

Abstract

The mandate of the Task Force Hemispheric Transport of Air Pollution (HTAP) under the Convention on Long-Range Transboundary Air Pollution (CLRTAP) is to improve the scientific understanding of the intercontinental air pollution transport, to quantify
5 impacts on human health, vegetation and climate, to identify emission mitigation options across the regions of the Northern Hemisphere, and to guide future policies on these aspects.

The harmonization and improvement of regional emission inventories is imperative to obtain consolidated estimates on the formation of global-scale air pollution. An emissions
10 dataset has been constructed using regional emission gridmaps (annual and monthly) for SO₂, NO_x, CO, NMVOC, NH₃, PM₁₀, PM_{2.5}, BC and OC for the years 2008 and 2010, with the purpose of providing consistent information to global and regional scale modelling efforts.

This compilation of different regional gridded inventories, including the Environmental Protection Agency (EPA)'s for USA, EPA and Environment Canada's for Canada, the European Monitoring and Evaluation Programme (EMEP) and Netherlands Organisation for Applied Scientific Research (TNO)'s for Europe, and the Model Inter-comparison Study in Asia (MICS-Asia)'s for China, India and other Asian countries, was gap-filled with the emission gridmaps of the Emissions Database for Global Atmospheric Research (EDGARv4.3) for the rest of the world (mainly South-America,
20 Africa, Russia and Oceania). Emissions from seven main categories of human activities (power, industry, residential, agriculture, ground transport, aviation and shipping) were estimated and spatially distributed on a common grid of 0.1° × 0.1° longitude–latitude, to yield monthly, global, sector-specific gridmaps for each substance and year.

The HTAP_v2.2 air pollutant gridmaps are considered to combine latest available regional information within a complete global dataset. The disaggregation by sectors, high spatial and temporal resolution and detailed information on the data sources and references used will provide the user the required transparency. Because HTAP_v2.2
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contains primarily official and/or widely used regional emission gridmaps, it can be recommended as a global baseline emission inventory, which is regionally accepted as a reference and from which different scenarios assessing emission reduction policies at a global scale could start.

5 An analysis of country-specific implied emission factors shows a large difference between industrialised countries and developing countries for all air pollutant emissions from the energy and industry sectors, but not from the residential one. A comparison of the population weighted emissions for all world countries, grouped into four classes of similar income, reveals that the per capita emissions are, with increasing income
10 group of countries, increasing in level but also in variation for all air pollutants but not for aerosols.

1 Introduction

Intercontinental transport of air pollution occurs on timescales of days to weeks and, depending on the specific type of pollutant, may contribute substantially to local scale
15 pollution episodes (HTAP, 2010). Common international understanding of global air pollution and its influence on human health, vegetation and climate, is imperative for providing a basis for future international policies and is a prime objective for the Task Force Hemispheric Transport of Air Pollution (TF HTAP)¹. While nowadays many countries and regions report their air pollutant emissions, these estimates may not be readily
20 accessible, or may be difficult to interpret without additional information, and their quality may differ widely, having various degrees of detail and being presented in different formats.

The UN Framework Convention on Climate Change (UNFCCC) requires official inventory reporting that complies with the TACCC principles of quality aiming at Trans-

¹More info on www.htap.org.

parency, Accuracy, Consistency, Comparability and Completeness², reviewed by UNFCCC roster experts and made available at their website (UNFCCC, 2013). Under the CLRTAP the parties need to report emissions to the EMEP Centre for Emission Inventories and Projections (CEIP), which also reviews data on completeness and consistency. Responsibility of providing emission inventories to several international bodies is often distributed within a particular country and so an inventory for, for example, methane can be provided by different organisations and although they represent the same region they might be different, in fact often are, leading to confusion we need to work with (Janssens-Maenhout et al., 2012).

