Long-term monitoring of persistent organic pollutants (POPs) at the Norwegian Troll station in Dronning Maud Land, Antarctica

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Abstract

A first long-term monitoring of selected persistent organic pollutants (POPs) in Antarctic air has been conducted at the Norwegian Research station Troll (Dronning Maud Land). As target contaminants 32 PCB congeners, α- and γ-hexachlorocyclohexane (HCH), trans- and cis-chlordane, trans- and cis-nonachlor, p,p’- and o,p-DDT, DDD, DDE as well as hexachlorobenzene (HCB) were selected. The monitoring program with weekly samples taken during the period 2007–2010 was coordinated with the parallel program at the Norwegian Arctic monitoring site (Zeppelin mountain, Ny-Ålesund, Svalbard) in terms of priority compounds, sampling schedule as well as analytical methods. The POP concentration levels found in Antarctica were considerably lower than Arctic atmospheric background concentrations. Similar as observed for Arctic samples, HCB is the predominant POP compound with levels of around 22 pg m\(^{-3}\) throughout the entire monitoring period. In general, the following concentration distribution was found for the Troll samples analyzed: HCB > Sum HCH > Sum PCB > Sum DDT > Sum chlordanes.

Atmospheric long-range transport was identified as a major contamination source for POPs in Antarctic environments. Several long-range transport events with elevated levels of pesticides and/or compounds with industrial sources were identified based on retroplume calculations with a Lagrangian particle dispersion model (FLEXPART).

The POP levels determined in Troll air were compared with concentrations found in earlier measurement campaigns at other Antarctic research stations from the past 18 yr. Except for HCB for which similar concentration distributions were observed in all sampling campaigns, concentrations in the recent Troll samples were lower than in samples collected during the early 1990s. These concentration reductions are obviously a direct consequence of international regulations restricting the usage of POP-like chemicals on a worldwide scale.
1 Introduction

Polar regions are considered today as important sentinels for global environmental processes for many scientific disciplines involved in environmental research. Polar locations are still characterized by minimum anthropogenic presence and, thus, are suitable regions for comprehensive baseline studies including global circulation systems, hemispheric transport of anthropogenic pollution, radiation as well as cryosphere related polar research (ACIA, 2005). Furthermore, the physico-chemical mechanisms controlling release, hemispheric transport, distribution and deposition of target contaminants can be examined in great detail with the help of long-term monitoring of pollutants (Ma et al., 2011; AMAP, 2009). Long-term atmospheric pollution monitoring in the polar regions of our globe is today considered as valuable and versatile scientific tool for assessing anthropogenic influences on the environment as well as controlling international regulation measures (Kallenborn and Berg, 2009; UNEP, 2011; Hung et al., 2010).

The current comprehensive Arctic long-term atmospheric POP monitoring programs, in particular at the Zeppelin station and the Canadian Alert research station, proved impressively the importance of these national commitments for the continuous operation of long-term atmospheric POPs monitoring (POPs = Persistent Organic Pollutants). They are considered as the central scientific basis for the empirical investigation of atmospheric transport and distribution processes (Dutchak and Zuber, 2010; UNEP, 2008; Aas et al., 2008) and are important for the evaluation of models simulating global distribution processes (Wania, 2003; Wania and Mackay, 1999; Armitage et al., 2006). In combination with meteorological modeling, the continuous monitoring reveals also atmospheric long-range transport events as well as potential source regions, where regulative measures may be appropriate when international mitigation strategies are assessed. This information is today an important resource for the global POP regulations in order to provide the scientific basis for appropriate counter measures for global restriction of POP usage and distribution (Rodan et al., 1999; Clapp, 2003). The Arc-
tic monitoring data are continuously reported to regional and international monitoring programs such as the Arctic Monitoring and Assessment Programme (AMAP), United Nations Economic Commission for Europe European Monitoring and Evaluation Programme (UNECE-EMEP), Global Atmospheric Watch (GAW) and others (Holoubek et al., 2001). Thus, the experience with the past 20 yr of monitoring in the Arctic proved that this type of data is urgently required for a scientifically sound evaluation of hemispherical transport and fate of legacy POPs and emerging so-called new POPs.

