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Mercury oxidation from bromine chemistry in the free troposphere over the southeastern US

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Abstract

The elevated deposition of atmospheric mercury over the Southeastern United States is currently not well understood. Here we measure partial columns and vertical profiles of bromine monoxide (BrO) radicals, a key component of mercury oxidation chemistry, to better understand the processes and altitudes at which mercury is being oxidized in the atmosphere. We use the data from a ground-based MAX-DOAS instrument located at a coastal site ~ 1 km from the Gulf of Mexico in Gulf Breeze, FL, where we had previously detected tropospheric BrO (Coburn et al., 2011). Our profile retrieval assimilates information about stratospheric BrO from the WACCM chemical transport model, and uses only measurements at moderately low solar zenith angles (SZA) to estimate the BrO slant column density contained in the reference spectrum (SCD_{Ref}). The approach has 2.6 degrees of freedom, and avoids spectroscopic complications that arise at high SZA; knowledge about SCD_{Ref} helps to maximize sensitivity in the free troposphere (FT). A cloud-free case study day with low aerosol load (9 April 2010) provided optimal conditions for distinguishing marine boundary layer (MBL: 0–1 km) and free tropospheric (FT: 1–15 km) BrO from the ground. The average daytime tropospheric BrO vertical column density (VCD) of $\sim 2.3 \times 10^{13}$ molec cm^{-2} (SZA $< 70^\circ$) is consistent with our earlier reports on other days. The vertical profile locates essentially all tropospheric BrO above 4 km, and shows no evidence for BrO inside the MBL (detection limit < 0.5 pptv). BrO increases in the FT. The average FT-BrO mixing ratio was ~ 0.9 pptv between 1–15 km, consistent with recent aircraft observations. We find that the oxidation of gaseous elemental mercury (GEM) by bromine radicals to form gaseous oxidized mercury (GOM) is the dominant pathway for GEM oxidation throughout the troposphere above Gulf Breeze. The column integral oxidation rates range from 3.0 – 3.4×10^5 molec $cm^{-2} s^{-1}$ for bromine, while contributions from ozone (O_3) and chlorine (Cl) were 0.9×10^5 and 0.2×10^5 molec $cm^{-2} s^{-1}$, respectively. The GOM formation rate is sensitive to recently proposed atmospheric scavenging reactions of the HgBr adduct by nitrogen dioxide (NO_2), and to a lesser extent also HO_2 radicals. Using a

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3-D chemical transport model, we find that surface GOM variations are typical also of other days, and are mainly derived from the free troposphere. Bromine chemistry is active in the FT over Gulf Breeze, where it forms water-soluble GOM that is subsequently available for wet scavenging by thunderstorms or transport to the boundary layer.

1 Introduction

Measurements of tropospheric halogen species are an area of increasing research interest due to the ability of halogens to destroy tropospheric ozone (O_3) (Read et al., 2008; Saiz-Lopez et al., 2012), oxidize atmospheric mercury (Holmes et al., 2009; Hynes et al., 2009), and modify oxidative capacity (Parella et al., 2012). Most assessments of the impacts of halogen chemistry are based on measurements of halogen oxides (bromine monoxide (BrO) and iodine monoxide (IO)), since these radicals are typically found at higher concentrations throughout the troposphere than the corresponding halogen atom radicals. Many of these studies take place in the planetary boundary layer (PBL) given that this region of the atmosphere is easily accessible from measurements located at the surface, and is also the most directly impacted by anthropogenic activities. However, halogen chemistry in the free troposphere (FT), albeit more challenging to measure, has the potential to affect an even larger air volume and mass. In particular, the colder temperatures of the free troposphere accelerate the bromine oxidation of gaseous elemental mercury (GEM) (Donohoue et al., 2006). Satellite-borne measurements represent a powerful resource for assessing global distributions and tropospheric vertical column densities (VCD) of BrO (GOME – van Roozendael et al., 2002; GOME-2 – Theys et al., 2011; Sihler et al., 2012). However, satellite retrievals rely on assumptions made about the vertical distribution of BrO, and uncertainties in these assumptions can lead to over/under predictions in the derived tropospheric VCD. The most direct method for measuring trace gas vertical distributions is through the use of aircraft (Prados-Ramon et al., 2011; Volkamer et al., 2015), or balloons (Fitzenberger et al., 2000; Dorf et al., 2006; Pundt et al., 2002). However, this type of measurement

