

Response to Reviewer Comments:

We thank the reviewer for the suggestions on how to improve our paper. Our responses follow below, with reviewer comments in **black** and our responses in **blue**. Alterations to the manuscript text are in quotes with new additions in *italics*.

Reviewer #1

1. The authors suggest in Page 8 line 25 “.. the expected local nocturnal chemical production of nitrate in the RL should exhibit relatively minor vertical variation (due to variations in temperature and RH) (Figure S5). In other words, without loss or dilution processes, it is expected that the NO₃-(p) concentration would increase to a similar extent at all RL altitudes.” and attribute the observed vertical distribution of NO₃-(p), which shows enhancements ~ 100 – 150 m agl at ~ 9:30 AM, to the differential horizontal advection in the RL. I am not 100% convinced that the horizontal differential advection is the main driver of the observed morning profiles of NO₃-(p). Figure 3a does NOT show maximum wind speed at ~ 250 m, which is also the altitude at which morning NO₃-(p) is minimum. Figure 3 suggests to me the strongest winds appear to be at 150 m very frequently, whereas on average they (b2 and b3) are basically the same through a deep layer and very light winds (~1.5 m/s). This is consistent with the idea that the horizontal transport aloft is responsible for the valley-wide distribution of pollutants/precursor gases, but it does not necessarily suggest its direct role in the vertical distribution of NO₃-(p). Also the average vertical profiles shown in Figure S3 and Figure 2A look very similar to the average nocturnal profiles of NO₂, O₃ and N₂O₅ reported by Brown et al. 1, indicating the altitude dependent chemical composition due to stable conditions could account for the observed NO₃-(p) profile if heterogeneous uptake of N₂O₅ is the key source of NO₃-(p). Consider re-phrasing some of the statements.

While we appreciate the reviewer’s concern, we disagree with at least the first part of the reviewer’s argument. Figure 3A (now Figure 4A) very clearly shows a maximum in horizontal wind speed at 250 m. This is apparent in the Figure, redrawn on the next page to explicitly show lines for 150 m and 250 m. The mean horizontal wind speed at 250 m is clearly larger than at 150 m. We suspect there was some confusion as to what altitude is associated with each arrow. The altitude is indicated by the little circles at the start of the arrow, not the arrow head. Given that both reviewers had some difficulty with this figure, we have modified the caption to explicitly indicate that the little circles are the altitude, not the arrow heads: “The length of the arrows corresponds to the wind speed and the direction to the average wind direction, *with the measurement height indicated by the small circle on the tail of the arrow.*” Our hope is that this change in the caption clarifies the altitude at which each measurement was made.

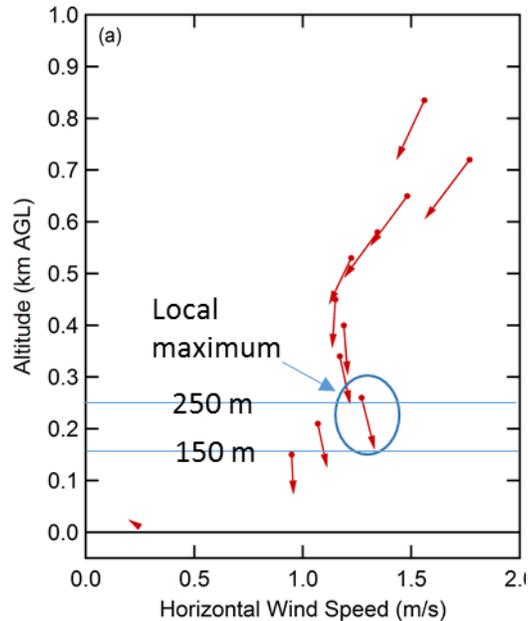


Figure 1. Annotated reproduction of Figure 3A from the original manuscript, showing the maximum horizontal wind speed is observed at 250 m.

Regarding the Brown et al. (2007) measurements, it should first be made clear that those measurements were made in the fall in a rural area of Colorado, not in wintertime CA over a medium-sized city such as Fresno. Regardless, they observed notable vertical structure in NO_3 and N_2O_5 , indicative of “distinct chemical regimes as a function of altitude.” One potentially important difference is that, unlike in Fresno, they observed substantial differences in the relative humidity with altitude. This could contribute to some of the vertical variation, although not likely the highly resolved structure that was occasionally seen by them in their series of overnight profiles.