Until today, the establishment of long-term continuous atmospheric monitoring of POPs in Antarctica has been seriously hampered by political, economic and logistical restrictions (e.g. remotesness and cost intensive logistics). During the past two decades, several campaign-based studies on short-term (weeks–month) atmospheric monitoring of POPs in Antarctica have been reported in the literature (Kallenborn et al., 1995, 1998; Dickhut et al., 2005; Gambaro et al., 2005; Xie et al., 2011a,b; Möller et al., 2011). Thus, only scattered scientific information has been available for the evaluation of atmospheric long-range transport of POPs in the high-latitude Southern Hemisphere and in particular for the Antarctic continent. The above-mentioned earlier studies indicated that potential primary sources (agriculture, industrial releases etc.) in the Southern Hemisphere contribute to the POP levels in the Antarctic atmosphere. Nevertheless, due to the remoteness of Antarctica and few potential industrial sources in the Southern Hemisphere, levels of POPs have been previously considered as extremely low – much lower than in the Arctic. However, Antarctica attracts more and more scientific interest as a location for pollutant research. Studies on potential local contamination issues in Antarctic environments reveal clearly that the anthropogenic impact on Antarctic pollution levels is not negligible despite of considerable international efforts to minimize this human impact by means of international regulations, bans and restrictions (SCAR = Scientific Committee on Antarctic Research, GOSEAC = SCAR Group of Specialists on Environmental Affairs and Conservation, etc.).

Therefore, in 2007, the Norwegian Institute for Air Research (NILU) in collaboration with the Norwegian Polar Institute (NPI) initiated the first long-term atmospheric moni-
Long-term monitoring of persistent organic pollutants (POPs)
R. Kallenborn et. al

2 Methods

Sample location. POP sampling was performed at the Troll Atmospheric Observatory, managed from the Norwegian Troll station (TRS; Dronning Maud Land/Queen Maud Land, Antarctica), located at 72° 01′ S, 2° 32′ E, at 1270 m.a.s.l. in the Nunatak area of Jutulsessen glacier, about 235 km south from the Antarctic coast (Fig. 1). Its closest neighbor stations are the South African SANAЕ IV station, 190 km west-north-west, and the German Neumeyer station, 420 km east-north-east of TRS. TRS is situated on snow-free bedrock and accessible by air transport during Antarctic summer, facilitated by a blue-ice airfield on the glacier 7 km north of the main station. In June 2003, the Norwegian Ministry of the Environment decided to upgrade an existing summer research station and provide it with a runway for airplanes (the Troll Airfield) enabling year-around operation. The new station was officially opened in 2005 and a comprehensive atmospheric monitoring program was established including pollutant monitoring in 2006/2007 (e.g. POPs). Data from the first four years of continuous air sampling for POP monitoring are presented here.
Sample collection. Samples were collected from February 2007 until December 2010 on a weekly basis. For the monitoring, continuous integrated weekly/7-days samples were collected with a flow rate of 10–15 m$^3$ h$^{-1}$ resulting in sample volumes between 2200 and 2500 m$^3$. A commercially available high volume air sampling device (DHA-80, DIGITEL, Hegenau, CH) for separate particle and gas phase collection was used for the monitoring of POPs. The gas phase was collected on pre-cleaned polyurethane foam plugs (PUF) whereas the particle phase was collected on glass fiber filters (GFF: 10 cm id, cut-off 50 µm, Schleicher-Schuell, Dassel, GER). Gas- and particle phase samples were collected simultaneously. Flow-rate and sampling conditions were digitally monitored and documented (e.g. power failures, etc.) as integrated part of the sampling and quality control procedure. Sector controlled sampling was performed where air masses passing over the main research station was avoided/not collected in order to minimize the potential for local contamination.

Sample storage. After the sampling was completed, the exposed filters (GFF and PUF) were sealed in a gas tight container for storage and transport to NILU’s laboratory for further processing and quantification. No manual treatment/exposure is required prior to analysis of the filters. In addition, extensive numbers of field and transport blank samples followed the yearly sample batch in order to control potential contamination risks (as a part of the extensive quality control procedure of the NILU monitoring program). All exposed filters were registered and stored frozen (−20°C) prior to analysis and quantification. The yearly sample set was shipped on the first flight every austral summer (November) to the analytical laboratory in Norway by combined ship and airfreight transport.