Currently available emission inventories differ in spatial and temporal resolution (“consistency”), in coverage of geographical area, time period and list of compounds (“completeness”) and in the sector-specific details of the source calculation (“transparency”). Moreover the official inventories submitted by countries have at least one year time lag, are updated with different frequency and with or without review of the historical time series. The work of Lamarque et al. (2010) provides a unique example of a comprehensive “composite” historical emissions dataset spanning from 1850 to 2000, mainly based on scientific estimates. The dataset also provided harmonized base-year (2000) emissions that were used as a starting point for the development of the so-called RCP (Representative Concentration Pathways) emission scenarios (e.g. Moss et al., 2010; van Vuuren et al., 2011). For other years and specific model domains covering multiple regions, atmospheric modellers often compile their own emission inputs drawing upon different pieces of the available inventories. These compilations involve sometimes arbitrary choices, and are often not clearly described or evaluated. For example, the atmospheric modelling groups, which contributed to the HTAP multi-model experiments described in HTAP (2010), used their own best estimates for emissions for the year 2001, obtaining in some cases comparable global emissions (e.g. for NO_x and SO₂), and sometimes getting larger differences (e.g. for NMVOC emissions).

²Timeliness is recently also considered.

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Moreover, Streets et al. (2010) evaluated the consistency of the emissions used in the various models and nationally reported emissions. For a follow-up study in HTAP Phase 2, it was recommended to provide a harmonised emissions dataset for the years 2008 and 2010 in line with the following 4 major objectives:

1. To facilitate development mitigation policies by making use of well documented national inventories;
2. To identify missing (anthropogenic) sources and gap-fill them with scientific inventories for a more complete picture at global scale;
3. To provide a reference dataset for further emission compilation activities (benchmarking or scenario exercises);
4. To provide a single entry point for consistent global and regional modelling activities focusing on the contribution of long-range (intercontinental) air pollution to regional air quality issues.

A harmonized global, gridded, air pollution emission dataset has been compiled with officially reported, gridded inventories at the national scale, to the extent possible and complemented with science-based inventories for regions and sectors where nationally reported data were not available.

Whereas for a preceding dataset³ of EDGAR-HTAP_v1 the nationally reported emissions, combined with regional scientific inventories and gapfilled with the global set originating from EDGARv4.2 were all gridded with geospatial data from EDGAR

³EDGAR-HTAP_v1 completed in October 2010 comprises sector-specific annual gridmaps for the six years from 2000 to 2005 and covers air pollutants (CH₄, CO, NH₃, NMVOC, SO₂ and NO_x) and particulate matter with its carbonaceous speciation (PM₁₀, PM_{2.5}, BC and OC). The annual gridmaps of 0.1° × 0.1° resolution are made available via http://edgar.jrc.ec.europa.eu/national_reported_data/htap.php and the CIERA and ECCAD servers. Documentation is available in the HTAP_v1 EUR25229EN report of Janssens-Maenhout et al. (2012) (http://edgar.jrc.ec.europa.eu/htap/EDGAR-HTAP_v1_final_jan2012.pdf).

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establishment of the 7 categories: Aircraft, International Shipping, Power Industry, Industry, Ground Transport, Residential and Agriculture (described in Table 1 underneath).

HTAP_v2.2 focusses only on anthropogenic emissions, in a comprehensive way, but excludes large-scale biomass burning (forest fires, peat fires and their decay) and agricultural waste or field burning. We refer to inventories such as GFED3 (van der Werf, 2010) for the forest, grassland and Savannah fires (IPCC categories 5A + C + 4E) and to the $1^\circ \times 1^\circ$ gridmaps of Yevich et Logan (2003) or the $0.1^\circ \times 0.1^\circ$ EDGARv4.2 gridmaps (EC-JRC/PBL, 2011) for the agricultural waste burning (4F). Moreover, only NH_3 emissions from the agricultural sector were taken up in the htap_8_Agriculture sector of HTAP_v2.2 inventory, so that the occasionally reported NO_x from agricultural waste burning or from biological N-fixation and crop residues (which is typically considered under S10 for Europe) are excluded.

2.2 Gridded input datasets for HTAP_v2.2

As explained earlier, the goal of the HTAP_v2.2 inventory is to provide consistent and highly resolved information (see Fig. 1a) to global and regional modelling. It is important to realize that in the HTAP modelling exercise both global and regional models are participating. The HTAP global modelling is coordinated with the regional modelling exercise of Air Quality Model Evaluation International Initiative AQMEII (Galmarini et al., 2012, 2015) that manages regional scale activities for Europe and North America, and the regional modelling exercise of the Model Intercomparison study for Asia MICS-Asia (Carmichael et al., 2008) that manages the regional modeling over Asia. Hence, the regional inventories used for HTAP_v2.2 are constructed and used in accordance with these regional activities.