To relate this to the Fresno measurements here, we emphasize that the chemical regime is not fundamentally separable from advection. This is especially so since both O_3 and NO_2 will start in the residual layer with approximately altitude-independent concentrations. If advection serves to bring in cleaner air (with respect to NO_x and O_3) from outside the city then the N_2O_5 production would be reduced. Conversely, if O_3 and NO_2 are higher outside the city then N_2O_5 production might actually be enhanced. Brown et al. (2007) did not explicitly discuss the potential role of advection, but it is implicit in serving to create vertical variability in concentrations. We also note that the surface and 50 m nocturnal wind speeds reported by Brown et al. tended to be higher than the wind speeds observed during DISCOVER-AQ. This could potentially explain some of the greater variability observed by them, especially since higher wind speeds would also tend to produce comparably higher nocturnal boundary layers, consistent with their observations of potential temperature. For Fresno, NO_2 is clearly lower outside the city region (Pusede et al., 2014; Pusede et al., 2016). Measurements of O_3 and NO_2

made over the DISCOVER-AQ period are available for Madera (located NW of Fresno), Fresno and Parlier (located SE of Fresno) from the California Air Resources Board. (These cities were selected for comparison, based on the comprehensive summertime analysis of Pusede et al. (2014).) O₃ concentrations are slightly higher in the more rural cities around Fresno (Figure 2; added to the supplemental as Figure S6). The particle nitrate production efficiency scales approximately as [NO₂][O₃] (all other factors being equal). This product is substantially higher in Fresno than in Madera and Parlier in the late afternoon, when boundary layer decoupling occurs (Figure 2; added to the supplemental as Figure S6). This suggests that advection from surrounding areas within the residual layer is likely to decrease the overall overnight particulate nitrate production over Fresno, as experienced the next morning when the residual layer air is entrained to the surface. However, we note that this assumes otherwise equivalent conditions (in particular, the loss rate of N₂O₅ with respect to uptake on particles). We also point to Figure S6 (now S7), which explicitly shows an inverse relationship between the nitrate concentration in the early-morning vertical profile and the overnight average wind speed at that altitude.

The above is all to say that we believe it is fully consistent with the observations of Brown et al. (2007) to make an argument that advection is a driving force in shaping the overnight the local vertical profile, since here the calculations of nitrate production based on observationally constrained initial conditions indicate that chemical or physical (T & RH) differences alone are unlikely to explain the observations. Thus, advection can have a secondary impact of changing the chemical conditions in addition to the direct impact of importing air with a lower particulate nitrate concentration. This was originally discussed on P8,L25 through P9,L11. We have revised this section to try and make these points more clearly, as:

“Box model calculations indicate that the expected local nocturnal chemical production of nitrate in the RL should exhibit relatively minor vertical variation due to variations in temperature and RH *alone* (Figure S5). In other words, without *advective* loss or dilution processes of *either* NO_{3⁻(p)} or the precursor gases it is expected that the NO_{3⁻(p)} concentration would increase to a similar extent at all RL altitudes.

The substantial changes observed in the shape of the vertical profile overnight indicate that night time differential advection in the RL is a major factor in determining the shape of the morning NO_{3⁻(p)} vertical profile during this pollution episode. Differential horizontal advection serves to *directly* export NO_{3⁻(p)} from the urban area and import cleaner air from surrounding areas. *Secondarily*, as NO_x concentrations are also lower outside of the Fresno urban area (Pusede et al., 2014), this differential advection will also influence the over-city concentrations of precursor gases (NO_x and O₃; Figure S3-S4) and consequently the *altitude-specific* nitrate production, with decreases likely. *This is supported by surface-level measurements of NO_x and O₃ made in Fresno and in the nearby and much more rural cities of Parlier (located 35 km SE of Fresno) and Madera (located 40 km NW of Fresno). The NO_x and*

NO₂ concentrations are higher and the O₃ lower in Fresno compared to the surrounding cities throughout the day, and the instantaneous nitrate production rate ([NO₂][O₃]) is substantially higher in Fresno in the late afternoon, when decoupling occurs (Figure S6). The important implication is that overnight advection both directly and indirectly alters the vertical NO₃⁻(p) profile and decreases the over-city NO₃⁻(p) concentrations in the morning, which will consequently serve to limit the extent of localized pollution build-up during events.”

