Supporting Information for

Aqueous-phase photochemical oxidation and direct photolysis of vanillin—a model compound of methoxy-phenols from biomass burning

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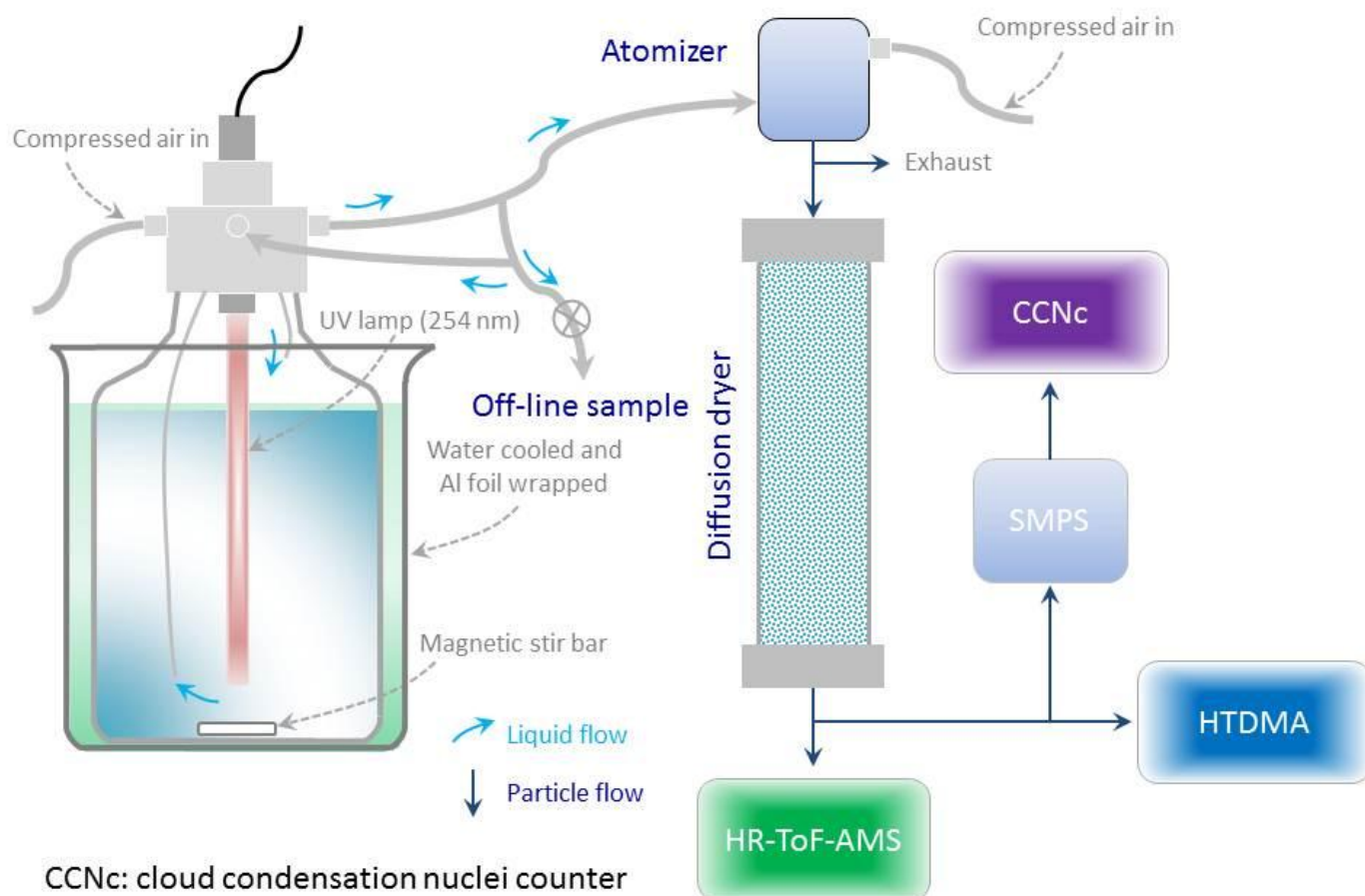[2]Department of Chemistry, University of Toronto, Toronto, Canada
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1 **1. Experimental setup**

2
3 Figure S1. Experimental setup for the aqueous-phase reactions.

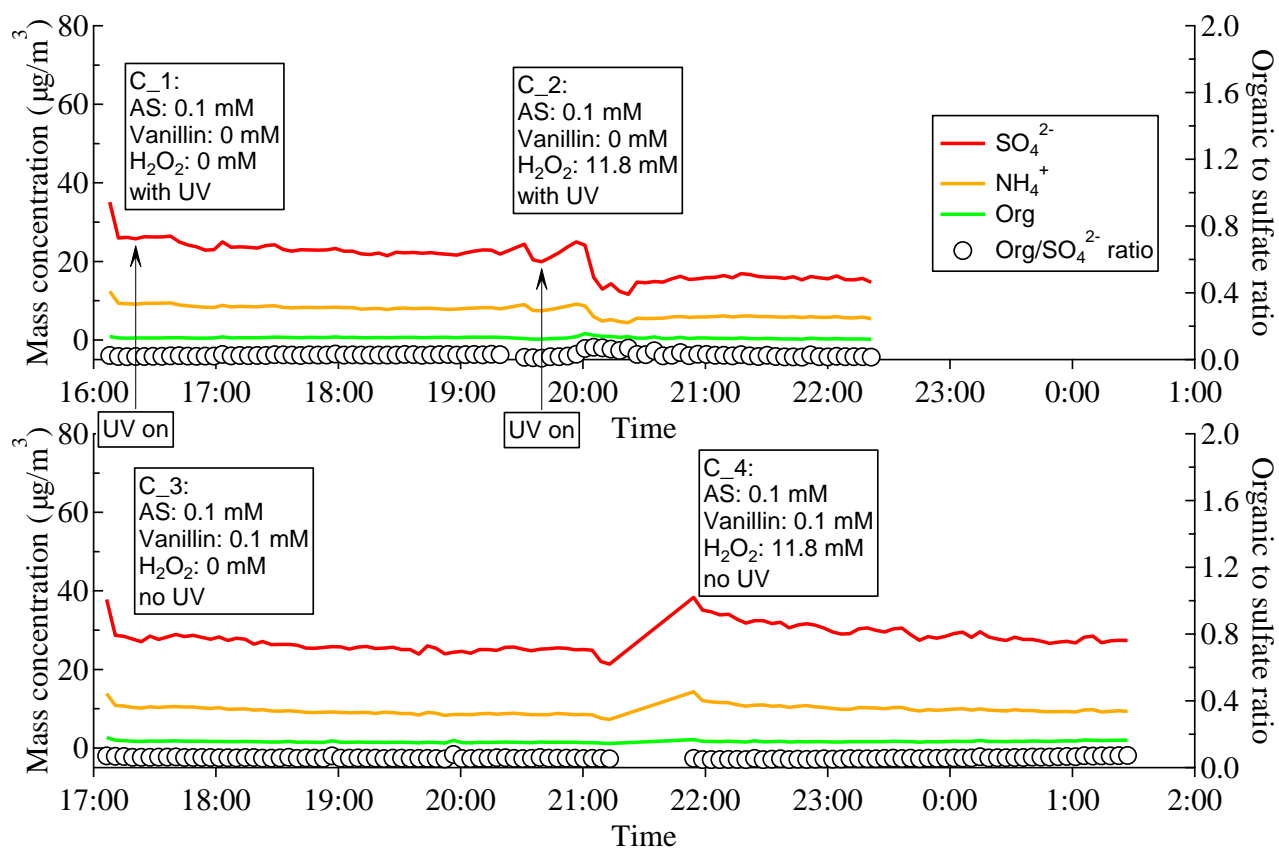1 2. **Examples of control experiments (C_1 to C_4)**