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2.2.1 USA and Canada: EPA and Environment Canada gridmaps and EPA temporal profiles

EPA (2013) provides the 2008 and 2010 areal and point source emissions for the complete North American domain at $0.1^\circ \times 0.1^\circ$ resolution, covering USA with a grid ranging from $180\text{--}63^\circ$ W in longitude and $75\text{--}15^\circ$ N in latitude and covering Canada with a grid from $142\text{--}47.8^\circ$ W in longitude and $85\text{--}41^\circ$ N in latitude. Mexico is not covered by these latitudes and it is gapfilled with EDGARv4.3 data (see Sect. 2.2.4). For the northern latitudes above 45° N, Environment Canada provided the 2008 basis and an update of the point sources for 2010, from which US EPA prepared the full set of detailed gridmaps also for 2010. The temporal profiles of US EPA were applied for USA and Canada with identical monthly distributions per sector for 2008 and 2010. More details about the US inventory are given by Pouliot et al. (2014, 2015).

2.2.2 Europe: TNO gridmaps and EMEP temporal trends

Countries that are parties to the CLRTAP (www.unece.org/env/lrtap/lrtap_h1.html) need to report anthropogenic emissions of air pollutants and particulate matter, but neither BC nor OC. These reported/official inventories are reported on the national level to EMEP-CEIP⁶ which provides the annual emission inventory data for CO , NH_3 , NMVOC, NO_x , SO_x , PM_{10} and $\text{PM}_{2.5}$ (not BC and not OC). However, the currently used EMEP grid uses a polar-stereographic projection with about $50\text{ km} \times 50\text{ km}$ grid cells centered over the European region and converting to a Mercator projection implied a loss of spatial accuracy. These reported data are incomplete according to the CEIP annual report of Mareckova et al. (2013) and for evaluation with the EMEP unified model further gapfilling is needed, resulting in a semi-official emission dataset. To overcome the problems of inconsistent emissions time series and fulfil the need for a higher spatial resolution to support AQ modelling in Europe in the European FP7 project Monitoring

⁶More info on www.ceip.at.

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Atmospheric Composition and Climate (MACC), TNO established a scientifically complete and widely accepted dataset, which is fully documented by Kuenen et al. (2014). This so-called TNO-MACC-II inventory of Kuenen et al. (2014) covers the same European domain with areal and point source emission gridmaps at $1/8^\circ \times 1/16^\circ$ resolution for SO_2 , NO_x , CO , NMVOC, NH_3 , PM_{10} , $\text{PM}_{2.5}$ with point sources allocated to their exact location. The grid-domain ranges from 30°W – 60°E in longitude and 72° – 30°N in latitude. The geographical area covered all EU-28 countries, Switzerland, Norway, Iceland and Liechtenstein, Albania, Bosnia-Herzegovina, Serbia, Macedonia, 6 Newly Independent States (Armenia, Azerbaijan, Belarus, Georgia, Moldova, Ukraine) and Turkey. Countries with partial coverage (Russia, Turkmenistan, Kazakhstan and Uzbekistan) were not used in the HTAP_v2 inventory because of inconsistencies with other datasets (see Sect. 2.2.4). Sector-specific data (given by SNAP-code, see Table 1) are used for all countries with complete coverage of their territory and for each substance the contribution from each sector is compared to EMEP and EDGARv4.3 estimates. Standard re-sampling is applied to obtain gridmaps at the common resolution of $0.1^\circ \times 0.1^\circ$. Point-source, ground-level airport emissions in the transport sector (under SNAP 8) were taken out, in order to avoid a double counting with the aviation sector (HTAP_1), for which the same geospatial dataset taken from EDGAR_v4.3 was used globally.

For NH_3 , the reporting of emissions from the energy, industry and residential sectors was apparently negligible for some countries⁷ compared to the agricultural emissions and was therefore not gapfilled by EMEP and/or TNO.