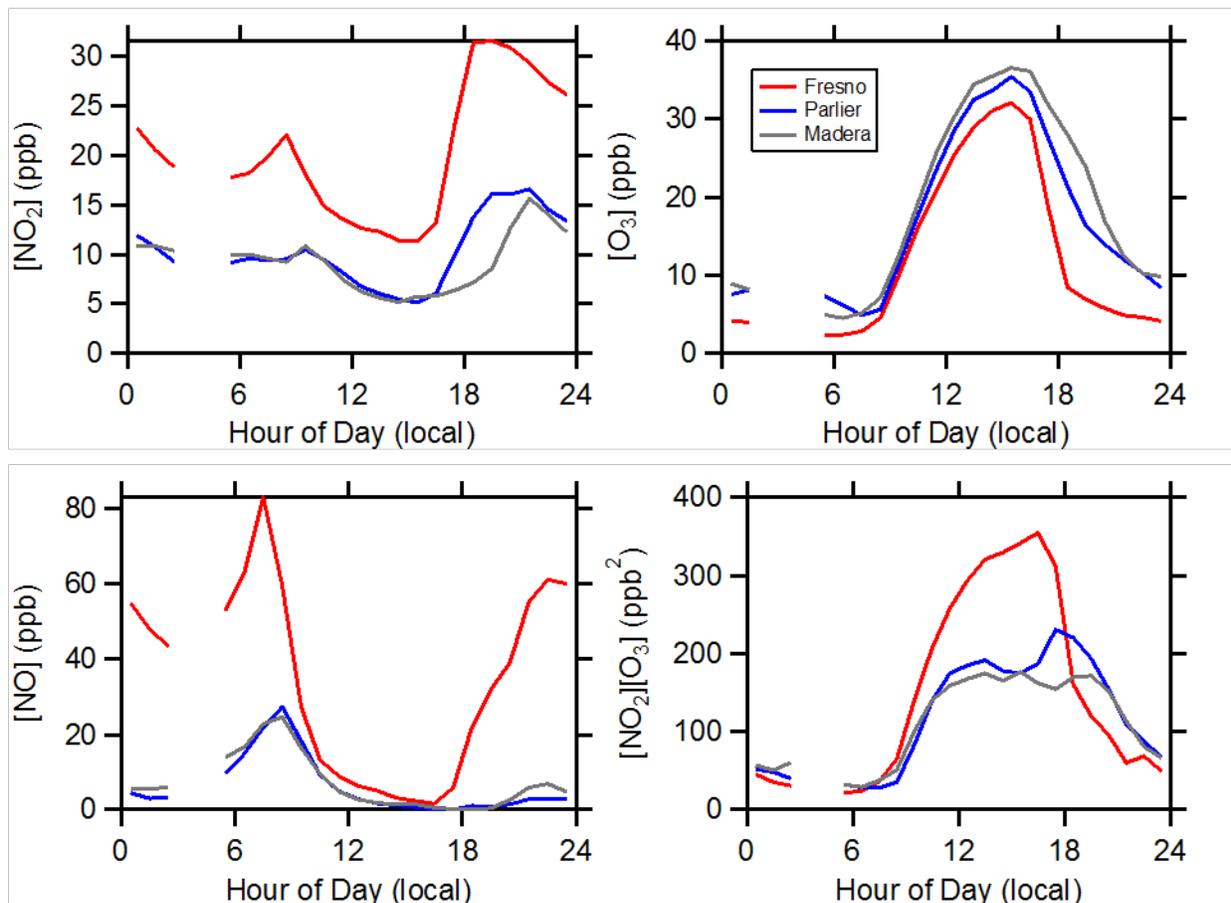


Figure 2. Diurnal profiles in NO₂, O₃, NO and NO₂ x O₃ concentrations for Fresno, Parlier and Madera.

2a. The authors assume morning chemistry due to ClNO₂ and/or HONO is negligible even though the vertical profiling takes place 9:30 AM, after ~1 – 2 after sunrise and Brown et al. 2 reports enhancements in NO₃⁻(p) during the night and early morning. More justification or comment on morning chemistry is needed.

As stated by (Brown et al., 2006), “Surface aerosol nitrate data were not of sufficient

quality and are not used.” Thus, they only report on measurements of particulate nitrate made at 90 m AGL, not at the surface. This is an important distinction since we find that the nocturnal boundary layer height was \ll 90 m AGL for the DISCOVER-AQ measurements and we are reporting on diurnal variation in surface concentrations. This key difference makes it difficult to compare the Brown et al. (2006) observations with our observations. Their comparison between O_3 and NO concentrations measured at 90 m and 7 m AGL certainly suggests that they were sampling particles above the surface layer when sampling at 90 m. As such, we would fully expect to observe different diurnal behavior between their study and ours. Their observing of overnight “enhancements” in particulate nitrate are fully consistent with the mechanism that we are proposing. We also note that Brown et al. (2006) do not report diurnal profiles for their particulate nitrate measurements, despite reporting them for many other parameters (e.g. particle number, black carbon, O_3 , NO). It is thus quite difficult to discern any specific timing with respect to time of day from their measurements.

Regarding HONO, (also relevant to 2b below), we note that the estimated peak OH includes contributions of HONO photolysis, which can actually dominate the OH production (Kim et al., 2014a). We now indicate this specifically as:

“Photochemical production of HNO_3 is calculated based on the oxidation of NO_2 by hydroxyl radicals, with wintertime concentrations estimated to peak around $[OH] = 10^6$ molecules cm^{-3} at noon in the region, *with contributions from $O(1D) + H_2O$ (from NO_2 photolysis), HONO photolysis and CH_2O photolysis* (Pusede et al., 2016).”

2b. In the same vein, the authors derive [OH] by using peak [OH] of 1×10^6 molecules/ cm^3 at noon, a relatively large number under wintertime conditions (although conditions in California might be different than other areas with typical winter season) and by scaling it with the observed solar radiation and estimate the photochemical contribution to be large, $\sim 60\%$ at 3 PM. Discussion on the effects of snow cover on the ground, cloud cover and fog during these episodes will be helpful since typical cold pool conditions often include those, which in turn limit the photochemical production.

