



*Supplement of*

## **Evaluating BC and NO<sub>x</sub> emission inventories for the Paris region from MEGAPOLI aircraft measurements**

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## 1 **Supplementary material**

### 2 **S.1) PSAP absorption coefficient measurements**

3 Particles are sampled through an isokinetic and isoaxial aircraft community aerosol inlet (CAI) based  
4 on the University of Hawaii shrouded solid diffuser inlet designed by A. Clarke and modified by  
5 Meteo France. The CAI inlet allows for entirely sampling the submicron particles and partly  
6 sampling of supermicron particles with an upper 50% sampling efficiency for super micron particle  
7 sizes at diameter around 5  $\mu\text{m}$  (McNaughton et al., 2007).

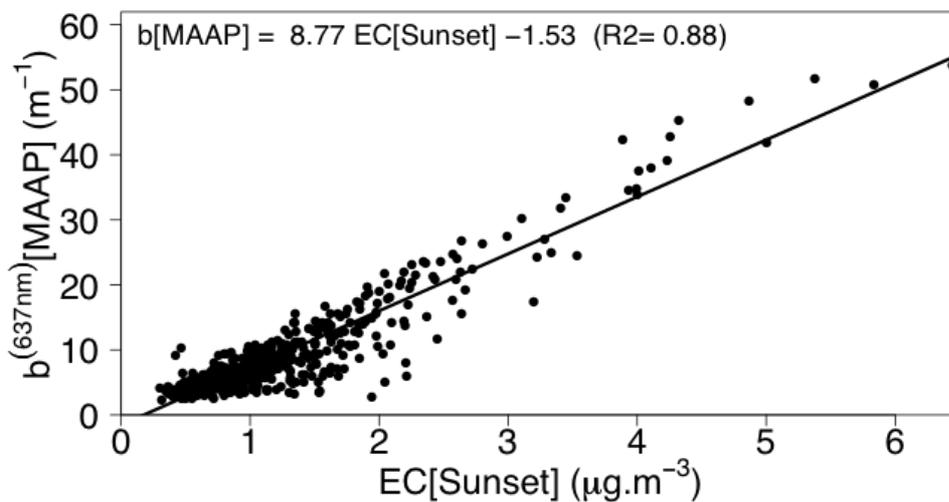
8 The PSAP is then used to measure in near real time the light absorption coefficient. The method is  
9 based on the integrating plate technique in which the change in optical transmission of a filter caused  
10 by particle deposition is related to the light absorption coefficient of the deposited particles using  
11 Beers law. During the MEGAPOLI campaign, the 3 wavelength (467, 530 and 660 nm) version has  
12 been operated on board the ATR-42 research aircraft. Several corrections have been applied to the  
13 PSAP measurements to obtain absorption coefficients at the three wavelengths and to deduce the BC  
14 content from the absorption coefficient extrapolated to 637 nm, thus, comparable to BC derived from  
15 state-of-the-art MAAP instruments. The PSAP calibration and correction methods are described in  
16 detail in Bond et al. (1999), Virkkula et al. (2005), and Müller et al. (2011). They include corrections  
17 for the PSAP spot size, for aerosol particle scattering (from simultaneous TSI nephelometer  
18 measurements) to take into account the decrease in filter transmittance due to scattering, for the  
19 absolute transmittance (filters have been changed after every flight, transmittance never decreased  
20 below 0.9). The PSAP flow had been calibrated as a function of upstream pressure. From the above  
21 described PSAP correction methods, an upper limit of the uncertainty in the derived absorption  
22 coefficient has been estimated as 30%.