BC and OC emission data are not available as emission gridmaps within the MACC-II dataset, but the PM gridmaps are accompanied by a recommendation on the PM com-

⁷No NH_3 emissions are reported in the energy sector: for the countries Albania, Bosnia-Herzegovina, Cyprus, Estonia, Greece, Ireland, Iceland, Luxembourg, Latvia, FRY Macedonia, Malta, Norway, Poland, Romania, Slovakia, and Slovenia; in the industry sector for the countries Albania, Bosnia-Herzegovina, Greece, Ireland, Iceland, and FRY Macedonia; and in the residential sector for the countries Greece, Iceland and Slovenia.

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position describing the carbonaceous profiles per SNAP code and country. This so-called PM split table (per SNAP code and country) of TNO (TNO, 2009) is used to derive the BC and OC from PM_{10} and $\text{PM}_{2.5}$ emission gridmaps (see Kuenen et al. (2014) for details).

Finally, to derive the monthly gridmaps we used the country-specific and sector-specific data monthly profiles per substance for the EMEP model (M. Schulz, personal communication, 2013 and A. Nyiri personal communication, 2013).

2.2.3 Asia: monthly gridmaps from MIX

For Asia, a different challenge is faced, because no countries except Japan are legally required to yearly report detailed emission inventories under the LRTAP, UNFCCC or similar conventions. However, in Asia many scientific efforts aimed at establishing a detailed emission inventory, accepted by the different regions, using official or semi-official statistics collected at county level (by provinces for China). Under the Model Inter-comparison Study for Asia Phase III (MICS-Asia III), a mosaic Asian anthropogenic emission inventory was developed for 2008 and 2010 (Li et al., 2015). The mosaic inventory, named MIX, incorporated several local emission inventories including the Multi-resolution Emission Inventory for China (MEIC), NH_3 emission inventory from Peking University (Huang et al., 2012), Korean emissions from the Clean Air Policy Support System (CAPSS) (Lee et al., 2011), Indian emissions from the Argonne National Laboratory (Lu et al., 2011), and fill the gap where local emission data are not available using REAS2.1⁸ developed by Kurokawa et al. (2013).

MEIC is developed by Tsinghua University under an open-access model framework that provides model-ready emission data over China to support chemical transport

⁸The REAS2.1 inventory for Japan includes the data developed by Ministry of the Environment of Japan (MOEJ, 2009) for NMVOC evaporative emissions from stationary sources, the database developed by the Ocean Policy Research Foundation (OPRF, 2012) for the maritime sector, and the Japan Auto-Oil Program Emission Inventory-Data Base (JEI-DB) developed by Japan Petroleum Energy Center (JPEC, 2012a–c) for other sources.

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and steel industry according to IEA (2013). Finally, international shipping contributes ~ 10% to the global SO₂ emissions. SO₂ gridmaps clearly show the ship emission tracks connecting Asia and Europe with Africa and America.

3.1.2 NO_x

5 Figure 2b shows that the major sources of NO_x are ground transport and power generation and these source contributions show a rather uniform feature for all the considered regions. In Central and South America major emissions are attributed to the transportation sector and just to a minor extent to the energy sector. Those industrialised
10 centres (such as Canada) show relatively high emissions of NO_x. International shipping and, in particular, aviation contribute together more than 10 % of global NO_x emissions.

3.1.3 CO

CO is a product of incomplete combustion, which can therefore be emitted by any fuel combustion (ground transport, industrial processes involving combustion, as well as
15 domestic heating). As presented in Fig. 2c, the power generation sector emits less CO than the residential one because of higher combustion efficiency and higher temperatures compared to domestic burners. In Africa, there are large emissions of CO from the residential sector, mainly due to the use of wood and charcoal for cooking activities. As shown in Fig. 2c, some industrial activities emit CO, like the production
20 of non-metallic minerals and crude steel and iron, which is particularly relevant for India and China, while non-ferrous metal and iron and steel production are dominant in Oceania.