The assumption of $[OH] = 1e6$ molecules cm^{-3} was based on the more detailed chemical calculations of Pusede et al. (2016), and includes contributions from $O(1D) + H_2O$ (from NO_2 photolysis), HONO photolysis and CH_2O photolysis. We note that there was no snow on the floor of the SJV during the campaign (and almost never is). There was little to no fog during DISCOVER-AQ, a contrast from some other years and a result of the regional climatology (e.g. el Nino vs. la Nina) (Young et al., 2016). Cloud cover was minimal. We have added the following to the “Materials and Methods” section of the manuscript: “*Local conditions during DISCOVER-AQ were relatively cool ($T_{avg} = 7.9$ °C) and dry ($RH_{avg} = 69\%$) with frequent sunshine and no visible fog.*”

2c. In general, more extended discussion contrasting pollution episodes in other areas will improve the overall quality of this manuscript.

We have added a new section “3.4 Linking to other regions.” Page 20, Lines 1-5 have been moved to this section and additional discussion is now provided. The new section is provided here:

3.4 Linking to other regions

“Production of $\text{NO}_3^-_{(p)}$ in the RL can vary widely based on initial concentrations of its precursor gases, as well as the rate of heterogeneous uptake of N_2O_5 by particles. It may be that production of $\text{NO}_3^-_{(p)}$ via the N_2O_5 hydrolysis pathway may be significant in the aloft RL in other regions with similar geographical and meteorological conditions, such as Salt Lake Valley, Utah (Kuprov et al., 2014; Baasandorj et al., 2017). However, in valley regions with lower NO_x or O_3 the nocturnal PNO_3^- may be lower, thus limiting the importance of this pathway (Akira et al., 2005; Bigi et al., 2012). Among other factors, the extent to which nocturnal $\text{NO}_3^-_{(p)}$ formation occurs more-so in the surface layer versus in layers aloft will depend importantly on the extent of NO_x emissions at the surface (which titrate O_3 , suppressing particulate nitrate formation), the absolute and relative height of the nocturnal boundary layer (which affects the rate HNO_3 deposition and the air volumes in which nitrate production occurs), and gradients in RH, T and NH_3 (Kim et al., 2014b).

For example, Baasandorj et al. (2017) observe at their valley wall and valley floor sites in wintertime Utah that O_3 concentrations near the surface remain well-above zero even during pollution episodes, thus allowing for surface-level $\text{NO}_3^-_{(p)}$ formation overnight, substantiated by direct measurements of N_2O_5 , in addition to formation aloft. Nitrate-specific diurnal profiles were not reported. In Shanghai, China, Wang et al. (2009) observed in fall 2007 that both O_3 and NO_2 remain elevated at night at the surface, with a concomitant increase in surface $\text{NO}_3^-_{(p)}$. And in wintertime Seoul, Korea, Kim et al. (2017) observed relatively limited diurnal variability in O_3 and NO_2 concentrations measured at 60 m, with both remaining elevated throughout the night. However, they did not observe any notable build up in $\text{NO}_3^-_{(p)}$ overnight, but did observe $\text{NO}_3^-_{(p)}$ to increase and peak in the morning, as here. In contrast, in Fresno here the night time surface O_3 levels during Episode 1 were nearly zero, suppressing surface $\text{NO}_3^-_{(p)}$ formation. This near-zero nocturnal O_3 is similar to observations by Kuprov et al. (2014) made a few years before Baasandorj et al. (2017) at one of the same valley floor sites in Utah, reflecting year-to-year differences. Such differences can influence the extent to which a notable increase in $\text{NO}_3^-_{(p)}$ is observed to occur in the early morning as air is entrained from the residual layer to the surface. This is because if surface production and production in the residual layer are similar in magnitude the contrast between the two will be reduced and entrainment

will appear to have a less apparent impact on the diurnal profile. However, because the effective volume of the residual layer is typically much larger than the nocturnal boundary layer (as is the case here), even without an observed increase in $\text{NO}_3^-_{(p)}$ at the surface in the morning the $\text{NO}_3^-_{(p)}$ produced in the residual layer can still dominate the overall $\text{NO}_3^-_{(p)}$ burden during the day.

Additionally, comparison between the Baasandorj et al. (2017) observations of late afternoon surface NO_2 and O_3 (which reflect the initial conditions within the residual layer) with the Fresno observations indicates differences can exist in how nocturnal production in layers aloft influences the build-up and sustaining of $\text{PM}_{2.5}$ in pollution episodes. They observed during a strong $\text{PM}_{2.5}$ episode a slow build-up of $\text{PM}_{2.5}$ followed by a plateau lasting multiple days. During this period, late-afternoon O_3 concentrations decreased over time while late-afternoon NO_2 was approximately constant (in the daily average). Consequently, the nitrate radical production rate in the residual layer, and thus the N_2O_5 and $\text{NO}_3^-_{(p)}$ production rates, decreased over time in their study. In contrast, for Episode 1 here, the late-afternoon nitrate radical production rate increased over time across the episode (by $0.25 \mu\text{g m}^{-3} \text{day}^{-1}$), with only a moderate decrease in the daytime O_3 over time (Figure S9). These differences reflect the different photochemical conditions between the regions, and illustrate the coupling between the daytime photochemical conditions (i.e. O_3 production) and night time $\text{NO}_3^-_{(p)}$ formation above the surface.”