### 23 **S.2) Determination of the mass-specific absorption coefficient**

24 The Multi-Angle Absorption Photometer (MAAP) instrument provides EBC concentrations  
25 deduced from the absorption coefficient measurement converted by a mass-specific absorption  
26 coefficient (MAC) of  $6.6 \text{ m}^2 \text{ g}^{-1}$  (Petzold and Schönlinner, 2004). This value had been initially  
27 derived from the comparison of 121 ambient EC concentrations measured at one rural  
28 background and three traffic impacted urban sites following the German thermal reference  
29 method (Schmid et al, 2001; VDI-1 method) with the corresponding MAAP absorption  
30 coefficient measurements (at 670 nm). However, considering the high uncertainties that exist  
31 on this coefficient, that MAC value may not be adapted to the Paris aerosol. Indeed, while  
32 Bond and Bergstrom (2006) have found only slightly variable MAC values for freshly emitted  
33 aerosol, around  $7.5 \pm 1.2 \text{ m}^2 \cdot \text{g}^{-1}$  (one sigma confidence) at 550 nm, an enhancement of this  
34 coefficient during the transition from external to internal mixing (notably due to organic  
35 coatings acting as a prism) has been noticed by several authors based on theoretical models

1 (around +50% in Bond et al., 2006) and laboratory experiments (between +80 and +110% in  
2 Schnaiter et al., 2005). It is likely to partly explain the large scatter (up to a factor of 4) of  
3 MAC values in ambient conditions (i.e. after a few hours) emphasized by Bond and Bergstrom  
4 (2006).

5 Considering such a large variability of MAC values from one location to the other, one must  
6 derive a MAC value relevant for the Paris megacity. A comparison between MAAP absorption  
7 coefficient  $b^{(637\text{ nm})}$  and Sunset Field instrument EC observations (co-located at the LHVP site)  
8 is thus performed over the July period. A scatter plot of MAAP versus Sunset observations is  
9 shown on Fig. S1. Both are highly correlated ( $R^2$  of 0.88), and the linear regression gives a  
10 slope of  $8.8 \pm 0.3 \text{ m}^2 \text{ g}^{-1}$  (at a 95% confidence interval). It is close to the values given by Bond  
11 and Bergstrom (2006) for freshly-emitted soot but in the upper range. It is also coherent with  
12 the MAC values of  $7.3 \pm 0.1 \text{ m}^2 \text{ g}^{-1}$  obtained during wintertime by Sciare et al. (2011) at Gif-  
13 sur-Yvette, a suburban location at 20 km south-west from Paris. Note that the higher (but  
14 older) MAC value ( $12.0 \pm 0.7 \text{ m}^2 \text{ g}^{-1}$ ) obtained at Gif-sur-Yvette by Liousse et al. (1993) may  
15 highlight some changes in the soot particles characteristics (e.g. size, shape, organic coating)  
16 or more probably reflect the uncertainties associated to the method to determine it (and, in  
17 particular, the thermo-optical protocol used to measure EC).