3.1.4 NMVOC

NMVOCs (non-methane volatile organic compounds) are emitted from chemical and
25 manufacturing industries, as well as fuel transformation processes, the production
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of primary fuels, the use of solvents and from the residential sector, inclusive waste (Fig. 2d). Important sources of NMVOCs include also evaporative emissions from road transport, specifically gasoline engines and the use of biofuels. Major emission sectors in the USA emitting NMVOCs include oil refineries, oil and gas production, several
5 industrial processes and motor vehicles. Most of the NMVOC emissions in Europe are due to solvent use, road transport, and the use of primary solid biomass in the residential sector. In the Middle East NMVOC sources include oil production and in South-Eastern Asia charcoal production. In China, particular high emissions are also associated with the high use of solvents in paints and in Brazil with the use of bio-
10 fuels. NMVOC speciation is not provided by the HTAP_v2.2 emission database; however TNO has produced a breakdown into 23 VOC species, which has been used for the RETRO project and the RCP scenarios of IPCC AR5. Recommendations for the NMVOC splits are given on the HTAP wiki site <http://iek8wikis.iek.fz-juelich.de/HTAPWiki/WP1.1>.

3.1.5 NH₃

NH₃ is mainly emitted by the agricultural sector, including management of manure and agricultural soils (application of nitrogen fertilizers, incl. animal waste), as Fig. 2i shows, while a relatively small amount is emitted by the deployment of catalysts in gasoline cars. Minor contributions are also observed for Asian countries from the residential
20 sector due to dung and vegetal waste burning and coal combustion. For industrialized regions, especially for countries using low sulphur fuel, Mejía-Centeneo et al. (2007) reported that the deployment of catalytic converters in gasoline cars enhanced the NH₃ emissions from this source since mid-2000. This is also observed by the larger NH₃ with increased transport activity and corresponding increased consumption of low sulphur fuels. In the USA gasoline vehicle catalysts represent ca 6 % of total NH₃
25 emissions, while a lower contribution is found for Europe due to the high deployment of diesel vehicles.

heavy fuel oil (in the Middle East power sector according to IEA (2013) activity data). The PM emissions from the industry of the other developing countries show a decrease from 2008 to 2010, indicating slow penetration of end-of-pipe abatement.

3.6 Qualitative assessment of the uncertainty of emission gridmaps

5 Even though the HTAP_v2.2 data sources are quite different, and lead to inconsistencies over borders, a bottom-up methodology with activity data and emission factors is applied to calculate emission totals and distribute these on the grid. The propagation of uncertainty is given by the effect of variables' uncertainties (or errors) on the uncertainty, i.e. the variance of the activity data and that of the emission factor. Table 3
10 provides some insight in the estimation of the uncertainty range, however the approach followed in HTAP v2.2 inhibits an overall consistent uncertainty assessment because it is not a bottom-up inventory.

We can only compare the HTAP v2.2 with the ECLIPSEv5 dataset of Klimont et al. (2015), which is a fully consistently built global bottom-up inventory and serves
15 as base year for the HTAP scenarios. At global level, a relatively good agreement is found with small relative emission differences $(ECLIPSEv5 - HTAPv2.2)/HTAPv2.2$ for the aggregated sectors in 2010. The relative difference for NO_x and CO is only -4% respectively $+5\%$. For SO_2 a larger difference of -8% reflects the recent important S-reductions for the non-ferrous metal smelters in ECLIPSEv5 (Klimont et al., 2013).
20 For NH_3 a relative difference of $+17\%$ is acceptable because of the larger uncertainty in emission factors driven by lack of information about manure management practices and also by incomplete data on the agricultural activities. For NMVOC a difference of -27% stems primarily from the assumptions about emissions from solvent use. The information about activity levels is scarce and even less is known about the emission
25 factors for some important sources. Both regional inventory compilers and modellers often make assumptions about per capita or per GDP solvent use NMVOC emissions from particular sectors. Here assumptions employed in the ECLIPSEv5 lead to lower emissions from these activities. As anticipated (and reflected in Table 3) larger differ-

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ences of 48 and 29% are present for $PM_{2.5}$ and BC, respectively. While for $PM_{2.5}$, assumptions about penetration and efficiency of filters in industrial and small-scale residential boilers as well as emission factors and activity data for biomass used in
5 cooking stoves play a key role, for BC assumptions about coal consumption in East Asia are of relevance since ECLIPSEv5 relied on provincial statistics for China which results in higher coal consumption than reported in national statistics and IEA. Additionally, ECLIPSEv5 includes emissions from kerosene wick lamps, especially relevant for South Asia and parts of Africa according to Lam et al. (2012), gas flaring and high emitting vehicles, which together result in about 30% higher emissions.