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3 Figure S2. Concentrations of species and ratios of organics to sulfate during control experiments C_1 to C_4
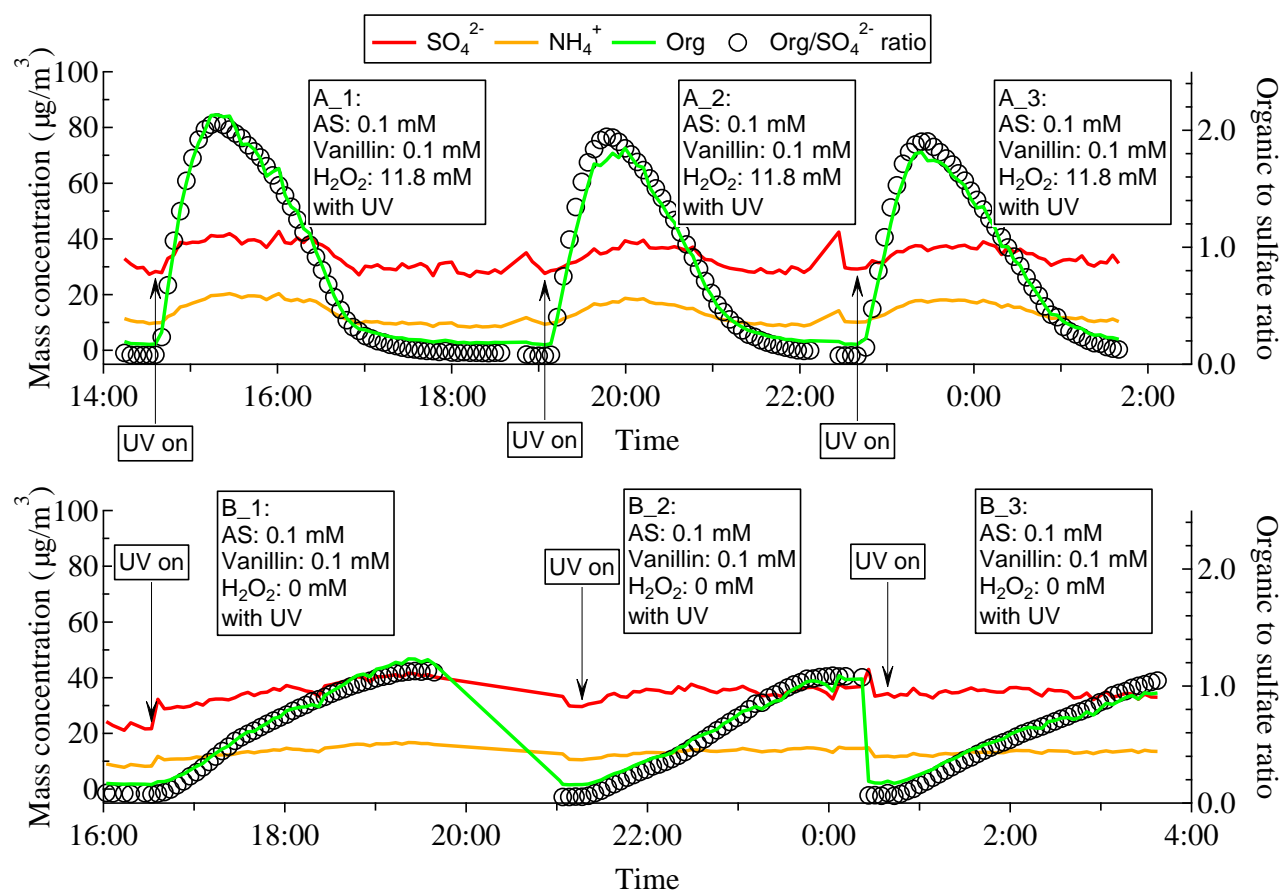
4 (Table 1).

5

Figure S3. Species concentrations and organics to sulfate ratios in experiments A_1 to A_3, and B_1 to B_3. The experimental conditions are specified in each box indicating the experiment number.

3. AMS analysis

3.1 Modification of the fragmentation table

In the default fragmentation table (Allan et al., 2004; Aiken et al., 2007), the signals of ions H_2O^+ and CO^+ are related to that of CO_2^+ with coefficients of 0.225 and 1.0, respectively. These coefficients work well for ambient data, but were found to be somewhat different in laboratory studies (Sun et al., 2010; Chen et al., 2011; Li et al., 2011). Several approaches can be used to obtain more representative coefficients for these ions (Li et al., 2011). Briefly, one can use the particle time-of-flight (pToF) data to exclude signal contribution from gaseous H_2O and CO (Zhang et al., 2005), and use high-resolution data from W mode to exclude interfering ions such as those from NH_4^+ (Li et al., 2011). For our experiments in which a dryer was used (little contribution from gaseous water) and no excess CO would be generated (as in burning experiments), the pToF correction was not necessary. But the correlations of high-resolution data would be helpful in determining the coefficients. Shown below in Figure S4 are those correlations in different runs of experiments. For H_2O^+ vs. CO_2^+, slopes close to unity were determined. But it is believed that approximately half of these H_2O^+ signals come from strongly bond particulate water, leaving the other half from organics (Chen et al., 2011). Therefore, a factor of 0.5 was applied to these slopes to obtain final coefficients between H_2O^+ and CO_2^+. For CO^+ vs. CO_2^+, the slopes were directly used as the coefficients. Table S1 summarizes those coefficients for different runs, with the default values also given for reference.
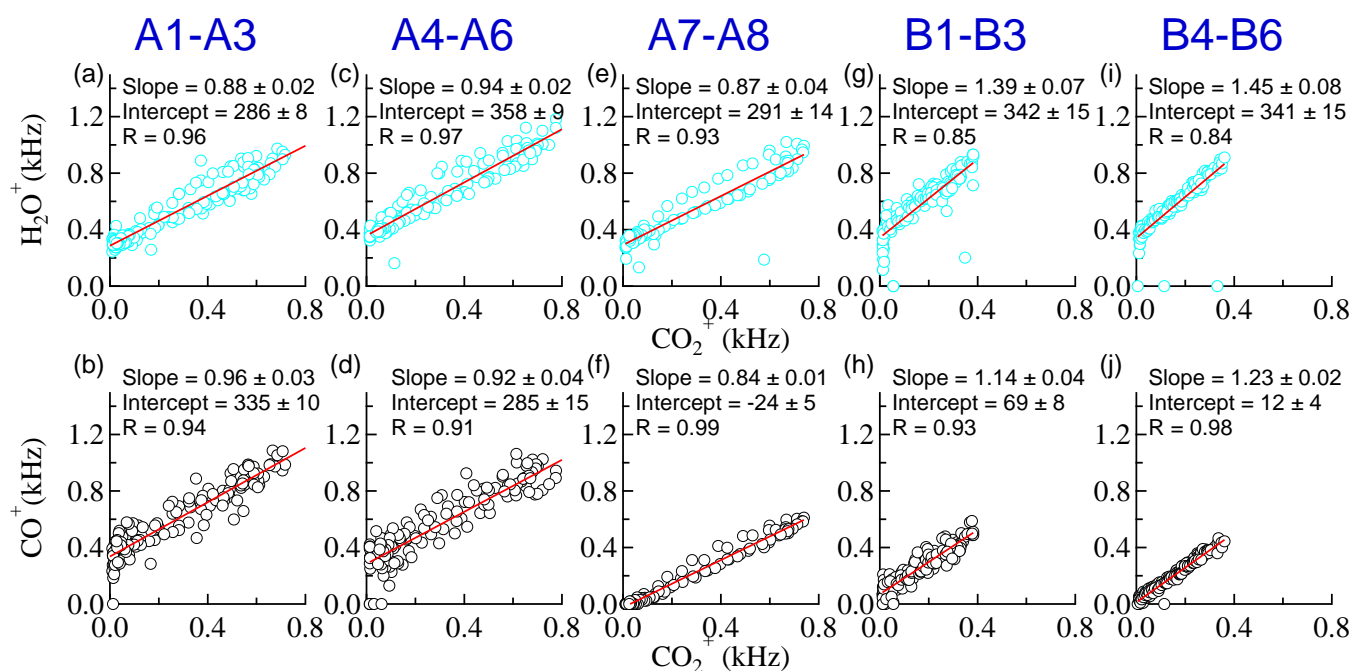