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is costly and potentially impractical if the goal is to establish long term trends in the FT. Ground-based measurements are typically more straightforward to deploy and maintain for extended periods of time, but optimizing ground-based capabilities to observe the FT remains an area of active research (Schofield et al., 2006; Theys et al., 2007; Hendrick et al., 2007; Coburn et al., 2011). Specifically, ground-based Multi-Axis Differential Optical Absorption Spectroscopy (MAX-DOAS) measurements are uniquely suited for this type of study since this technique also assesses vertical distributions, and derived VCDs can be directly compared with models and satellites. Additionally, the DOAS retrieval allows for the detection of not only BrO, but also other trace gases that have significant impacts on the chemical cycling of bromine species in the atmosphere, such as NO₂ and some volatile organic compounds (VOCs). However, measurements of FT BrO from ground-based MAX-DOAS are not straightforward for several reasons: (1) stratospheric BrO represents a large portion of the measured signal and creates a background that has to be accounted for when attempting to assess the FT, (2) ozone absorption structures are strongly present in the same wavelength region as BrO and can create interferences due to stratospheric ozone absorption, in particular at high solar zenith angles (SZA) (Aliwell et al., 2002; van Roozendael et al., 2002), and (3) the sensitivity of this technique peaks at the instrument altitude and decreases with increasing altitude. Recent advances with testing stratospheric BrO profiles in atmospheric models (Liang et al., 2014) provide opportunities to properly account for #1 by assimilating information from atmospheric models. Further, retrievals that avoid SZA larger 70° do not suffer from #2. Moreover, certain measurement geometries retain information about the FT. Figure S1 (Supplement) depicts the box Air Mass Factors (bAMFs), which represent the sensitivity of the slant column density (SCD) measurement geometry to BrO concentrations at different altitudes, for two pointing directions (or elevation angles, EA = 25 and 90° upwards) at several SZA; at SZA < 70° the sensitivity of these EAs peaks between 2–15 km. A more comprehensive view of the bAMFs for different EAs over a wider SZA range is shown in Fig. S2 (Supplement).

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Van Roozendael et al. (2002) compared ground-based and balloon-borne measurements to VCDs of BrO from the space-borne Global Ozone Monitoring Experiment (GOME) and found all platforms were consistent with a rather widespread tropospheric BrO VCD of 1–3 × 10¹³ molec cm⁻² once appropriate radiative transfer effects were taken into consideration. Salawitch et al. (2005) and Theys et al. (2011) also report satellite derived tropospheric BrO VCDs (GOME and GOME-2, respectively) for the mid-latitudes of 2 × 10¹³ and 1–3 × 10¹³ molec cm⁻², respectively. Ground based measurements (Theys et al., 2007; Coburn et al., 2011) in the mid-latitudes have reported BrO VCDs of 1–2 × 10¹³ molec cm⁻² that are comparable to the findings from satellites. Volkamer et al. (2015) recently reported 1.6 × 10¹³ molec cm⁻² BrO VCD in the tropics measured by limb-observations from aircraft. All of these studies point to the widespread presence of BrO in the FT, corresponding to a VCD of 1–3 × 10¹³ molec cm⁻². Based on these reports, tropospheric BrO could account for 20–30 % of a total BrO VCD ~ 5–6 × 10¹³ molec cm⁻² as seen from satellite (van Roozendael et al., 2002; Theys et al., 2011), and significantly impact the lifetime of tropospheric O₃ and atmospheric GEM (Wang et al., 2015).

1.1 Atmospheric Hg in the southeastern US

Mercury in the atmosphere exists in three forms: gaseous elemental mercury (Hg⁰, GEM), gaseous oxidized mercury in the form of either Hg²⁺ or Hg¹⁺ (GOM), and particle-bound mercury (PBM). Understanding the processes that cycle mercury between its various forms (GEM ↔ GOM ↔ PBM) is of great importance because this speciation controls the deposition of mercury to the environment, i.e. GOM and PBM are more readily removed from the atmosphere via wet and dry deposition than GEM. (Lindberg and Stratton, 1998; Bullock, 2000). Once deposited, biological processes can methylate Hg²⁺ to form the neurotoxin methyl mercury, which bioaccumulates in fish. Enhancement factors for methyl mercury of up to 10⁶ relative to water have been measured in predatory fish tissues (Schroeder and Munthe, 1998; Selin et al., 2010).

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A better understanding of the processes controlling atmospheric mercury oxidation, and therewith removal, is particularly relevant for regions that experience high levels of mercury deposition, such as the southeastern United States (SE US). Figure 1 shows a map of the Total Mercury Wet Deposition in the US from 2013 (http://nadp.sws.uiuc.edu/maplib/pdf/mdn/hg_dep_2013.pdf). The high deposition levels experienced in the SE US cannot be explained by regional anthropogenic sources of mercury alone, which are mainly located within the Ohio River Valley where the prevailing winds carry emissions northeast. This indicates that a regional emission-deposition pattern is most likely not the major source–receptor relationship for mercury entering the environment over Florida, in the SE US, meaning that other possibilities, such as enhanced atmospheric oxidation followed by deposition, need to be explored.

2 Experimental

2.1 Atmospheric conditions

A case study during 9 April 2010 provided optimal conditions for assessing the ability of a ground-based MAX-DOAS instrument to measure FT trace gases (see Supplement for a brief overview of instrumentation and measurement site). Figure 2 shows a time series of trace gas differential slant column densities (dSCD) of BrO, IO, NO₂, and O₄ for the week surrounding the case study day; with 9 April highlighted by the blue box. The IO measurements are assimilated and used in the modeling portion of this study (Sect. 3.3), while the NO₂ measurements give an indication on the amount of influence from anthropogenic activities in the lowest layers of the BL. This day provides an excellent case study for two reasons: (1) consistent shape of the O₄ dSCDs across elevation angles as well as the clear splitting between the values is a good indicator for a cloud-free day, and (2) the relatively high O₄ dSCDs values (compared with other days) indicates a low aerosol load, enabling the instrument to realize longer light paths (increased sensitivity due to fewer extinction events), and an unobstructed view of the

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FT. An inspection of webcam pictures for the instrument site proved the day to be free of visual clouds, and precursory look at the aerosol load confirmed the low values. Figure 2 also contains in-situ O₃ measurements (from both the US EPA site and a nearby Mercury Deposition Network (MDN) site) as well as wind direction measurements from a WeatherFlow, Inc. monitoring station located in Gulf Breeze, FL near the US EPA site.