10 In addition, the spatial allocation is subject to other types of errors, with a spatial variance for point sources and a more important systematic error when a spatial proxy is used to distribute the emissions. Geo-spatial consistency is lower in the HTAP_v2.2 database than if the national totals would have been spatially redistributed with one harmonised spatial proxy dataset.

15 Another type of inconsistency occurs when speciation of a substance is done with gridmaps of different data sources. This was another reason not to use the PM gridmaps of EMEP, as no BC and OC speciation is available from the same EMEP data source. Instead we used the gridmaps of TNO for all PM components (PM_{10} and $PM_{2.5}$) and the TNO speciation file for BC and OC. In addition a check was performed
20 to ensure that the sum of BC and OC emissions in every grid cell is smaller than the $PM_{2.5}$ emission in that grid cell. Thereto a re-allocation of the emissions of some point sources (industrial facilities) was needed within Europe (e.g. Poland) and performed in consultation with TNO.

25 Another check was to estimate per grid cell the change in emission from 2008 to 2010 and allowed to find missing sources. However, global consistency cannot be guaranteed and a comparison of different countries or of different years cannot be conclusive. In particular point sources are very important input, but their strengths and locations are subject to input errors with larger consequences and cannot be extrapolated in time. (Closure of power plants as large point sources can change the emission distri-

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Table 1. Sectors in the HTAP_v2.2 inventory (only anthropogenic sources are included) and the corresponding Nomenclature for Reporting (NFR) and the Selected Nomenclature for Sources of Air Pollution (SNAP) codes as spelled out in the EMEP (2002) Reporting Guidelines.

Tag	Description	IPCC level (NFR code)	EMEP SNAP code
htap_1_Aircraft	International and domestic aviation	1.A.3a(i) + (ii)	S8 ^a
htap_2_International Shipping	International shipping	1.A.3d(ii)	
htap_3_Power industry	Power generation	1.A.1a	S1
htap_4_Industry	industrial non-power but large-scale combustion emissions and emissions of industrial processes ^b and product use inclusive solvents.	1.A.1b + c, 1.A.2, 1.B.1 + 2, 2.A + B + C + D + G, 3	S3 + S4 + S5 + S6 ^f
htap_5_Ground transport	Transport by road, railway, inland waterways, pipeline and other ground transport of mobile machinery ^d . Htap_5 does not include re-suspended dust from pavements or tyre and brake wear.	1.A.3b + c + d(ii) + e	S71 + S72 + S73 + S74 + S75 + S8 ^e
htap_6_Residential	Small-scale combustion, including heating, cooling, lighting, cooking and auxiliary engines to equip ^f residential, commercial buildings, service institutes, and agricultural facilities and fisheries; solid waste (landfills/incineration) and wastewater treatment.	1.A.4 + 5 6.A + B + C + D	S2 + S9
htap_8_Agriculture	Agricultural emissions from livestock, crop cultivation but not from agricultural waste burning and not including Savannah burning.	4.A + B + C + D	S10

Notes: ^a S8 (point source) includes local emissions of aircrafts around the airport only below 3000 ft.

^b Product testing by the manufacturer inside is not considered an emission of the building (htap_6) but taken up under the industry (htap_4). The oil production sector is completely covered in htap_6 and includes the fugitive (evaporative) emissions (mainly NMVOC) during the oil and gas exploration and production and transmission. As such, there are NMVOC emissions along the oil tanker tracks visible under the htap_4 sector.

^c Note that S34 = S3 + S4 in the TNO-MACC-II inventory (Kuenen et al., 2014). Fuel transformation processes (and refineries) are included here.

^d The pipeline transport does not include transmission of natural gas and crude oil, because the latter is included in the oil and gas production industry under htap_4 but it does include the transport of refined products (motorgasoline, diesel, liquefied petroleum gas) or goods. The other ground transport includes all mobile (non-stationary) machinery (as used in the agriculture, forestry or construction sector).

^e For the split-up of SNAP7 into S71, S72, S73, S74 and S75 we refer to the definitions used for the TNO-MACCII inventory documented in Kuenen et al. (2014).