Figure S4. Relationships of ion signal intensities (kHz) of H_2O^+ vs. CO_2^+ and CO^+ vs. CO_2^+ from high-resolution mass spectra obtained in W mode.

1 Table S1. Coefficients of H_2O^+ and CO^+ signals as related to the CO_2^+ signal used in the fragmentation table.

2 The default values are also shown for reference. Experimental runs indicated by experiment numbers (e.g.,

3 A_1-A_3) are listed in Table 1.

Experiment	Default	A_1-A_3	A_4-A_6	A_7-A_8	B_1-B_3	B_4-B_6
Org18/Org44 or HROrgH$_2$O$^+$/HROrgCO$_2$$^+$	0.225	0.44	0.57	0.44	0.69	0.72
Org28/Org44 or HROrgCO$^+$/HROrgCO$_2$$^+$	1.00	0.96	0.92	0.82	1.14	1.23

4

5 3.2 Background subtraction

6 Trace amounts of impurities in water can contribute to organic mass spectra. Although their mass is

7 generally less than 5%, their contribution to specific organic ions in the mass spectra could interfere with

8 real signals from reaction products. For example, the peak at 137 as a tracer of the ring-retaining structure in

9 Figure 2 needs to be scrutinized to make sure it is not from the background. Thus, for mass spectra obtained

10 and shown in Figure 2, background subtraction was performed. Assuming the impurities also evolved during

11 the photochemical oxidation and direct photolysis as did the target reactant, time-specific background

12 subtraction was performed. For example, the mass spectrum obtained at 40 minutes in condition (A) was

13 obtained by subtracting the mass spectrum in experiment A_2 at 40 minutes from a mass spectrum in

14 condition (A) and that in experiment C_1 at 40 minutes from a mass spectrum in condition (A), and so on.

15

4. Estimation of the OH radical concentration

During the photochemical oxidation experiments, OH radical was generated by photolysis of H_2O_2. To determine the H_2O_2 decay rate during photolysis, the H_2O_2 concentration was measured by a colorimetric method (Allen et al., 1952). A calibration curve is shown in Figure S5.

The H_2O_2 decay rate was determined in "neutral" (ammonium sulfate, 0.1 mM) or acidic (ammonium bisulfate, 0.1 mM) conditions. The decay rate was relatively independent of pH, as suggested by Tan et al. (Tan et al., 2009). From these decay experiments, a decay rate (k_{obs}) of 1.9×10^{-4} s^{-1} was determined (Figure S6, green dotted curve), which is comparable to that in a previous study (see Table S2) that used the same brand and model of UV lamp (An et al., 2001). However, if this decay rate is taken as the photolysis rate constant (k_1, Table S2), the modeled H_2O_2 decay (see below) after taking into account reactions with other reactive oxygen species (such as OH, HO_2 radicals etc., reactions 2-7 in Table S2) would not match the experimental decay of H_2O_2 (see Figure S6, red curve). After a few iterations, it was determined that a photolysis rate constant (k_1) of 1.0×10^{-4} s^{-1} should be used instead (Figure S6, blue curve).

Rate constants (k_2 to k_6) for reactions of other reactive oxygen species are adopted from the literature as shown in Table S2. The rate constant (k_7) of the oxidation reaction of organics (Org) by OH radicals is assumed to be 1×10^8 M^{-1}s^{-1} (Lee et al., 2012). Reaction rate expressions were thus set up for these reactions. The concentrations of the species H_2O_2, HO_2, OH, and Org were then solved with the help of the POLYMATH (Willimantic, CT) ordinary differential equation (ODE) solver via two approaches: (1) assuming a pseudo-steady-state for OH radicals, and (2) solving stiff ODEs without a constraint on OH radicals.

For approach (1), the expression for OH concentration is shown in Eq. S1 and set to 0.

$$\frac{d[OH]}{dt} = 2 \times k1[H_2O_2] - k2[OH][H_2O_2] + k3[HO_2][H_2O_2] - k5[OH][HO_2] - 2 \times k6[OH][OH] - k7[Org][OH] = 0 \quad \text{Eq. S1}$$

Thus

$$2 \times k6[OH]^2 + (k2[H_2O_2] + k5[HO_2] + k7[Org]) \times [OH] - (2 \times k1[H_2O_2] + k3[HO_2][H_2O_2]) = 0 \quad \text{Eq. S2}$$

Solving this quadratic equation:

$$[OH] = \frac{-(k2[H_2O_2] + k5[HO_2] + k7[Org]) \pm \sqrt{(k2[H_2O_2] + k5[HO_2] + k7[Org])^2 + 4 \times 2 \times k6 \times (2 \times k1[H_2O_2] + k3[HO_2][H_2O_2])}}{2 \times 2 \times k6} \quad \text{Eq. S3}$$

The expressions of other species (H_2O_2, HO_2, and Org) are shown as below.

$$\frac{d[H_2O_2]}{dt} = -k1[H_2O_2] - k2[OH][H_2O_2] - k3[HO_2][H_2O_2] + k4[HO_2][HO_2] + k6[OH][OH] \quad \text{Eq. S4}$$

$$\frac{d[HO_2]}{dt} = k2[OH][H_2O_2] - k3[HO_2][H_2O_2] - 2 \times k4[HO_2][HO_2] - k5[OH][HO_2] \quad \text{Eq. S5}$$

$$\frac{d[Org]}{dt} = -k7[Org][OH] \quad \text{Eq. S6}$$

Replacing all the [OH] in Eq. S4 − Eq. S6 with the positive solution of the quadratic equation for the pseudo-steady-state assumption for OH radicals (Eq. S3), the concentrations of H_2O_2, HO_2, and Org as a function of time were obtained by solving the ODEs (Eq. S4 − Eq. S6). The OH concentration was then

7

back-calculated from Eq. S3. Initial concentrations of 10^{-3} to 10^{-6} M were tested for Org, and the OH concentration ranged from $6.5 - 7.8 \times 10^{-12}$ M. An OH radical concentration of 7.2×10^{-12} M was then chosen, with the Org concentration of 10^{-4} M (as the initial concentration of VL added).

For approach (2), Eq. S1 and Eq. S4 – Eq. S6 were included in the calculation without any constraint on OH radicals. The set of ODEs was stiff and thus an option of stiff ODE calculation was used in POLYMATH was used to solve the species concentrations of H_2O_2, HO_2, Org, as well as OH, as a function of time. OH concentration was estimated to be 7.0×10^{-12} M, showing little difference from that obtained by approach (1).

