi, E., Baltensperger, U.,
1019 Beddows, D. C. S., Beukes, J. P., Collins, D., Ding, A., Harrison, R. M., Henzing, B., Hooda, R.,
1020 Hu, M., Hörrak, U., Kivekäs, N., Komsaare, K., Krejčí, R., Kristensson, A., Laakso, L., Laaksonen,
1021 A., Leaitch, W. R., Lihavainen, H., Mihalopoulos, N., Németh, Z., Nie, W., O'Dowd, C., Salma, I.,
1022 Sellegri, K., Svenningsson, B., Swietlicki, E., Tunved, P., Ulevicius, V., Vakkari, V., Vana, M.,
1023 Wiedensohler, A., Wu, Z., Virtanen, A., and Kulmala, M.: Global analysis of continental boundary
1024 layer new particle formation based on long-term measurements, *Atmos. Chem. Phys.*, 18, 14737–
1025 14756, 2018.

1026 Oberdörster, G., Oberdörster, E., and Oberdörster, J.: Nanotoxicology: an emerging discipline
1027 evolving from studies of ultrafine particles, *Environ. Health Perspect.*, 113, 823–839, 2005.

1028 O'Dowd, C. D., Jimenez, J. L., Bahreini, R., Flagan, R. C., Seinfeld, J. H., Hämeri, K., Pirjola, L.,
1029 Kulmala, M., Jennings, S. G., and Hoffmann, Th.: Marine aerosol formation from biogenic iodine
1030 emissions, *Nature* 417, 632–636, 2002.

1031 Paasonen, P., Kupiainen, K., Klimont, Z., Visschedijk, A., Denier van der Gon, H. A. C., and Amann,
1032 M.: Continental anthropogenic primary particle number emissions, *Atmos. Chem. Phys.*, 16, 6823–
1033 6840, 2016.

1034 Paasonen, P., Peltola, M., Kontkanen, J., Junninen, H., Kerminen, V.-M., and Kulmala, M.:
1035 Comprehensive analysis of particle growth rates from nucleation mode to cloud condensation
1036 nuclei in Boreal forest, *Atmos. Chem. Phys. Discuss.*, <https://doi.org/10.5194/acp-2018-169>, in
1037 review, 2018.

1038 Petäjä, T., Mauldin, III, R. L., Kosciuch, E., McGrath, J., Nieminen, T., Paasonen, P., Boy, M.,
1039 Adamov, A., Kotiaho, T., and Kulmala, M.: Sulfuric acid and OH concentrations in a boreal forest
1040 site, *Atmos. Chem. Phys.*, 9, 7435–7448, 2009.

1041 Putaud, J.-P., Van Dingenen, R., Alastuey, A., Bauer, H., Birmili, W., Cyrus, J., Flentje, H., Fuzzi, S.,
1042 Gehrig, R., Hansson, H. C., Harrison, R. M., Herrmann, H., Hitztenberger, R., Hüglin, C., Jones,
1043 A.M., Kasper-Giebl, A., Kiss, G., Koussa, A., Kuhlbusch, T. A. J., Löschau, G., Maenhaut, W.,
1044 Molnár, A., Moreno, T., Pekkanen, J., Perrino, C., Pitz, M., Puxbaum, H., Querol, X., Rodriguez,
1045 S., Salma, I., Schwarz, J., Smolík, J., Schneider, J., Spindler, G., ten Brink, H., Turšič, J., Viana,
1046 M., Wiedensohler, and A., Raes, F.: A European Aerosol Phenomenology - 3: physical and
1047 chemical characteristics of particulate matter from 60 rural, urban, and kerbside sites across
1048 Europe, *Atmos. Environ.*, 44, 1308–1320, 2010.

1049 Riccobono, F., Schobesberger, S., Scott, C., Dommen, J., Ortega, I., Rondo, L., Almeida, J., Amorim,
1050 A., Bianchi, F., Breitenlechner, M., David, A., Downard, A., Dunne, E., Duplissy, J., Ehrhart, S.,
1051 Flagan, R., Franchin, A., Hansel, A., Junninen, H., Kajos, M., Keskinen, H., Kupc, A., Kurten, A.,
1052 Kvashin, A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Nieminen, T., Onnela, A.,
1053 Petäjä, T., Praplan, A., Santos, F., Schallhart, S., Seinfeld, J., Sipila, M., Spracklen, D., Stozhkov,
1054 Y., Stratmann, F., Tome, A., Tsagkogeorgas, G., Vaattovaara, P., Viisanen, Y., Vrtala, A., Wagner,
1055 P., Weingartner, E., Wex, H., Wimmer, D., Carslaw, K., Curtius, J., Donahue, N., Kirkby, J.,
1056 Kulmala, M., Worsnop, D., and Baltensperger, U.: Oxidation products of biogenic emissions
1057 contribute to nucleation of atmospheric particles, *Science*, 344, 717–721, 2014.

1058 Riipinen, I., Pierce, J. R., Yli-Juuti, T., Nieminen, T., Häkkinen, S., Ehn, M., Junninen, H., Lehtipalo,
1059 K., Petäjä, T., Slowik, J., Chang, R., Shantz, N. C., Abbatt, J., Leitch, W. R., Kerminen, V.-M.,
1060 Worsnop, D. R., Pandis, S. N., Donahue, N. M., and Kulmala, M.: Organic condensation: a vital
1061 link connecting aerosol formation to cloud condensation nuclei (CCN) concentrations, *Atmos.*
1062 *Chem. Phys.*, 11, 3865–3878, 2011.

1063 Salma, I., Borsós, T., Weidinger, T., Aalto, P., Hussein, T., Dal Maso, M., and Kulmala, M.:
1064 Production, growth and properties of ultrafine atmospheric aerosol particles in an urban
1065 environment, *Atmos. Chem. Phys.*, 11, 1339–1353, 2011.

- 1066 Salma, I., Borsós, T., Németh, Z., Weidinger, T., Aalto, T., and Kulmala, M.: Comparative study of
1067 ultrafine atmospheric aerosol within a city, *Atmos. Environ.*, 92, 154–161, 2014.
- 1068 Salma, I., Füre, P., Németh, Z., Farkas, Á., Balásházy, I., Hofmann, W., and Farkas, Á.: Lung burden
1069 and deposition distribution of inhaled atmospheric urban ultrafine particles as the first step in their
1070 health risk assessment, *Atmos. Environ.*, 104, 39–49, 2015.
- 1071 Salma, I., Németh, Z., Weidinger, T., Kovács, B., and Kristóf, G.: Measurement, growth types and
1072 shrinkage of newly formed aerosol particles at an urban research platform, *Atmos. Chem. Phys.*,
1073 16, 7837–7851, 2016a.
- 1074 Salma, I., Németh, Z., Kerminen, V. M., Aalto, P., Nieminen, T., Weidinger, T., Molnár, Á., Imre, K.,
1075 and Kulmala, M.: Regional effect on urban atmospheric nucleation, *Atmos. Chem. Phys.*, 16,
1076 8715–8728, 2016b.
- 1077 Salma, I., Varga, V., and Németh, Z.: Quantification of an atmospheric nucleation and growth process
1078 as a single source of aerosol particles in a city, *Atmos. Chem. Phys.*, 17, 15007–15017, 2017.
- 1079 Schobesberger, S., Junninen, H., Bianchi, F., Lonn, G., Ehn, M., Lehtipalo, K., Dommen, J., Ehrhart,
1080 S., Ortega, I. K., Franchin, A., Nieminen, T., Riccobono, F., Hutterli, M., Duplissy, J., Almeida, J.,
1081 Amorim, A., Breitenlechner, M., Downard, A. J., Dunne, E. M., Flagan, R. C., Kajos, M.,
1082 Keskinen, H., Kirkby, J., Kupc, A., Kurten, A., Kurten, T., Laaksonen, A., Mathot, S., Onnela, A.,
1083 Praplan, A. P., Rondo, L., Santos, F. D., Schallhart, S., Schnitzhofer, R., Sipilä, M., Tome, A.,
1084 Tsagkogeorgas, G., Vehkamäki, H., Wimmer, D., Baltensperger, U., Carslaw, K. S., Curtius, J.,
1085 Hansel, A., Petäjä, T., Kulmala, M., Donahue, N. M., and Worsnop, D. R.: Molecular
1086 understanding of atmospheric particle formation from sulfuric acid and large oxidized organic
1087 molecules, *Proc. Natl. Acad. Sci. U.S.A.*, 110, 17223–17228, 10.1073/pnas.1306973110, 2013.
- 1088 Sihto, S.-L., Mikkilä, J., Vanhanen, J., Ehn, M., Liao, L., Lehtipalo, K., Aalto, P. P., Duplissy, J.,
1089 Petäjä, T., Kerminen, V.-M., Boy, M., and Kulmala, M.: Seasonal variation of CCN concentrations
1090 and aerosol activation properties in boreal forest, *Atmos. Chem. Phys.*, 11, 13269–13285, 2011.
- 1091 Sipilä, M., Berndt, T., Petäjä, T., Brus, D., Vanhanen, J., Stratmann, F., Patokoski, J., Mauldin, R. L.
1092 3rd, Hyvärinen, A. P., Lihavainen, H., and Kulmala, M.: The role of sulfuric acid in atmospheric
1093 nucleation, *Science*, 327(5970), 1243–6. doi: 10.1126/science.1180315, 2010.
- 1094 Spracklen, D. V., Carslaw, K. S., Merikanto, J., Mann, G. W., Reddington, C. L., Pickering, S., Ogren,
1095 J. A., Andrews, E., Baltensperger, U., Weingartner, E., Boy, M., Kulmala, M., Laakso, L.,
1096 Lihavainen, H., Kivekäs, N., Komppula, M., Mihalopoulos, N., Kouvarakis, G., Jennings, S. G.,
1097 O'Dowd, C., Birmili, W., Wiedensohler, A., Weller, R., Gras, J., Laj, P., Sellegri, K., Bonn, B.,
1098 Krejčí, R., Laaksonen, A., Hamed, A., Minikin, A., Harrison, R. M., Talbot, R., and Sun, J.: The
1099 contribution of boundary layer nucleation events to total particle concentrations on regional and
1100 global scales, *Atmos. Chem. Phys.*, 6, 5631–5648, 2006.
- 1101 Tröstl, J., Chuang, W. K., Gordon, H., Heinritzi, M., Yan, C., Molteni, U., Ahlm, L., Frege, C.,
1102 Bianchi, F., Wagner, R., Simon, M., Lehtipalo, K., Williamson, C., Craven, J. S., Duplissy, J.,
1103 Adamov, A., Almeida, J., Bernhammer, A. K., Breitenlechner, M., Brilke, S., Dias, A., Ehrhart, S.,

1104 Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Gysel, M., Hansel, A., Hoyle, C. R., Jokinen, T.,
1105 Junninen, H., Kangasluoma, J., Keskinen, H., Kim, J., Krapf, M., Kürten, A., Laaksonen, A.,
1106 Lawler, M., Leiminger, M., Mathot, S., Möhler, O., Nieminen, T., Onnela, A., Petäjä, T., Piel, F.,
1107 M., Miettinen, P., Rissanen, M. P., Rondo, L., Sarnela, N., Schobesberger, S., Sengupta, K., Sipilä,
1108 M., Smith, J. N., Steiner, G., Tomè, A., Virtanen, A., Wagner, A. C., Weingartner, E., Wimmer, D.,
1109 Winkler, P. M., Ye, P. L., Carslaw, K. S., Curtius, J., Dommen, J., Kirkby, J., Kulmala, M.,
1110 Riipinen, I., Worsnop, D. R., Donahue, N. M., and Baltensperger, U.: The role of low-volatility
1111 organic compounds in initial particle growth in the atmosphere, *Nature*, 533, 527,
1112 10.1038/nature18271, 2016.

1113 Vakkari, V., Tiitta, P., Jaars, K., Croteau, P., Beukes, J. P., Josipovic, M., Kerminen, V.-M., Kulmala,
1114 M., Venter, A. D., van Zyl, P. G., Worsnop, D. R., and Laakso, L.: Reevaluating the contribution of
1115 sulfuric acid and the origin of organic compounds in atmospheric nanoparticle growth, *Geophys.*
1116 *Res. Lett.*, 42, 10486–10493, 2015.

1117 Vuollekoski, H., Sihto, S.-L., Kerminen, V.-M., Kulmala, M., and Lehtinen, K. E. J.: A numerical
1118 comparison of different methods for determining the particle formation rate, *Atmos. Chem. Phys.*,
1119 12, 2289–2295, 2012.

1120 Wehner, B., Wiedensohler, A., Tuch, T. M., Wu, Z. J., Hu, M., Slanina, J., and Kiang, C. S.:
1121 Variability of the aerosol number size distribution in Beijing, China: new particle formation, dust
1122 storms, and high continental background, *Geophys. Res. Lett.*, 31, L22108, 2004.

1123 Wiedensohler, A., Cheng, Y. F., Nowak, A., Wehner, B., Achtert, P., Berghof, M., Birmili, W., Wu, Z.
1124 J., Hu, M., Zhu, T., Takegawa, N., Kita, K., Kondo, Y., Lou, S. R., Hofzumahaus, A., Holland, F.,
1125 Wahner, A., Gunthe, S. S., Rose, D., Su, H., and Pöschl, U.: Mobility particle size spectrometers:
1126 harmonization of technical standards and data structure to facilitate high quality long-term
1127 observations of atmospheric particle number size distributions, *Atmos. Meas. Tech.*, 5, 657–685,
1128 2012.

1129 Woo, K. S., Chen, D. R., Pui, D. Y. H., and McMurry, P. H.: Measurement of Atlanta aerosol size
1130 distributions: observations of ultrafine particle events, *Aerosol Sci. Technol.*, 34, 75–87, 2001.

1131 Xiao, S., Wang, M. Y., Yao, L., Kulmala, M., Zhou, B., Yang, X., Chen, J. M., Wang, D. F., Fu, Q.
1132 Y., Worsnop, D. R., and Wang, L.: Strong atmospheric new particle formation in winter in urban
1133 Shanghai, China, *Atmos. Chem. Phys.*, 15, 1769–1781, 2015.

1134 Yli-Juuti, T., Riipinen, I., Aalto, P. P., Nieminen, T., Maenhaut, W., Janssens, I. A., Claeys, M.,
1135 Salma, I., Ocskay, R., Hoffer, A., Imre, K., and Kulmala, M.: Characteristics of new particle
1136 formation events and cluster ions at K-pusztá, Hungary. *Boreal Environ. Res.*, 14, 683–698, 2009.

1137 Zhang, R., Wang, G., Guo, S., Zamora, M. L., Ying, Q., Lin, Y., Wang, W., Hu, M., and Wang, Y.:
1138 Formation of urban fine particulate matter, *Chem. Rev.*, 115, 3803–3855, 2015.

1139 **Supplementary material**

1140

1141 **Table S1.** Relative contributions of particle number concentration increment ($dN_{\text{nuc}}/dt=dN_{6-25}/dt-$
 1142 $dN_{\text{Ai}, <25}/dt$), coagulation scavenging loss (F_{coag}) and growth out of particles from the diameter interval
 1143 of 6–25 nm (F_{growth}) relative to the formation rate J_6 in the near-city background and city centre
 1144 separately for 1-year long measurement time intervals. The measurement year and number of
 1145 quantifiable (class 1A) new aerosol particle formation and growth events (n) are also shown.