3. Transport out or entrainment from free troposphere in the afternoon leads to the loss of $\text{PM}_{2.5}$, but it also introduces O_3 and dilutes NO_x , impacting the PNO_3 formation at night. What about the later effects? Figure S3c indicates significant differences in O_3 above and below the inversion cap during this episode.

Indeed, entrainment of FT air alters concentrations of O_3 and NO_x in the mixed boundary layer, which then becomes the residual layer. However, as we use the observations of O_3 and NO_2 to constrain the formation model, and since the daytime boundary layer is well-mixed, the influence of daytime entrainment on at least the initial conditions in the residual layer is accounted for. To the extent that there is much entrainment of FT air into the residual layer at night this would impact chemical production in the residual layer. Given the strong atmospheric stability at night, nighttime entrainment of FT air into the residual layer is not expected to be substantial, just as mixing between the residual layer and the nighttime boundary layer is strongly limited. (This point was already noted on P8 when we stated that “In the absence of a strong jet aloft and no convective mixing, night time entrainment of cleaner FT air into the RL is expected to be considerably slower than horizontal advection.”) To make this point further, we have now added the following to the main text description of the box model: *“While entrainment of FT air also alters the NO_2 and O_3 concentrations in the mixed layer, since these are constrained by the surface observations (within the mixed layer) this is accounted for.”*

4. Page 13 line 17: “The peak and turnover in surface-level NO₃-(p) occurs when higher RL layers, where [NO₃-(p)]_{RL} < [NO₃-(p)]_{ML}, are entrained.”. Shouldn’t it be “[NO₃-(p)]_{RL} > [NO₃-(p)]_{ML}”?

No. The increase up to the peak occurs due to entrainment of air with [NO₃-(p)]_{RL} > [NO₃-(p)]_{ML}. However, the actual peak occurs when air above the especially nitrate rich layer is entrained to the surface and where this higher-level air has [NO₃-(p)]_{RL} < [NO₃-(p)]_{ML}. The help clarify this, we now indicate that this “occurs when *even* higher RL layers” are entrained.

5. Using increase rate in 24-h PM_{2.5} running average and % nitrate contribution might be more suitable for estimating the nocturnal nitrate production as it already take into consideration the transport term.

We assume this comment is in reference to using the early-morning peak values (Figure 6B) to characterize the increase. Since we have direct measurements of particulate nitrate concentrations at the surface, it seems more straightforward to use those measurements, rather than PM_{2.5} and the nitrate contribution (since these are all related). As an alternative to the peak analysis, we have fit the 24-h average particulate nitrate concentrations over the episode. This gives a slope of 0.66 ug m⁻³ d⁻¹, as opposed to 1.32 ug m⁻³ d⁻¹ for the peak analysis. This is now stated on P18 as: “*For comparison, the 24-h average surface-level NO₃ (p) increases by 0.66 μg m⁻³ day⁻¹.*”

Reviewer #2:

We thank the reviewer for their helpful comments. Upon comparing the line numbers referenced by the reviewer with the line numbers in the ACPD paper, it appears that the reviewer may have incidentally been reading the initial submission that did not include changes in response to the Quick Review process. We note this simply to avoid confusion in terms of line numbers/figure numbers. We will refer to the published ACPD version in our response.

1. It would be helpful to have a better discussion related to how this paper fits into previous studies from Idaho, Washington, and Utah. The authors state in the abstract (page 2, lines 17-19) that the results from their paper “provide general insights into the evolution of pollution episodes in wintertime environments”. However, the discussion that relates their findings to previous research in other areas is limited to two sentences (page 20, lines 1-5).

We have added a new section “3.4 Linking to other regions.” Page 20, Lines 1-5 have been moved to this section and additional discussion is now provided.

2. The surface heat flux and friction velocity are mentioned in the manuscript and appendix related to the calculation of mixing layer height, it is not clear what instrument this data come from, or if they came from a model. (page 12 line 10, page 25, lines 16-19)

We accidentally failed to state that the surface heat flux and friction velocity were both constrained by sonic anemometer measurements made at the Huron site, in addition to the radiosonde measurements. This information has now been added.

3. The dataset might be too limited, but can the overnight advection be quantified and related to local sources? This would benefit the discussion on page 9 (lines 5-8) and the discussion about the advection from a nearby source on page 17 (lines 9-11).

It is difficult to be much more quantitative regarding the particular role of advection from this dataset. The wind profiler data available from nearby Huron provide guidance, but as they are not co-located with the ground site it is difficult to be more explicitly quantitative. We do note here that we found that our back-calculated N_2O_5 uptake coefficient was lower than the expected uptake coefficient based on composition, which is consistent with advection having decreased local concentrations relative to the total potential nitrate formation overnight in the absence of advection. Regarding the discussion on P17, we have worked to understand the particular source of this evening “bump” in greater detail, for example using HYSPLIT back trajectories to identify likely source locations, but no clear answer has developed from this. One of the challenges here is that we do not know, for example, whether this late evening increase is likely the result of (for example) some source located close and transported a short distance over a short period of time or located further away and transported over a longer distance and time. Assessment is further complicated by the strong shift in wind direction between the surface (10 m) and higher altitudes (Figure 3 in the initial manuscript, now Figure 4). Since the focus of the manuscript is on the early morning and afternoon behavior, rather than this late evening phenomenon, we have limited the discussion regarding the nature or location of the nearby source.