Figure S7 shows the OH concentrations estimated by approach (1) with a pseudo-steady-state assumption for OH radicals and approach (2) without any constraint on OH radicals. Two k_1 values were used for approach (1): the one determined from the H_2O_2 decay experiment ($k_1 = k_{obs} = 1.9 \times 10^{-4}$ s^{-1}, red curve in Figure S7), and (b) the one determined from the model ($k_1 = 1.0 \times 10^{-4}$ s^{-1}, blue curve in Figure S7) by taking into account other reactions (reactions 2-6 in Table S2). A k_1 value of 1.0×10^{-4} s^{-1} was used for approach (2) (green curve in Figure S7). With a k_1 value of 1.0×10^{-4} s^{-1}, the two approaches resulted in very similar and quite steady OH concentrations, while using the H_2O_2 decay constant as the photolysis constant leads to a factor of two overestimation of the OH radical concentration. An OH concentration of 7.0×10^{-12} M, which is one order of magnitude higher than that in cloud water, but close to that in wet aerosol particles (Ervens et al., 2013), was thus used.

By using the OH concentration of 7.0×10^{-12} M, the bimolecular reaction rate constant between VL and OH was determined to be 3.9×10^8 $M^{-1}s^{-1}$ by simply dividing the decay rate of VL in condition (A) (2.7×10^{-3} s^{-1}, see Figure 4) by the steady-state OH concentration (7.0×10^{-12} M). Alternatively, the same model with all the reactions in Table S2 was also run by varying k_7 (as the k_{OH} of organics, VL in this case), as in Figure S8. Regardless of which approach was used (OH concentration constrained or not), the modeled VL decay matched the measured one when k_7 was set to 4×10^8 $M^{-1}s^{-1}$. Therefore, the bimolecular reaction rate constant of VL and OH reaction (k_{OH}) was determined to be $\sim 4 \times 10^8$ $M^{-1}s^{-1}$.

Table S2. Reactions and rate constants for aqueous-phase reactions involving reactive oxygen species (H_2O_2, HO_2, OH, etc.) as well as organics.

| Reaction	Rate constant	References
1 $H_2O_2 + h\nu \rightarrow 2OH$	$k_1 = 1.1 \times 10^{-4}$ s^{-1}	Tan et al. 2009 (Heraeus Noblelight Hg lamp, 254 nm)
	$k_1 = 1.6 \times 10^{-4}$ s^{-1}	Perri et al. 2009 (Strahler Hg lamp, 254 nm)
	$k_{obs} = 1.9 \times 10^{-4}$ s^{-1} $k_1 = 1.0 \times 10^{-4}$ s^{-1}	This study (UVP Pen-Ray lamp, 254 nm)
	$k_{obs} = 2.0 \times 10^{-4}$ s^{-1}	An et al. 2001 (UVP Pen-Ray lamp, 254 nm)
2 $OH + H_2O_2 \rightarrow HO_2 + H_2O$	$k_2 = 2.7 \times 10^7$ M^{-1}s^{-1}	Sander and Crutzen, 1996; Warneck, 1999; Perri et al. 2009
3 $HO_2 + H_2O_2 \rightarrow OH + H_2O + O_2$	$k_3 = 3.7$ M^{-1}s^{-1}	Tan et al. 2009; Lee et al. 2012
4 $HO_2 + HO_2 \rightarrow H_2O_2 + O_2$	$k_4 = 9.7 \times 10^5$ M^{-1}s^{-1}	Sander and Crutzen, 1996; Warneck, 1999; Perri et al. 2009
5 $OH + HO_2 \rightarrow H_2O + O_2$	$k_5 = 7.1 \times 10^9$ M^{-1}s^{-1}	Sander and Crutzen, 1996; Warneck, 1999; Perri et al. 2009
6 $OH + OH \rightarrow H_2O_2$	$k_6 = 5.5 \times 10^9$ M^{-1}s^{-1}	Sander and Crutzen, 1996; Warneck, 1999; Perri et al. 2009
7 $Org + OH \rightarrow oxyOrg$	$k_7 = 1 \times 10^8$ M^{-1}s^{-1}	Lee et al. 2012

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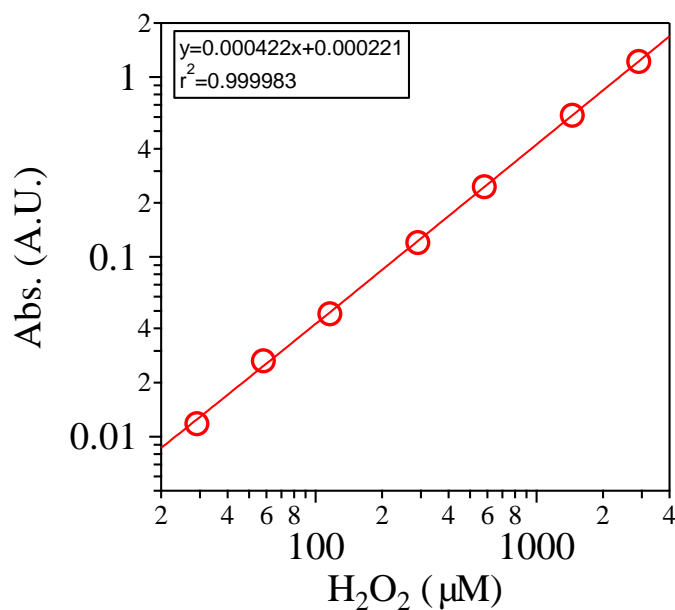
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3   Figure S5. A typical calibration curve for H_2O_2.
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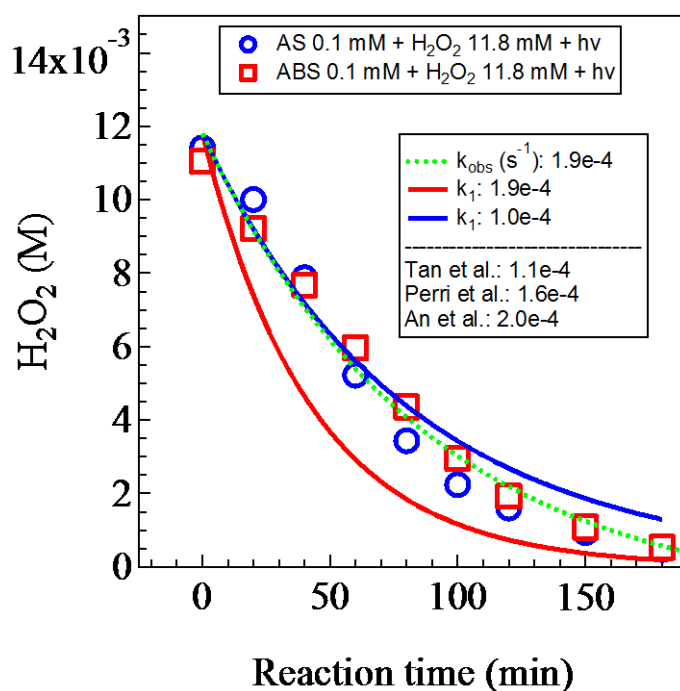
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6   Figure S6. H_2O_2 decay experiments under "neutral" (blue circles) and acidic (red squares) conditions. The
7   green dotted curve is the fitted (exponential decay) curve for data points under both "neutral" and acidic
8   conditions, resulting in a decay rate of $k_{obs} = 1.9 \times 10^{-4}$ s^{-1}. The red solid curve is the calculated H_2O_2 decay
9   from reactions 1-6 with $k_1 = k_{obs}$. The blue solid curve is the calculated H_2O_2 decay from reactions 1-6 with
10     $k_1 = 1.0 \times 10^{-4}$ s^{-1}.
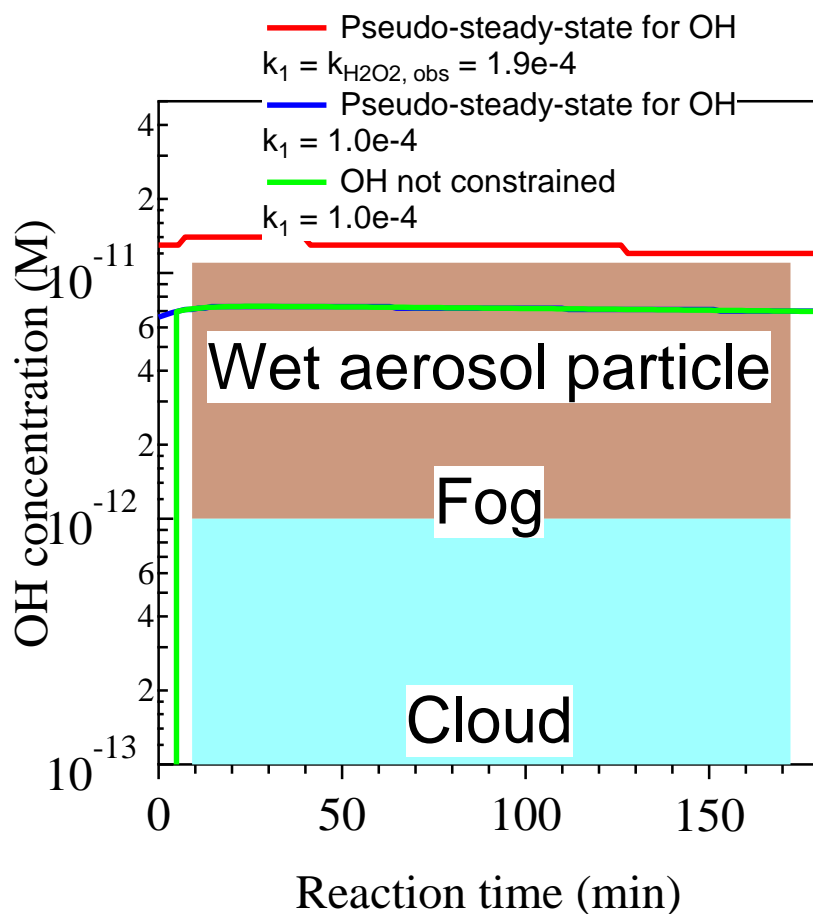
Figure S7. OH concentrations estimated from two approaches and with k_1 set to different values. Also shown are typical ranges of OH concentration under ambient conditions (Warneck, 2003; Ervens et al., 2011; Ervens et al., 2013).