1146

Environment and year/ statistics	Contribution in %		
	dN_{nuc}/dt	F_{coag}	F_{growth}
Background, 2012–2013, $n=43$			
Minimum	45	4	2
Maximum	93	38	26
Mean	76	14	10
St. deviation	12	9	5
Centre, 2008–2009, $n=31$			
Minimum	32	13	3
Maximum	84	44	38
Mean	54	29	18
St. deviation	13	8	9
Centre, 2013–2014, $n=48$			
Minimum	43	9	3
Maximum	86	37	30
Mean	63	22	15
St. deviation	11	7	7
Centre, 2014–2015, $n=56$			
Minimum	45	6	2
Maximum	91	46	32
Mean	70	17	14
St. deviation	12	7	8
Centre, 2015–2016, $n=17$			
Minimum	50	4	2
Maximum	92	43	30
Mean	74	14	11
St. deviation	11	9	8
Centre, 2017–2018, $n=52$			
Minimum	44	4	3
Maximum	93	41	31
Mean	70	17	13
St. deviation	11	8	7

1147

1148 **Table S2.** Ranges, averages and standard deviations of condensation sink value during the nucleation
 1149 process, daily maximum gas-phase H₂SO₄ proxy, daily mean air temperature and daily mean relative
 1150 humidity on quantifiable (class 1A) new particle formation and growth events in the near-city
 1151 background and city centre separately for the 1-year long measurement time intervals and for the joint
 1152 5-year long city centre data set.
 1153

Environment	Background		Centre				
	2012– 2013	2008– 2009	2013– 2014	2014– 2015	2015– 2016	2017– 2018	All 5 years
Condensation sink, CS (10 ⁻³ s ⁻¹)							
Minimum	1.63	3.1	2.0	2.4	1.69	2.1	1.69
Median	5.6	9.5	9.9	8.6	5.0	8.4	8.9
Maximum	14.6	21	17.8	21	18.4	18.5	21
Mean	6.2	11.0	10.4	9.4	6.8	8.7	9.4
St. deviation	3.1	4.9	3.7	4.2	4.2	4.6	4.3
Gas-phase H ₂ SO ₄ proxy (10 ⁴ μg m ⁻⁵ W s)							
Minimum	40	10.9	12.2	5.8	34	7.3	5.8
Median	93	39	40	38	79	46	41
Maximum	163	96	139	135	190	134	190
Mean	93	39	45	42	82	50	48
St. deviation	32	17	27	23	38	31	29
Air temperature, <i>T</i> (°C)							
Minimum	-5.2	-0.46	-1.78	-1.19	-1.07	1.21	-1.78
Median	11.5	17.1	16.8	15.3	14.2	16.7	16.1
Maximum	27	23	28	28	28	27	28
Mean	11.5	16.3	15.7	15.0	13.6	16.4	15.5
St. deviation	8.1	5.6	6.9	7.2	8.3	6.5	6.8
Relative humidity, RH (%)							
Minimum	41	32	41	31	39	36	31
Median	63	49	60	50	55	52	53
Maximum	91	74	78	77	89	73	89
Mean	64	51	60	50	56	52	54
St. deviation	12	11	10	9	12	9	11

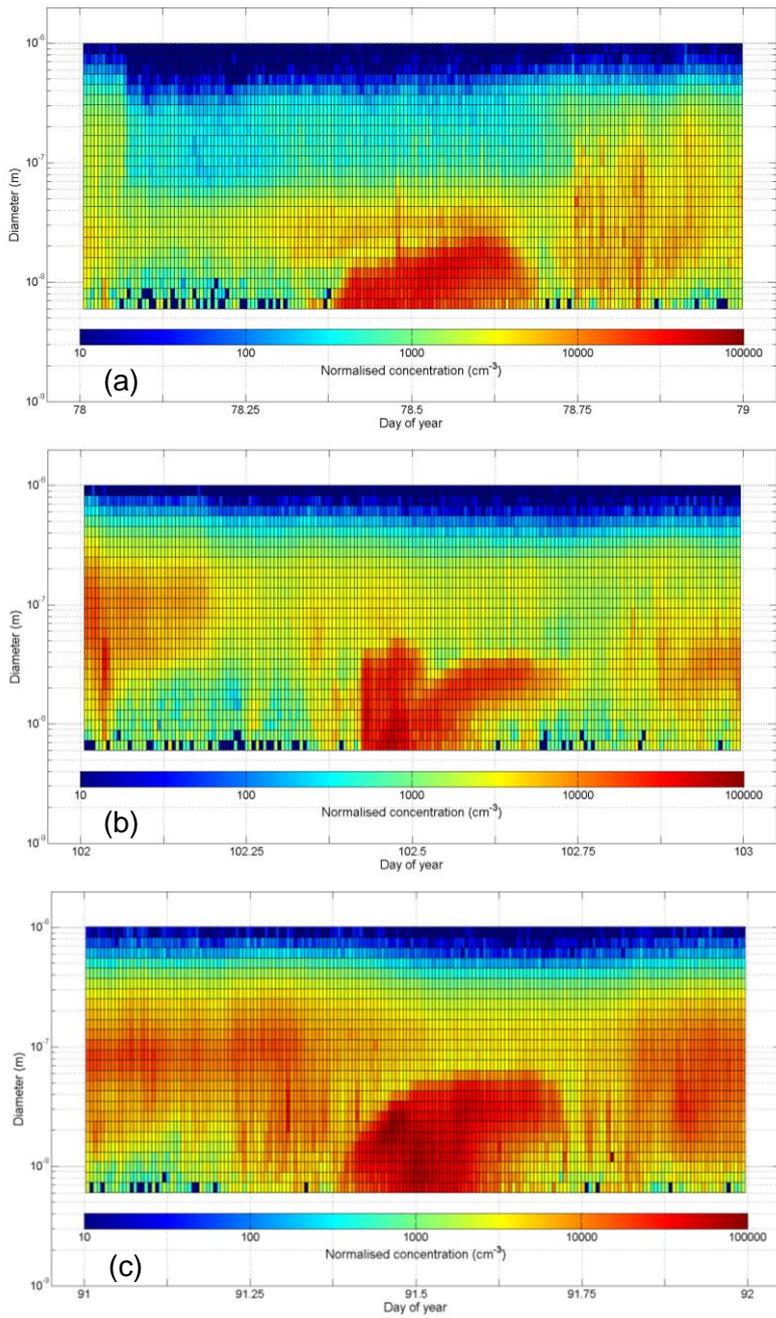
1154

1155 **Table S3.** Ranges, averages and standard deviations of daily median concentrations of SO₂, O₃, NO_x
 1156 and CO gases on quantifiable (class 1A) new particle formation and growth event days in the near-city
 1157 background and city centre separately for the 1-year long measurement time intervals and for the joint
 1158 5-year long city centre data set.
 1159

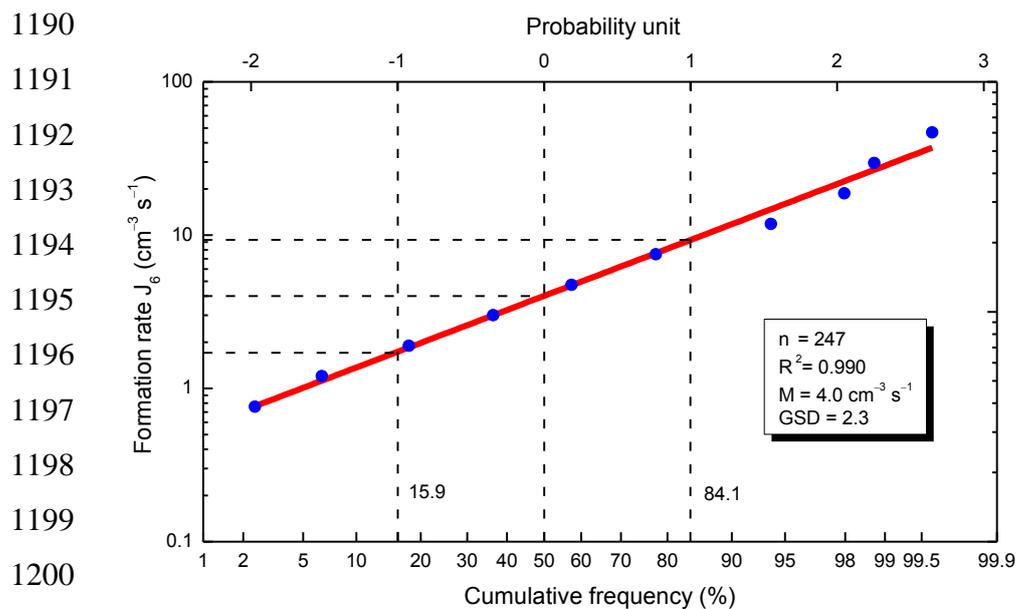
Environment	Background		Centre				
	2012– 2013	2008– 2009	2013– 2014	2014– 2015	2015– 2016	2017– 2018	All 5 years
SO ₂ concentration (µg m ⁻³)							
Minimum	4.4	3.4	2.0	0.90	3.3	0.80	0.80
Median	6.2	5.3	5.1	3.9	5.2	3.7	4.8
Maximum	11.7	8.3	8.2	10.4	11.4	7.0	11.4
Mean	6.5	5.4	5.1	4.4	5.9	3.9	4.7
St. deviation	1.4	1.2	1.8	2.4	2.4	1.8	2.1
O ₃ concentration (µg m ⁻³)							
Minimum	8.7	1.80	0.80	10.3	13.0	3.7	0.80
Median	61	44	25	35	36	29	31
Maximum	85	93	67	66	61	68	93
Mean	55	39	28	33	37	31	33
St. deviation	21	28	19	14	14	17	19
NO _x concentration (µg m ⁻³)							
Minimum	4.9	13.0	34	32	30	17.8	13.0
Median	12.2	49	72	87	72	75	74
Maximum	66	213	143	186	120	167	213
Mean	15.8	62	77	96	76	79	81
St. deviation	12.1	42	28	41	24	33	38
CO concentration (mg m ⁻³)							
Minimum	0.167	0.26	0.30	0.26	0.29	0.20	0.198
Median	0.31	0.48	0.56	0.54	0.42	0.52	0.51
Maximum	0.87	0.76	0.79	0.95	0.88	0.86	0.95
Mean	0.38	0.47	0.54	0.55	0.46	0.51	0.52
St. deviation	0.18	0.13	0.14	0.16	0.16	0.15	0.15

1160

1161
1162
1163
1164
1165
1166
1167
1168
1169
1170
1171
1172
1173
1174
1175
1176
1177
1178
1179
1180
1181
1182
1183
1184
1185



1186 **Figure S1.** Size distribution surface plots for new aerosol particle formation and consecutive particle
1187 diameter growth process as banana-shape plot with limited growth of particles on 19–03–2017 (a), with
1188 an emission interference on 12–04–2015 (b) and with a broad unresolvable onset on 01–04–2017 (c) in
1189 the city centre.



1202 **Figure S2.** Log-probability graph of the formation rate J_6 and its cumulative frequency distribution for

1203 n individual data in the joint overall 6-year long data set. The linear line in red represents the apparent

1204 fit to the data. Coefficient of determination (R^2), median J_6 value (M) and its geometric standard

1205 deviation (GSD) obtained from the fitted line are also shown.

Consequences of dynamic and timing properties of new aerosol particle formation and consecutive growth events

Imre Salma and Zoltán Németh

Institute of Chemistry, Eötvös University, H-1518 Budapest, P.O. Box 32, Hungary

Correspondence to: Imre Salma (salma@chem.elte.hu)

Abstract. A variety of contributions to the emerging research field of urban atmospheric new particle formation (NPF) and consecutive particle diameter growth based on gradually generating, several-year long, semi-continuous, critically evaluated, complex and coherent data sets are presented here. Dynamic properties, i.e. particle formation rate J_6 and particle diameter growth rate GR_{10} , and timing properties, i.e. starting time (t_1) and duration time interval (Δt) of 247 quantifiable atmospheric NPF and growth events identified in the city centre and near-city background of Budapest over 6 full measurement years together with related gas-phase H_2SO_4 proxy, condensation sink (CS) of vapours, basic meteorological data and concentrations of criteria pollutant gases were derived, evaluated, discussed and interpreted. In the city centre, nucleation ordinarily starts at 09:15 UTC+1, and it is maintained for approximately 3 h. The NPF and growth events produce 4.6 aerosol particles with a diameter of 6 nm in 1 cm^3 of air in 1 s, and cause the particles with a diameter of 10 nm to grow with a typical rate of 7.3 nm h^{-1} . Nucleation starts approximately 1 h earlier in the near-city background, it shows substantially smaller J_6 (with a median of $2.0\text{ cm}^{-3}\text{ s}^{-1}$) and GR_{10} values (with a median of 5.0 nm h^{-1}), while the duration of nucleation is similar to that in the centre. Monthly distributions of the dynamic properties and daily maximum H_2SO_4 proxy do not follow the mean monthly pattern of the event occurrence frequency. The factors that control the event occurrence and that govern the intensity of particle formation and growth are not directly linked. ~~Condensing atmospheric chemical species and/or their processes in the city centre seem to contribute equally to new particle formation and particle growth. In the near city background, however, chemical compounds available and their processes power particle growth more than particle formation.~~ New particle formation and growth processes advance in a different manner in the city and its close environment. This could mainly be related to diversities in atmospheric composition, chemistry and physics. We showed that there is a minimum growth rate (1.8 nm h^{-1} is our case) that is required for nucleated particles to reach the lower end of the diameter interval measured (in our case 6 nm). Monthly distributions and relationships among the properties mentioned

32 provided indirect evidence that chemical species other than H₂SO₄ largely influence the particle
33 growth and possibly atmospheric NPF process as well. The J_6 , GR₁₀ and Δt can be described
34 by log-normal distribution function. Most of the extreme dynamic properties could not be
35 explained by H₂SO₄ proxy, CS, meteorological data or pollutant gas concentrations.
36 Approximately 40% of the NPF and growth events exhibited broad beginning, which can be an
37 urban feature. For 9% of all quantifiable days, it was feasible to calculate 2 separate sets of
38 dynamic properties. The later onset frequently shows more intensive particle formation and
39 growth than the first onset by a typical factor of approximately 1.4. The first event is attributed
40 to regional type, while the second event, superimposed on the first, is often associated with sub-
41 regional, thus urban NPF and growth process.

42

43 **1 Introduction**

44

45 Molecules and molecular fragments in the air collide randomly and can form electrically neutral
46 or charged clusters. Most clusters decompose shortly. Chemical stabilising interactions among
47 certain components within a cluster can enhance its lifetime, during which it can grow further
48 by additional molecular collisions through some distinguishable size regimes (Kulmala et al.,
49 2014). If the diameter of these clusters reaches a critical value of 1.5 ± 0.3 nm (Kulmala et al.,
50 2013), they become thermodynamically stable, and their further growth turns into a spontaneous
51 process. Supersaturation is a necessary atmospheric condition for this principal transformation.
52 It is essentially a phase transition, which takes place in a dispersed manner in the atmosphere,
53 so it generates an aerosol system. The newly formed particles grow further by condensation to
54 larger sizes in most cases due to the existing supersaturation. Photochemical oxidation products
55 such as H₂SO₄ (Sipilä et al., 2010), extremely low-volatile organic compounds (ELVOCs, Ehn
56 et al., 2014; Jokinen et al., 2015) and highly oxygenated molecules (HOMs, Bianchi et al., 2016;
57 Kirkby et al., 2016; Tröstl et al., 2016) together with H₂O vapour, NH₃ (Kirkby et al., 2011),
58 amines (Almeida et al., 2013), other oxidation products of volatile organic compounds (VOCs;
59 Metzger et al., 2010; Schobesberger et al., 2013; Riccobono et al., 2014) and NO₂ can play an
60 important role in both the particle formation and growth. The VOCs include compounds of both
61 anthropogenic and biogenic origin, mainly isoprenoids such as α -pinene (Kirkby et al., 2016).
62 In some specific coastal regions, iodine oxides produced from marine biota are involved
63 (O'Dowd et al., 2002). Atmospheric concentration of these key compounds at a level that is
64 smaller by 12–14 orders of magnitude than the concentration of air molecules is already
65 sufficient for the phenomenon (Kulmala et al., 2014). Relative importance of the organics

66 increases with particle size (Riipinen et al., 2011; Ehn et al., 2014), and their supersaturation is
67 maintained by fast gas-phase autooxidation reactions of VOCs (Crouse et al., 2013). The
68 overall phenomenon is ordinarily confined in time for 1 day or so, and, therefore, it can be
69 regarded as an event in time, and is referred as new aerosol particle formation (NPF) and
70 consecutive particle growth event.

71
72 Such events appear to take place almost everywhere in the world and anytime (Kulmala et al.,
73 2004; Kerminen et al., 2018; Nieminen et al., 2018). Their occurrence frequency and, more
74 importantly, their contribution to particle number concentrations were found to be substantial
75 or determinant in the global troposphere (Spracklen et al., 2006; Kulmala et al., 2014).
76 Moreover, their contribution to the number of cloud condensation nuclei (CCN) can be 50% or
77 even more (Makkonen et al., 2009; Merikanto et al., 2009; Sihto et al., 2011), which links the
78 events to climate system, and emphasizes their global relevance (Kerminen et al., 2012;
79 Makkonen et al., 2012; Carslaw et al., 2013; Gordon et al., 2016). New particle formation and
80 growth events were proved to be common in polluted air of large cities as well with a typical
81 relative occurrence frequency between 10% and 30% (Woo et al., 2001; Baltensperger et al.,
82 2002; Alam, et al., 2003; Wehner et al., 2004; Salma et al., 2011; Dall'Osto et al., 2013; Xiao
83 et al., 2015; Zhang et al., 2015; Kulmala et al., 2017, Nieminen et al., 2018). The coupling and
84 relationships between regional and urban (sub-regional) NPF were demonstrated at least under
85 favourable orographic conditions (Salma et al., 2016b). New particle formation can increase
86 the existing particle number concentrations in city centres by a factor of approximately 2 on
87 nucleation days, while it can produce approximately 28% of ultrafine (UF) particles on a longer
88 (e.g. annual) time scale (Salma et al., 2017). Particle concentrations from NPF are also
89 important when compared to the (primary) particles emitted by their dominant source in cities,
90 namely by road vehicles with internal combustion engines (Paasonen et al., 2016). These results
91 jointly suggest that particles from NPF and growth events in cities can influence not only the
92 urban climate but can contribute to the public's excess health risk from particle number
93 exposures (Oberdörster et al., 2005; Braakhuis et al., 2014; Salma et al., 2015), and,
94 furthermore, could be linked to the role of human actions in all these effects.