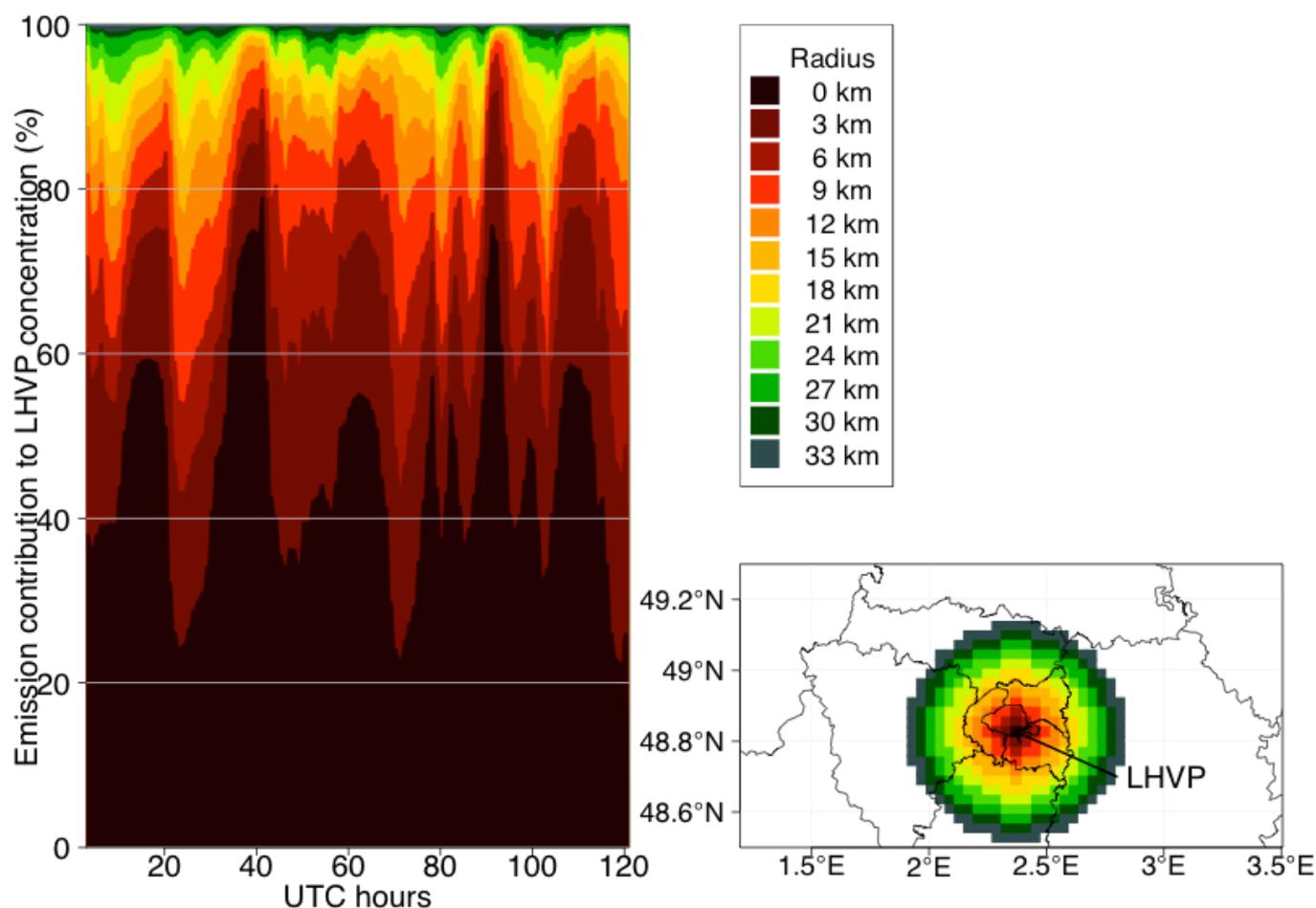
18  
19 Figure S1 : Hourly MAAP absorption coefficients at 637 nm against Sunset Field EC  
20 concentrations at the LHVP site during July.

21 In addition, there may be discrepancies between the aircraft PSAP and the ground based  
22 MAAP instrument. An intercomparison has been performed on July 11 during a few hours,  
23 this PSAP giving around 10% higher EBC concentrations, but these results do not appear as  
24 representative enough for the whole July month. Müller et al. (2011) have reported PSAP  
25 versus MAAP slopes of  $0.79 \pm 0.07$  during the GAW2005 inter-comparison campaign, and  
26  $1.05 \pm 0.08$  and  $0.99 \pm 0.10$  during the EUSAAR2007 one. Applied to our situation, this would

1 lead to a PSAP EBC over Sunset EC ratio between  $1.0 \pm 0.08$  ( $1.27 * 0.79$ ) and  $1.3 \pm 0.13$   
2 ( $1.27 * 1.05$ ). In order to be conservative, that latter value is retained for the MAC uncertainty,  
3 thus estimated to 40%.

### 4 **S.3) LHVP site representativeness**

5 In order to assess the area impacting results obtained at the LHVP site, a simulation is  
6 performed over 5 days (1-5 July) in which  $\text{NO}_x$  emissions at various distances from the site  
7 from 0 to 33 km are colored by inert tracers (Fig. S2). Tracer emission rates are taken from the  
8 TNO inventory. For simplification, no diurnal profile is assigned to these emissions, but this is  
9 not expected to modify results since most of emissions share the same diurnal variability  
10 (associated to the dominant traffic source). From this approach, it is possible to determine the  
11 contribution of emissions at a given distance to the concentration at the LHVP site.