Sample preparation. The sample preparation method was optimized based on an earlier described method (Kallenborn et al., 2006). Details on the preparation methods are given in the Supplement of the paper. All solvents, adsorbents and gases used were of highest quality available, well suited and selected for ultra-trace analysis. Solvents and adsorbents were purchased from Merck (Spurenanalyse quality, Darmstadt, GER), all gases were provided by Hydro (N$_2$ and He, 6.0 quality, Porsgrunn, NO).
All samples were Soxhlet extracted for 12 h with n-hexane/acetone (50:50, v:v). An amount of 10 ng of $^{13}$C labeled Internal Standards (IST) for all congener and isomer groups analyzed was added prior to extraction (for details see Appendix A of the Supplement). The GFF and PUF extracts were unified after extraction. Volume reduction to ca. 500 µL was performed on a Turbovap (Caliper-Zymark, Hopkinton, MA, USA). After volume reduction, the extract was transferred to a 15 cm glass column (with stop cock) filled with precleaned and preconditioned Silica (2-%-w water) and a top layer of pre-cleaned sodium sulfate for cleanup. After cleanup, volume of the eluate was reduced to 500 µL (Turbovap) and further reduced to 100 µL under a gentle N$_2$ stream. Before quantitative analysis, 10 ng of unlabeled tetrachloronaphthalene (TCN) was added as recovery standard (RecStd, Tables S1–S3).

Quantitative analysis. The same priority target analytes were chosen for the Troll POP monitoring as measured at the Zeppelin research station (Ny-Ålesund, Svalbard, Arctic) for the past 20 yr. The respective monitoring programs (Zeppelin and Troll) were harmonized with respect to sampling and priority compounds in order to allow good comparability of the obtained data for future global assessments. In total, 32 PCB congeners and 13 organochlorine pesticides (OCP) were quantified (for details on compounds and analytical methods, see Appendix A including Tables S1–S3) of the Supplement.

After sample preparation, the eluates were quantitatively analyzed using high-resolution gas chromatography coupled to a high-resolution sector field mass spectrometer as detector (HRGC/HRMS). A 6890N Agilent gas chromatograph (Agilent, Palo Alto CA, USA) was coupled to an Autospec Ultima sector field mass spectrometer (Waters, Milford, MA, USA). An aliquot of 2 µL of the eluate was injected into the gas chromatograph in splitless mode (280°C and 2 min splitless time). The analyzes were performed with maximum resolution (ca. 10 000) in Electron Impact (EI: PCBs, HCB, DDT derivatives) and Negative Ion Chemical Ionization mode NICI: OCP) for the respective compounds groups (for details on separation parameters incl. temperatures for injection and transfer line, temperature program, mass spectrometer parameters...
incl. reference and quantifier ions selected, lock mass calibration, see Appendix A, Supplement). Isotope dilution was applied for identification and quantification of the target chlorinated chemicals.

Quality control. The responsible technicians at the Troll research station followed detailed routines developed by NILU in order to avoid unwanted contamination during fieldwork. A set of field blank samples (1 field blank/month + transport blanks) was added to the field samples in order to control unintended contamination during storage and transport.

The analytical procedure used for the study was accompanied by a comprehensive quality control program based on the requirements of NILU’s accreditation, according to EU standard EN 45001. The instrument limit of detection (LOD) was determined by calculating the signal to noise ratio ($S/N > 3$ for solvent blanks (using $n$-hexane). Extensive field and laboratory blank values were analyzed in order to monitor and control possible contamination during sample transport and laboratory work. No evidence for contamination during sample processing was found for the data presented here (no elevated values in field and laboratory blanks). Based upon average blank concentrations (for both laboratory and field blanks) the limit of quantification (LOQ) was calculated for all compounds with $LOQ = \text{average blank value plus 3 standard deviation (STDev) of the blank concentrations.}$

All samples $LOQ > x > LOD$ are considered with high uncertainties and reported $< LOQ$ (Table S7). All raw data are openly accessible from the NILU database (http://ebas.nilu.no) for thorough examinations. However, all original data are also listed in Appendix B of the Supplement (Table S8). All values below LOD were excluded from further statistical treatment (treated as not detected = n.d.).

Transport model calculations. To relate measured concentrations to transport from potential source regions, 3-hourly backward simulations from the Troll station were made with the Lagrangian particle dispersion model FLEXPART (Kallenborn et al., 2006; Stohl et al., 2006, 2007, 2010) for the entire measurement period (see http://zardoz.nilu.no/~andreas/BACKWARD_PRODUCTS/STATIONS/TROLL/). FLEXPART
was driven with 3-hourly operational meteorological data from the European Centre for Medium-Range Weather Forecasts with $1^\circ \times 1^\circ$ resolution. During every 3-hour interval, 60,000 particles were released at the measurement point and followed backward for 20 days to calculate emission sensitivity ($S$) on a $1^\circ \times 1^\circ$ grid, under the assumption that removal processes can be neglected. $S$ (in units of skg$^{-1}$) in a particular grid cell is proportional to the particle residence time in that cell and measures the simulated mixing ratio at the receptor that a source of unit strength (1 kg s$^{-1}$) in the cell would produce. The $S$ distribution in a 100 m layer adjacent to the surface (so-called footprint layer) was used to characterize the potential influence from Southern Hemisphere potential POP surface sources during the 20-days preceding the 3-hourly particle release period. For the purpose of this paper, model results were averaged for the periods when POP samples were taken.