95
96 Despite these potentials, conclusive interpretation of the data obtained, and results derived
97 specifically for cities remained hindered so far. Several-year long, semi-continuous, critically
98 evaluated, complex and coherent data sets are required for this purpose, which have been
99 generating gradually. As part of this international progress, investigations dedicated to urban

100 NPF and growth events in Budapest have been going on since November 2008. Measurements
101 for 5 full years were realised in the city centre at a fixed location, 1 full year was devoted to
102 measurements in a near-city background environment, and some other measurements were
103 accomplished in different urban microenvironments for time intervals of a few months. The
104 main objectives of this study are to determine, present and analyse the dynamic properties, i.e.
105 particle formation rate and particle diameter growth rate, timing properties, i.e. starting time
106 and duration time interval of nucleation process of NPF and growth events together with the
107 major sources and sink of condensing vapours, basic meteorological data and criteria pollutant
108 gases for 6 years, to investigate and interpret their relationships, to discuss their monthly
109 distributions, to evaluate and detect some of their features specific for urban atmospheric
110 environments, and to demonstrate some specific urban influence on the calculation of the
111 properties. These quantities and relationships are of basic importance in many atmospheric
112 processes for several reasons. **Our goals are in line with the research needs for global
113 atmospheric nucleation studies (Kerminen et al., 2018; Nieminen et al., 2018).**

114

115 **2 Experimental methods**

116

117 The measurements took place at two urban locations in Budapest, Hungary. Most measurements
118 were realised at the Budapest platform for Aerosol Research and Training (BpART) facility (N
119 47° 28' 29.9", E 19° 3' 44.6", 115 m above mean sea level (a.s.l.; Salma et al., 2016a). This site
120 represents a well-mixed, average atmospheric environment for the city centre. The other
121 location was situated at the NW border of Budapest in a wooded area of the Konkoly
122 Astronomical Observatory of the Hungarian Academy of Sciences (N 47° 30' 00.0", E 18° 57'
123 46.8", 478 m a.s.l.). This site characterises the air masses entering the city since the prevailing
124 wind direction in the area is NW. The measurements were accomplished for 6 full-year long
125 time intervals, i.e. from 03–11–2008 to 02–11–2009, from 19–01–2012 to 18–01–2013, from
126 13–11–2013 to 12–11–2014, from 13–11–2014 to 12–11–2015, from 13–11–2015 to 12–11–
127 2016 and from 28–01–2017 to 27–01–2018. In the measurement year 2012–2013, the
128 instruments were set up in the near-city background, while in all other years, they were installed
129 in the city centre. Local time (LT=UTC+1 or daylight-saving time, UTC+2) was chosen as the
130 time base of the data unless otherwise indicated because **it had been observed in earlier
131 investigations that** the daily activity time pattern of inhabitants substantially influences the
132 atmospheric processes in cities (Salma et al., 2014).

133

134 The main measuring system was a flow-switching type differential mobility particle sizer
135 (DMPS). It consists of a radioactive (^{60}Ni) bipolar charger, a Nafion semi-permeable membrane
136 dryer, a 28-cm long Vienna-type differential mobility analyser and a butanol-based
137 condensation particle counter (TSI, model CPC3775). The sample flow was 2.0 L min^{-1} in the
138 high-flow mode, and 0.31 L min^{-1} in the low-flow mode with sheath air flow rates 10 times
139 larger than for the sample flows. The DMPS measures particle number concentrations in an
140 electrical mobility diameter range from 6 to 1000 nm in the dry state of particles (with a relative
141 humidity of $\text{RH} < 30\%$) in 30 channels, which finally yields 27 channels after averaging 3
142 overlapping channels when joining the data for the 2 flow modes. The time resolution of the
143 measurements was approximately 10 min till 18-01-2013, and 8 min from 13-11-2013 (after
144 a planned update of the DMPS system). There was no upper size cut-off inlet applied to the
145 sampling line, and a weather shield and insect net were only attached. The sampling inlet was
146 installed at a height of 12.5 m above the street level in the city centre, and of approximately 1.7
147 m in the near-city background. The measurements were performed according to the
148 international technical standard (Wiedensohler et al., 2012). The availability of the DMPS data
149 over 1-year long time intervals are summarised in Table 1. Synoptic meteorological data for air
150 temperature (T), RH, wind speed (WS) and wind direction (WD) were obtained from a
151 measurement station of the Hungarian Meteorological Service (HMS, no. 12843) by
152 standardised methods with a time resolution of 1 h. Global solar radiation (GRad) data were
153 measured by the HMS at a distance of 10 km in E direction with a time resolution of 1 h.
154 Meteorological data were available in $>90\%$ of the possible cases in each year. Concentrations
155 of SO_2 , O_3 , NO_x and CO were obtained from measurement stations of the National Air Quality
156 Network in Budapest (in a distance of 4.5 km from the urban site, and of 6.9 km from the near-
157 city background site) located in the upwind prevailing direction from the measurement sites.
158 They are measured by UV fluorescence (Ysselbach 43C), UV absorption (Ysselbach 49C),
159 chemiluminescence (Thermo 42C) and IR absorption methods (Thermo 48i), respectively with
160 a time resolution of 1 h. The concentration data were available in $>85\%$ of the yearly time
161 intervals, and $>98\%$ of them were above the limit of determinations. It is worth mentioning that
162 concentration of SO_2 in the Budapest area is ordinarily distributed without larger spatial (and
163 temporal) gradients (Salma et al., 2011). For the present study, this was proved by evaluating
164 the concentration ratios from 2 different municipal stations which are in the closest distance
165 from the BpART in 2 different directions with an angle of 60° between them. The mean SO_2
166 concentration ratio and standard deviation (SD) for the 2 stations were $81 \pm 20\%$ over the 5-year

167 long measurement time interval. The assumption can also be justified indirectly by a conclusion
168 on the monthly distribution of SO₂ concentration in Sect. 4.2.

169

170 **3 Data treatment**

171

172 The measured DMPS data were evaluated according to the procedure protocol recommended
173 by Kulmala et al. (2012) with some refinements that are related to urban features (see Sect. 3.1).
174 Particle number concentrations in the diameter ranges from 6 to 1000 nm (N), from 6 to 25 nm
175 (N_{6-25}) and from 6 to 100 nm (N_{6-100} or UF particles) were calculated from the measured and
176 inverted DMPS concentrations. Particle number size distribution surface plots showing jointly
177 the variation in particle diameter and particle number concentration density in time were also
178 derived. Identification and classification of NPF and growth events was accomplished on these
179 surface plots (Dal Maso et al., 2005; Németh et al., 2018) on a daily basis into the following
180 main classes: NPF event days, non-event days, days with undefined character, and days with
181 missing data (for more than 4 h during the midday). Relative occurrence frequency of events
182 was determined for each month and year as the ratio of the number of event days to the total
183 number of relevant (i.e. all–missing) days. A subset of NPF events with uninterrupted evolution
184 in time, which are called quantifiable (class 1A) events, were further separated because the time
185 evolution of their size distribution functions was utilised to determine the dynamic and timing
186 properties with good accuracy and reliability.

187

188 **3.1 Dynamic and timing properties**

189

190 Growth rate (GR) of nucleation-mode particles was calculated by mode-fitting method
191 (Kulmala et al., 2012). Particle number median mobility diameter (NMMD) of the nucleation
192 mode were obtained from fitting the individual size distributions by DoFit algorithm (Hussein
193 et al., 2004). The growth rate was determined as the slope of the linear line fitted to the time
194 series of the NMMD data within a time interval around a diameter d , where the dependency
195 could be satisfactorily approximated by linear fit. Since the nucleation mode was mostly
196 estimated by N_{6-25} in the calculations of the formation rate (see below), and since the median
197 of the related diameter interval (from 6 to 25 nm) is close to $d=10$ nm, GRs for particles with a
198 diameter of 10 nm were determined (GR₁₀). This type of GR can be interpreted as an average
199 GR as far as the given particle diameter range is concerned, but it actually expresses the

200 beginning of the growth process only, which may have considerable effects on the formation
201 rate calculations in specific cases (see later).

202

203 Time evolution of an aerosol population is described by the general dynamic equation which
204 was rearranged, simplified and approximated by several quantities (Kulmala et al., 2001; Dal
205 Maso et al., 2002; Kulmala et al., 2012; Cai and Jiang, 2017) to express the formation rate J_6
206 of particles with the smallest detected diameter of $d_{\min}=6$ nm in a form utilised in the present
207 evaluation as

208

$$209 \quad J_6 = \frac{dN_{6-25}}{dt} + \text{CoagS}_{10} N_{6-25} + \frac{\text{GR}_{10}}{(25-6)} N_{6-25} - \frac{dN_{\text{Ai}, <25}}{dt}. \quad (1)$$

210

211 The first term on the right side of Eq. 1 expresses the concentration increment. The particle
212 number concentration in the size range from 6 to 25 nm (i.e. N_{6-25}) was selected to approximate
213 the nucleation-mode particles $N_{\text{nuc}} \approx N_{6-25}$. This is a usual and reasonable choice because it was
214 proved to be advantageous and effective way in handling fluctuating data sets since N_{6-25} often
215 exhibits less sensitivity and smaller scatter in time than the fitted area of the nucleation mode.
216 It is implicitly assumed that the intensity of the NPF is constant for a certain time interval, and,
217 therefore, dN_{6-25}/dt can be determined as the slope of the linear function of N_{6-25} versus time t
218 within an interval where the dependence could be satisfactorily approximated by linear fit. A
219 limitation of the relatively wide size range (6–25 nm) selected can be manifested by
220 disturbances from primary particles particularly in urban environments. This is taken into
221 account by the last term of Eq. 1 and is discussed later.

222

223 The second term on the right side of Eq. 1 represents the loss of particles due to coagulation
224 scavenging (with pre-existing particles). The coagulation scavenging efficiency for particles
225 with a diameter of 10 nm (CoagS_{10}) was selected to approximate the mean coagulation
226 efficiency of nucleation-mode particles ($\text{CoagS}_{\text{nuc}}$). This diameter was chosen by considering
227 the median of the related diameter range, which was discussed above for GR. The coagulation
228 efficiency was calculated from classical aerosol mechanics with adopting a mass
229 accommodation coefficient of 1 and utilizing the Fuchs' transition-regime correction factor
230 (Kulmala et al., 2001; Dal Maso et al., 2005; Kulmala et al., 2013) by using computation scripts
231 developed at the University of Helsinki. Self-coagulation within the nucleation mode was

232 neglected due to its limited concentration. Hygroscopic growth of particles was not considered
233 since this depends on chemical composition of particles, which is unknown.

234

235 The third term on the right side of Eq. 1 expresses the particle growth out of the considered size
236 range by condensation of vapours. The GR_{10} was selected to approximate a representative value
237 at the median of the particle diameter range considered (Vuollekoski et al., 2012). It is implicitly
238 assumed that GR_{10} can be regarded to be constant over the time interval under consideration.
239 Nevertheless, the growth of nucleation-mode particles in time is occasionally limited (Fig. S1a).
240 In these specific cases, the mean relative area of the nucleation mode below 25 nm was
241 determined by fitting the individual size distributions around the time of the maximum
242 nucleation-mode NMMD, and the ratios were averaged. A correction in form of the mean
243 relative area was adopted as a multiplication factor for the growth out term in Eq. 1. On very
244 few days, the growth of newly formed particles was followed by a decrease in nucleation-mode
245 NMMD (Salma et al., 2016a). In these cases, the shrinkage rate (with a formal $GR_{10}<0$) was
246 derived and adopted in Eq. 1.

247

248 The fourth term on the right side of Eq. 1 expresses the contribution of high-temperature
249 emission sources, usually of vehicular road traffic (Paasonen et al., 2016; Salma et al., 2017) to
250 N_{6-25} , which can provisionally disturb the assumption of $N_{nuc}\approx N_{6-25}$. A typical example of such
251 a situation is shown in Fig. S1b from 10:09 to 12:23 LT. In these specific cases, the contribution
252 of primary emissions was estimated from the slope of the time series of the fitted peak area of
253 the Aitken mode below $d<25$ nm ($N_{Ai,<25}$) in the time region under consideration. Reliable
254 separation of the nucleation and Aitken modes from each other was hindered or was not possible
255 for a few individual size distributions due to overlapping modes and the scatter in the
256 concentration data, and these individual Aitken-mode areas were excluded from or skipped in
257 the time series. Relative contributions of the concentration increment coagulation loss and
258 growth out from the diameter interval to J_6 are decreasing in this order with mean values of
259 71%, 17% and 12%, respectively (Table S1).

260

261 The formation and growth rates for the measurement years of 2008–2009 and 2012–2013 were
262 calculated earlier by a slightly different way and neglecting the urban features discussed above
263 (Salma et al., 2011, 2016b). To obtain consistent data sets, the dynamic properties for these 2
264 years were re-evaluated by adopting the present improved protocol and implementing the
265 experience gained over the years. The mean new-to-old rate ratios with SDs for the GR_{10} and

266 J_6 were 1.06 ± 0.32 and 1.23 ± 0.37 , respectively in the city centre (2008–2009) and 1.04 ± 0.21
267 and 1.20 ± 0.35 , respectively in the near-city background (2012–2013). It was the smaller rates
268 that were primarily and sometimes substantially impacted. The modifications were
269 simultaneously adopted. The subtraction of particle number concentrations emitted by road
270 traffic from N_{6-25} , which usually leads to a decrease in the coagulation loss and loss due to
271 growth out from the diameter range of 6–25 nm, and which can influence the slope of the
272 concentration change in time (dN_{6-25}/dt) in a positive or negative manner depending on the
273 actual time evolution of perturbing emission source. In addition to that, the time interval in
274 which this slope is considered to be constant was set in a new treatment. It is mentioned that
275 the relative contributions of the concentration increment, coagulation loss and growth out from
276 the diameter interval to J_6 have different weights in propagating their effects. Furthermore, J_6
277 itself also depends on the GR_{10} , which makes the relationships even more complex. These
278 explain why the changes resulted in overall increments. The re-calculation is considered as a
279 methodological improvement over the years of research.

280

281 The assumptions and estimations above usually represent a reasonable approximation to reality.
282 The N_{6-25} is derived from the experimental data in a straightforward way, the GR_{10} and the
283 corrections for primary particles and limited particle growth depend on the quality of the size
284 distribution fitting as well, while the $CoagS_{10}$ is determined by using a theoretical model. The
285 resulting accuracies of the dynamic properties, in particular of J_6 , look rather complicated. They
286 also depend on the spatial heterogeneity in the air masses measured particularly for the
287 observations performed at the fixed site, size and time resolution of the concentrations
288 measured, diameter range of the size distributions, fluctuations in the experimental data,
289 selection of the particle diameter interval, choice of the time interval of interest (for linear fits),
290 sensitivity of the models to the input uncertainties (Vuollekoski et al., 2012), and also on the
291 extent of the validity of the assumptions applied under highly polluted conditions (Cai and
292 Jiang, 2017). The situation is further complicated with the fact that the dynamic (and also the
293 timing) properties are connected to each other. Finally, it is important to recognise that some
294 NPF and growth curves on the surface plots have rather broad starting time interval (Fig. S1a
295 and S1c). They occur in a considerable abundance in cities, e.g. in 40% of all quantifiable events
296 in Budapest (Sect. 4.4). This may yield badly defined or composite dynamic properties, whose
297 uncertainty can have principle limitations which can prevail on the experimental and model
298 uncertainties.

299

300 Timing properties of NPF and growth events are increasingly recognised, and they can provide
301 valuable information even if they are estimated indirectly from the observed diameter interval
302 >1.5 nm (Sect. 1). The earliest estimated time of the beginning of a nucleation (t_1) and the latest
303 estimated time of the beginning of a nucleation (t_2) were derived by a comparative method
304 (Németh and Salma, 2014) based on the variation in the content of the first size channel of the
305 DMPS system. Both time parameters include a time shift that accounts for the particle growth
306 from the stable neutral cluster mode at approximately 2 nm to the smallest detectable diameter
307 limit of the DMPS systems (6 nm in our case) by adopting the GR value in the size window
308 nearest to it in size space. The difference $\Delta t=t_2-t_1$ is considered as the duration time interval of
309 the nucleation process. It represents the time interval during which new aerosol particles are
310 generated in the air. **The timing properties are expressed in UTC+1**, and their uncertainty is
311 regarded to be ca. 30 min under ordinary NPF and growth situations.