Oxidized mercury measurements ($\text{Hg}^{\text{II}} = \text{GOM} + \text{PBM}$) at the Pensacola MDN site during study period (Edgerton et al., 2006) are also shown in the bottom panel of Fig. 2 (also Fig. 9, Sect. 3.3). On 9 April, Hg^{II} concentrations at the MDN site were rising from near zero on 8 April (due to rain) to peak values of 15–40 pgm^{-3} on the following days, which is above average for the season. In prior years, average daily peak concentrations at this site in spring were 15 pgm^{-3} , which is higher than during any other season (Weiss-Penzias et al., 2011; Nair et al., 2012). Observed GEM concentrations are persistently around 1.4 ngm^{-3} throughout early April, as expected for this season, and therefore not shown.

2.2 External model overview

The Whole Atmosphere Community Climate Model version 4 (WACCM4) (Garcia et al., 2007; Marsh et al., 2013) has been extensively evaluated for its representation of the stratosphere, including stratospheric BrO (SPARC CCMVal, 2010). The model does not represent tropospheric bromine sources from very short lived species (VLSL, bromocarbons); this is an active choice to assure a-priori information about tropospheric BrO represents a lower limit (see Sect. 2.4.1). However, CHBr₃ and CH₂Br₂ concentrations are fixed at the cold point and add about 5–6 pptv stratospheric Br_y (stratospheric Br_y loading is 21–22 pptv). In this work, WACCM is run with specified (external) meteorological fields. This is achieved by relaxing the horizontal winds and temperatures to reanalysis fields. The reanalysis fields used are taken from the NASA Global Modeling and Assimilation Office (GMAO) Modern-Era Retrospective Analysis for Research and Applications (MERRA) (Rienecker et al., 2011). The horizontal resolution is 1.9° × 2.5°

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individual molecules (colored lines) are shown along with the total removal rate (black lines). Panel (c) shows the vertical profile of the ratio of the “revised” total rate to “traditional” total rate, which demonstrates the enhanced oxidation of HgBr when considering the additional scavenging reactions. In the “traditional” model the percent contributions to the column integrated rate of oxidation of HgBr are 71.3 and 28.7% for OH and Br, respectively; and in the “revised” model the percent contributions are as follows: 86.1% (NO₂); 10.6% (HO₂); 1.8% (BrO); 0.9% (OH); 0.3% (Br); 0.2% (IO); and < 0.1% (I). Note that mercury oxidation is initiated by reaction between Br radicals and GEM in both reaction schemes, and the additional scavenging reactions in the “revised” scheme primarily increase the overall rate of oxidation at altitudes where HgBr decomposition is fast. The greatest enhancement is seen below 8 km, where the overall rate of oxidation is ~ 20 times faster, primarily because of the reaction of HgBr with NO₂ and HO₂.

Figure 8 also illustrates the increased number of species produced from the additional oxidation mechanisms, some of which may have physical and chemical properties that differ from the two products of the “traditional” mode, which are also products in the “revised” scenario but are present at much lower concentrations. In the “traditional” scenario at 1 km, the scavenging products HgBrOH and HgBr₂ account for 96 and 4% of the total HgBrX, respectively, and these values drop to 0.2% and << 0.1% in the “revised” scenario where HgBrNO₂ accounts for 95% of HgBrX. In the “revised” scenario, HgBrNO₂ remains the major product throughout the atmosphere, but at free tropospheric altitudes HgBrHO₂ also contributes significantly at 35%, compared to HgBrNO₂ at 59%. There are currently no observations of the molecular composition of GOM with which to evaluate these simulated product distributions.

Mercury lifetime with respect to oxidation: the above oxidation rates correspond to a minimum lifetime of GEM with respect to oxidation by bromine radicals of ~ 40 days in the FT (based on the MAX-DOAS measurements). The total tropospheric column average lifetimes are 62, 88, and 63 days for the scenarios including BrO profiles from the MAX-DOAS measurements, WACCM, and GEOS-Chem, respectively, where in

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each case the contribution from the reactions with O₃ and Cl are the same and the differences are owing to the differing amounts of Br radicals. This is much shorter than the currently expected atmospheric lifetime on the order of several months. However, the box model only accounts for partitioning of the GOM species between the gas phase and aerosols; once they are in the aqueous phase they can be photo-reduced to GEM that can then subsequently return to the gas phase (Costa and Liss, 1999), thus extending the effective lifetime significantly beyond that calculated above. The kinetic coefficients for reactions involving HgBr come mainly from quantum chemical calculations, which have significant uncertainties, so the simulated GEM lifetime could be extended by reducing the rate coefficients within their uncertainties. Another possible mechanism is the photo-dissociation of HgBrX products containing species that have significant absorption cross-sections in the ultra-violet/visible (UV/Vis) region of the electromagnetic radiation spectrum, e.g. HgBrNO₂, HgBrHO₂, which could reproduce HgBr. This HgBr could then thermally decompose to re-form GEM, or be oxidized again. Our observations are consistent with previous findings (Wang et al., 2015), that establish GEM as a chemically highly dynamic component of the FT. It is expected that GOM species will go through additional processes also in the aqueous phase, which could significantly impact the ultimate fate of the mercury.