3 Results and discussions

3.1 Hexachlorobenzene (HCB)

Hexachlorobenzene (HCB; CAS No. 118-74-1) was used as fungicide. Currently HCB is predominantly released as an industrial by-product from various chemical processes. HCB was identified as the most abundant target contaminant in this study.

3.2 Polychlorinated biphenyls (PCB)

PCB (CAS No. 1336-36-3) is a technical mixture consisting of 209 different congeners with 1 to 10 chlorine atoms attached to a biphenyl molecular structure. PCB has been used in a variety of technical applications (Breivik et al., 2002). Of the total of 32 PCB congeners analyzed, only 14 congeners were quantified in 7 samples between February 2008 and September 2007. Although PCBs were identified in all Troll air samples, due to elevated blank levels measured (Table S7), for most of the samples concentrations below the limit of quantification (LOQ) were determined. Therefore, no in-depth
pattern elucidation for PCBs was performed in this study. Based upon this information, possible local contamination is assumed. Therefore, the station will be moved to a more remote location for the next sampling season (C. R. Lunder, personal communication, 2013).

As a general feature, in six samples, PCB 47 was found to be the most abundant congener (< LOQ – between 0.09–1.54 pg m\(^{-3}\)). The congeners PCB 28, 33, 47 and 52 were found in all seven samples above LOQ. The seven samples with quantified PCB levels (> LOQ) are listed in Table 1.

The low PCB concentrations measured at Troll, relative to what has been reported for the Arctic (Breivik et al., 2002), could be expected as only an estimated 3% of the historical use of PCBs took place in the Southern Hemisphere according to a global inventory of global PCB production and use (Hung et al., 2010). For the countries of Chile and Argentina, the estimated historical use of total PCBs is 1270 and 7200 tons, respectively, which translates into 0.1% and 0.5% of the global historical production, although these numbers are affected by significant uncertainties (Breivik et al., 2002). For comparison, the compound specific limit of quantification (LOQ) values can be found in Appendix A of the Supplement (Table S7).

### 3.3 Chlorinated pesticides

A representative group of organochlorine pesticides (OCP) was selected as target analytes for the Troll monitoring program (12 compounds). All of the here selected OCP are target chemicals for the global monitoring plan of the UNEP (United Nations Environmental Programme) Stockholm convention for the global regulation of POPs. The \(\alpha\)- and \(\gamma\)-hexachlorocyclohexane (HCH), four chlordane-related cyclodiene pesticides (\(trans\)-, \(cis\)-chlordane and \(trans\)-, \(cis\)-nonachlor), as well as the six major constituents of the DDT group (\(o\),\(p\)-,\(p\)-DDT, DDE, DDD) were analyzed (see Table S6).
4 Discussions

Although considerable concentration variations were identified for the entire data set, several features were found characteristic for the POP distribution in the Troll atmospheric samples. HCB was the POP with the highest concentrations, followed by the sums of HCHs, PCBs, DDTs and chlordanes (HCB > Sum HCH > Sum PCB > Sum DDT > Sum chlordanes). For single compounds other than HCB, \( a\)-HCH and \( p,p'\)-DDE were the most prominent constituents.

HCB was identified as the major POP in the Antarctic air samples with concentration levels varying around 22 pg m\(^{-3}\) (average concentration) for the entire monitoring period. HCB was positively identified in all 196 samples quantified. No temporal trend could be established for HCB at Troll during the 4 yr measuring period. However, in contrast to the Arctic data from the Zeppelin station (Hung et al., 2010), a characteristic seasonal pattern was found for Troll air samples with highest levels during the austral winter period (week 25–45) and lower levels during the summer seasons (Fig. 2). This is likely related to seasonally varying atmospheric transport from Australia, Africa and South America, which is much more effective in winter than in summer (Gambaro et al., 2005; Kallenborn et al., 1998). However, increased OH-radical concentrations in combination with higher ambient temperatures may also lead to increased transformation and, thus, lower HCB air concentrations during the summer season. The HCB concentrations found at Troll station are generally 3–4 times lower than those observed at the Zeppelin station (Svalbard, Norwegian Arctic).