312

313 **3.2 Sources and sink**

314

315 Relative effects and role of gas-phase H_2SO_4 were estimated by its proximity measure (proxy
316 value) containing both its major source and sink terms under steady-state conditions according
317 to Petäjä et al. (2009). It was calculated for $\text{GRad} > 10 \text{ W m}^{-2}$. Formally, it is possible to convert
318 the H_2SO_4 proxy values to H_2SO_4 concentrations by an empirical scaling factor of $k=1.4 \times 10^{-7}$
319 $\times \text{GRad}^{-0.70}$, where GRad is expressed in a unit of W m^{-2} (Petäjä et al., 2009). The factor was,
320 however, derived for a remote boreal site, and, therefore, we prefer not to perform the
321 conversion since urban areas are expected to differ from the boreal regions, and adopting the
322 factor could distort the dynamic relationships or time trends investigated. The conversion was
323 applied only to estimate the order of average H_2SO_4 atmospheric concentration levels. The
324 results derived by utilising the proxy are subject to larger uncertainties than for the other
325 properties because of these limitations, but they may indicate well gross tendencies.

326

327 Condensation sink for vapour molecules onto the surface of existing aerosol particles was
328 computed for discrete size distributions as described in earlier papers (Kulmala et al., 2001; Dal
329 Maso et al., 2002, 2005) and summarised by Kulmala et al. (2013). The equilibrium vapour
330 pressure of the condensing species was assumed to be negligible at the surface of the particles,
331 so similar to sulfuric acid. Dry particle diameters were considered in the calculations.

332

333 **4 Results and discussion**

334

335 Annual median particle number concentrations based on the individual data in the near-city
336 background in 2012–2013, and in the city centre for the separate measurement years of 2008–
337 2009, 2013–2014, 2014–2015, 2015–2016 and 2017–2018 were 3.4×10^3 , and 11.5×10^3 ,
338 9.7×10^3 , 9.3×10^3 , 7.5×10^3 and $10.6 \times 10^3 \text{ cm}^{-3}$, respectively. The data for the city centre indicate
339 a decreasing trend. **The first 4 values unambiguously show a decrease, while the last data point**
340 **may look somewhat differently. Rigorous statistical evaluation of the joint data set of particle**
341 **number concentrations in various size fractions over a decennial time interval from November**
342 **2008 to November 2018 is in progress, and its preliminary results in the one hand, confirm the**
343 **decreasing tendency, and in the other hand, reveal some fine structure to this dependency.** The
344 mean UF/N ratio with SD for the same measurement time intervals were $67 \pm 14\%$, and $79 \pm 6\%$,
345 $75 \pm 10\%$, $75 \pm 11\%$, $76 \pm 11\%$ and $80 \pm 10\%$, respectively. The values correspond to ordinary
346 urban atmospheric environments in Europe (Putaud et al., 2010).

347

348 An overview on the number of classified days separately for the 1-year long measurement time
349 intervals is given in Table 1. The availability of the daily size distribution surface plots with
350 respect to all days ensures that the data are representative on yearly and monthly time scales,
351 except for the months August and September 2015, when there were missing days in larger
352 ratios. The number of quantifiable event days (248 cases) is also considerable, which establishes
353 to arrive at firm conclusion on the NPF and growth events as well.

354

355 **Table 1.** Number of days with new aerosol particle formation and growth event, quantifiable (class 1A)
 356 event days, non-event days, undefined days, missing days and the coverage of relevant days with respect
 357 to all days in the near-city background and city centre separately for the 1-year long measurement time
 358 intervals.
 359

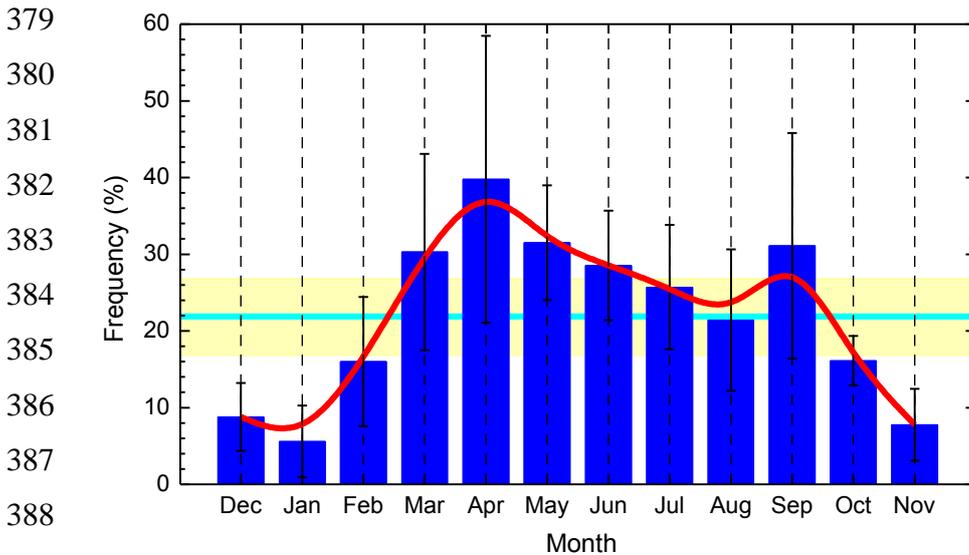
Environment	Background		Centre				
	Time interval	2012–2013	2008–2009	2013–2014	2014–2015	2015–2016	2017–2018
Event days		96	83	72	81	35	83
Quantifiable days		43	31	48	56	18	52
Undefined days		19	34	24	25	8	23
Non-event days		231	229	267	240	226	257
Coverage (%)		95	95	99	95	73	99
Missing days		20	19	2	19	97	2

360

361

362 It was previously shown that the NPF and growth events observed in the city centre of Budapest
 363 and its background ordinarily happen above a larger territory or region in the Carpathian Basin
 364 (Salma et al., 2011; Németh and Salma, 2014), and they are linked to each other (Salma et al.,
 365 2016b). From the point of the occurrence frequency distribution, they can, therefore, be
 366 evaluated jointly in the first approximation. An overall monthly mean relative occurrence
 367 frequency of nucleation days derived for all 6 measurement years is shown in Fig. 1. The annual
 368 mean frequency with SD was $22\pm 5\%$, which is considerable and is in line with other urban sites
 369 (Sect. 1). The monthly mean frequency has a temporal variation, which can be characterised by
 370 a noteworthy pattern. The mean monthly dependency exhibits an absolute and a local minimum
 371 in January (5.6%) and August (21%), respectively, and an absolute and a local maximum in
 372 April (40%), and September (31%), respectively. Nevertheless, the SDs of the monthly (and
 373 annual) means indicate prominent variability from year to year. The pattern can be related to
 374 multivariate relationships and complex interplay among the influencing factors, which include
 375 the air temperature (January is the coldest month, while August is the warmest month in the
 376 Carpathian Basin) and enhanced emission of biogenic VOCs in springtime (March–April) and
 377 early autumn (September) as well (Salma et al., 2016b).

378



379
380
381
382
383
384
385
386
387
388
389
390 **Figure 1.** Monthly mean relative occurrence frequency of new aerosol particle formation and
391 consecutive particle diameter growth events with respect to the number of relevant days for the joint 6-
392 year long data set. The error bars show ± 1 standard deviation, the horizontal line in cyan indicates the
393 overall annual mean frequency, the yellow bands represent ± 1 standard deviation of the annual mean,
394 and the smooth curve in red serves to guide the eye.

395
396 The properties and variables studied were derived in full time resolution. They were averaged
397 in several ways for different conditions and for various purposes to obtain typical average
398 descriptive characteristics. In 1 case (31–08–2016), the NPF and growth event could reliable
399 be identified, while the measured absolute particle number concentrations could not be
400 validated due to some experimental troubles, and, therefore, it was left out from the further
401 calculations. Similarly, there were 1 and 4 events with unusually/extraordinarily large dynamic
402 properties in the measurement years 2014–2015 and 2017–2018, respectively. More
403 specifically, 5 individual J_6 data when expressed in a unit of $\text{cm}^{-3} \text{s}^{-1}$ and 1 individual GR_{10}
404 data when given in nm h^{-1} were >20 (Table 3). These extremes were left out from the overview
405 statistics to maintain the representativity (they could be influenced by some unknown extra or
406 very local sources) and to fulfil better the basic requirements of correlation analysis. If an event
407 showed a double beginning then the dynamic properties for the first onset were considered in
408 the basic overview since this onset is of regional relevance (Salma et al., 2016b). The extreme
409 NPF and growth events and the characteristics for the second onsets were, however, evaluated
410 separately and are discussed in detail and interpreted in Sect. 4.4.

411

412 **4.1 Ranges and averages**

413

414 Ranges and averages with SDs of formation rate J_6 , growth rate GR_{10} , starting time of
415 nucleation (t_1) and duration time interval of nucleation (Δt) are summarised in Table 2 for
416 separate measurement years and for the joint 5-year long city centre data set. In the city centre,
417 nucleation generally starts at 09:15 UTC+1, and it is typically maintained for approximately 3
418 h. The NPF and growth events ordinarily produce 5.6 new aerosol particles with a diameter of
419 6 nm in 1 cm³ of air in 1 s, and cause the particles with a diameter of 10 nm to grow with a
420 typical rate of 7.6 nm h⁻¹. The statistics for J_6 and GR_{10} are based on 199 and 203 events,
421 respectively. The corresponding data for the separate years show considerable variability
422 without obvious trends or tendencies. The differences between the years can likely be related
423 to changes in actual atmospheric chemical and physical situations and conditions, and to the
424 resulting modifications in the sensitive balance and delicate coupling among them from year to
425 year. Spread of the individual data for GR_{10} is smaller than for J_6 ; the relative SDs for the joint
426 5-year long city centre data set were 38% and 68%, respectively, while the (external) relative
427 SDs calculated from the annual mean values were 4.2% and 14.0%, respectively.

428

429 The dynamic properties and t_1 data tend to be smaller in the near-city background than in the
430 city centre. In general, nucleation starts 1 h earlier in the background, and the events typically
431 show significantly smaller J_6 (with a median of 2.0 cm⁻³ s⁻¹) and GR_{10} (with a median of 5.0
432 nm h⁻¹). Duration of the nucleation is very similar to that in the city centre. All starting times
433 of nucleation were larger than (in a few cases, very close to) the time of the sunrise. This implies
434 that no nocturnal NPF and growth event has been identified in Budapest so far. The particle
435 growth process (the so-called banana curve) could be traced usually for a longer time interval
436 (up to 1.5 d) in the background than in the centre.

437

438 These results are in line with the ideas on atmospheric nucleation and consecutive particle
439 growth process (e.g. Kulmala et al., 2014; Zhang et al., 2015; Kerminen et al., 2018). It was
440 observed in a recent overview study (Nieminen et al., 2018) that the formation rate of 10–25
441 nm particles increased with the extent of anthropogenic influence, and in general, it was 1–2
442 orders of magnitude larger in cities than at sites in remote and clean environments. This
443 highlights the importance of some anthropogenic vapours such as SO₂, NH₃ and amines to NPF
444 and growth. The data also confirm our earlier findings with respect to Budapest and its regional

445 background within the Carpathian Basin achieved with shorter, 2-year long data sets (Salma et
 446 al., 2016b)

447
 448 **Table 2.** Ranges, averages and standard deviations of aerosol particle formation rate J_6 , particle diameter
 449 growth rate GR_{10} , starting time (t_1) and duration time interval ($\Delta t=t_2-t_1$) of nucleation process of
 450 quantifiable (class 1A) new particle formation and growth events in the near-city background and city
 451 centre separately for the 1-year long measurement time intervals and for the joint 5-year long city centre
 452 data set.

453

Environment	Background		Centre				
	2012– 2013	2008– 2009	2013– 2014	2014– 2015	2015– 2016	2017– 2018	All 5 years
Formation rate J_6 ($\text{cm}^{-3} \text{s}^{-1}$)							
Minimum	0.48	1.47	1.13	0.81	1.19	1.60	0.81
Median	2.0	4.2	3.5	4.4	4.6	6.3	4.6
Maximum	5.6	15.9	17.8	18.0	15.3	17.3	18.0
Mean	2.2	4.7	5.2	5.6	5.0	6.6	5.6
St. deviation	1.3	2.6	3.7	4.2	3.7	3.3	3.8
Growth rate GR_{10} (nm h^{-1})							
Minimum	3.0	3.7	3.1	2.8	3.2	3.3	2.8
Median	5.0	7.6	6.6	6.5	8.0	7.5	7.3
Maximum	9.8	17.4	19.0	18.0	15.5	19.8	19.8
Mean	5.2	7.8	7.2	7.3	7.7	8.0	7.6
St. deviation	1.4	2.6	2.8	3.2	3.0	2.8	2.9
Starting time, t_1 (HH:mm UTC+1)							
Minimum	05:51	07:14	06:44	05:48	07:31	05:57	05:48
Median	08:19	09:26	09:22	08:48	09:45	09:18	09:15
Maximum	11:09	11:38	12:21	11:23	12:45	12:15	12:45
Mean	08:17	09:27	09:25	08:49	10:02	09:24	09:19
St. deviation	01:11	01:05	01:26	01:22	01:23	01:36	01:26
Duration time, Δt (HH:mm)							
Minimum	01:23	00:52	00:42	00:31	01:03	01:26	00:31
Median	03:16	02:36	02:04	03:53	02:31	03:49	02:57
Maximum	06:44	06:04	05:34	07:46	06:05	07:55	07:55
Mean	03:30	02:44	02:14	03:52	02:58	03:57	03:18
St. deviation	01:40	01:11	01:01	01:40	01:47	01:39	01:40

454
 455 Ranges and averages with SDs of some related atmospheric properties, namely of mean CS
 456 averaged for the time interval from t_1 to t_2 , daily maximum gas-phase H_2SO_4 proxy, daily mean
 457 T and RH (Table S2), and of daily median concentrations of SO_2 (as the major precursor of gas-

458 phase H₂SO₄), O₃ (as an indicator of photochemical activity), NO_x and CO gases (as indicators
459 of anthropogenic combustion activities and road vehicle emissions) (Table S3) were also
460 derived for quantifiable NPF and growth event days, and are further evaluated. The annual
461 mean CS values exhibited decreasing tendency in the city centre over the years (as can be
462 expected from the particle number concentrations as well). The individual values remained
463 below approximately $20 \times 10^{-3} \text{ s}^{-1}$, which agrees well with the results of our earlier study (Salma
464 et al., 2016b) according to which the CS suppresses NPF above this level in the Carpathian
465 Basin. Maximum H₂SO₄ proxy values reached substantially higher levels (by a factor of
466 approximately 2) in the near-city background than in the city centre due mainly to the
467 differences in the CS and [SO₂]. The differences between the 2 sites are particularly evident
468 when considering their smallest values. The largest variability in the annual average values
469 were observed for the proxy. Median concentration of H₂SO₄ molecules was roughly estimated
470 to be approximately $5 \times 10^5 \text{ cm}^{-3}$ by adopting the scaling factor (Sect. 3.2). ~~The dynamic~~
471 ~~properties seem to be not very sensitive to air T~~ The air *T* displayed quite similar and comparable
472 values over the years at both sites. **The discussion of its overall effect on the dynamic properties**
473 **is accomplished in Sec. 4.2, where the monthly distributions are presented.** Some events
474 happened at daily mean temperatures below zero. The daily mean RH and its SD for the city
475 centre and near-city background were $54 \pm 11\%$ and $64 \pm 12\%$, respectively. There were events
476 that occurred at RHs as high as 90%. Relationships of the dynamic properties with *T* and RH
477 are also obscured with strong seasonal cycle of these meteorological data and with the fact that
478 air masses arriving to the receptor site in different trajectories are often characterised by distinct
479 levels of meteorological data.

480

481 As far as the pollutant gases are concerned (Table S3), SO₂ showed somewhat smaller daily
482 median values, and O₃ exhibited substantially smaller levels on event days in the city centre
483 than in the near-city background, while concentrations of NO_x and CO were obviously larger
484 in the city than in its close background. The differences can primarily be explained by the
485 intensity and spatial distribution of their major sources and atmospheric chemical reactions, and
486 the joined concentration data resembles typical situations without photochemical smog
487 episodes in cities. There was no obvious decrease in SO₂ concentration during these years in
488 contrast with an earlier decreasing trend from mid-1980s till about 2000. ~~No evident or sensitive~~
489 ~~effect of atmospheric gases on the dynamic or timing properties could be deduced from the~~
490 ~~averaged data. This can probably be explained by a dedicated balance between the intensifying~~

491 ~~and suppressing effects, which were averaged out on a yearly time scale. Relationships on~~
492 ~~shorter scales are further investigated and discussed in more detail in the following sections.~~

493

494 **4.2 Monthly distributions**

495

496 Distributions of the monthly mean J_6 , GR₁₀, daily maximum gas-phase H₂SO₄ proxy, mean CS,
497 daily mean air T and RH, and daily median SO₂, O₃, NO_x and CO concentrations for quantifiable
498 NPF and growth events for the joint 5-year long city centre data sets are shown in Fig. 2. The
499 distributions – eminently for J_6 , GR₁₀, H₂SO₄ proxy and SO₂ – do not follow the monthly pattern
500 of the event occurrence frequency at all (cf. Fig. 1). Instead, the J_6 , GR₁₀ and H₂SO₄ proxy tend
501 to exhibit larger values in summer months, and they temporal changes over the other months
502 are smooth and do not show distinctive features. The elevations are substantial; the estimated
503 maximum level was larger than the baseline by a factor of 2.1 for the J_6 , and by a factor of
504 approximately 1.4 for the GR₁₀ and H₂SO₄ proxy. The intensity of solar radiation at the surface,
505 its seasonal cycling, concentration of atmospheric precursors in different months, biogenic
506 processes, anthropogenic activities and the fact that rate coefficients of many thermal
507 chemical/physicochemical processes in the nature (including GR, Paasonen et al., 2018)
508 increase with T could play an important role in explained the distributions. A more
509 comprehensive study involving chemicals and their photochemistry is required for more
510 detailed explanation.