14 Figure S2 : Emission contributions to LHVP concentrations at several distances from the site  
15 (left panel) and map of these emissions (right panel). Distances are those of the centers of the  
16 grid points.

17 As expected, the largest contribution originates from emissions in the cell where LHVP is  
18 located. It is quite variable in time with values ranging from 25 to 75%, notably depending on

1 the wind, with highest values (at hour ~90) associated to stagnant conditions (wind speed  
2 below 1 m.s<sup>-1</sup>). Conversely, the most distant emissions beyond 21 km contribute to less than  
3 10%. These results indicate that emission error factors obtained at LHVP are strongly  
4 influenced by nearby sources and are not representative for the whole agglomeration.

5

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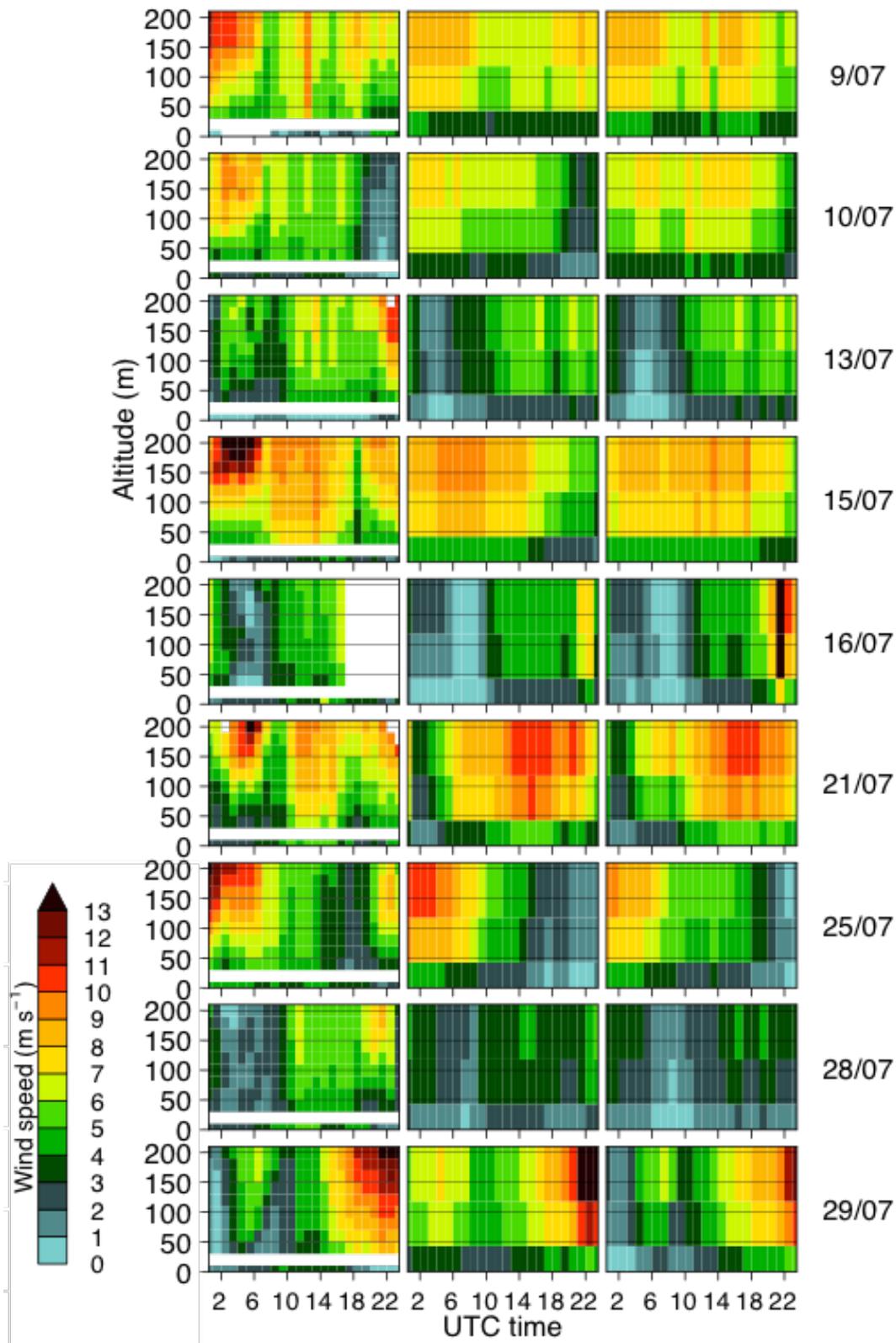
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2 **S.4) Supplementary figures and tables**