The most abundant OCP compound in Troll atmospheric samples was \( \alpha\)-HCH with concentrations varying between 0.02 and 0.46 pg m\(^{-3}\). The \( \alpha\)-HCH isomer was identified in all 196 samples analyzed. As also shown for the Arctic Zeppelin station (AMAP; 2009; Hung et al., 2010; Kallenborn and Berg, 2006; Ma et al., 2011), \( \alpha\)-HCH had higher concentrations in Troll air samples than the related \( \gamma\)-HCH isomer (\( \alpha\)/\( \gamma\)-HCH ratio: 2–24) in all Troll air samples. For the HCH isomers, \( \alpha\)-HCH was usually 3–4 times higher concentrated than the sister isomer, \( \gamma\)-HCH. Thus a ratio \( \alpha\)/\( \gamma\)–HCH in the
range of 3–5 is considered as an expected ratio indicating no unusually contaminated air masses. Variations between 2 and 5 were commonly observed in the Troll air samples. The chlordane compound cis-chlordane (cis-CD) was the most abundant of those compounds (identified in 162 samples, concentration levels: 0.02–0.20 pg m$^{-3}$) whereas for $p,p'$-DDE higher concentration levels were found (identified in 156 samples; concentration levels: 0.02–0.42 pg m$^{-3}$). Surprisingly, the presence of $p,p'$-DDT and $o,p'$-DDT was confirmed in about 30 % of all samples. In the technical DDT mixture the major constituent is $p,p'$-DDT and $o,p'$-DDT is present with 15–20 % dependent on the production pathways (Metcalf, 1995). The ratio between the major constituent $p,p'$-DDT and $o,p'$-DDT (the major byproduct) is about 3–5. Thus, a ratio between 1 and 3 ($p,p'$-DDT/$o,p'$-DDT) indicates fresh, not weathered technical DDT as potential source for the observed air contamination. However, since DDT is considered an important by-product during dicofol production, a high proportion of $o,p'$-DDT in air samples may also indicate dicofol application in the respective source regions. For the Troll air samples, in 35 samples $p,p'$-DDT/$o,p'$-DDT ratios between 1 and 6 could be established, confirming a contribution of fresh technical DDT/dicofol in the respective Troll air samples (mainly for 2007, see Fig. 3). In the period 2008–2010, $p,p'$-DDT and $o,p'$-DDT were detected only occasionally.

The long-range atmospheric transport potential among the individual compounds selected may be further discriminated by their potential to be scavenged by aerosols, rain and snow (Lei and Wania, 2004). To illustrate, we have estimated the percentage sorbed onto aerosols for selected POPs from 5°C to −50°C as detailed in Appendix C, which covers the observed temperature range (∼0°C to −40°C) previously reported at Troll (Hansen et al., 2009). To illustrate this, we have evaluated the estimated percentage sorbed onto aerosols for selected POPs from +5°C to −50°C based on previously reported data from Troll (∼ 0°C to −40°C), according to Hansen et al. (2009).
At −20 °C, HCB, HCHs and PCBs 28, 52 and 101 are predicted to dominate in the gaseous state among the substances included, which mirrors the overall results of the measurements. There are also some notable differences predicted within groups of substances in terms of gas/particle partitioning. At −15 °C, 87 % of p,p′-DDD, 60 % of p,p′-DDT and 29 % of p,p′-DDE is predicted to be sorbed onto aerosols which could help to explain why p,p′-DDE appears to be occasionally elevated at Troll. Thus, p,p′-DDE is associated more to the vapour phase, whereas the other isomers are expected to be deposited more rapidly by particle related deposition processes.

For other related substances which were predominantly present in the gas phase almost through the entire temperature range, such as HCHs, differences in rain or snow scavenging may be more relevant to elucidate potential differences in atmospheric deposition. While γ-HCH may be more easily washed out by rain than α-HCH (Lei and Wania, 2004), any potential differences in snow scavenging between these two isomers are more difficult to predict (Lei and Wania, 2004). For the one-week sampling period at Troll the measured air concentrations are often very low or close to detection limits. Therefore one may argue that any potential long-range atmospheric transport events are best identified by elevated air concentrations of more volatile substances, while samples showing occasionally elevated air concentrations of less volatile substances alone may rather indicate influence of local sources within the Antarctic.

For HCHs direct applications of technical HCH or emissions from technical HCH waste deposited in southern African or South American regions may also directly contribute to the observed elevated levels.