Figure S8. VL decay modeled from two approaches: pseudo-steady-state for OH radicals (a), and no constraint on OH radicals (b). Different values of k_7 in Table S2 were tested in both approaches to see which one would give the experimentally observed VL decay (red circles).

5. Off-line analyses

5.1 Ultra-performance liquid chromatography with diode array detector (UPLC-DAD)
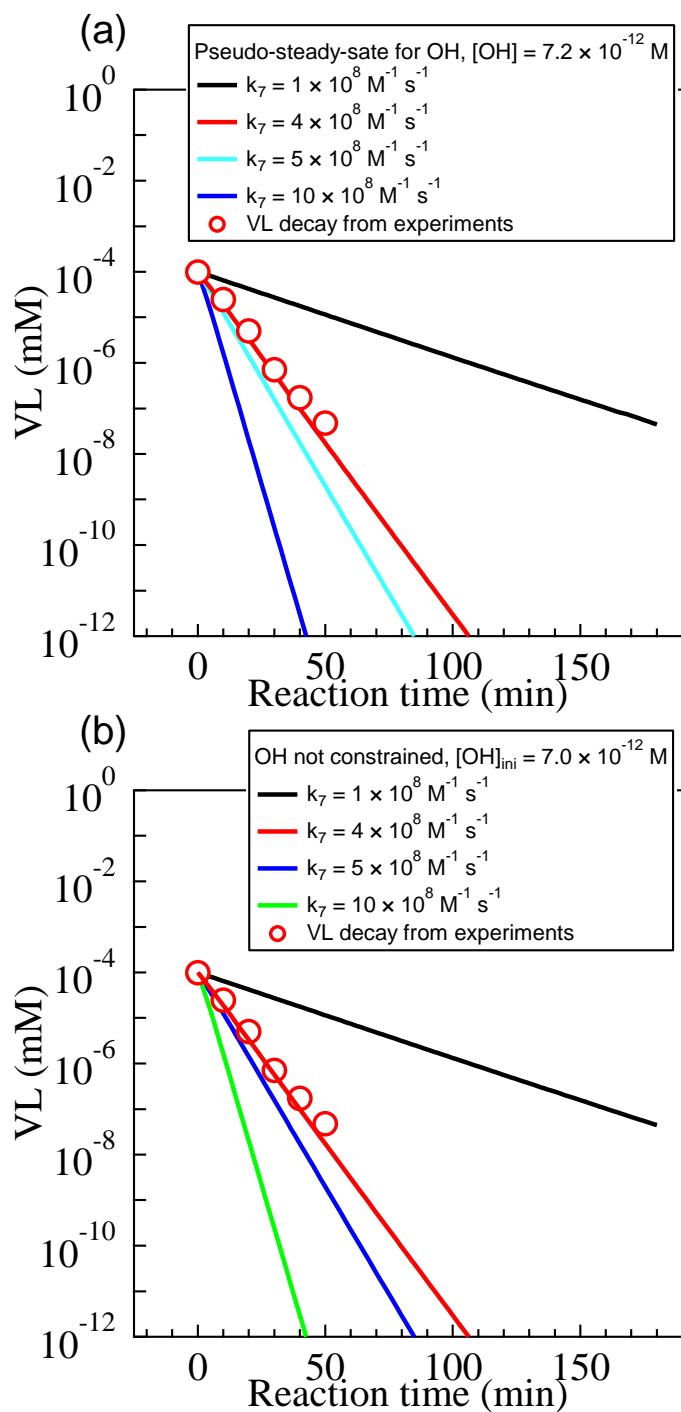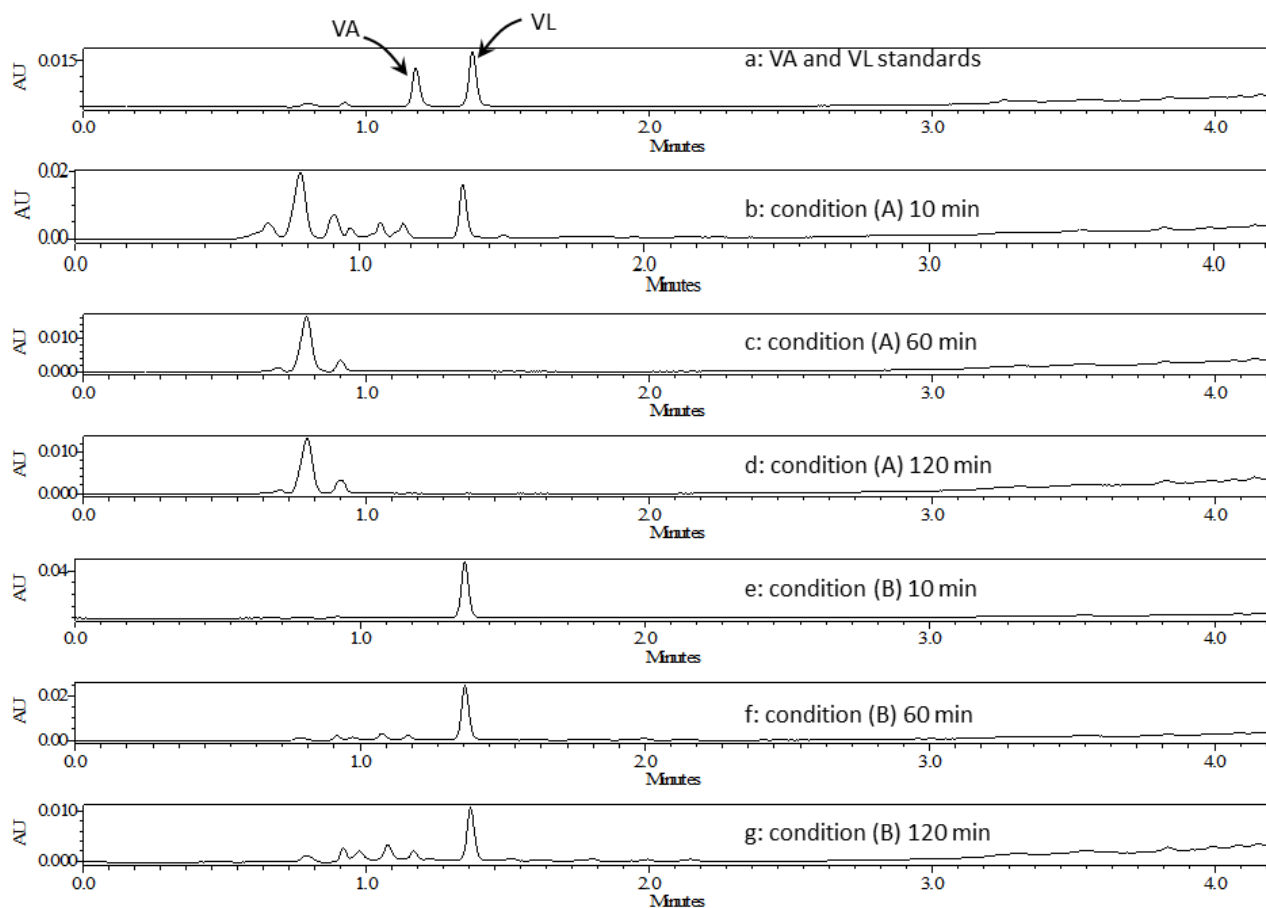
Aqueous samples at different time intervals from experiments A_1 to A_3, and those from B_1 to B_3 (see Table 1), as well as VL and vanillic acid (VA) aqueous solutions with known concentrations were analyzed by UPLC-DAD (Waters, Milford, MA). A Waters ACQUITY UPLC HSS C18 column (1.8 μg, 2.1 mm × 100 mm) was used. A solvent gradient with solvent A as 0.02 vol% trifluoroacetic acid (TFA) in water and solvent B as 0.02 vol% TFA in acetonitrile was used for the separation. The gradient was 35% B in the first 0.2 minutes, increased to 90% B in the next 2.8 minutes, stayed at 90% B for 1 minute, and then decreased to 35% B in the next 0.2 minutes. The flow rate was 0.45 mL/min. Two channels with UV absorption of 254 nm and 300 nm were used for the detection. VL was quantified with the 300 nm channel while VA was quantified with the 254 nm channel. Typical chromatograms (UV absorption at 300 nm) and calibration curves are shown in Figure S9 and Figure S10, respectively.