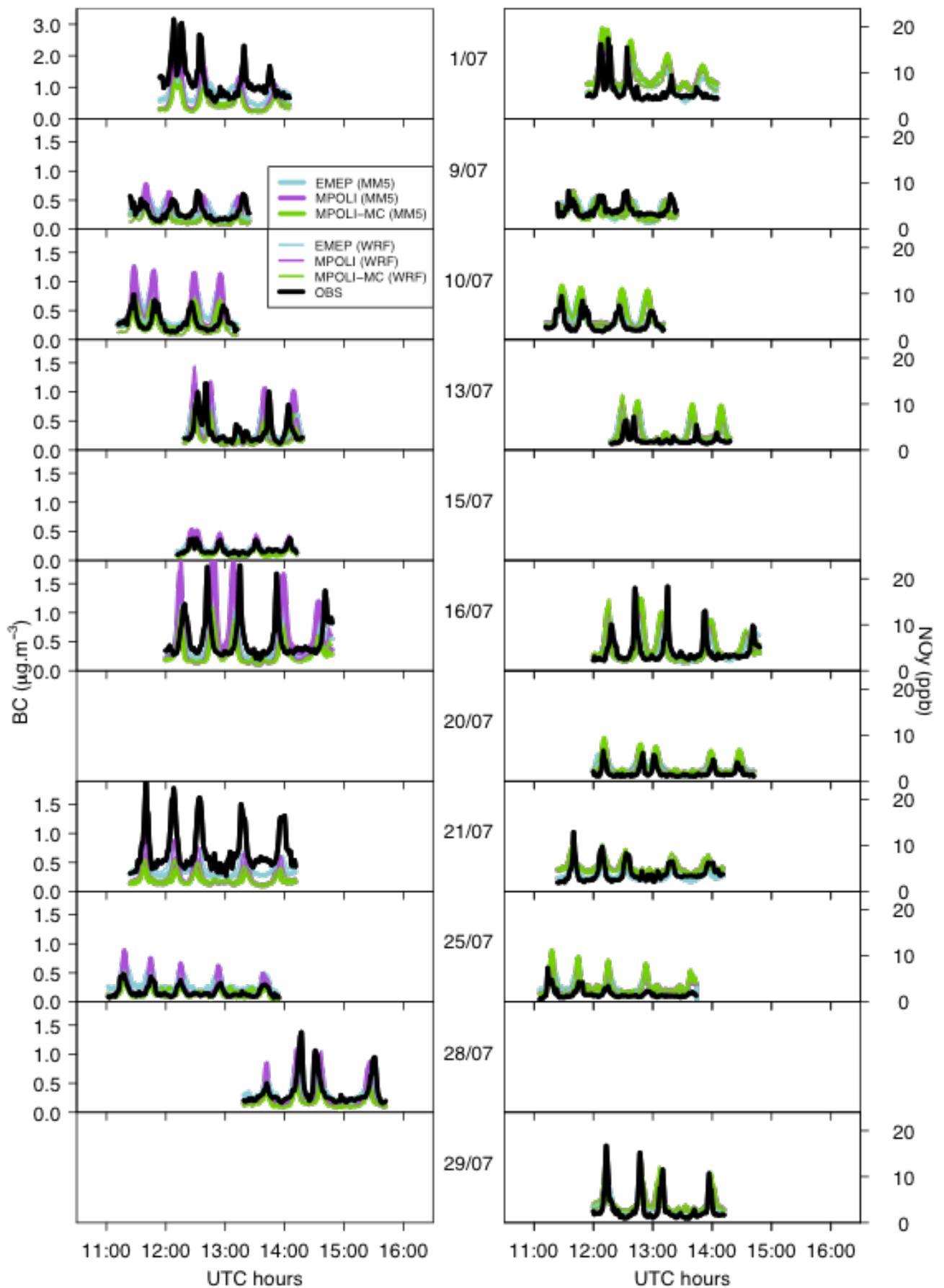
3 Table S1 : July BC emissions in the Ile-de-France region.