4.1 Atmospheric long-range transport

For the elucidation of potential contamination sources, also general pattern differences were compared. For the identification of different emission sources, the concentrations of a typical agrochemical (γ-HCH) were plotted against the concentrations of an industrial chemical (HCB), see Fig. 4. HCB is the most abundant contaminant in Troll air samples with levels between 10 and 80 pg m⁻³ whereas γ-HCH is found in concentra-
tions of around 0.1 to 0.5 pg m$^{-3}$. An unusually high concentration of at least one of the compounds indicates atmospheric transport of contaminated air mass to the Troll station. Relatively high concentrations of $\gamma$-HCH are found for samples taken during week 27/2009 and 33/2010 (Fig. 4).

Figure 5 shows the corresponding footprint emission sensitivity obtained from the FLEXPART 20-day backward simulation for the period when this air sample was taken. The emission sensitivity over Chile and Argentina for this sample (Fig. 5) is enhanced by nearly an order of magnitude compared to the seasonal-mean 20-day-backwards footprint emission sensitivity, indicating unusually strong transport from these potential source regions to the Troll station.

In Fig. 4, significantly elevated levels for the agrochemical $\gamma$-HCH were observed during week 27/2009. For this week, the FLEXPART calculations show that high emission sensitivity values extend much further north over Africa (lower part, Fig. 5) than for seasonally averaged transport (upper part, Fig. 5). The emission sensitivity over Southern Africa is nearly one order of magnitude higher than in the seasonal mean, indicating much stronger transport from Southern Africa to Troll than normally. Again, the increase in emission sensitivity is consistent with the measured $\gamma$-HCH concentration enhancement. According to a recent review by Vijgen et al. (2011), Lindane® ($>99\%$ $\gamma$-HCH) as well as the technical HCH mixture was produced in South Africa in significant amounts. It is also reported that about 70,000 tons of HCH waste isomers are still stored and/or deposited in this country.

Also several other samples were identified with elevated contribution of either HCB and/or $\gamma$-HCH (Fig. 4), and for several of those enhanced transport from continental source regions (mainly from South America) can be seen in the FLEXPART calculations. Thus, the atmospheric long-range transport of polluted air masses is considered as the main source for the legacy POPs monitored at Troll.
4.2 Comparison with historical data

Previously, only a few campaign-based data sets of atmospheric POP concentrations were collected with active high-volume sampling devices in Antarctica. An eleven weeks sampling campaign was performed in 1993 (October–December) at the Italian Research station Terra-Nova-Bay, now Zuchelli station (Victoria Land; Kallenborn et al., 1995). As a follow up of the 1993 sampling program and in cooperation with the British Antarctic Survey (BAS), NILU performed a first year-around measuring campaign for POPs in Antarctic air at the British research station in Signy Island (South Orkney Islands) in the period 1994–1995 (Kallenborn et al., 1998). During both campaigns similar priority compounds were quantified and analytical methods were used as studied in this paper. HCB and HCHs were also measured during a process study for air/water exchange of selected chlorinated pesticides in the Ross Sea region at Terra Nova Bay in 2003 (9 samples; Cincinelli et al., 2009). Finally, a campaign-based study on PCBs in Terra-Nova Bay air samples was recently published (Gambaro et al., 2005). The here-discussed results are summarized for comparison in Table 2. For almost all compounds analyzed, considerable concentration reductions are registered compared to the above mentioned previous atmospheric POP measurement campaigns. The only exception is HCB (Table 2) for which similar concentration levels were found as registered in the 1993 sampling campaign at Terra-Nova-Bay (Victoria Land; Kallenborn et al., 1995).

The average levels measured in Terra-Nova-Bay air from 2003 (Cincinelli et al., 2009) are based on nine weekly summer samples only (December to March 2003). A factor of 2 lower concentrations is, thus, considered as well within the range of variations compared to the 2010 whole-year sampling set from Troll, given the observed seasonal variation with a summer minimum (Fig. 2).

In the early studies (sampling 1993, 1995; Kallenborn et al., 1995, 1998) a slight predominance of \( \gamma \)-HCH was identified in atmospheric samples. However, in the data set presented here, \( \alpha \)-HCH was the most abundant HCH isomer analyzed in all samples. The most obvious reason for this pattern shift along with significant overall concentra-
tion reductions of about a factor 100 compared to the 1993 and 1995 campaigns (see Table 2) may be sought in the national ban of lindane (> 99 % γ-HCH) in Asian countries (i.e. China and India) around the year 2000. However, other potential reasons may contribute (incl. increased emissions from γ-HCH deposited in South Africa previously mentioned) and should thus be included in the current discussion on potential γ-HCH sources for Antarctic atmospheric samples.