Atmospheric implications: the rapid oxidation of mercury in the lower FT is of potential relevance in the SE US, where there have been several studies linking deep convective activity to the elevated levels of mercury found in rainwater (Guentzel et al., 2001; Landing et al., 2010; Nair et al., 2013). The bio-accumulation of methyl mercury in fish tissues is particularly relevant in this region, where it has been deemed unsafe to eat fish harvested from many lakes in the region (Engle et al., 2008; Liu et al., 2008). Wet deposition measurements of mercury exceed what can be explained through regional sources in the southeast. In fact, Guentzel et al. (2001) estimated that < 46% of the mercury deposited in Florida was a result of local emissions, the other > 50% was attributed to long-range transport of mercury in the atmosphere; the transported fraction may have increased since those data were collected in the 1990s because regional

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US mercury emissions have declined while global emissions have risen. This attribution, coupled with mountaintop and aircraft studies locating elevated levels of GOM in the free troposphere (Swartzendruber et al., 2006; Fain et al., 2009; Swartzendruber et al., 2009; Lyman and Jaffe, 2012; Brooks et al., 2014; Weiss-Penzias et al., 2015),
5 strongly suggests the presence of a global “pool” of mercury in the upper atmosphere that contributes to widespread mercury deposition on a local to regional scale.

On the case study day, Hg^{II} (= GOM + PBM) concentrations at the nearby Pensacola MDN site reached 25 pgm^{-3} around midday and nearly 40 pgm^{-3} on 10 April (Fig. 9). High Hg^{II} events in Pensacola are frequently consistent with emissions from a nearby
10 coal-fired power plant, however about 25 % of such events have significant contributions from the free troposphere (Weiss-Penzias et al., 2011). The high $\text{Hg}^{\text{II}} : \text{SO}_2$ ratios ($10\text{--}20 \text{ pgm}^{-3} \text{ ppb}^{-1}$) and high $\text{NO}_y : \text{SO}_2$ ratios ($3.5\text{--}5.2 \text{ ppb ppb}^{-1}$) recorded on 9–10 April make the power plant an unlikely source of the Hg^{II} on these days (Fig. S8); for comparison, Weiss-Penzias et al. (2011) reported $\text{Hg}^{\text{II}} : \text{SO}_2$ and $\text{NO}_y : \text{SO}_2$ ratios
15 of 3.5 and 1.0, respectively, in power plant plumes. Like Hg^{II} concentrations, ozone concentrations and Hg^{II} dry deposition in Florida also peak in spring, all of which are consistent with a significant source of surface Hg^{II} being from the upper troposphere (Lyman et al., 2009; Sexaur Gustin et al., 2012). We use the GEOS-Chem model to further probe the sources of this Hg^{II} . Figure 9 shows that the model generally reproduces
20 the day-to-day variability of Hg^{II} observations during April 2010 (but not extremes), such as the low concentrations during 7–8 April and relative maxima on 5 and 10 April. In both the model and observations, Hg^{II} concentrations rise abruptly in the morning, consistent with entrainment of Hg^{II} aloft, and unlike the other combustion tracers (e.g. NO_y , SO_2) that reach high concentrations at night (Fig. S8 in the Supplement). The
25 model does not reproduce the abrupt drop in Hg^{II} around 18:00 CST, however, which may be partly due to local sea breeze circulations that are evident in wind observations but not simulated at the $4^\circ \times 5^\circ$ model resolution. In addition, mixing depth errors in the driving meteorology are known to affect other species, particularly at night (Lin et al.,

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2010; McGrath-Spangler and Molod, 2014). To assess the contribution of the free troposphere to surface Hg^{II} during April 2010, we conduct an additional model simulation with zero anthropogenic emissions in North America and no mercury redox chemistry in the lower troposphere ($> 700 \text{ hPa}$). The simulation is initialized from the base run
5 on 1 April. Due to its fast deposition, all boundary-layer Hg^{II} in the sensitivity simulation after about 1 day originates from oxidation of GEM in the free troposphere or, less likely, from intercontinental transport of anthropogenic Hg^{II} . Figure 9 shows there is little difference between the base and sensitivity simulations, meaning that the FT is the main source of boundary-layer Hg^{II} in the model and that variability in the FT
10 component explains most of the day-to-day Hg^{II} variability. Thus, the 3-D model shows that conditions on 9 April are favorable for Hg^{II} transport to the boundary layer, exactly when the DOAS observations find substantial amounts of BrO are present in the free troposphere.

The GEM oxidation mechanism in the 3-D model corresponds to the “traditional”
15 scheme in the box model and bromine concentrations in GEOS-Chem are lower than recent aircraft observations (Volkamer et al., 2015), so the GEM oxidation may be faster than simulated. Greater Hg^{II} production in the free troposphere might help correct the model’s 20 % low bias in mercury wet deposition over the Southeast US (Zhang et al., 2012), but this would depend on the rate of any compensating reduction reaction, as
20 discussed above. The findings of this study indicate that the amount of bromine located in the FT is sufficient to quickly oxidize GEM.