VA had a retention time of ~1.2 minutes, while VL had a retention time of ~1.4 minutes (Figure S9-a). Sample from condition (A) showed a strong VL peak at 10 minutes (Figure S9-b), but the peak became very small at later time intervals (Figures S9-c, and –d). There was a very small amount of VA at 10 minutes, and its peak at later time intervals became undetectable in 300-nm chromatograms and very small in 254-nm chromatograms (the wavelength for its quantification, not shown). There were a number of peaks with retention times shorter than 1.2 minutes, probably due to the formation of other products (e.g., small carboxylic acids). Since most of them were not well separated, they were not quantified. Samples from condition (B) showed a strong VL peak at 1.4 minutes throughout the course of the 3-hour experiments, in accordance with the slow reactions of VL under condition (B). Some small peaks at 1.2 minutes (VA) and at even shorter retention times were observed. They might also be those small carboxylic acids formed at later time intervals, although their amounts would be much smaller than in condition (A). This also supports the analysis of the O:C ratio (see main text Section 3.2), which concludes that the identified high-molecular-weight products (see discussion later on UPLC-ToF-MS) cannot explain the relatively high O:C ratio (~1.0 at the end of the experiments, see Figure 3-B) and certain amounts of small and highly oxygenated products were also formed later in the experiments under condition (B).

Figure S9. Typical chromatograms from the standards (a), samples in condition (A) at different time intervals (b-d), and samples in condition (B) at different time intervals (e-g).
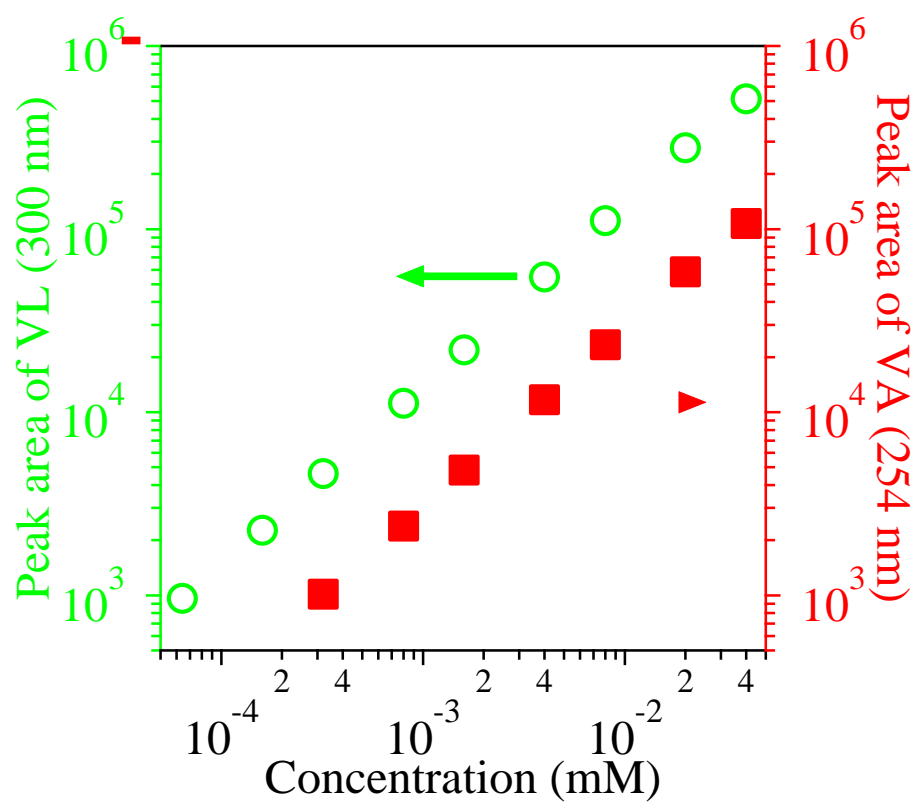
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2 Figure S10. Typical calibration curves for VL (green) and VA (red) using UPLC-DAD.