SNAP sector	Description	July factor*	July emissions (tons)		
			EMEP	TNO	TNO-MP
1	Public electricity and other energy transformation	0.234	1	2	2
2	Small combustion plants	0.261	22	77	18
3	Industrial combustion and processes with contact	0.939	4	6	1
4	Industrial process emissions	1.026	7	30	9
5	Fossil fuel production	1.001	7	0	0
6	Solvent and product use	1.077	2	2	0
7	Road transport	1.029	353	316	234
8	Other non-road transport and mobile machinery	1.030	131	57	22
9	Waste disposal	1.000	3	50	2
10	Agriculture	0.593	0	0	0
Total		-	530	540	288

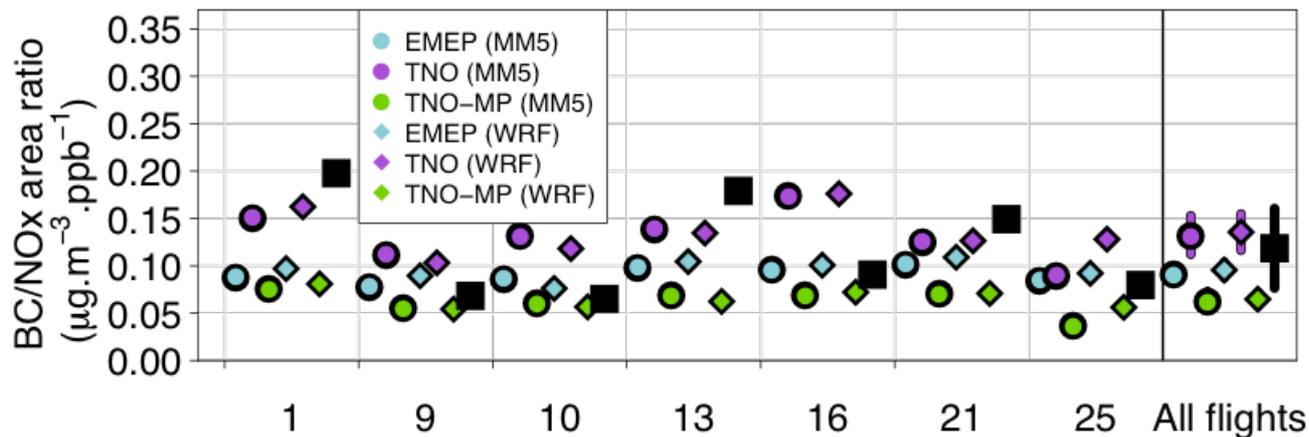
4 \*For each SNAP sector: [July emission] = [annual emission]\*[July factor]/12



1  
 2 Figure S3 : Hourly wind speed at various altitudes (above ground) at the SIRTA site for each  
 3 flight day (no available data for the 1 July), from Lidar observations (left column), MM5  
 4 (middle column) and WRF model simulations (right column). Previously discussed wind  
 5 speed observations at ground are also reported in the left column.