4 Conclusions

We show the benefits of determining SCD_{Ref} to maximize the sensitivity of ground based MAX-DOAS measurements to detect BrO in the FT, and improve the overall
25 consistency of time-resolved BrO tropospheric VCDs. The retrieval can also be applied to other trace gases. Knowledge of SCD_{Ref} allows retrievals at higher altitudes,

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i.e., the derivation of one vertical profile for each MAX-DOAS scan throughout the day (SZA < 70°) and the assessment of the diurnal variation of the partial BrO vertical columns. Our retrieval is complementary to previous studies that have characterized the stratosphere using zenith-sky measurements under twilight conditions (Theys et al., 2007; Hendrick et al., 2007), and minimizes the influence of O₃ absorption, and the contribution of stratospheric BrO to the overall BrO signal by using CTM output to constrain stratospheric BrO in combination with BrO dSCDs measured at low SZA instead. The FT VCDs reported here are in good agreement with the previously cited values for BrO, with the average BrO FT VCD ($\sim 2.3 \times 10^{13}$ molec cm⁻²) falling within the range reported by other studies ($1\text{--}3 \times 10^{13}$ molec cm⁻²) (see Wang et al., 2015, and references therein). These measurements all point to the presence of background amounts of BrO in the FT that is larger than current models predict.

The presented box model studies indicate that bromine radicals are the dominant oxidant for atmospheric GEM throughout the FT above the studied region. Our results confirm that mercury is rapidly oxidized by bromine, and a chemically highly dynamic species in the atmosphere. The chemical lifetime of GEM is ~ 40 days in the tropical FT based on calculations presented in this study; longer GEM global lifetimes should thus be regarded to indicate “effective lifetimes”, i.e., are the result of rapid chemical cycling of GOM back to GEM. Mercury measurements during our study period show high surface Hg^{II} concentrations that likely originate in the FT, meaning that we have observed substantial BrO columns under conditions favorable for Hg^{II} transport to the boundary layer. Additionally, this study suggests that the experimental observation of elevated GOM in the FT may be linked to our incomplete understanding about tropospheric bromine sources (Swartzendruber et al., 2006; Faïn et al., 2009; Lyman and Jaffe, 2012; Wang et al., 2015; Shah et al., 2015). The findings of this study indicate that the amount of bromine located in the FT above the (coastal regions) of the SE US is sufficient to quickly oxidize GEM to GOM, which in turn can be wet deposited, and as such can help explain the observed elevated wet deposition pattern in this region. Our results highlight the need to understand BrO vertical profiles in the FT, and represent

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them in atmospheric models to understand the location where mercury is oxidized in the atmosphere, and available for wet and dry deposition. More studies are needed to test and represent the bromine sources in atmospheric models, test atmospheric GOM abundances by field data, clarify the chemical identity of GOM, its global distributions, and dry and wet removal.

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Table 2. Summary of mercury reactions and rate coefficients used in box-model.

Reaction	Rate or equilibrium ¹ Coefficient ²	Reference
$\text{Hg}^0 + \text{O}_3 \rightarrow \text{HgO} + \text{O}_2$	3×10^{-20}	Hall (1995)
$\text{HgO}_{(g)} \leftrightarrow \text{HgO}_{(aq)}$	K_{eq}^1	Rutter and Schauer (2007)
$\text{HgO}_{(aq)} \rightarrow \text{Hg}_{(g)}^0$	1.12×10^{-5}	Costa and Liss (1999)
$\text{Hg}^0 + \text{Cl} \xrightarrow{M, Br} \text{HgClBr}^3$	$2.2 \times 10^{-32} \times \exp\left(680 \times \left(\frac{1}{T} - \frac{1}{298}\right)\right) \times [M]$	Donohue (2005)
$\text{HgClY}_{(g)} \leftrightarrow \text{HgClY}_{(aq)}$	K_{eq}^1	Rutter and Schauer (2007)
$\text{HgClY}_{(aq)} \rightarrow \text{Hg}_{(g)}^0$	1.12×10^{-5}	Costa and Liss (1999)
$\text{Hg}^0 + \text{Br} \xrightarrow{M} \text{HgBr}$	$1.46 \times 10^{-32} \times \left(\frac{T}{298}\right)^{-1.86} \times [M]$	Donohue (2006)
$\text{HgBr} + M \rightarrow \text{Hg}^0 + \text{Br} + M$	$4.0 \times 10^9 \times \exp\left(\frac{-7292}{T}\right)$	Goodsite et al. (2012)
$\text{HgBr}_{(g)} \leftrightarrow \text{HgBr}_{(aq)}$	K_{eq}^1	Rutter and Schauer (2007)
$\text{HgBr} + Y^4 \rightarrow \text{HgBrY}$	$2.5 \times 10^{-10} \times \left(\frac{T}{298}\right)^{-0.57}$	Goodsite et al. (2004)
$\quad \rightarrow \text{Hg}^0 + \text{Br}_2$	3.9×10^{-11}	Balabanov et al. (2005)
$\text{HgBrY}_{(g)} \leftrightarrow \text{HgBrY}_{(aq)}$	K_{eq}^1	Rutter and Schauer (2007)
$\text{HgBrY}_{(aq)} \rightarrow \text{Hg}_{(g)}^0$	1.12×10^{-5}	Costa and Liss (1999)
$\text{HgBr} + Y'^5 \rightarrow \text{HgBrY}'$	1×10^{-10}	Dibble et al. (2012)
$\text{HgBrY}'_{(g)} \leftrightarrow \text{HgBrY}'_{(aq)}$	K_{eq}^1	Rutter and Schauer (2007)
$\text{HgBrY}'_{(aq)} \rightarrow \text{Hg}_{(g)}^0$	1.12×10^{-5}	Costa and Liss (1999)

¹ Equilibrium coefficient is parameterized according to Rutter and Schauer (2007): $K_{eq} = (SA - PM)/10^{((-4250/T)+10)}$, where SA = the specific aerosol surface area, and PM = the particulate mass.