1 5.2 Ultra-performance liquid chromatography electrospray ionization time-of-flight mass

2   spectrometry (UPLC-ESI-ToF-MS)

3   Samples from A_9 and B_7 (Table 1) were collected at different time intervals. 10 μL of 1%

4 hydrogen peroxide catalase were added to each sample from A_9 to destroy unreacted H_2O_2. A

5 Waters ACQUITY UPLC system plus a Bruker Daltonics MiroOTOF MS with an electrospray

6 ionization (ESI) source was used for sample analysis. The UPLC column was a Waters ACQUITY

7 UPLC HSS C18 column (2.1 $\times$ 100 mm, 1.8 μm). Mobile phases were 0.1% acetic acid in water (A)

8 and 0.1% acetic acid in methanol (B). The gradient was 100% A in the first 2 minutes, decreased to

9 10% A in 28 minutes, stayed at 10% A until 42.2 minutes, and increased back to 100% A in 45

10 minutes. The flow rate was 0.2 mL/min. The ESI-MicrOTOF MS was operated under negative mode
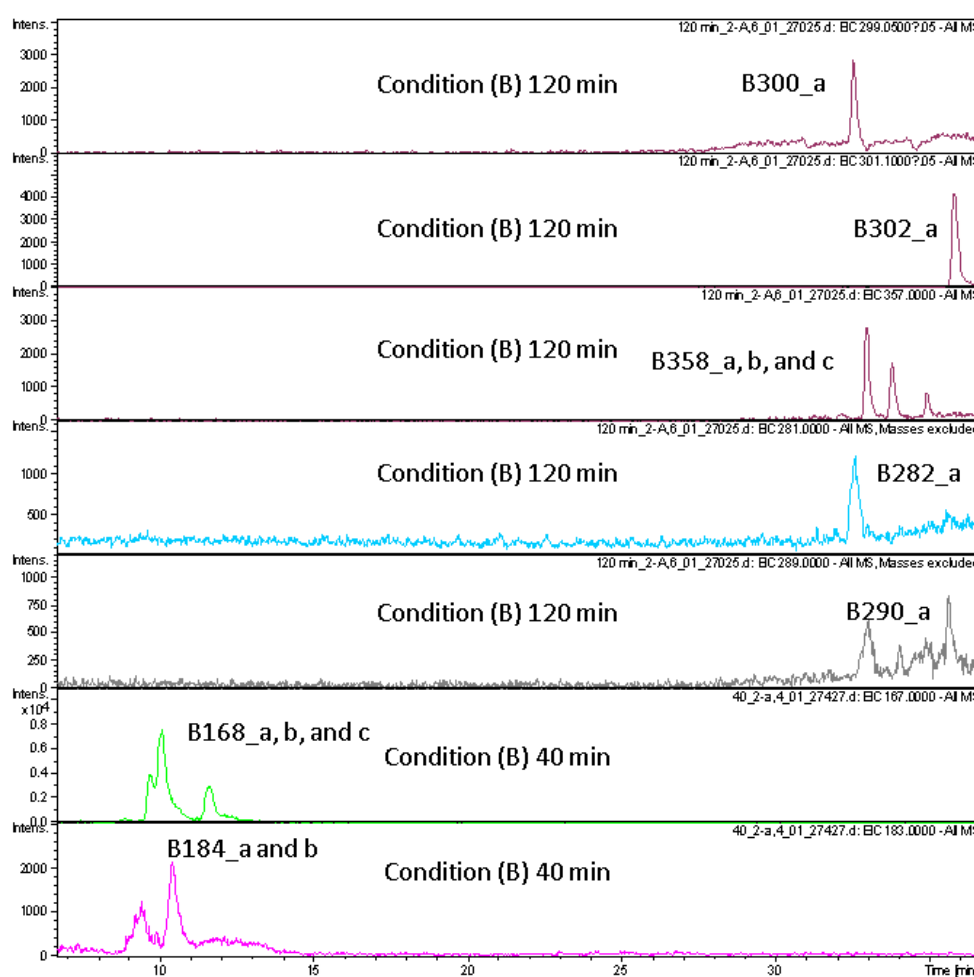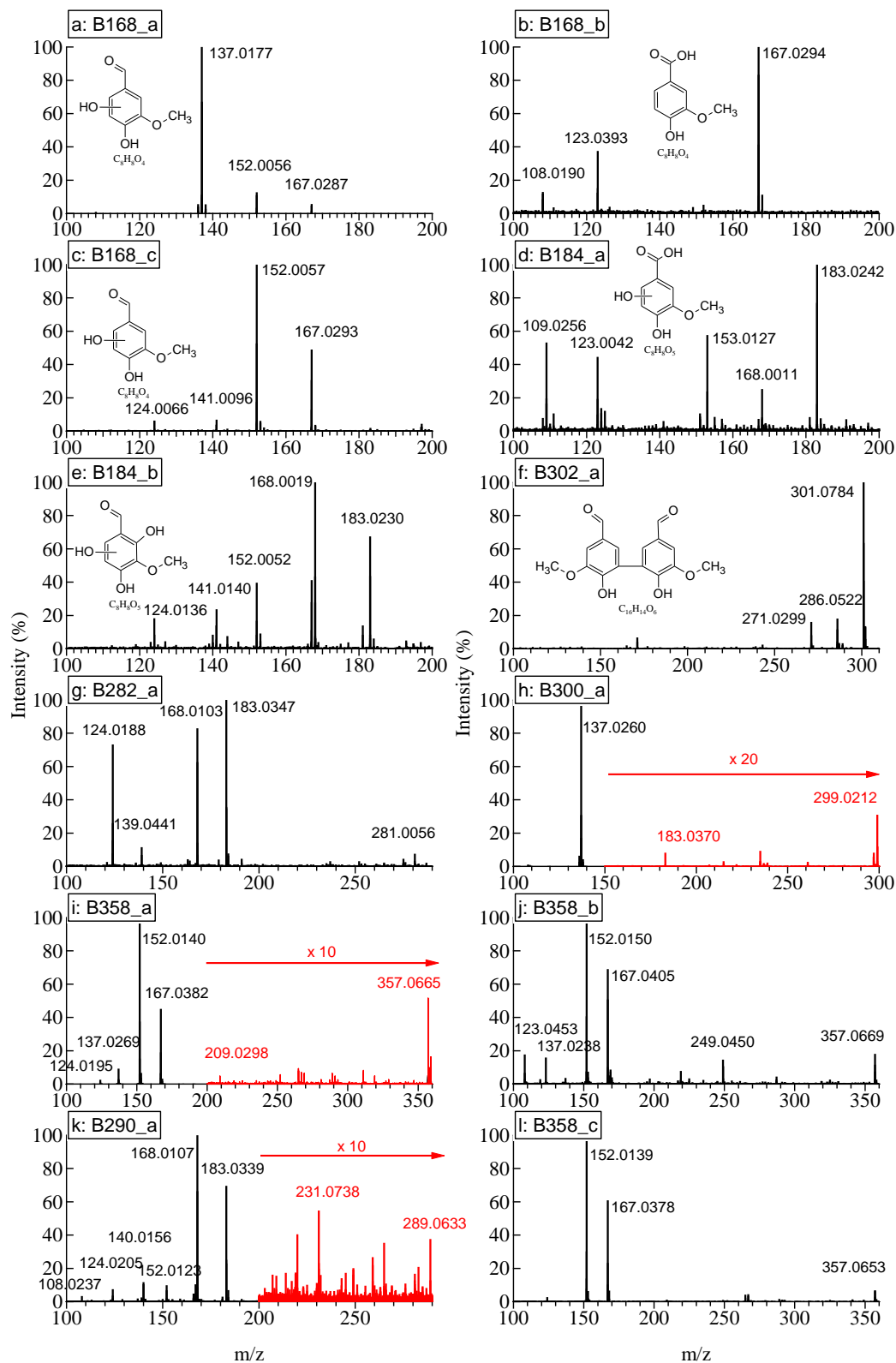
11 with a scanning *m/z* range of 50-1000.