4. Page 6, lines 15-16: “differential horizontal transport in the RL” sounds a bit awkward; the term advection might be suitable here.

We have changed this to “altitude-dependent advection” so as to continue to emphasize the idea that the influence of advection is not the same at all layers.

5. Page 7, lines 22-24: “The derived, observationally constrained...” Referring to the estimated versus observed profiles of nitrate. While in general there is good agreement, there is not good agreement in the morning profiles close to the surface (<75m), especially in regard to the shape of the profiles. It would be worth mentioning that here.

We have clarified the comparison to be more precise as follows:

“The derived, observationally constrained NO₃ (p) profiles based on the estimated NO₃ (p) exhibit generally good correspondence with the sparser direct measurements of NO₃-(g+p), although on one of the two days available for comparison the total NO₃⁻ somewhat exceeds the estimated NO₃⁻(p) below ~75 m (Figure 2).”

6. Figure 4 is referenced in the text before Figure 3, these two figures should be switched.

We have switched the order of these figures.

7. Page 9, line 17-18: The maximum wind speed occurring at 250m is not clear based on the data shown in Figure 3a.

We have added the word “local” to indicate that there is a local, not global, maximum at ~250 m height. Also, in Figure 1 above in this response we point out the local maximum. We have also modified the caption as follows: *“The length of the arrows corresponds to the wind speed and the direction to the average wind direction, with the measurement height indicated by the small circle on the tail of the arrow.”*

8. Page 12, lines 24-26: “good agreement” while the timing of the morning peak between the model and observations is captured the model does not capture increases in the surface nitrate concentrations after the ML decreases. While this is discussed later in the paper, it should be mentioned here that the model does not have good agreement with observations during the evening.

We now state: *“The model predictions for the individual flight days also exhibit generally good agreement with the NO₃ (p) observations except in the late evening, discussed further below (Figure S7).”*

9. Page 17, line 9: Should this be referencing Figure S6 instead of Figure 8?

This should have been Figure S7, not Figure 8. This has been corrected.

10. Page 19, lines 18-19: “vertical mixing has a particularly large impact on the ...” Is this vertical mixing really entrainment and dry deposition? The paper does not quantify or discuss modeling results for vertical mixing throughout the boundary layer; the box model focus is dry deposition and entrainment.

We have clarified this statement in the revised manuscript. What was intended here is an indication that the very shallow nocturnal boundary layer means that vertical mixing in the

morning, which entrains air from the residual layer, can strongly control the surface concentrations. We now state:

“vertical mixing in the early morning, which entrains air from the residual layer into the surface mixed layer, has a particularly large impact on the surface concentrations here due to the nocturnal boundary layer being exceptionally shallow.”

11. The sampling times for the instruments might be beneficial to the reader (Table A1), for example “Fast measurements: : :” (page 7, line 16) does not really have a context.

We have revised this to say: “Fast measurements of total NO_3^- (gas + particle, NO_3^- (g+p)) were only available for a subset of flights (Pusede et al., 2016), and particulate-only NO_3^- measurements were not made with sufficient time resolution, *less than about a minute*, to allow for robust characterization of the NO_3^- (p) vertical profile.”

12. Page 25, lines 25-27: Is the assumption for the boundary layer to linearly drop over a 1-hour period reasonable, it seems too quick, and how does this assumption impact the results? This is vaguely referred to on page 17 (lines 19-22) where the decoupling in the model occurs very rapidly while the temperature and RH changes from observations appear to be more gradual. Is this, or could it potentially, have an impact on the evening increase in nitrate concentrations?

The fall in the boundary layer height occurs much faster than the temperature and RH changes, as it is strongly related to the input of solar radiation. The decrease in the boundary layer height can occur very rapidly as the solar flux decreases since there is then a distinct lack of thermal forcing. Thus, despite the temperature remaining elevated for many hours after the sun goes down, there is not an energy input to sustain a high boundary layer height. This general phenomenon is evident in the classic figure from Stull (redrawn on [Wikipedia](#), pasted below for reference). More specific to the SJV, Bianco et al. (2011) observed diurnal profiles of boundary layer heights across summer-winter seasons for one year using wind profiles at various sites in the SJV. At the site located nearest to Fresno, Chowchilla, they observed that the daytime BLH was approximately constant during the winter, with perhaps a slight drop ~2 h before sunset followed by a rapid drop in the hour right around sunset. This behavior is generally consistent with our model framework. Certainly the linear decline is an approximation to the true behavior. We have added the following to Appendix B to indicate this: *“A relatively rapid (~1 h) decline in the mixed layer height is consistent with wintertime observations of diurnal BLH profiles (Bianco et al., 2011).”*

In terms of the implications of this approximation/assumption, this would, however, not have an impact on the late evening increase in particulate nitrate. But, if we assumed a slower decline (such that the BLH remained higher for longer) then the calculated late afternoon decrease in surface concentrations of particulate nitrate due to the “gas-phase pump” and

HNO₃ deposition would be decreased (c.f. equation C9, which has BLH in the denominator). Alternatively, if the BLH fell more rapidly, then the calculated late afternoon decrease in particulate nitrate would have been greater. We aimed to make this point in the original manuscript with the parenthetical that stated (P17,L4-7):

“(In the model here, the decoupling is assumed to occur very rapidly while the temperature and RH changes are from observations and occur more gradually. If the decoupling was actually slower the influence of the gas-phase pump at this point in time would be reduced and the modelled decrease in NO₃ (p) that occurs around 3-5 pm would be less than shown.)”