² Rate coefficients are given in either $\text{cm}^3 \text{molec}^{-1} \text{s}^{-1}$ or s^{-1} .

³ Assumes that the reaction between Hg^0 and Cl is the rate limiting step to form HgCl which will then quickly react with Br to form HgClBr .

⁴ Y = Br, OH.

⁵ Y' = HO_2 , NO_2 , BrO, IO, I.

28351

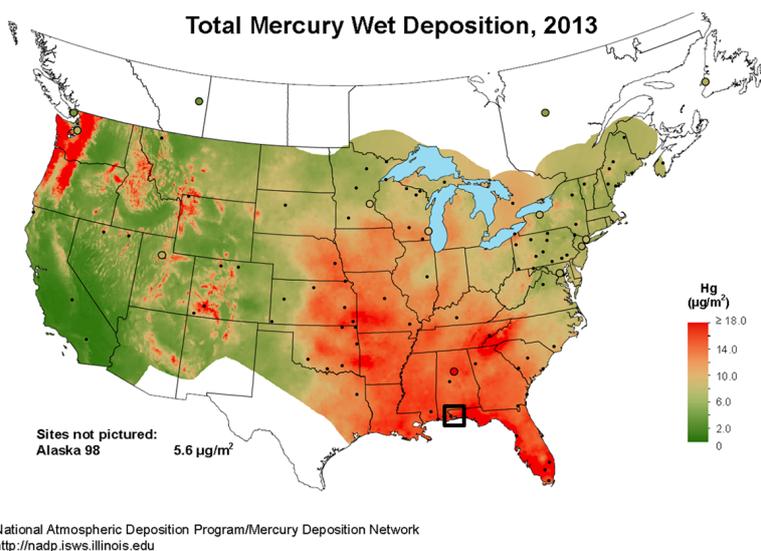


Figure 1. Total mercury wet deposition in the US for 2013. The highest levels of Hg deposition are observed in the southeastern US, where no local and regional sources are located immediately upwind. The black square indicates the measurement location.

28352

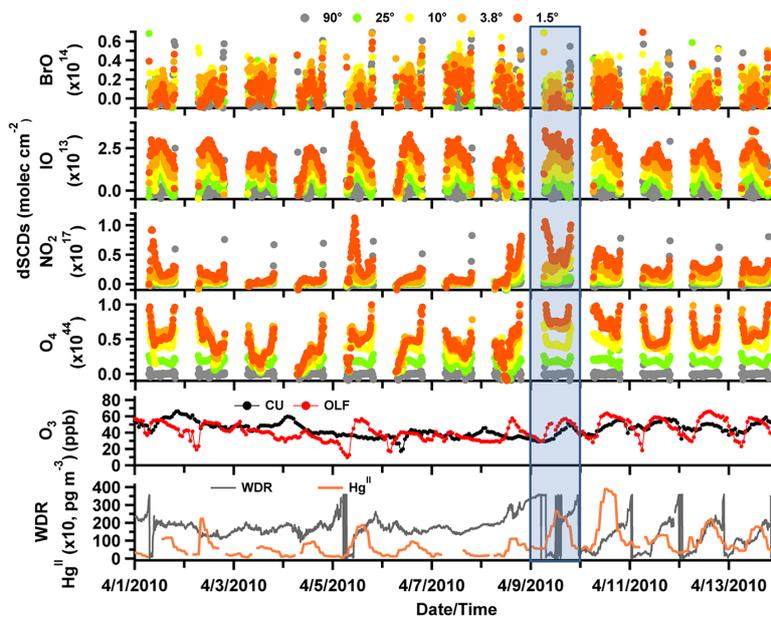


Figure 2. Overview of the MAX-DOAS measurements for the week surrounding the case study day (9 April, highlighted with blue box). Also included are O_3 measurements from a monitor collocated with the MAX-DOAS instrument (label “CU”, black trace) and a monitor located at an Mercury Deposition Network (MDN) site ~ 30 km northwest of the EPA site (label “OLF”, red trace), and wind direction measurements (grey trace) from a site near the EPA facility along with Hg^{II} measurements from the MDN site (scaled by a factor of 10, orange trace).