In order to gain scientific understanding on global distribution processes for anthropogenic pollutants, the establishment of a global atmospheric monitoring network is considered an important tool both for regulators and environmental scientists (AMAP, 2009; Kallenborn and Berg, 2004; Kallenborn et al., 2012). Therefore regional based monitoring programs are encouraged to coordinate their monitoring priorities in order to establish global networks (UNEP, 2009). However, the first global effectiveness assessment of the POPs Stockholm conventions revealed that most of atmospheric POPs monitoring data (relevant for regulative purposes) is reported from the Northern Hemisphere. The long-term POPs monitoring networks in Southern Hemisphere are not yet on the same levels of sophistication as compared to/with networks in the northern regions (UNEP, 2009). The obvious differences in available monitoring infrastructure, cost for establishing and maintaining the infrastructure, expert support, national regulation programs and research funding priorities is considered as important reason for the still large lack of atmospheric POP monitoring data from the Southern Hemisphere and especially in Antarctica (compared to Arctic regions). Therefore, the here presented POP monitoring and research study on POP distribution in Antarctic air will hopefully contribute significantly to the ongoing coordinated efforts for the continuous surveillance of priority POPs in Antarctica, and thus, help to understand the complex processes leading to the currently observed global distribution of this type of anthropogenic pollution.
5 Conclusions

The first long-term monitoring program for atmospheric transported persistent organic pollutants is presented. The monitoring program has been established (2007) at the Norwegian Troll station (Dronning Maud Land). The concentrations found were considerably lower than the monitoring data measured in the Arctic illustrating the combination of remoteness and lack of potential local sources for Antarctic sampling sites. By combining modern modelling tools with pattern evaluation of available concentration data, several atmospheric long range transport events were identified and discussed here. The influence of the southern African continent for the transport of legacy organochlorine pesticides (i.e. HCHs) was confirmed in this study.

Supplementary material related to this article is available online at: http://www.atmos-chem-phys-discuss.net/13/6219/2013/acpd-13-6219-2013-supplement.pdf.

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Long-term monitoring of persistent organic pollutants (POPs)

R. Kallenborn et al


Long-term monitoring of persistent organic pollutants (POPs)

R. Kallenborn et al


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Table 1. PCB concentrations in Troll air samples (2007–2010). The concentration levels are given in [pg m\(^{-3}\)].

<table>
<thead>
<tr>
<th>Structure</th>
<th>PCB</th>
<th>Sampling week/year conc. [pg m(^{-3})]</th>
</tr>
</thead>
<tbody>
<tr>
<td>2,4,4’-TriCB</td>
<td>28</td>
<td>30/08 0.25 48/08 0.16 10/09 0.13 21/09 0.21 28/10 0.18 32/10 0.45 37/10 0.15</td>
</tr>
<tr>
<td>2,4’,5-TriCB</td>
<td>31</td>
<td>30/08 0.23 48/08 0.22 &lt; LOQ 21/09 0.21 28/10 0.19 32/10 0.48 &lt; LOQ</td>
</tr>
<tr>
<td>2’,3,4-TriCB</td>
<td>33</td>
<td>30/08 0.18 48/08 0.11 10/09 0.10 21/09 0.15 28/10 0.13 32/10 0.35 37/10 0.11</td>
</tr>
<tr>
<td>3,4,4’-TriCB</td>
<td>37</td>
<td>30/08 0.04 48/08 0.02 10/09 0.01 21/09 0.02 28/10 0.01 32/10 0.06 37/10 0.02</td>
</tr>
<tr>
<td>2,2’,4,4’-TetraCB</td>
<td>47</td>
<td>30/08 &lt; LOQ 48/08 0.14 10/09 0.29 21/09 0.80 28/10 0.56 32/10 1.54 37/10 0.47</td>
</tr>
<tr>
<td>2,2’,5,5’-TetraCB</td>
<td>52</td>
<td>30/08 0.18 48/08 0.17 10/09 0.15 21/09 0.22 28/10 0.19 32/10 0.45 37/10 0.17</td>
</tr>
<tr>
<td>2,3’,4,4’-TetraCB</td>
<td>66</td>
<td>30/08 0.04 48/08 0.03 &lt; LOQ 10/09 &lt; LOQ 21/09 &lt; LOQ 28/10 &lt; LOQ 32/10 &lt; LOQ 37/10 &lt; LOQ</td>
</tr>
<tr>
<td>2,4,4’,5-TetraCB</td>
<td>74</td>
<td>30/08 0.03 &lt; LOQ 48/08 &lt; LOQ 10/09 &lt; LOQ 21/09 &lt; LOQ 28/10 &lt; LOQ 32/10 &lt; LOQ</td>
</tr>
<tr>
<td>2,2’,4,4’,5-PentaCB</td>
<td>99</td>
<td>30/08 0.11 48/08 0.04 &lt; LOQ 10/09 &lt; LOQ 21/09 &lt; LOQ 28/10 &lt; LOQ</td>
</tr>
<tr>
<td>2,2’,4,5,5’-PentaCB</td>
<td>101</td>
<td>30/08 &lt; LOQ 48/08 &lt; LOQ 10/09 0.10 21/09 0.07 28/10 0.07 32/10 0.24 37/10 0.09</td>
</tr>
<tr>
<td>2,3,3’,4,4’-PentaCB</td>
<td>105</td>
<td>30/08 &lt; LOQ 48/08 &lt; LOQ 10/09 &lt; LOQ 21/09 &lt; LOQ 28/10 &lt; LOQ</td>
</tr>
<tr>
<td>2,3,4,4’,5-PentaCB</td>
<td>114</td>
<td>30/08 0.02 48/08 0.02 &lt; LOQ 10/09 &lt; LOQ 21/09 &lt; LOQ 28/10 &lt; LOQ</td>
</tr>
</tbody>
</table>