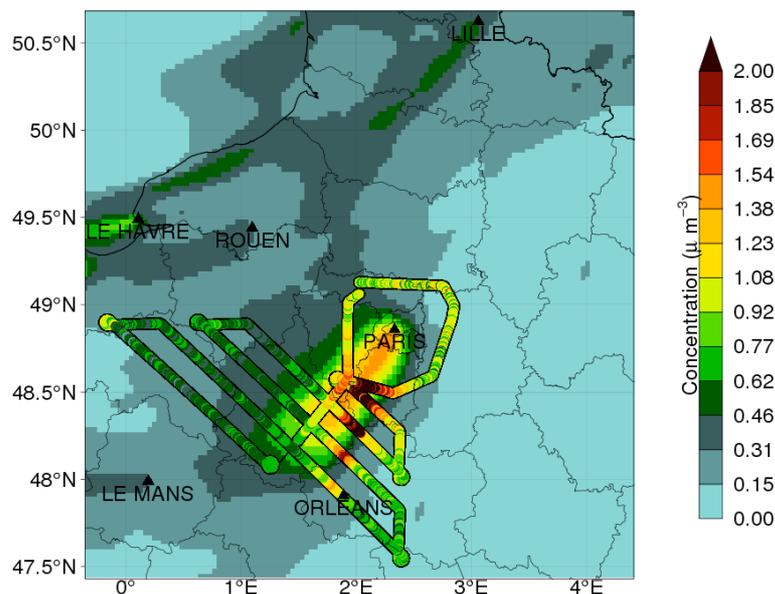


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 2 Figure S4: Observed and simulated BC (left column) and NO<sub>y</sub> (right column) concentrations  
 3 along aircraft trajectories, for each flight.

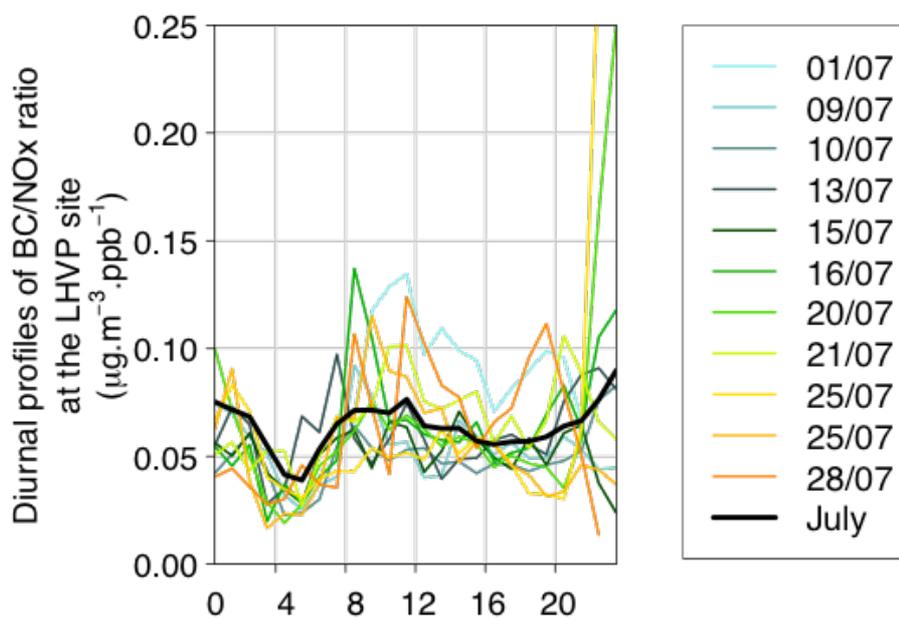


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2 Figure S5: Ratio of the BC over NO<sub>y</sub> peak area for observations and simulations.