^f In particular industrial, commercial and/or agricultural buildings can be more extensively equipped with auxiliary stationary (non-mobile) infrastructure in and around the building (e.g. lifting devices).

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Table 2. Comparison of per capita emissions in 2010 for USA, Germany and China from HTAP_v2.2.

Substance	USA	Germany	China
kg SO _x yr ⁻¹ cap ⁻¹	32.6	5.2	20.9
kg NO _x yr ⁻¹ cap ⁻¹	43.6	14.2	20.8
kg VOC yr ⁻¹ cap ⁻¹	43.1	11.9	16.7
kg CO yr ⁻¹ cap ⁻¹	148.3	35.6	125.6
kg NH ₃ yr ⁻¹ cap ⁻¹	11.6	7.3	6.7
kg PM _{2.5} yr ⁻¹ cap ⁻¹	5.3	1.1	12.2
kg BC yr ⁻¹ cap ⁻¹	0.9	0.2	1.3

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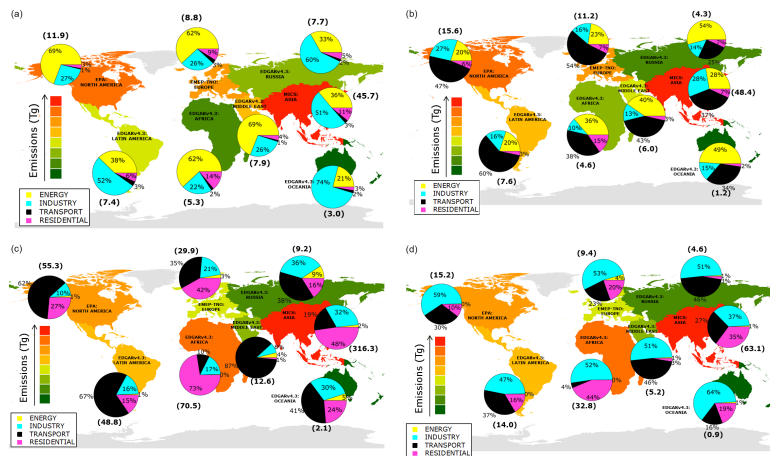


Figure 2. (a) Total Tg SO₂ emissions for 2010 (in brackets) and sector-specific composition for world regions. **(b)** Total Tg NO_x emissions for 2010 (in brackets) and sector-specific composition for world regions. **(c)** Total Tg CO emissions for 2010 (in brackets) and sector-specific composition for world regions. **(d)** Total Tg NMVOC emissions for 2010 (in brackets) and sector-specific composition for world regions.

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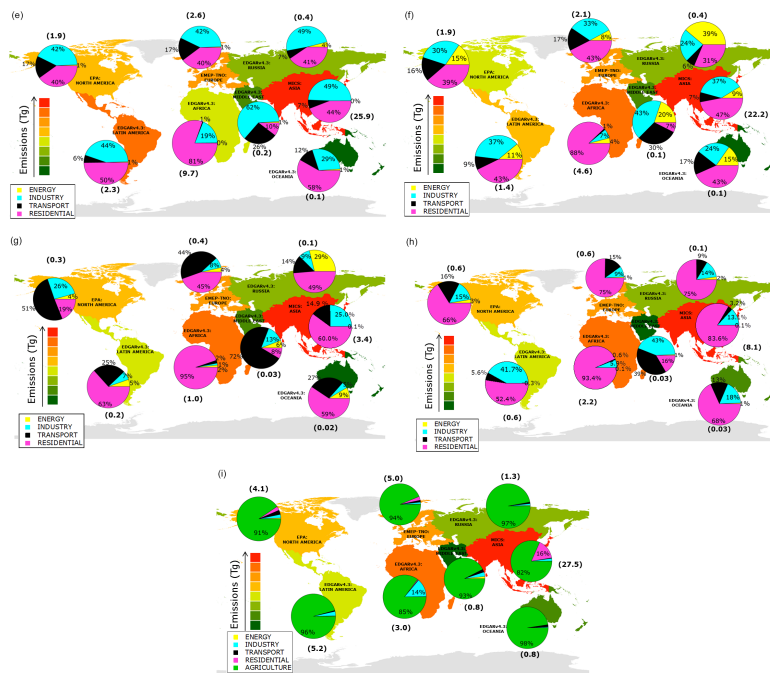
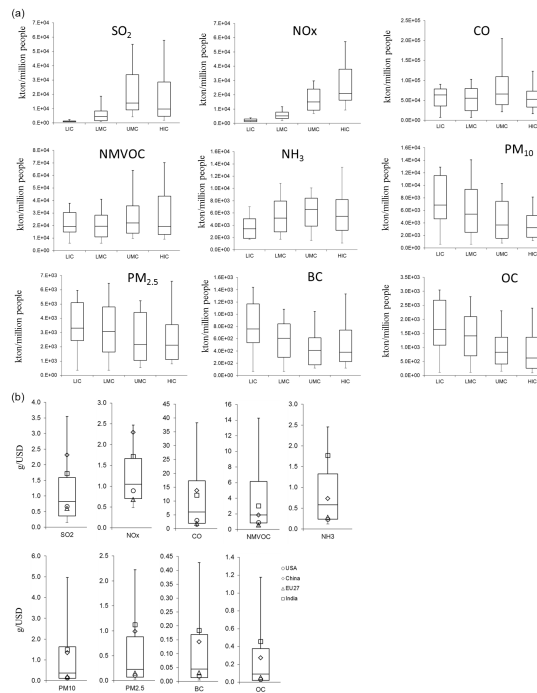