28353

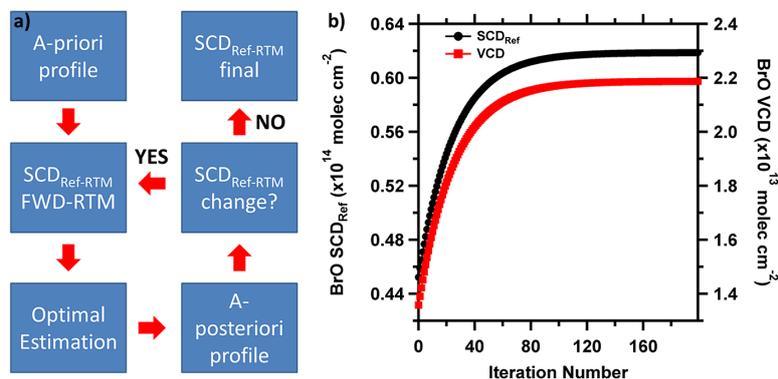


Figure 3. Conceptual sketch of the SCD_{Ref} retrieval (a), and resulting SCD_{Ref} values (black trace). The sensitivity of the a-posteriori BrO VCD (red trace) to SCD_{Ref} is also shown (b).

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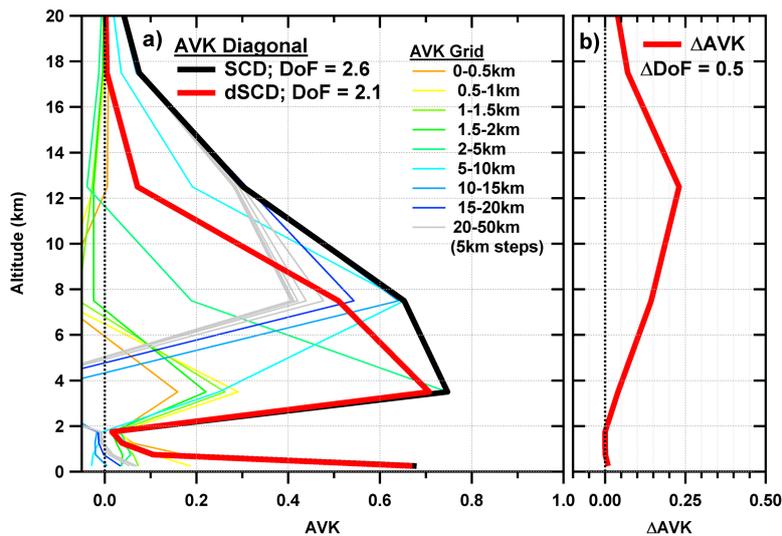


Figure 4. Characterizing the retrieval averaging kernel (AVK) and Degrees of Freedom (DoF). **(a)** AVK output for an inversion of BrO SCDs (at SZA = 25°) as a function of altitude. The colored traces represent the individual altitude grids of the inversion accounting for SCD_{Ref}; the thick black trace is the diagonal of the AVK matrix; the thick red trace is the diagonal of the AVK matrix from the inversion not accounting for SCD_{Ref}. **(b)** shows the difference between the two AVK diagonals as a function of altitude.

28355

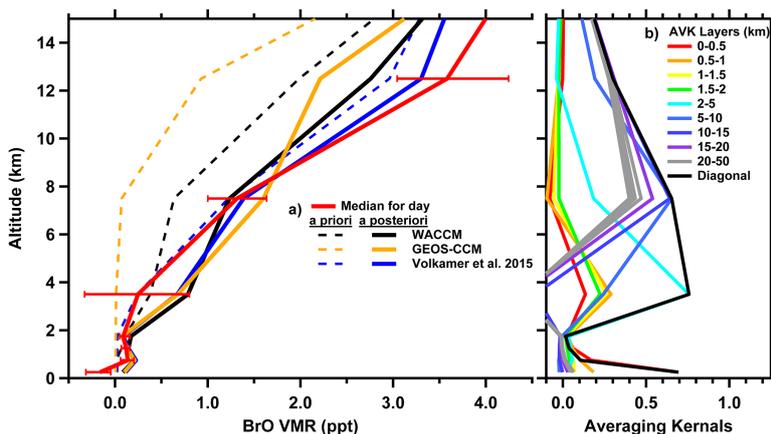


Figure 5. BrO profiles for a single MAX-DOAS measurement scan at SZA ~ 23° around solar noon. **(a)** contains three of the a-priori profiles tested (dashed lines) and the respective a-posteriori profiles (solid lines); colors correspond to use of WACCM (black); GEOS-CCM (orange); and TORERO RF12 (blue) as a-priori information. **(b)** contains the AVK from the inversion using the WACCM profile as the a-priori, and gives an indication on the amount of information coming from the measurements as a function of altitude.