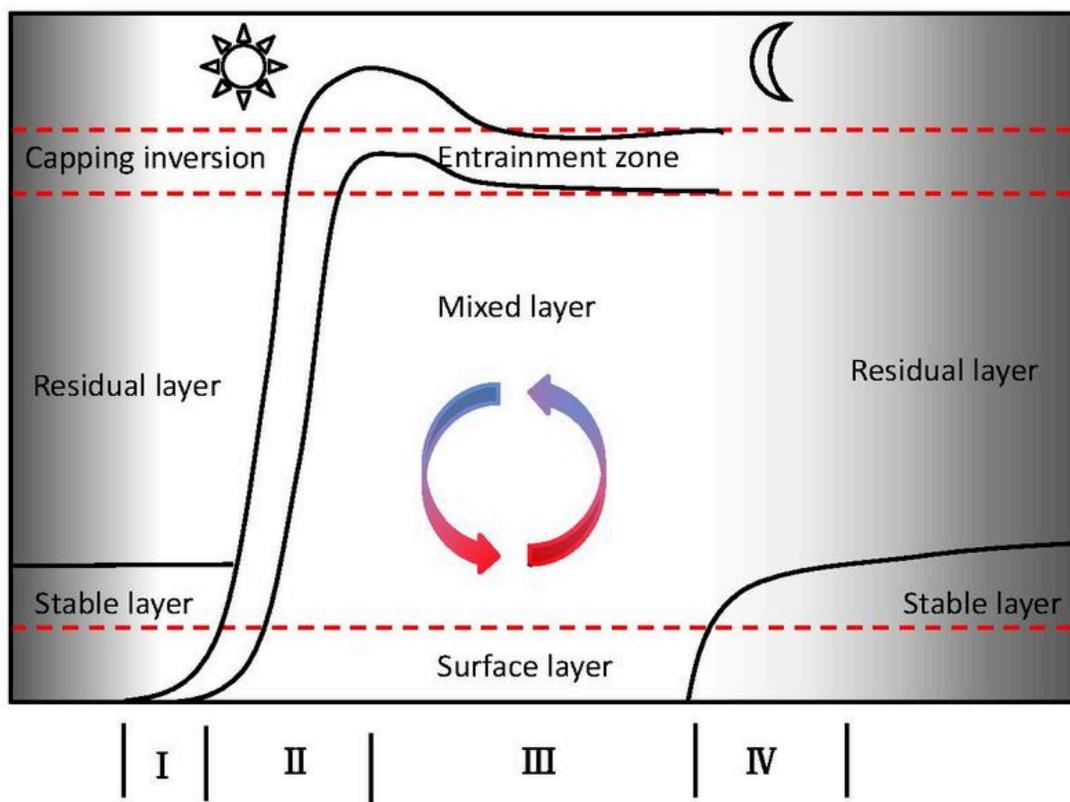


Figure 3. Evolution of the near surface atmosphere over time, taken from Wikipedia and after Stull. Period I = sunrise, Period II = growth of the daytime boundary layer, Period III = sustaining of the daytime boundary layer, Period IV = decoupling of the surface layer and residual layer near sunset.

13. There are a few (~20-30) minor typos or grammatical errors.

We have read through and done our best to identify all of these and fix them.

References:

- Akira, K., Akikazu, K., Kiyoshi, I., Yoshio, I., Masahiko, S., Lal, S. M., and Balkrishan, S.: Investigation of air pollution concentration in Kalthmandu valley during winter season, *Journal of Environmental Sciences*, 17, 1008-1013, doi:1001-0742(2005)06-1008-06, 2005.
- Baasandorj, M., Hoch, S. W., Bares, R., Lin, J. C., Brown, S. S., Millet, D. B., Martin, R., Kelly, K., Zarzana, K. J., Whiteman, C. D., Dube, W. P., Tonnesen, G., Jaramillo, I. C., and Sohl, J.: Coupling between Chemical and Meteorological Processes under Persistent Cold-Air Pool Conditions: Evolution of Wintertime PM_{2.5} Pollution Events and N₂O₅ Observations in Utah's Salt Lake Valley, *Environ. Sci. Technol.*, doi: 10.1021/acs.est.6b06603, 2017. doi:10.1021/acs.est.6b06603, 2017.
- Bianco, L., Djalalova, I. V., King, C. W., and Wilczak, J. M.: Diurnal Evolution and Annual Variability of Boundary-Layer Height and Its Correlation to Other Meteorological Variables in California's Central Valley, *Boundary-Layer Meteorology*, 140, 491-511, doi:10.1007/s10546-011-9622-4, 2011.
- Bigi, A., Ghermandi, G., and Harrison, R. M.: Analysis of the air pollution climate at a background site in the Po valley, *J. Environ. Monitoring*, 14, 552-563, doi:10.1039/C1EM10728C, 2012.
- Brown, S. G., Hyslop, N. P., Roberts, P. T., McCarthy, M. C., and Lurmann, F. W.: Wintertime Vertical Variations in Particulate Matter (PM) and Precursor Concentrations in the San Joaquin Valley during the California Regional Coarse PM/Fine PM Air Quality Study, *J. Air & Waste Manage. Assoc.*, 56, 1267-1277, doi:10.1080/10473289.2006.10464583, 2006.
- Brown, S. S., Dube, W. P., Osthoff, H. D., Wolfe, D. E., Angevine, W. M., and Ravishankara, A. R.: High resolution vertical distributions of NO₃ and N₂O₅ through the nocturnal boundary layer, *Atmos. Chem. Phys.*, 7, 139-149, 2007.
- Kim, H., Zhang, Q., Bae, G. N., Kim, J. Y., and Lee, S. B.: Sources and atmospheric processing of winter aerosols in Seoul, Korea: insights from real-time measurements using a high-resolution aerosol mass spectrometer, *Atmos. Chem. Phys.*, 17, 2009-2033, doi:10.5194/acp-17-2009-2017, 2017.
- Kim, S., VandenBoer, T. C., Young, C. J., Riedel, T. P., Thornton, J. A., Swarthout, B., Sive, B., Lerner, B., Gilman, J. B., Warneke, C., Roberts, J. M., Guenther, A., Wagner, N. L., Dubé, W. P., Williams, E., and Brown, S. S.: The primary and recycling sources of OH during the NACHTT-2011 campaign: HONO as an important OH primary source in the wintertime, *Journal of Geophysical Research-Atmospheres*, 119, 6886-6896, doi:10.1002/2013JD019784, 2014a.
- Kim, Y. J., Spak, S. N., Carmichael, G. R., Riemer, N., and Stanier, C. O.: Modeled aerosol nitrate formation pathways during wintertime in the Great Lakes region of North America, *Journal of Geophysical Research-Atmospheres*, 119, 12420-12445, doi:10.1002/2014jd022320, 2014b.
- Kuprov, R., Eatough, D. J., Cruickshank, T., Olson, N., Cropper, P. M., and Hansen, J. C.: Composition and secondary formation of fine particulate matter in the Salt Lake Valley: Winter 2009, *J. Air & Waste Manage. Assoc.*, 64, 957-969, doi:10.1080/10962247.2014.903878, 2014.
- Pusede, S. E., Duffey, K. C., Shusterman, A. A., Saleh, A., Laughner, J. L., Wooldridge, P. J., Zhang, Q., Parworth, C. L., Kim, H., Capps, S. L., Valin, L. C., Cappa, C. D., Fried, A., Walega, J., Nowak, J. B.,

Weinheimer, A. J., Hoff, R. M., Berkoff, T. A., Beyersdorf, A. J., Olson, J., Crawford, J. H., and Cohen, R. C.: On the effectiveness of nitrogen oxide reductions as a control over ammonium nitrate aerosol, *Atmos. Chem. Phys.*, 16, 2575-2596, doi:10.5194/acp-16-2575-2016, 2016.

Pusede, S. E., Gentner, D. R., Wooldridge, P. J., Browne, E. C., Rollins, A. W., Min, K. E., Russell, A. R., Thomas, J., Zhang, L., Brune, W. H., Henry, S. B., DiGangi, J. P., Keutsch, F. N., Harrold, S. A., Thornton, J. A., Beaver, M. R., St. Clair, J. M., Wennberg, P. O., Sanders, J., Ren, X., VandenBoer, T. C., Markovic, M. Z., Guha, A., Weber, R., Goldstein, A. H., and Cohen, R. C.: On the temperature dependence of organic reactivity, nitrogen oxides, ozone production, and the impact of emission controls in San Joaquin Valley, California, *Atmos. Chem. Phys.*, 14, 3373-3395, doi:10.5194/acp-14-3373-2014, 2014.

Wang, X., Zhang, Y., Chen, H., Yang, X., Chen, J., and Geng, F.: Particulate Nitrate Formation in a Highly Polluted Urban Area: A Case Study by Single-Particle Mass Spectrometry in Shanghai, *Environ. Sci. Technol.*, 43, 3061-3066, doi:10.1021/es8020155, 2009.

Young, D. E., Kim, H., Parworth, C., Zhou, S., Zhang, X., Cappa, C. D., Seco, R., Kim, S., and Zhang, Q.: Influences of emission sources and meteorology on aerosol chemistry in a polluted urban environment: results from DISCOVER-AQ California, *Atmos. Chem. Phys.*, 16, 5427-5451, doi:10.5194/acp-16-5427-2016, 2016.