LOQ: Limit of Quantification
### Table 2. Atmospheric POP distribution in Antarctic air; a comparison (concentration range in parenthesis).

<table>
<thead>
<tr>
<th>Compounds</th>
<th>Terra Nova Bay; 1993 (average conc.)(^{16})</th>
<th>Signy Island; 1994–1995 (average conc.)(^{18})</th>
<th>Terra-Nova-Bay; 2005 (average conc.)(^{17})</th>
<th>Troll; 2010 (this study)</th>
</tr>
</thead>
<tbody>
<tr>
<td>HCB</td>
<td>21 (n.d.–28)</td>
<td>–</td>
<td>11.4 (6–20)(^2)</td>
<td>22.9</td>
</tr>
<tr>
<td>Sum α- and γ-HCH</td>
<td>13 (5–20)</td>
<td>27 (3–33)</td>
<td>0.8 (0.3–1.2)(^2)</td>
<td>0.2</td>
</tr>
<tr>
<td>Sum 4 CD (^a)</td>
<td>3.5 (n.d.–6)</td>
<td>0.4 (.004–0.9)</td>
<td>–</td>
<td>0.1</td>
</tr>
<tr>
<td>Sum 6 DDT (^b)</td>
<td>–</td>
<td>0.2 (0.07–0.4)</td>
<td>–</td>
<td>0.03</td>
</tr>
<tr>
<td>Sum PCB (^c)</td>
<td>9 (n.d.–15)</td>
<td>7 (0.02–17)</td>
<td>0.7 (0.3–1.3)(^d)</td>
<td>0.5</td>
</tr>
</tbody>
</table>

\(^{a}\) Sum trans/cis – chlordane and nonachlor;

\(^{b}\) Sum DDT = \(p,p’/o,o’\) – DDT, DDD, DDE,

\(^{c}\) SUM PCB, 18, 28/31, 52, 99, 101, 118, 149, 138, 153, 180; n.d. = not detected;

\(^{d}\) Only gas phase compounds quantified.
Fig. 1. The position of the Antarctic Troll Atmospheric observatory (Norwegian Troll research Station, dronning Maud Land, Antarctica). For comparison the British research stations Signy Island and the Italian Terra Nova Bay stations are shown. Map: courtesy of Wikipedia.
Fig. 2. Concentration distribution of hexachlorobenzene (HCB) in atmospheric samples from Troll (Antarctica), period: 2007–2010, x-axis: sample weeks, y-axis: conc. [pgm$^{-3}$]
**Figure 3.** $p,p'-/o,p$-DDT ratios (y-axis) for weekly atmospheric samples (week No.) from the Troll station (Antarctica); y-axis = conc./conc.
**Fig. 4.** Concentration comparison of HCB (y-axis) and \( \gamma \)-HCH (x-axis) in air samples from Troll (Antarctica: 2007–2010).
Fig. 4: Concentration comparison of HCB (y-axis) and γ-HCH (x-axis) in air samples from Troll (Antarctica: 2007–2010).

Fig. 5: Footprint emission sensitivities calculated with FLEXPART. Upper line: average monthly simulations (June and August) for the entire sampling period (2007–2010) as examples for the transport climatology for the respective months. Lower line: 20-day backward simulations for two weekly POP samples taken at Troll station during weeks 27/2009 (left) and 33/2010 (right).