511

512 **It is worth mentioning that [SO₂] did not change substantially for the NPF event and non-event**
513 **days, while GRad was typically larger by a factor of ca. 2 and CS was smaller by approximately**
514 **30% on event days than on non-event days. The differences in the GRad (and some other**
515 **properties) are, however, biased by the seasonal cycle of solar electromagnetic radiation via the**
516 **seasonal variation of NPF occurrence frequency, while the CS indicated a modest seasonal**
517 **dependency. Interpretation of their joint effect should be approached by care, requires further**
518 **evaluations and is to be realised fully in a further study.** Nevertheless, the misalignment among
519 the monthly distributions of NPF and growth event occurrence frequency and all the other
520 properties indicates that the occurrence or its basic causes are not linked with the dynamic
521 properties in a straightforward or linear manner, ~~and 2) gas-phase H₂SO₄ does not seem to be~~
522 ~~the controlling factor of NPF occurrence in the Carpathian Basin including Budapest.~~

523

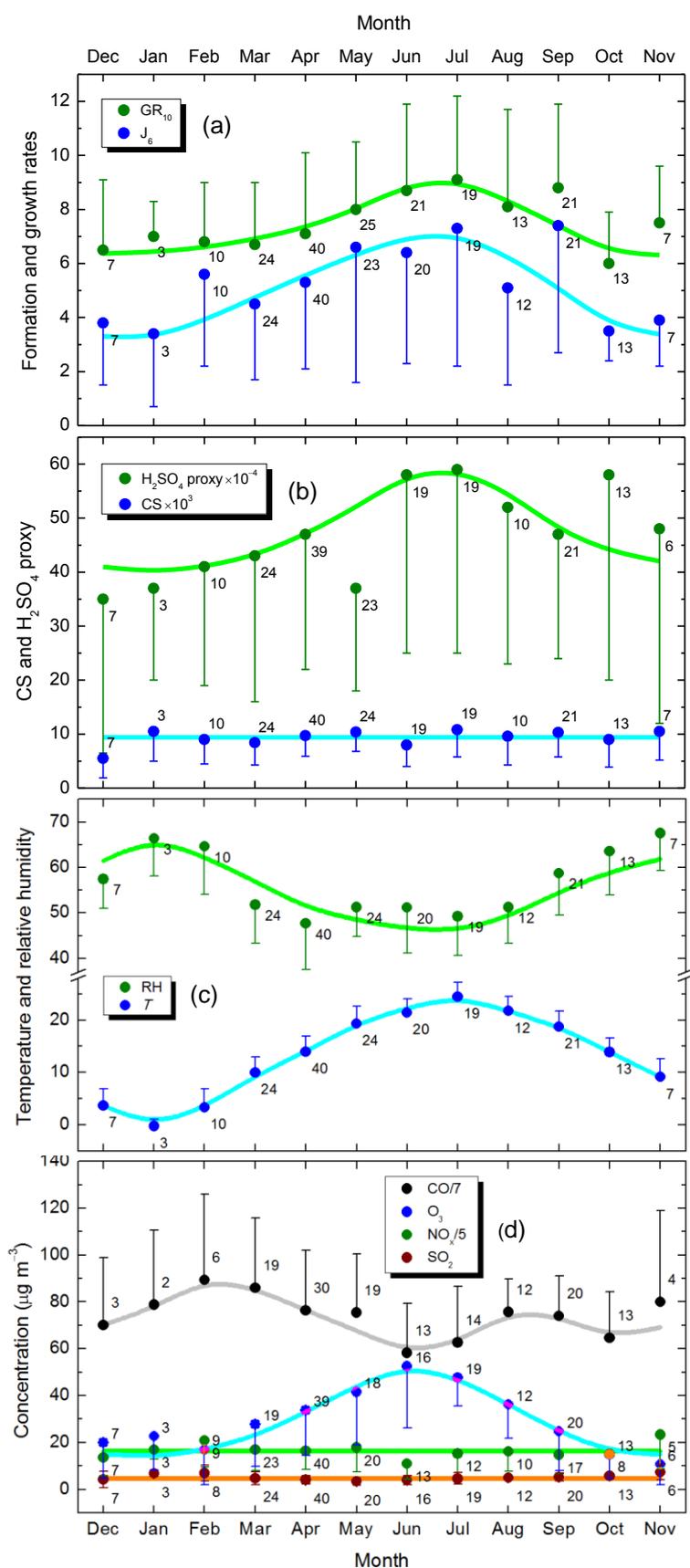
524 Some of our results are in line with other observations according to which GR exhibited almost
525 exclusively a summer maximum, while some other finding are different in the sense that the
526 seasonal variability in particle formation rate was quite modest and could not be established
527 earlier (Nieminen et al., 2018). There is one more aspect which may be worth realising in this
528 respect. A large fraction of compounds contributing to NPF and growth in cities can originate
529 from anthropogenic precursors (Vakkari et al., 2015). Their emissions may peak any time of
530 year depending on human habits and requirements (Nieminen et al., 2018). Nevertheless, the
531 fact that our monthly distributions of the dynamic properties in urban environments follow the
532 universal summer maximum behaviour may indicate the overall prevailing role of atmospheric
533 photochemistry coupled with biogenic emissions of aerosol precursor vapours.

534

535 The monthly mean J_6 , GR_{10} and H_2SO_4 proxy data still have considerable uncertainty, which
536 makes their interpretation not yet completely conclusive. The uncertainties are influenced by
537 inherent fluctuations in the primary data sets, enhancing effects caused by combining some
538 individual primary data into compound variables (such as H_2SO_4 proxy), number of data items
539 available for different properties and months, variations in other or unknown relevant
540 environmental conditions, and by the variability in relative nucleation occurrence frequency
541 from year to year. The resulting uncertainties are expected to decrease with the length of the
542 available data sets, which emphasized the need to continue the measurements.

543

544 The monthly distributions of CS, and SO_2 and NO_x concentrations could be represented by
545 constant values of the overall means and SDs of $(9.4 \pm 4.3) \times 10^{-3} \text{ s}^{-1}$, $4.7 \pm 2.1 \mu\text{g m}^{-3}$ and 81 ± 38
546 $\mu\text{g m}^{-3}$, respectively with an acceptable accuracy. This suggests that CS, SO_2 and NO_x in
547 Budapest do not critically or substantially affect either the dynamic properties (or the event
548 occurrence). Monthly distributions of air T and O_3 concentration showed a maximum over
549 summer months, while RH reflected the T tendency. The monthly variation of T on event days
550 and on non-event days were similar. More importantly, higher biogenic emissions and typically
551 stronger photochemistry are expected during the summer, which both enhance the production
552 rate of nucleating and condensing vapours, while there is practically nothing extra that would
553 suppress the dynamical properties (Kerminen et al., 2018). As a result of these, the dynamic
554 rates show a summer maximum. This is completely consistent with the results from other urban
555 and non-urban studies (Nieminen et al., 2018). Distribution of CO was more changing and
556 without obvious tendentious temporal structure or feature than for the other gases, and,
557 therefore, its interpretation is encumbered so far. However, it doesn't seem to substantially
558 affect the dynamic properties.



560 **Figure 2.** Distribution of
 561 monthly mean aerosol
 562 particle formation rate J_6 in
 563 a unit of $\text{cm}^{-3} \text{s}^{-1}$ and
 564 particle diameter growth
 565 rate GR_{10} in a unit of nm h^{-1}
 566 (a), mean condensation sink
 567 for vapours (CS) in a unit of
 568 s^{-1} averaged over the
 569 nucleation time interval (t_1 ,
 570 t_2) and daily maximum gas-
 571 phase H_2SO_4 proxy in a unit
 572 of $\mu\text{g m}^{-5} \text{W s}$ (b), daily
 573 mean air temperature (T) in
 574 a unit of $^\circ\text{C}$ and daily mean
 575 relative humidity (RH) in %
 576 (c), and daily median
 577 concentrations of SO_2 , O_3 ,
 578 NO_x and CO for
 579 quantifiable (class 1A) new
 580 particle formation and
 581 growth events in the city
 582 centre for the joint 5-year
 583 long time interval. The error
 584 bars are shown for one side
 585 for clarity and indicate 1
 586 standard deviation. Number
 587 of the individual data
 588 averaged in each month is
 589 displayed next to the
 590 symbols. The horizontal
 591 lines indicate the overall
 592 mean. The nonlinear curves
 593 assist to guide the eye.

595 4.3 Relationships

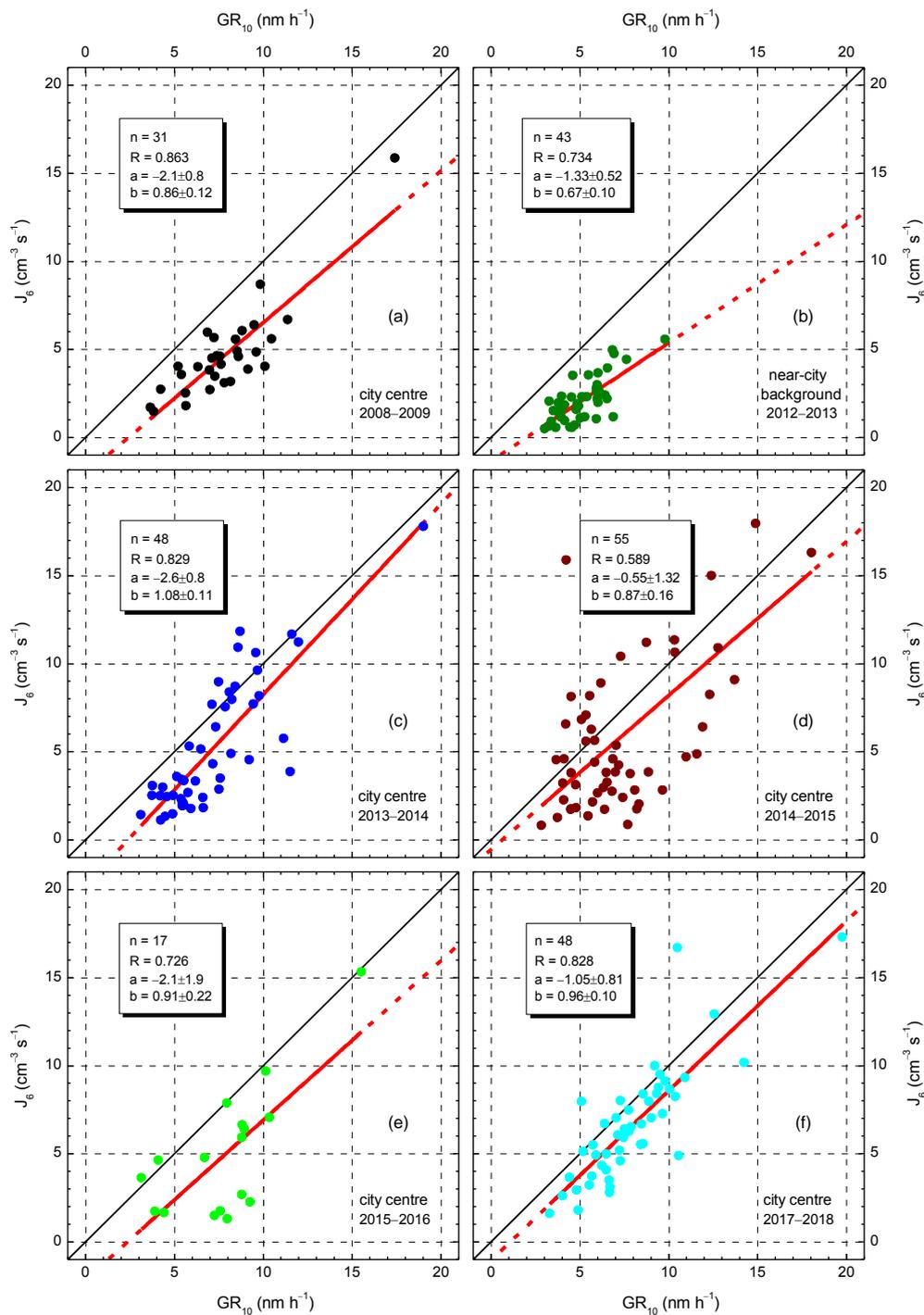
596

597 Pearson's coefficients of correlation (R) between J_6 and GR_{10} revealed significant linear
598 relationship between them for all annual data sets (the mean R and SD were 0.768 ± 0.099 ,
599 number of data pairs $n=243$). This confirms that formation of new aerosol particle and their
600 growth to larger sizes in the atmosphere are tightly and positively linked together. It should be
601 noted that J_6 and GR_{10} are not completely independent variables (see Eq. 1 and Table S1). The
602 linear relationship between the dynamic properties was observed under different atmospheric
603 conditions in many environments (Nieminen et al., 2018). At some sites, this relationship could
604 not be proved due to the weak variability in the variables.

605

606 The dynamic properties can also be coupled to the concentrations of aerosol precursor
607 compounds and properties of a pre-existing particle population, thus to atmospheric
608 environment (Kerminen et al., 2018). It is, therefore, sensible to investigate the city centre and
609 near-city background data separately. Scatter plots between J_6 and GR_{10} for the 1-year long
610 measurement time intervals are shown in Fig. 3. For the city centre, the regression lines follow
611 the line with a slope of 1 in all 5 years. The mean slope (b) with SD for the joint 5-year long
612 city centre data set was $b=0.94 \pm 0.07$ expressed formally in a unit of $\text{cm}^{-3} \text{s}^{-1} \text{nm}^{-1} \text{h}$. This
613 ~~implies that the relevant chemical species and/or their processes in the air of the city centre~~
614 ~~contribute equally to the formation of 6 nm particles and to their growth process.~~ At the same
615 time, the regression line for the near-city background deviated significantly with a $b=0.67 \pm 0.10$
616 $\text{cm}^{-3} \text{s}^{-1} \text{nm}^{-1} \text{h}$ from the J_6 vs. GR_{10} dependency for the city centre. This can imply that NPF
617 and growth processes advance in a different manner in these 2 environments, ~~and that the~~
618 ~~chemical compounds available and their processes power particle growth more than new~~
619 ~~particle formation in the near-city background.~~ This can be related to the differences between
620 the city and its close environment as far as the atmospheric composition (for instance, the VOC
621 and NO_x concentrations), chemistry and physics, and other delicate conditions are concerned
622 (Paasonen et al., 2018). The narrower range and smaller number of individual dynamic
623 properties available for the near-city background relative to those in the city centre represent
624 some inherent limitation or weakness in the explanation, and, therefore, it can strictly be
625 regarded as a working hypothesis because a rigorous statistical treatment would require larger
626 variability in the near-city background data.

627
 628
 629
 630
 631
 632
 633
 634
 635
 636
 637
 638
 639
 640
 641
 642
 643
 644
 645
 646
 647
 648
 649
 650
 651
 652
 653



654 **Figure 3.** Scatter plots for aerosol particle formation rate J_6 and consecutive particle diameter growth
 655 rate GR_{10} in city centre (a and c–f) and near-city background (b) separately for the 1-year long
 656 measurement time intervals. Number of data point (n), their coefficient of correlation (R) and the
 657 intercept (a) and slope (b) of the regression line with standard deviations are also indicated. The lines in
 658 black represent the line with a slope of 1, the solid lines in red show the regression lines, while the
 659 dashed parts in red are extrapolated from the regression line.