3

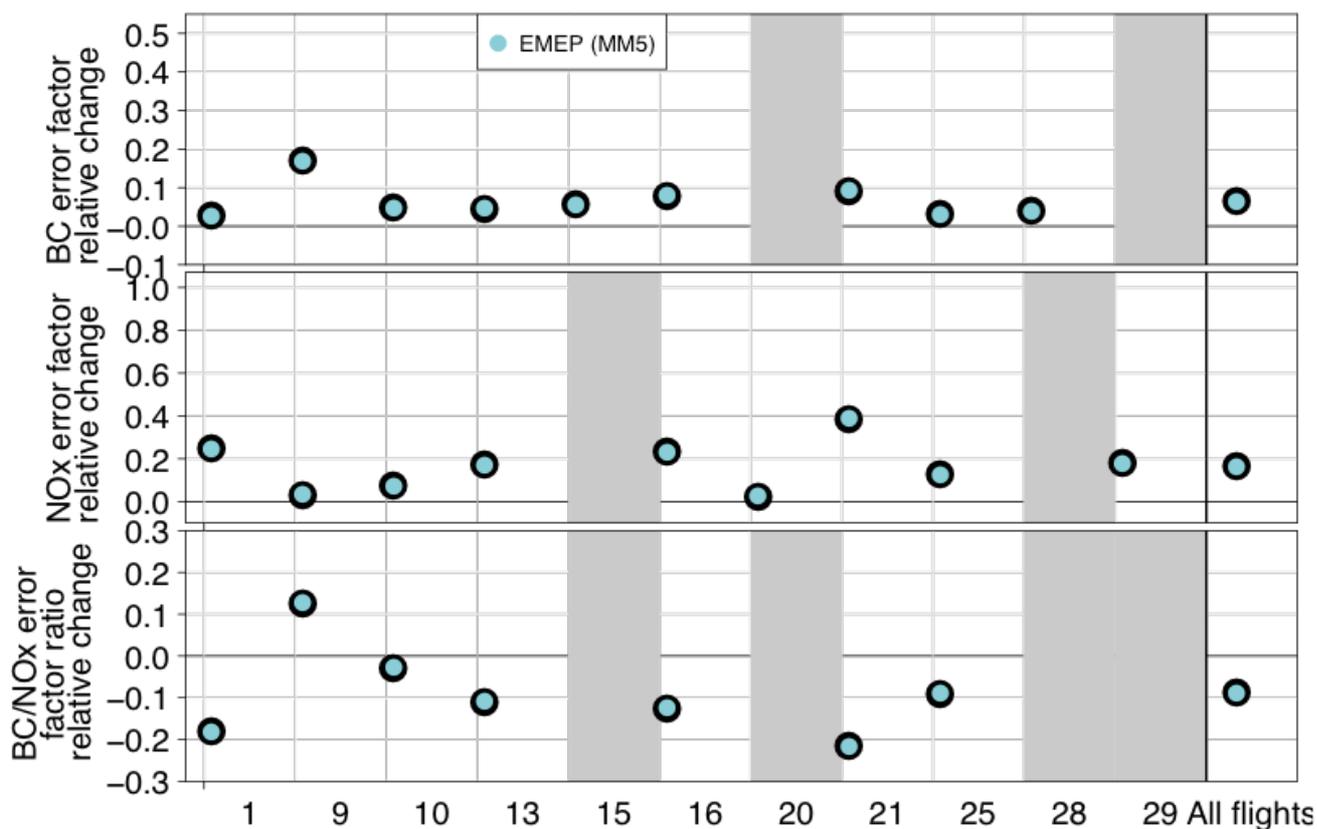


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5 Fig. S6 : Aircraft trajectory and observed (along the trajectory) and modeled (in background,  
6 with the TNO-MM5 case) BC concentration for the 1 July.



7

1 Figure S7 : BC/NO<sub>x</sub> ratio diurnal profiles for each flight day and in average over the whole  
 2 July month at the LHVP site.



3  
 4 Figure S8 : BC, NO<sub>x</sub> error factor and BC/NO<sub>x</sub> error factor ratio changes after removing any  
 5 deposition on both species (note that the scale is no longer logarithmic).

6  
 7 **S.5) Additional references**

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