28356

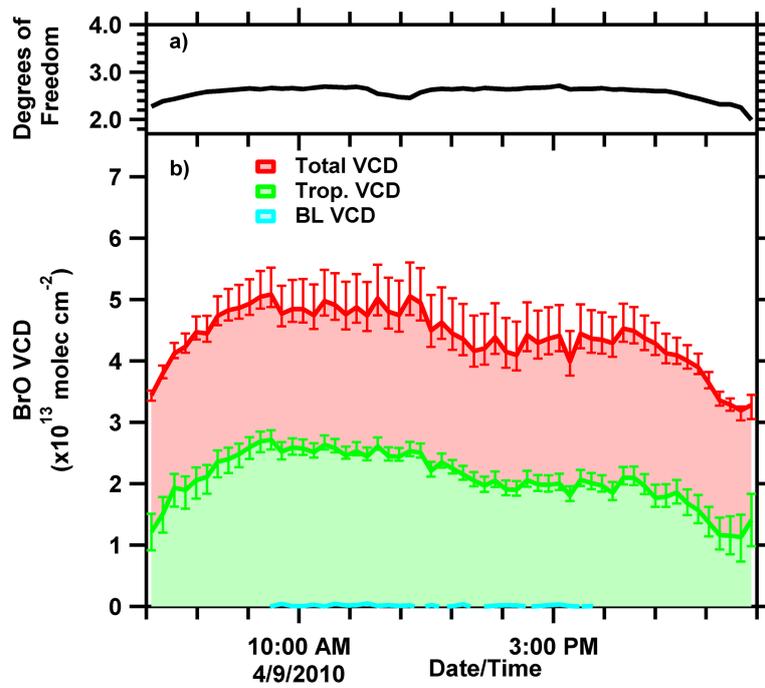


Figure 6. Graphic of the diurnal variation in the BrO vertical profiles represented as partial VCDs for the BL (0–1 km), troposphere (0–15 km), total (0–50 km) **(b)**, and the total DoFs from the inversion for each profile in **(a)**. The error bars on the VCDs indicate the range of values retrieved using the three different a-priori profiles.

28357

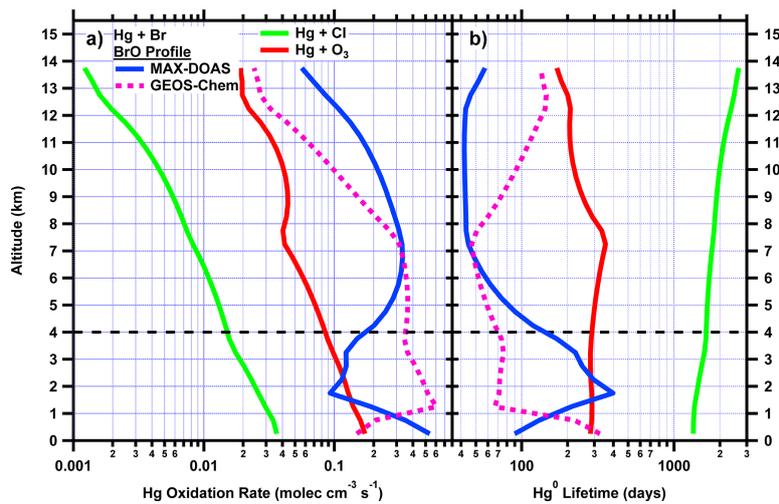


Figure 7. Box model results of the rate of mercury oxidation as a function of altitude for three species: (1) bromine radicals (green), (2) chlorine radicals (red), and (3) bromine radicals (solid blue: MAX-DOAS; dashed pink: GEOS-Chem) **(a)**; the corresponding lifetimes are found in **(b)**. The black dashed line at 4 km shows where measurement sensitivity starts to drop because of the decreasing amount of BrO (the measured parameter) in the lower layers of the atmosphere.

28358

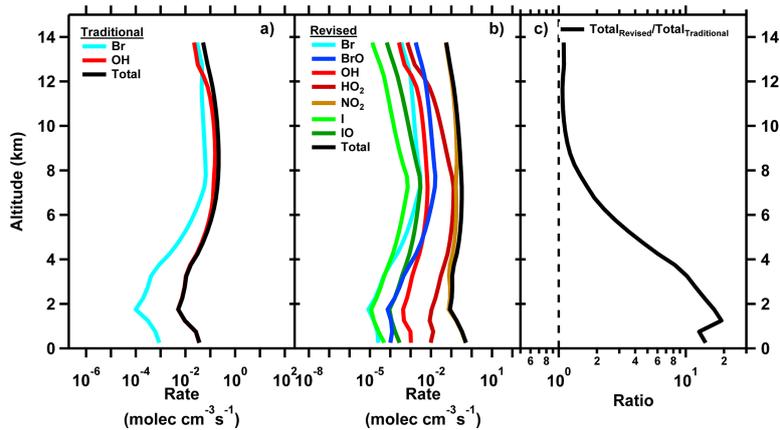


Figure 8. Box model results for the scavenging of the HgBr adduct as a function of altitude for two different reaction schemes: traditional (a) and revised (b). (c) contains the ratio of the total rates (black traces in panels a and b) to show the enhancement in the rate of the scavenging reaction when other reactants are taken into consideration.

28359

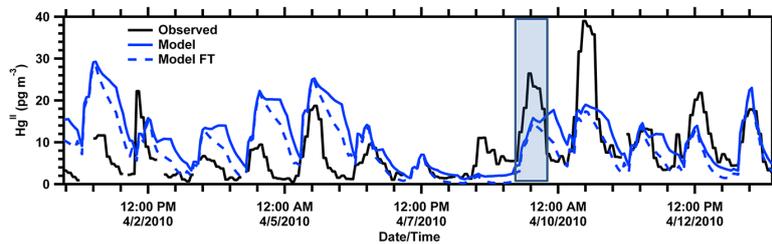


Figure 9. Observed (black solid trace) and simulated (blue traces) concentrations of oxidized mercury ($\text{Hg}^{\text{II}} = \text{GOM} + \text{PBM}$) during April 2010 (9 April highlighted). Model results from GEOS-Chem (solid) include a sensitivity test, labeled FT (dashed trace), in which North American anthropogenic emissions are zero and no GEM oxidation occurs in the lower troposphere (> 700 hPa); therefore all Hg^{II} in the “FT” model is derived from oxidation of in the free troposphere or long-range transport.

28360