660

661 The intercepts (a) of the regression lines were identical for all data sets within their uncertainty
662 interval. The mean intercept and SD were estimated to be $-1.7 \pm 0.8 \text{ cm}^{-3} \text{ s}^{-1}$. This finding is
663 interpreted as the existence of a minimum GR or more exactly of a minimally required GR that
664 leads to $J_6 > 0$. Particles that exhibit at least this level of GR can escape coagulation mainly with
665 larger particles and reach the detectable diameter (6 nm in our case) by condensational growth.
666 The minimal GR was derived as $\text{GR}_{\min} = -a/b$, and its mean and SD are $1.8 \pm 1.0 \text{ nm h}^{-1}$ for the
667 conditions ordinarily present in the Budapest air. Nucleation processes which are initiated under
668 circumstances that cause the newly formed particle with a diameter of 10 nm to grow with a
669 rate $< \text{GR}_{\min}$ are normally not observed. Anyway, these are expected to be events with relatively
670 small J_6 (weak phenomena) due to the relationship between GR_{10} and J_6 . The events with GR
671 larger but close to this limit could be still masked by fluctuating experimental data. Their
672 identification and evaluation can be made feasible by decreasing the lower measurement
673 diameter limit of DMPS systems down to 3 nm, or by different instruments such as particle size
674 magnifier or neutral cluster and air ions spectrometer.

675

676 Correlations between individual H_2SO_4 proxy values on one side and J_6 or GR_{10} on the other
677 side were not significant. This is consistent with the corresponding conclusion of Sect. 4.2 and
678 with the earlier results according to which the mean contribution of H_2SO_4 condensation to the
679 particle GR_{10} was only 12.3% in Budapest (Salma et al., 2016b). The lack of correlation and
680 the average concentrations of SO_2 derived separately for the NPF and growth event and non-
681 event days suggest that this precursor gas is ordinarily available in excess and, therefore, the
682 formation of H_2SO_4 is likely governed by photochemical conditions, and that other chemical
683 species than H_2SO_4 can have larger influence on the particle growth. The role of H_2SO_4 in the
684 nucleation process and early particle growth could be still relevant or determinant.

685

686 Coefficients of correlation between CS on one side and J_6 or GR_{10} on the other side for the joint
687 city centre data sets were modest ($R=0.41$ and 0.32 , respectively with $n=194$ and 197 ,
688 respectively). This is simply related to the fact that larger GR values are typical for polluted
689 urban air (Kulmala et al., 2017) since particles capable of escaping coagulation scavenging need
690 to grow faster in comparison to cleaner environments, and the enhanced requirements for the
691 growth are linked to increased formation rates as well. It should be noted here that the GR of
692 newly formed particles to larger sizes is primarily coupled to 1) CS, which is further linked to
693 the entire aerosol particle population (including the newly formed particles, thus the NPF itself),

694 2) to the total concentration and some physicochemical properties of non-volatile gaseous
695 compounds and 3) to their production rate in the gas phase from aerosol precursor compounds
696 (e.g. Kerminen et al., 2018). These couplings could result in rather complex behaviour, and
697 their understanding is essential when analysing atmospheric observations.

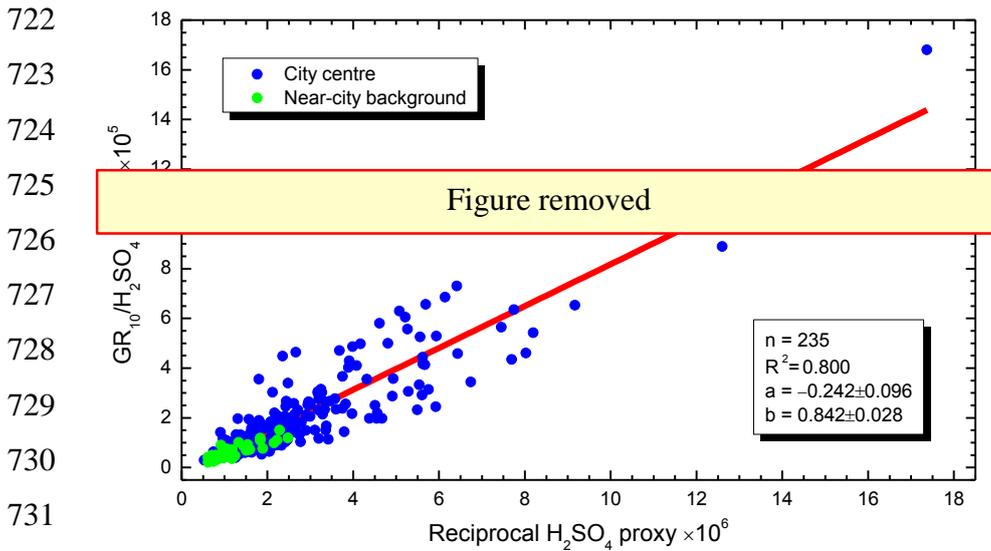
698

699 As far as the pollutant gases are concerned, no correlation could be identified between J_6 or
700 GR_{10} on one side and the gas concentrations on the other side. The coefficients of correlation
701 between CS and NO_x or CO were modest ($R=0.37$ and 0.42 , respectively with $n=164$ and 152 ,
702 respectively), while correlation of NO_x and CO on one side with WS was also modest but
703 negative ($R= -0.32$ and -0.42 , respectively with $n=167$ and 155 , respectively). The former
704 relationships can be explained by the fact that vehicular road traffic in cities is a considerable
705 and common source of NO_x , CO and primary particles (Paasonen et al., 2016), and the emitted
706 particles largely contribute to CS levels. The latter relationships are linked to the effect of large-
707 scale air mass transport (often connected to high WSs) on urban air pollution or air quality.

708

709 ~~Importance and contribution of condensing vapours other than H_2SO_4 are further demonstrated~~
710 ~~in Fig. 4. The data for the city centre and near-city background were presented in a linearized~~
711 ~~form and separately for the 2 sites. Nevertheless, the fitting of the correlation line was~~
712 ~~accomplished for the joint 6 year long data set. It can be demonstrated in particular on non-~~
713 ~~linearized plot of the GR_{10}/H_2SO_4 proxy as function of H_2SO_4 proxy (not shown here) that the~~
714 ~~2 data sets merge into each other without any relevant structure, and, therefore, that they can be~~
715 ~~regarded to be coherent. This approach seems sensible when considering also the limited~~
716 ~~accuracy of the values. The relationship between the 2 composite variables in Fig. 4 was~~
717 ~~significant ($R^2=0.800$, $p<0.05$). It can be interpreted as the increasing atmospheric concentration~~
718 ~~of gas phase H_2SO_4 can be related to larger contributions of other vapours than H_2SO_4 to~~
719 ~~particle growth. The other or competing compounds may include oxidation products and their~~
720 ~~dimers from photooxidation of VOC precursors from both biogenic and anthropogenic sources.~~

721



732 **Figure 4.** Dependence of the growth rate GR_{10} (in a unit of nm h^{-1}) of new particle formation and growth
 733 events normalised to the daily maximum gas phase H_2SO_4 proxy ($\mu\text{g m}^{-5} \text{W s}$) on the reciprocal proxy
 734 value in the city centre and near city background. The linear line in red represents the line fitted to the
 735 joint data set. Number of individual data considered (n), their coefficient of determination (R^2) and the
 736 intercept (a) and slope (b) of the fitted regression line with standard deviations are also shown.

738 4.4 Extreme and multiple events

739
 740 The joint 6-year long data sets of J_6 , GR_{10} and Δt containing all, 247 individual data each could
 741 be characterised by lognormal distribution function. This is demonstrated by log-probability
 742 graph for J_6 in Fig. S2 as example. The coefficient of determination, median and geometric
 743 standard deviation for J_6 , GR_{10} and Δt data sets were 0.990, 4.0 cm^{-3} and 2.3; 0.993, 6.8 nm h^{-1}
 744 and 1.46; and 0.998, 02:57 (0.123 d) and 1.74, respectively.

745
 746 One of the major properties of this distribution type is that it contains relatively large individual
 747 data with considerably high abundances. There were 5 individual J_6 and 5 individual GR_{10} data
 748 above the 98% percentile of the data sets, which belonged to 9 separate NPF and growth events
 749 (days). Their specifications, properties and parameters are summarised in Table 3. All these
 750 events occurred in the city centre from April to September. ~~Their number in the separate~~
 751 ~~consecutive measurement years (Sect. 2) were 1, 0, 1, 2, 0 and 5, respectively.~~ The medians of
 752 J_6 , GR_{10} , CS and air T for the subsets of these 9 extreme event days were larger by factors of
 753 5.2, 2.4, 1.5 and 1.4, respectively than for the city centre data. **At the same time**, the medians
 754 of the other atmospheric properties and concentrations in these 2 respective data sets agreed
 755 within approximately 10%. There was a single event associated with an extreme H_2SO_4 proxy

756 (of $23 \times 10^5 \mu\text{g m}^{-5} \text{ W s}$) and relatively low NO_x concentration ($44 \mu\text{g m}^{-3}$), which indicate
757 exceptionally favourable conditions for NPF and growth. In addition to this case, there were
758 only a few days that were characterised by an unusually large CS ($23 \times 10^{-3} \text{ s}^{-1}$) – which could
759 in turn be linked to higher dynamic rates (Sect. 4.3) – or by somewhat larger SO_2 ($8.1 \mu\text{g m}^{-3}$)
760 or lower NO_x concentration ($34 \mu\text{g m}^{-3}$). For all the other events, however, no simple or
761 compound property of the investigated variables could explain the extreme rates. Instead, they
762 may be related to some other chemical species and/or atmospheric processes, which were not
763 including in the present study. ~~Since the extreme NPF and growth events usually resembled the~~
764 ~~time evolution for regional events (well-developed banana curves) – sometimes with multiple~~
765 ~~onsets –, the missing atmospheric players in increased concentrations or their relevant processes~~
766 ~~are expected to appear on a larger horizontal spatial scale.~~

767

768 **Table 3.** Date (in a format of dd–MM–yyyy), new particle formation rate J_6 (in a unit of $\text{cm}^{-3} \text{ s}^{-1}$),
769 particle diameter growth rate GR_{10} (nm h^{-1}), starting time t_1 of nucleation (HH:mm UTC+1), duration
770 time interval $\Delta t = t_2 - t_1$ of nucleation (HH:mm), mean condensation sink CS during the nucleation process
771 (10^{-3} s^{-1}), daily maximum gas-phase H_2SO_4 proxy ($10^4 \mu\text{g m}^{-5} \text{ W s}$), daily mean air temperature T ($^\circ\text{C}$),
772 daily mean relative humidity RH (%), daily median concentrations of SO_2 , O_3 , NO_x ($\mu\text{g m}^{-3}$) and CO
773 (mg m^{-3}) gases, and the type of the onset for extreme quantifiable (class 1A) new particle formation and
774 growth events. The cells in yellow indicate the values which are above the 98% percentile of the
775 corresponding data sets. N.a.: not available.

776

Date/ property	15– 09– 2009	20– 04– 2014	19– 05– 2015	04– 07– 2015	28– 05– 2017	25– 06– 2017	02– 08– 2017	31– 08– 2017	09– 09– 2017
J_6	15.9	17.8	24	16.3	27	33	30	47	17.3
GR_{10}	17.4	19.0	12.2	18.0	9.2	17.0	11.8	21	19.8
t_1	10:20	08:52	08:52	09:38	06:34	10:18	07:39	10:06	11:38
Δt	01:23	01:42	03:57	02:06	07:15	02:46	06:58	06:19	02:06
Proxy	38	42	25	16	229	41	69	92	45
CS	13.4	8.9	13.7	11.9	6.9	10.5	23	18.2	15.5
T	20	13.0	22	26	20	24	29	23	19.1
RH	60	62	48	40	40	68	49	47	58
SO_2	6.1	2.5	4.4	2.3	3.4	3.1	5.6	8.1	6.6
O_3	16.3	43	n.a.	33	61	56	34	24	12.9
NO_x	69	34	174	70	44	66	n.a.	109	112
CO	0.42	n.a.	0.71	0.33	0.31	0.50	0.97	0.62	0.71
Onset	ordinary	double	broad	ordinary	broad	broad	broad	broad	ordinary

777

778

779 Each quantifiable NPF and growth event was labelled as ordinary or broad by visual inspection
780 of its beginning part. If the width of the beginning was smaller than approximately 2 h or there
781 was a determinant single growth curve (rib) on the size distribution surface plot then the onset
782 was labelled as ordinary, otherwise as broad (Fig. S1a and S1c for broad onsets). Broad onsets
783 can be generated by 1) long-lasting nucleation process, 2) disrupted and started over nucleation
784 due to changing atmospheric and meteorological conditions or 3) multiple nucleation processes
785 close to each other in time (Salma et al., 2016b). The broad onsets were specified as doublets
786 if the nucleation mode could be separated into 2 submodes by size distribution fitting.
787 Approximately 40% of all quantifiable events had a broad onset. This indicates that NPF and
788 growth events with broad/multiple onsets are abundant in the urban environment, which could
789 be an important difference from remote or clean atmospheres.

790

791 For 9% of all quantifiable days, it was feasible to calculate 2 sets of dynamic properties for
792 onsets 1 and 2 with a reasonable accuracy. In the near-city background, the medians of J_6 and
793 GR_{10} for the onset 1 were similar or somewhat smaller than the corresponding medians for the
794 whole near-city background data set, while for the onset 2, they were substantially larger,
795 namely $4.1 \text{ cm}^{-3} \text{ s}^{-1}$ and 10.0 nm h^{-1} , respectively (cf. Table 2). Actually, the latter values were
796 closer to the medians of the city centre than for the near-city background. ~~The dynamic~~
797 ~~properties for the city centre for both the onset 1 and onset 2 were somewhat larger than for the~~
798 ~~whole the city centre data set.~~ Approximately 75% of the doublets resulted in individual
799 onset2/onset1 ratios larger than unity. Their overall median ratios for J_6 and GR_{10} were similar
800 and approximately 1.4, while for the near-city background, they were about 2. The results are
801 in line with ~~and confirm~~ our earlier conclusion according to which the second onsets (if it is a
802 new formation process and not just a started over event) ~~often generate new particles more~~
803 ~~intensively~~ **are more intensive** than the first onsets (Salma et al., 2016b). These particles also
804 grow faster. This can be explained by the fact that the first event is ~~often or likely~~ of regional
805 scale since its dynamic properties resemble those of the regional background ~~process~~ (Yli-Juuti
806 et al., 2009), while the later event can be characterised by values typical for the city centre
807 (Salma et al., 2016b). The later event (or events) are mainly caused and governed by sub-
808 regional processes. These findings are also coherent with a previous observation of NPF and
809 growth events with multiple onsets in semi-clean savannah and industrial environments
810 (Hirsikko et al., 2013), and they also fit well into the existing ideas on mixing regional and

811 urban air parcels that exhibit different properties such as precursor concentrations, T and RH
812 (Kulmala et al., 2017).

813

814 **5 Conclusions**

815

816 Magnitude of the particle number concentration level produced solely by NPF and growth
817 (strength of the events) can roughly be estimated by considering the median J_6 , median duration
818 of nucleation Δt (their distribution function is lognormal; Table 2) and the mean coagulation
819 loss of these particles F_{coag} (0.17; Sect. 3.1 and Table S1) as: $J_6 \times \Delta t \times (1 - F_{\text{coag}})$. In central
820 Budapest, it yields a concentration of 10^4 cm^{-3} . This is in line with another result achieved by
821 nucleation strength factor (Salma et al., 2017). More importantly, the estimated concentration
822 from NPF and growth process is comparable to the annual median atmospheric concentrations
823 (Sect. 4). This simple example indicates that the phenomenon is not only relevant for aerosol
824 load and climate issues on regional or global spatial scales – which were first recognised – but
825 it can affect the urban climate and the health risk of inhabitants of cities as well.

826

827 Similar recognitions have led to the emerge of urban atmospheric nucleation studies. As part of
828 this international progress, we presented here a considerable variety of contributions, which
829 became feasible thank to gradually generating, multi-year long, critically evaluated, complex
830 and coherent data sets. Dynamic and timing properties of 247 NPF and growth events were
831 studied together with supporting aerosol properties, meteorological data and pollutant gas
832 concentrations in near-city background and city centre of Budapest for 6 years. The results and
833 conclusions derived form in important component that is based on atmospheric observations.
834 The present study can also be considered as the first step toward a larger statistical evaluation
835 process. The results are to be combine with results from laboratory experiments and finally,
836 with theoretical models to further improve our understanding on the atmospheric processes in
837 cities. Specialities and features of the urban atmospheric NPF and growth phenomena are finally
838 to be also considered when assessing their potentials to increase UF and CCN concentrations
839 or their health implications.

840

841 The present research based on ambient atmospheric measurements provided evidence that some
842 important chemical players in the NPF and growth events are still missing. Considering the
843 results and conclusions of cloud chamber experiments, these factors are expected to be related
844 mainly to oxidation products of VOCs and/or their processes. Further dedicated research

845 including sophisticated measurements, data evaluations and modelling studies is required to
846 find and identify these species and their processes, and to account their multifactorial role in
847 more detail. Such measurement campaign focusing on chemical composition of molecular
848 clusters, precursors and nucleating vapours by applying recent expedient instruments in
849 Budapest over the months of the highest expected event occurrence has been just realised within
850 a frame of an international cooperation. Its perspective results can hopefully provide additional
851 valuable information for some of the conclusion base on indirect evidence for the time being
852 and can further clarify the overall picture on urban multicomponent nucleation and growth
853 phenomenon.