Figure 2. (e) Total Tg PM₁₀ emissions for 2010 (in brackets) and sector-specific composition for world regions. **(f)** Total Tg PM_{2.5} emissions for 2010 (in brackets) and sector-specific composition for world regions. **(g)** Total Tg BC emissions for 2010 (in brackets) and sector-specific composition for world regions. **(h)** Total Tg OC emissions for 2010 (in brackets) and sector-specific composition for world regions. **(i)** Total Tg NH₃ emissions in 2010 (in brackets) and sector-specific composition for world regions.

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Figure 3. (a) 2010 per capita emissions per substance and per group of countries (in kg cap⁻¹ yr⁻¹): low income (LIC), lower middle income (LMC), upper middle income (UMC) and high income (HIC) with the maximum, and minimum and the percentiles reported in the box plot (10, 50, 90°) and the maximum and minimum in each group of countries. **(b)** Pollutant specific emissions divided by GDP (gUSD⁻¹) for the year 2010. Percentiles are reported in the box plots (10, 25, 50, 75, 90°) together with emission/GDP for specific regions (EU27, USA, China and India).

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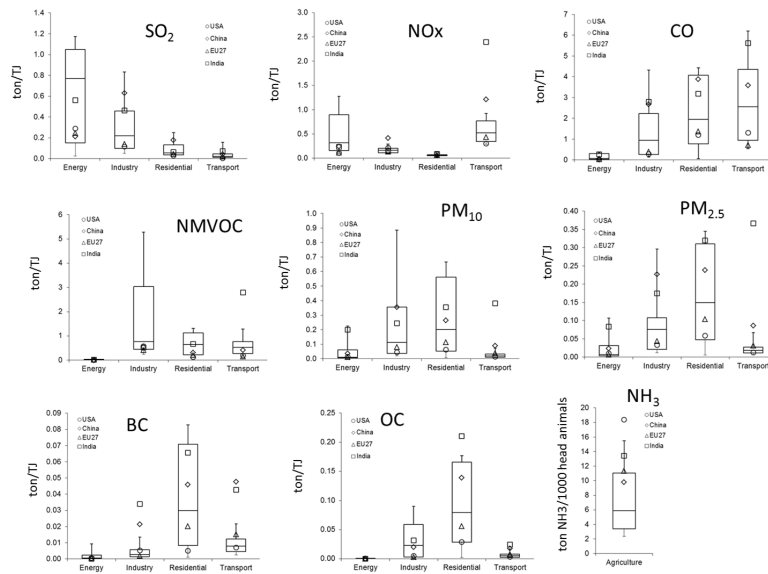


Figure 4. Sector specific implied emissions ($t(TJ)^{-1}$) for the year 2010. Percentiles are reported in the box plots (10, 25, 50, 75, 90^o) together with implied emission factors for specific regions (EU27, USA, China and India).