854

855 **Data availability.** The observational data used in this paper are available on request from the
856 corresponding author or at the website of the Budapest platform for Aerosol Research and Training
857 (<http://salma.web.elte.hu/BpART>).

858

859 **Competing interest.** The authors declare that they have no conflict of interest.

860

861 **Acknowledgements.** The authors thank Markku Kulmala and his research team at the University of
862 Helsinki for their cooperation. Financial support by the National Research, Development and Innovation
863 Office, Hungary (contracts K116788 and PD124283); by the European Regional Development Fund
864 and the Hungarian Government (GINOP-2.3.2-15-2016-00028) is gratefully acknowledged.

865

866 **References**

- 867 Alam, A., Shi, J. P., and Harrison, R. M.: Observations of new particle formation in urban air, J.
868 Geophys. Res., 108 (D3), 4093, doi:10.1029/2001JD001417, 2003.
- 869 Almeida, J., Schobesberger, S., Kurten, A., Ortega, I. K., Kupiainen-Maatta, O., Praplan, A. P.,
870 Adamov, A., Amorim, A., Bianchi, F., Breitenlechner, M., David, A., Dommen, J., Donahue, N.
871 M., Downard, A., Dunne, E., Duplissy, J., Ehrhart, S., Flagan, R. C., Franchin, A., Guida, R.,
872 Hakala, J., Hansel, A., Heinritzi, M., Henschel, H., Jokinen, T., Junninen, H., Kajos, M.,
873 Kangasluoma, J., Keskinen, H., Kupc, A., Kurten, T., Kvashin, A. N., Laaksonen, A., Lehtipalo,
874 K., Leiminger, M., Leppa, J., Loukonen, V., Makhmutov, V., Mathot, S., McGrath, M. J.,
875 Nieminen, T., Olenius, T., Onnela, A., Petäjä, T., Riccobono, F., Riipinen, I., Rissanen, M., Rondo,
876 L., Ruuskanen, T., Santos, F. D., Sarnela, N., Schallhart, S., Schnitzhofer, R., Seinfeld, J. H.,
877 Simon, M., Sipilä, M., Stozhkov, Y., Stratmann, F., Tome, A., Tröstl, J., Tsagkogeorgas, G.,
878 Vaattovaara, P., Viisanen, Y., Virtanen, A., Vrtala, A., Wagner, P. E., Weingartner, E., Wex, H.,
879 Williamson, C., Wimmer, D., Ye, P. L., Yli-Juuti, T., Carslaw, K. S., Kulmala, M., Curtius, J.,

880 Baltensperger, U., Worsnop, D. R., Vehkamäki, H., and Kirkby, J.: Molecular understanding of
881 sulphuric acid–amine particle nucleation in the atmosphere, *Nature*, 502, 359–363, 2013.

882 Baltensperger, U., Streit, N., Weingartner, E., Nyeki, S., Prévôt, A. S. H., Van Dingenen, R., Virkkula,
883 A., Putaud, J. P., Even, A., Brink, H., Blatter, A., Neftel, A., and Gaggeler, H. W.: Urban and rural
884 aerosol characterization of summer smog events during the PIPAPO field campaign in Milan, Italy,
885 *J. Geophys. Res.*, 107(D22), 8193, doi:10.1029/2001JD001292, 2002.

886 Bianchi, F., Tröstl, J., Junninen, H., Frege, C., Henne, S., Hoyle, C. R., Molteni, U., Herrmann, E.,
887 Adamov, A., Bukowiecki, N., Chen, X., Duplissy, J., Gysel, M., Hutterli, M., Kangasluoma, J.,
888 Kontkanen, J., Kürten, A., Manninen, H. E., Münch, S., Peräkylä, O., Petäjä, T., Rondo, L.,
889 Williamson, C., Weingartner, E., Curtius, J., Worsnop, D. R., Kulmala, M., Dommen, J., and
890 Baltensperger, U.: New particle formation in the free troposphere: A question of chemistry and
891 timing, *Science*, 352, 1109–1112, <https://doi.org/10.1126/science.aad5456>, 2016.

892 Braakhuis, H. M., Park, M. V., Gosens, I., De Jong, W. H., and Cassee, F. R.: Physicochemical
893 characteristics of nanomaterials that affect pulmonary inflammation, *Part. Fibre Toxicol.*, 11:18,
894 doi: 10.1186/1743-8977-11-18, 2014.

895 Cai, R. and Jiang, J.: A new balance formula to estimate new particle formation rate: reevaluating the
896 effect of coagulation scavenging, *Atmos. Chem. Phys.*, 17, 12659–12675, 2017.

897 Carslaw, K. S., Lee, L. A., Reddington, C. L., Pringle, K. J., Rap, A., Forster, P. M., Mann, G. W.,
898 Spracklen, D. V., Woodhouse, M. T., Regayre, L. A., and Pierce, J. R.: Large contribution of
899 natural aerosols to uncertainty in indirect forcing, *Nature*, 503, 67–71, 2013.

900 Crounse, J. D., Nielsen, L. B., Jørgensen, S., Kjaergaard, H. G., and Wennberg, P. O.: Autoxidation of
901 organic compounds in the atmosphere, *J. Phys. Chem. Lett.*, 4, 20, 3513–3520, 2013.

902 Dal Maso, M., Kulmala, M., Lehtinen, K. E. J., Mäkelä, J. M., Aalto, P. P., and O’Dowd, C.:
903 Condensation and coagulation sinks and formation of nucleation mode particles in coastal and
904 boreal forest boundary layers, *J. Geophys. Res.*, 107(19D), 8097, 10.1029/2001jd001053, 2002.

905 Dal Maso, M., Kulmala, M., Riipinen, I., Wagner, R., Hussein, T., Aalto, P. P., and Lehtinen, K. E. J.:
906 Formation and growth of fresh atmospheric aerosols: eight years of aerosol size distribution data
907 from SMEAR II, Hyytiälä, Finland, *Boreal Environ. Res.*, 10, 323–336, 2005.

908 Dall’Osto, M., Querol, X., Alastuey, A., O’Dowd, C., Harrison, R. M., Wenger, J., and Gómez-
909 Moreno, F. J.: On the spatial distribution and evolution of ultrafine particles in Barcelona, *Atmos.*
910 *Chem. Phys.*, 13, 741–759, 2013.

911 Ehn, M., Thornton, J. A., Kleist, E., Sipilä, M., Junninen, H., Pullinen, I., Springer, M., Rubach, F.,
912 Tillmann, R., Lee, B., Lopez-Hilfiker, F., Andres, S., Acir, I. H., Rissanen, M., Jokinen, T.,
913 Schobesberger, S., Kangasluoma, J., Kontkanen, J., Nieminen, T., Kurten, T., Nielsen, L. B.,
914 Jørgensen, S., Kjaergaard, H. G., Canagaratna, M., Dal Maso, M., Berndt, T., Petäjä, T., Wahner,
915 A., Kerminen, V. M., Kulmala, M., Worsnop, D. R., Wildt, J., and Mentel, T. F.: A large source of
916 low-volatility secondary organic aerosol, *Nature*, 506, 476–479, 2014.

917 Gordon, H., Sengupta, K., Rap, A., Duplissy, J., Frege, C., Williamson, C., Heinritzi, M., Simon, M.,
918 Yan, C., Almeida, J., Tröstl, J., Nieminen, T., Ortega, I. K., Wagner, R., Dunne, E. M., Adamov,
919 A., Amorim, A., Bernhammer, A. K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven,
920 J. S., Dias, A., Ehrhart, S., Fischer, L., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Hakala, J.,
921 Hoyle, C. R., Jokinen, T., Junninen, H., Kangasluoma, J., Kim, J., Kirkby, J., Krapf, M., Kürten,
922 A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Molteni, U., Monks, S. A., Onnela,
923 A., Peräkylä, O., Piel, F., Petäjä, T., Praplan, A. P., Pringle, K. J., Richards, N. A. D., Rissanen, M.
924 P., Rondo, L., Sarnela, N., Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sharma, S., Sipilä, M.,
925 Steiner, G., Stozhkov, Y., Stratmann, F., Tomé, A., Virtanen, A., Vogel, A. L., Wagner, A. C.,
926 Wagner, P. E., Weingartner, E., Wimmer, D., Winkler, P. M., Ye, P., Zhang, X., Hansel, A.,
927 Dommen, J., Donahue, N. M., Worsnop, D. R., Baltensperger, U., Kulmala, M., Curtius, J., and
928 Carslaw, K. S.: Reduced anthropogenic aerosol radiative forcing caused by biogenic new particle
929 formation, *Proc. Natl. Acad. Sci. U.S.A.*, 113, 12053–12058,
930 <https://doi.org/10.1073/pnas.1602360113>, 2016.

931 Hirsikko, A., Vakkari, V., Tiitta, P., Hatakka, J., Kerminen, V.-M., Sundström, A.-M., Beukes, J. P.,
932 Manninen, H. E., Kulmala, M., and Laakso, L.: Multiple daytime nucleation events in semi-clean
933 savannah and industrial environments in South Africa: analysis based on observations, *Atmos.*
934 *Chem. Phys.*, 13, 5523–5532, 2013.

935 Hussein, T., Puustinen, A., Aalto, P. P., Mäkelä, J. M., Hämeri, K., and Kulmala, M.: Urban aerosol
936 number size distributions, *Atmos. Chem. Phys.*, 4, 391–411, 2004.

937 Hussein, T., Martikainen, J., Junninen, H., Sogacheva, L., Wagner, R., Dal Maso, M., Riipinen, I.,
938 Aalto, P. P., and Kulmala, M.: Observation of regional new particle formation in the urban
939 atmosphere, *Tellus 60B*, 509–521, 2008.

940 Jokinen, T., Berndt, T., Makkonen, R., Kerminen, V.-M., Junninen, H., Paasonen, P., Stratmann, F.,
941 Herrmann, H., Guenther, A. B., Worsnop, D. R., Kulmala, M., Ehn, M. and Sipilä, M.: Production
942 of extremely low volatile organic compounds from biogenic emissions: Measured yields and
943 atmospheric implications, *Proc. Natl. Acad. Sci. U.S.A.*, 112, 7123–7128, 2015.

944 Kerminen, V.-M., Paramonov, M., Anttila, T., Riipinen, I., Fountoukis, C., Korhonen, H., Asmi, E.,
945 Laakso, L., Lihavainen, H., Swietlicki, E., Svenningsson, B., Asmi, A., Pandis, S. N., Kulmala, M.,
946 and Petäjä, T.: Cloud condensation nuclei production associated with atmospheric nucleation: a
947 synthesis based on existing literature and new results, *Atmos. Chem. Phys.*, 12, 12037–12059,
948 2012.

949 Kerminen, V.-M., Chen, X., Vakkari, V., Petäjä, T., Kulmala, M., and Bianchi, F.: Atmospheric new
950 particle formation and growth: review of field observations, *Environ. Res. Lett.*, 13 (2018) 103003,
951 2018.

952 Kirkby, J., Curtius, J., Almeida, J., Dunne, E., Duplissy, J., Ehrhart, S., Franchin, A., Gagné, S., Ickes,
953 L., Kürten, A., Kupc, A., Metzger, A., Riccobono, F., Rondo, L., Schobesberger, S.,
954 Tsagkogeorgas, G., Wimmer, D., Amorim, A., Bianchi, F., Breitenlechner, M., David, A.,

955 Dommen, J., Downard, A., Ehn, M., Flagan, R. C., Haider, S., Hansel, A., Hauser, D., Jud, W.,
956 Junninen, H., Kreissl, F., Kvashin, A., Laaksonen, A., Lehtipalo, K., Lima, J., Lovejoy, E. R.,
957 Makhutov, V., Mathot, S., Mikkilä, J., Minginette, P., Mogo, S., Nieminen, T., Onnela, A., Pereira,
958 A., Petäjä, T., Schnitzhofer, R., Seinfeld, J. H., Sipilä, M., Stozhkov, Y., Stratmann, F., Tome, A.,
959 Vanhanen, J., Viisanen Y., Vrtala, A., Wagner, P.E., Walther, H., Weingartner, E., Wex, H.,
960 Winkler, P.M., Carslaw, K. S., Worsnop, D. R., Baltensperger, U., and Kulmala, M.: The role of
961 sulfuric acid, ammonia and galactic cosmic rays in atmospheric aerosol nucleation, *Nature*, 476,
962 429–433, 2011.

963 Kirkby, J., Duplissy, J., Sengupta, K., Frege, C., Gordon, H., Williamson, C., Heinritzi, M., Simon,
964 M., Yan, C., Almeida, J., Tröstl, J., Nieminen, T., Ortega, I. K., Wagner, R., Adamov, A., Amorim,
965 A., Bernhammer, A.-K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven, J., Dias, A.,
966 Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Hakala, J., Hoyle, C. R., Jokinen, T.,
967 Junninen, H., Kangasluoma, J., Kim, J., Krapf, M., Kürten, A., Laaksonen, A., Lehtipalo, K.,
968 Makhmutov, V., Mathot, S., Molteni, U., Onnela, A., Peräkylä, O., Piel, F., Petäjä, T., Praplan, A.
969 P., Pringle, K., Rap, A., Richards, N. A. D., Riipinen, I., Rissanen, M. P., Rondo, L., Sarnela, N.,
970 Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sipilä, M., Steiner, G., Stozhkov, Y., Stratmann, F.,
971 Tomé, A., Virtanen, A., Vogel, A. L., Wagner, A., Wagner, P. E., Weingartner, E., Wimmer, D.,
972 Winkler, P. M., Ye, P., Zhang, X., Hansel, A., Dommen, J., Donahue, N. M., Worsnop, D. R.,
973 Baltensperger, U., Kulmala, M., Carslaw, K. S., and Curtius, J.: Ion-induced nucleation of pure
974 biogenic particles, *Nature*, 533, 521–526, <https://doi.org/10.1038/nature17953>, 2016.

975 Kulmala, M., Dal Maso, M., Mäkelä, J. M., Pirjola, L., Väkevä, M., Aalto, P., Miikkulainen, P.,
976 Hämeri, K., and O'Dowd, C. D.: On the formation, growth and composition of nucleation mode
977 particles, *Tellus B*53, 479–490, 2001.

978 Kulmala, M., Vehkamäki, H., Petäjä, T., Dal Maso, M., Lauri, A., Kerminen, V.-M., Birmili, W., and
979 McMurry, P.: Formation and growth rates of ultrafine atmospheric particles: a review of
980 observations, *J. Aerosol Sci.*, 35, 143–176, 2004.

981 Kulmala, M., Petäjä, T., Nieminen, T., Sipilä, M., Manninen, H. E., Lehtipalo, K., Dal Maso, M.,
982 Aalto, P. P., Junninen, H., Paasonen, P., Riipinen, I., Lehtinen, K. E. J., Laaksonen, A., and
983 Kerminen, V.-M.: Measurement of the nucleation of atmospheric aerosol particles, *Nat. Protoc.*, 7,
984 1651–1667, doi:10.1038/nprot.2012.091, 2012.

985 Kulmala, M., Kontkanen, J., Junninen, H., Lehtipalo, K., Manninen, H. E., Nieminen, T., Petäjä, T.,
986 Sipilä, M., Schobesberger, S., Rantala, P., Franchin, A., Jokinen, T., Järvinen, E., Äijälä, M.,
987 Kangasluoma, J., Hakala, J., Aalto, P.P., Paasonen, P., Mikkilä, J., Vanhanen, J., Aalto, J., Hakola,
988 H., Makkonen, U., Ruuskanen, T., Mauldin, R. L. III, Duplissy, J., Vehkamäki, H., Bäck, J.,
989 Kortelainen, A., Riipinen, I., Kurtén, T., Johnston, M. V., Smith, J. N., Ehn, M., Mentel, T. F.,
990 Lehtinen, K. E. J., Laaksonen, A., Kerminen, V.-M., and Worsnop, D. R.: Direct observations of
991 atmospheric aerosol nucleation, *Science*, 339, 943–946, 2013.

992 Kulmala, M., Petäjä, T., Ehn, M., Thornton, J., Sipilä, M., Worsnop, D. R., and Kerminen, V.-M.:
993 Chemistry of atmospheric nucleation: On the recent advances on precursor characterization and
994 atmospheric cluster composition in connection with atmospheric new particle formation, *Annu.*
995 *Rev. Phys. Chem.*, 65, 21–37, 2014.

996 Kulmala, M., Kerminen, V. M., Petäjä, T., Ding, A. J., and Wang, L.: Atmospheric gas-to-particle
997 conversion: why NPF events are observed in megacities, *Faraday Discuss.*,
998 doi:10.1039/C6FD00257A, 2017.

999 Makkonen, R., Asmi, A., Korhonen, H., Kokkola, H., Järvenoja, S., Räisänen, P., Lehtinen, K. E. J.,
1000 Laaksonen, A., Kerminen, V.-M., Järvinen, H., Lohmann, U., Bennartz, R., Feichter, J., and
1001 Kulmala, M.: Sensitivity of aerosol concentrations and cloud properties to nucleation and
1002 secondary organic distribution in ECHAM5-HAM global circulation model, *Atmos. Chem. Phys.*,
1003 9, 1747–1766, 2009.

1004 Makkonen, R., Asmi, A., Kerminen, V.-M., Boy, M., Arneth, A., Hari, P., and Kulmala, M.: Air
1005 pollution control and decreasing new particle formation lead to strong climate warming, *Atmos.*
1006 *Chem. Phys.*, 12, 1515–1524, 2012.

1007 Merikanto, J., Spracklen, D. V., Mann, G. W., Pickering, S. J., and Carslaw, K. S.: Impact of
1008 nucleation on global CCN, *Atmos. Chem. Phys.*, 9, 8601–8616, 2009.

1009 Metzger, A., Verheggen, B., Dommen, J., Duplissy, J., Prévôt, A. S. H., Weingartner, E., Riipinen, I.,
1010 Kulmala, M., Spracklen, D. V., Carslaw, K. S., and Baltensperger, U.: Evidence for the role of
1011 organics in aerosol particle formation under atmospheric conditions, *Proc. Natl. Acad. Sci. U. S.*
1012 *A.*, 107, 6646–6651, 2010.

1013 Németh, Z. and Salma, I.: Spatial extension of nucleating air masses in the Carpathian Basin, *Atmos.*
1014 *Chem. Phys.*, 14, 8841–8848, 2014.

1015 Németh, Z., Rosati, B., Ziková, N., Salma, I., Bozó, L., Dameto de España, C., Schwarz, J., Ždímal,
1016 V., and Wonaschütz, A.: Comparison of atmospheric new particle formation and growth events in
1017 three Central European cities, *Atmos. Environ.*, 178, 191–197, 2018.

1018 Nieminen, T., Kerminen, V.-M., Petäjä, T., Aalto, P. P., Arshinov, M., Asmi, E., Baltensperger, U.,
1019 Beddows, D. C. S., Beukes, J. P., Collins, D., Ding, A., Harrison, R. M., Henzing, B., Hooda, R.,
1020 Hu, M., Hörrak, U., Kivekäs, N., Komsaare, K., Krejčí, R., Kristensson, A., Laakso, L., Laaksonen,
1021 A., Leaitch, W. R., Lihavainen, H., Mihalopoulos, N., Németh, Z., Nie, W., O'Dowd, C., Salma, I.,
1022 Sellegri, K., Svenningsson, B., Swietlicki, E., Tunved, P., Ulevicius, V., Vakkari, V., Vana, M.,
1023 Wiedensohler, A., Wu, Z., Virtanen, A., and Kulmala, M.: Global analysis of continental boundary
1024 layer new particle formation based on long-term measurements, *Atmos. Chem. Phys.*, 18, 14737–
1025 14756, 2018.

1026 Oberdörster, G., Oberdörster, E., and Oberdörster, J.: Nanotoxicology: an emerging discipline
1027 evolving from studies of ultrafine particles, *Environ. Health Perspect.*, 113, 823–839, 2005.

1028 O'Dowd, C. D., Jimenez, J. L., Bahreini, R., Flagan, R. C., Seinfeld, J. H., Hämeri, K., Pirjola, L.,
1029 Kulmala, M., Jennings, S. G., and Hoffmann, Th.: Marine aerosol formation from biogenic iodine
1030 emissions, *Nature* 417, 632–636, 2002.

1031 Paasonen, P., Kupiainen, K., Klimont, Z., Visschedijk, A., Denier van der Gon, H. A. C., and Amann,
1032 M.: Continental anthropogenic primary particle number emissions, *Atmos. Chem. Phys.*, 16, 6823–
1033 6840, 2016.

1034 Paasonen, P., Peltola, M., Kontkanen, J., Junninen, H., Kerminen, V.-M., and Kulmala, M.:
1035 Comprehensive analysis of particle growth rates from nucleation mode to cloud condensation
1036 nuclei in Boreal forest, *Atmos. Chem. Phys. Discuss.*, <https://doi.org/10.5194/acp-2018-169>, in
1037 review, 2018.

1038 Petäjä, T., Mauldin, III, R. L., Kosciuch, E., McGrath, J., Nieminen, T., Paasonen, P., Boy, M.,
1039 Adamov, A., Kotiaho, T., and Kulmala, M.: Sulfuric acid and OH concentrations in a boreal forest
1040 site, *Atmos. Chem. Phys.*, 9, 7435–7448, 2009.

1041 Putaud, J.-P., Van Dingenen, R., Alastuey, A., Bauer, H., Birmili, W., Cyrus, J., Flentje, H., Fuzzi, S.,
1042 Gehrig, R., Hansson, H. C., Harrison, R. M., Herrmann, H., Hitzemberger, R., Hüglin, C., Jones,
1043 A.M., Kasper-Giebl, A., Kiss, G., Kousa, A., Kuhlbusch, T. A. J., Lösschau, G., Maenhaut, W.,
1044 Molnár, A., Moreno, T., Pekkanen, J., Perrino, C., Pitz, M., Puxbaum, H., Querol, X., Rodriguez,
1045 S., Salma, I., Schwarz, J., Smolík, J., Schneider, J., Spindler, G., ten Brink, H., Turšič, J., Viana,
1046 M., Wiedensohler, and A., Raes, F.: A European Aerosol Phenomenology - 3: physical and
1047 chemical characteristics of particulate matter from 60 rural, urban, and kerbside sites across
1048 Europe, *Atmos. Environ.*, 44, 1308–1320, 2010.

1049 Riccobono, F., Schobesberger, S., Scott, C., Dommen, J., Ortega, I., Rondo, L., Almeida, J., Amorim,
1050 A., Bianchi, F., Breitenlechner, M., David, A., Downard, A., Dunne, E., Duplissy, J., Ehrhart, S.,
1051 Flagan, R., Franchin, A., Hansel, A., Junninen, H., Kajos, M., Keskinen, H., Kupc, A., Kurten, A.,
1052 Kvashin, A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Nieminen, T., Onnela, A.,
1053 Petäjä, T., Praplan, A., Santos, F., Schallhart, S., Seinfeld, J., Sipila, M., Spracklen, D., Stozhkov,
1054 Y., Stratmann, F., Tome, A., Tsagkogeorgas, G., Vaattovaara, P., Viisanen, Y., Vrtala, A., Wagner,
1055 P., Weingartner, E., Wex, H., Wimmer, D., Carslaw, K., Curtius, J., Donahue, N., Kirkby, J.,
1056 Kulmala, M., Worsnop, D., and Baltensperger, U.: Oxidation products of biogenic emissions
1057 contribute to nucleation of atmospheric particles, *Science*, 344, 717–721, 2014.

1058 Riipinen, I., Pierce, J. R., Yli-Juuti, T., Nieminen, T., Häkkinen, S., Ehn, M., Junninen, H., Lehtipalo,
1059 K., Petäjä, T., Slowik, J., Chang, R., Shantz, N. C., Abbatt, J., Leitch, W. R., Kerminen, V.-M.,
1060 Worsnop, D. R., Pandis, S. N., Donahue, N. M., and Kulmala, M.: Organic condensation: a vital
1061 link connecting aerosol formation to cloud condensation nuclei (CCN) concentrations, *Atmos.*
1062 *Chem. Phys.*, 11, 3865–3878, 2011.

1063 Salma, I., Borsós, T., Weidinger, T., Aalto, P., Hussein, T., Dal Maso, M., and Kulmala, M.:
1064 Production, growth and properties of ultrafine atmospheric aerosol particles in an urban
1065 environment, *Atmos. Chem. Phys.*, 11, 1339–1353, 2011.

- 1066 Salma, I., Borsós, T., Németh, Z., Weidinger, T., Aalto, T., and Kulmala, M.: Comparative study of
1067 ultrafine atmospheric aerosol within a city, *Atmos. Environ.*, 92, 154–161, 2014.
- 1068 Salma, I., Fűri, P., Németh, Z., Farkas, Á., Balásházy, I., Hofmann, W., and Farkas, Á.: Lung burden
1069 and deposition distribution of inhaled atmospheric urban ultrafine particles as the first step in their
1070 health risk assessment, *Atmos. Environ.*, 104, 39–49, 2015.
- 1071 Salma, I., Németh, Z., Weidinger, T., Kovács, B., and Kristóf, G.: Measurement, growth types and
1072 shrinkage of newly formed aerosol particles at an urban research platform, *Atmos. Chem. Phys.*,
1073 16, 7837–7851, 2016a.
- 1074 Salma, I., Németh, Z., Kerminen, V. M., Aalto, P., Nieminen, T., Weidinger, T., Molnár, Á., Imre, K.,
1075 and Kulmala, M.: Regional effect on urban atmospheric nucleation, *Atmos. Chem. Phys.*, 16,
1076 8715–8728, 2016b.
- 1077 Salma, I., Varga, V., and Németh, Z.: Quantification of an atmospheric nucleation and growth process
1078 as a single source of aerosol particles in a city, *Atmos. Chem. Phys.*, 17, 15007–15017, 2017.
- 1079 Schobesberger, S., Junninen, H., Bianchi, F., Lonn, G., Ehn, M., Lehtipalo, K., Dommen, J., Ehrhart,
1080 S., Ortega, I. K., Franchin, A., Nieminen, T., Riccobono, F., Hutterli, M., Duplissy, J., Almeida, J.,
1081 Amorim, A., Breitenlechner, M., Downard, A. J., Dunne, E. M., Flagan, R. C., Kajos, M.,
1082 Keskinen, H., Kirkby, J., Kupc, A., Kurten, A., Kurten, T., Laaksonen, A., Mathot, S., Onnela, A.,
1083 Praplan, A. P., Rondo, L., Santos, F. D., Schallhart, S., Schnitzhofer, R., Sipilä, M., Tome, A.,
1084 Tsagkogeorgas, G., Vehkamäki, H., Wimmer, D., Baltensperger, U., Carslaw, K. S., Curtius, J.,
1085 Hansel, A., Petäjä, T., Kulmala, M., Donahue, N. M., and Worsnop, D. R.: Molecular
1086 understanding of atmospheric particle formation from sulfuric acid and large oxidized organic
1087 molecules, *Proc. Natl. Acad. Sci. U.S.A.*, 110, 17223–17228, 10.1073/pnas.1306973110, 2013.
- 1088 Sihto, S.-L., Mikkilä, J., Vanhanen, J., Ehn, M., Liao, L., Lehtipalo, K., Aalto, P. P., Duplissy, J.,
1089 Petäjä, T., Kerminen, V.-M., Boy, M., and Kulmala, M.: Seasonal variation of CCN concentrations
1090 and aerosol activation properties in boreal forest, *Atmos. Chem. Phys.*, 11, 13269–13285, 2011.
- 1091 Sipilä, M., Berndt, T., Petäjä, T., Brus, D., Vanhanen, J., Stratmann, F., Patokoski, J., Mauldin, R. L.
1092 3rd, Hyvärinen, A. P., Lihavainen, H., and Kulmala, M.: The role of sulfuric acid in atmospheric
1093 nucleation, *Science*, 327(5970), 1243–6. doi: 10.1126/science.1180315, 2010.
- 1094 Spracklen, D. V., Carslaw, K. S., Merikanto, J., Mann, G. W., Reddington, C. L., Pickering, S., Ogren,
1095 J. A., Andrews, E., Baltensperger, U., Weingartner, E., Boy, M., Kulmala, M., Laakso, L.,
1096 Lihavainen, H., Kivekäs, N., Komppula, M., Mihalopoulos, N., Kouvarakis, G., Jennings, S. G.,
1097 O'Dowd, C., Birmili, W., Wiedensohler, A., Weller, R., Gras, J., Laj, P., Sellegri, K., Bonn, B.,
1098 Krejčí, R., Laaksonen, A., Hamed, A., Minikin, A., Harrison, R. M., Talbot, R., and Sun, J.: The
1099 contribution of boundary layer nucleation events to total particle concentrations on regional and
1100 global scales, *Atmos. Chem. Phys.*, 6, 5631–5648, 2006.
- 1101 Tröstl, J., Chuang, W. K., Gordon, H., Heinritzi, M., Yan, C., Molteni, U., Ahlm, L., Frege, C.,
1102 Bianchi, F., Wagner, R., Simon, M., Lehtipalo, K., Williamson, C., Craven, J. S., Duplissy, J.,
1103 Adamov, A., Almeida, J., Bernhammer, A. K., Breitenlechner, M., Brilke, S., Dias, A., Ehrhart, S.,

1104 Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Gysel, M., Hansel, A., Hoyle, C. R., Jokinen, T.,
1105 Junninen, H., Kangasluoma, J., Keskinen, H., Kim, J., Krapf, M., Kürten, A., Laaksonen, A.,
1106 Lawler, M., Leiminger, M., Mathot, S., Möhler, O., Nieminen, T., Onnela, A., Petäjä, T., Piel, F.,
1107 M., Miettinen, P., Rissanen, M. P., Rondo, L., Sarnela, N., Schobesberger, S., Sengupta, K., Sipilä,
1108 M., Smith, J. N., Steiner, G., Tomè, A., Virtanen, A., Wagner, A. C., Weingartner, E., Wimmer, D.,
1109 Winkler, P. M., Ye, P. L., Carslaw, K. S., Curtius, J., Dommen, J., Kirkby, J., Kulmala, M.,
1110 Riipinen, I., Worsnop, D. R., Donahue, N. M., and Baltensperger, U.: The role of low-volatility
1111 organic compounds in initial particle growth in the atmosphere, *Nature*, 533, 527,
1112 10.1038/nature18271, 2016.

1113 Vakkari, V., Tiitta, P., Jaars, K., Croteau, P., Beukes, J. P., Josipovic, M., Kerminen, V.-M., Kulmala,
1114 M., Venter, A. D., van Zyl, P. G., Worsnop, D. R., and Laakso, L.: Reevaluating the contribution of
1115 sulfuric acid and the origin of organic compounds in atmospheric nanoparticle growth, *Geophys.*
1116 *Res. Lett.*, 42, 10486–10493, 2015.

1117 Vuollekoski, H., Sihto, S.-L., Kerminen, V.-M., Kulmala, M., and Lehtinen, K. E. J.: A numerical
1118 comparison of different methods for determining the particle formation rate, *Atmos. Chem. Phys.*,
1119 12, 2289–2295, 2012.

1120 Wehner, B., Wiedensohler, A., Tuch, T. M., Wu, Z. J., Hu, M., Slanina, J., and Kiang, C. S.:
1121 Variability of the aerosol number size distribution in Beijing, China: new particle formation, dust
1122 storms, and high continental background, *Geophys. Res. Lett.*, 31, L22108, 2004.

1123 Wiedensohler, A., Cheng, Y. F., Nowak, A., Wehner, B., Achtert, P., Berghof, M., Birmili, W., Wu, Z.
1124 J., Hu, M., Zhu, T., Takegawa, N., Kita, K., Kondo, Y., Lou, S. R., Hofzumahaus, A., Holland, F.,
1125 Wahner, A., Gunthe, S. S., Rose, D., Su, H., and Pöschl, U.: Mobility particle size spectrometers:
1126 harmonization of technical standards and data structure to facilitate high quality long-term
1127 observations of atmospheric particle number size distributions, *Atmos. Meas. Tech.*, 5, 657–685,
1128 2012.

1129 Woo, K. S., Chen, D. R., Pui, D. Y. H., and McMurry, P. H.: Measurement of Atlanta aerosol size
1130 distributions: observations of ultrafine particle events, *Aerosol Sci. Technol.*, 34, 75–87, 2001.

1131 Xiao, S., Wang, M. Y., Yao, L., Kulmala, M., Zhou, B., Yang, X., Chen, J. M., Wang, D. F., Fu, Q.
1132 Y., Worsnop, D. R., and Wang, L.: Strong atmospheric new particle formation in winter in urban
1133 Shanghai, China, *Atmos. Chem. Phys.*, 15, 1769–1781, 2015.

1134 Yli-Juuti, T., Riipinen, I., Aalto, P. P., Nieminen, T., Maenhaut, W., Janssens, I. A., Claeys, M.,
1135 Salma, I., Ocskay, R., Hoffer, A., Imre, K., and Kulmala, M.: Characteristics of new particle
1136 formation events and cluster ions at K-pusztá, Hungary. *Boreal Environ. Res.*, 14, 683–698, 2009.

1137 Zhang, R., Wang, G., Guo, S., Zamora, M. L., Ying, Q., Lin, Y., Wang, W., Hu, M., and Wang, Y.:
1138 Formation of urban fine particulate matter, *Chem. Rev.*, 115, 3803–3855, 2015.