12

13 Figure S11. Extracted ion chromatograms (EIC) for samples from condition (B).

16

Figure S12. Mass spectra of products identified by UPLC-ESI-ToF-MS.

Table S3. Formulas, double bond equivalency, major ions, and fragmentation routes proposed for products identified by UPLC-ESI-ToF-MS. Note: [a] DBE = double bond equivalency; [b] Δ = measured − calculated; [c] possible fragmentation routes.

Name	Formula	DBE[a]	Ion peak	Measured	Calculated	Δ^b (mDa)	Loss[c]
B168_a	$C_8H_8O_4$	4	$C_8H_7O_4^-$	167.0287	167.0344	-5.7	−H
			$C_7H_4O_4^-$	152.0056	152.0110	-5.4	−H−CH_3
			$C_7H_5O_3^-$	137.0177	137.0239	-6.2	−H−CH_2O
B168_b	$C_8H_8O_4$	4	$C_8H_7O_4^-$	167.0294	167.0344	-5.0	−H
			$C_7H_7O_2^-$	123.0393	123.0446	-5.3	−H−CO_2
			$C_6H_4O_2^-$	108.0190	108.0211	-2.1	−H−CH_3−CO_2
B168_c	$C_8H_8O_4$	4	$C_8H_7O_4^-$	167.0293	167.0344	-5.1	−H
			$C_7H_4O_4^-$	152.0057	152.0110	-5.3	−H−CH_3
			$C_6H_5O_4^-$	141.0096	141.0188	-9.2	−H−C_2H_2
			$C_6H_4O_3^-$	124.0066	124.0160	-9.4	−H−CH_3−CO
B184_a	$C_8H_8O_5$	4	$C_8H_7O_5^-$	183.0242	183.0293	-5.1	−H
			$C_7H_4O_5^-$	168.0011	168.0058	-4.7	−H−CH_3
			$C_7H_5O_4^-$	153.0127	153.0188	-6.1	−H−CH_2O
			$C_6H_3O_3^-$	123.0042	123.0082	-4.0	−H−$C_2H_4O_2$
			$C_6H_5O_2^-$	109.0256	109.0290	-3.4	−H−CH_2O−CO_2
B184_b	$C_8H_8O_5$	4	$C_8H_7O_5^-$	183.0230	183.0293	-6.3	−H
			$C_7H_4O_5^-$	168.0019	168.0058	-3.9	−H−CH_3
			$C_7H_4O_4^-$	152.0052	152.0110	-5.8	−H−CH_3O
			$C_6H_5O_4^-$	141.0140	141.0188	-4.8	−H−C_2H_2O
			$C_6H_4O_3^-$	124.0136	124.0160	-2.4	−H−CH_3O −CO
B302_a	$C_{16}H_{14}O_6$	10	$C_{16}H_{13}O_6^-$	301.0784	301.0712	7.2	−H
			$C_{15}H_{10}O_6^-$	286.0522	286.0477	4.5	−H−CH_3
			$C_{14}H_7O_6^-$	271.0299	271.0243	5.6	−H−CH_3−CH_3
B282_a	$C_{15}H_6O_6$	13	$C_{15}H_5O_6^-$	281.0056	281.0086	-3.0	−H
B300_a	$C_{15}H_8O_7$	12	$C_{15}H_7O_7^-$	299.0212	299.0192	2.0	−H
B358_a	$C_{16}H_{16}O_8$	9	$C_{16}H_{14}NaO_8^-$	357.0665	357.0586	7.9	−2H+Na^+
B358_b	$C_{16}H_{16}O_8$	9	$C_{16}H_{14}NaO_8^-$	357.0669	357.0586	8.3	−2H+Na^+
B290_a	$C_{15}H_{14}O_6$	9	$C_{15}H_{13}O_6^-$	289.0633	289.0712	-7.9	−H
B358_c	$C_{16}H_{16}O_8$	9	$C_{16}H_{14}NaO_8^-$	357.0653	357.0586	6.7	−2H+Na^+

5.3 Gas chromatography mass spectrometry (GC-MS) with derivatization

Using the BF$_3$/butanol method, samples from A_9 and B_7 or standard compounds (glyoxylic acid, oxalic acid, malonic acid and pyruvic acid, from Sigma-Aldrich) in aqueous phase were adjusted to pH = 8.5~9.5 using a 0.1 M NaOH (AR grade, Sigma-Aldrich) solution. The 5-mL samples were first concentrated to approximately 0.5 mL using a rotary evaporator under vacuum, then transferred to a 4-mL graduated conical vial and further concentrated to dryness under a nitrogen stream. A total volume of 0.25 mL of 10% BF$_3$/1-butanol (Supelco, Bellefonte, PA, USA) was immediately added to the sample. The samples and reagents were mixed with ultrasonication for 1 minute and then heated at 100 ℃ for 30 minutes to form butyl ester for the carboxyl group and acetal for the aldehyde group. Once the sample cooled down, the derivatives were extracted with hexane (95%, Mallinckrodt) (5 ×1 mL) after adding 0.3 mL of acetonitrile (ACN, 99.8%, Duksan) and 3 mL of deionized water. The fractions were combined and further washed with deionized water (3 × 3 mL). Then, the resulting solutions were dried over anhydrous Na$_2$SO$_4$ (AR grade, Sigma-Aldrich) and concentrated to about 250 μL by a nitrogen stream.

The resulting derivatives were injected into a GC-MS system (Perkin-Elmer Clarus 500) for analysis. The temperature program was set to a constant 40 ℃ for 4 minutes, increased to 200 ℃ at the rate of 8 ℃/min and held for 0 minutes, then further to 300 ℃ at 20 ℃/min and finally maintained at 300 ℃ for 1 minute. The injection volume was 10 μL. The MS data (total ion chromatogram) were acquired in the full scan mode (m/z of 50–600) using electron ionization. The solvent delay was set to 5 minutes.

Using the pentafluorophenylhydrazine (PFPH) method, the aqueous-phase samples from A_9 and B_7 or standard compounds (formaldehyde and glyoxal) were adjusted to pH=8.5~9.5 with the help of a 0.1 M NaOH solution. The 5-mL samples were first concentrated to approximately 1 mL using a rotary evaporator under vacuum, then transferred to a vial upon which 1 mL of 0.5 mM PFPH (Sigma-Aldrich, 97%) aqueous solution was immediately added. The mixtures were left to react at room temperature in the dark overnight. Later, 5 × 2 mL hexane was used to extract the derivatives, followed by drying with anhydrous Na$_2$SO$_4$ and filtering. The extract was concentrated to about 200 μL before the GC-MS analysis. The GC-MS conditions were the same as those when the BF$_3$/butanol method was used.
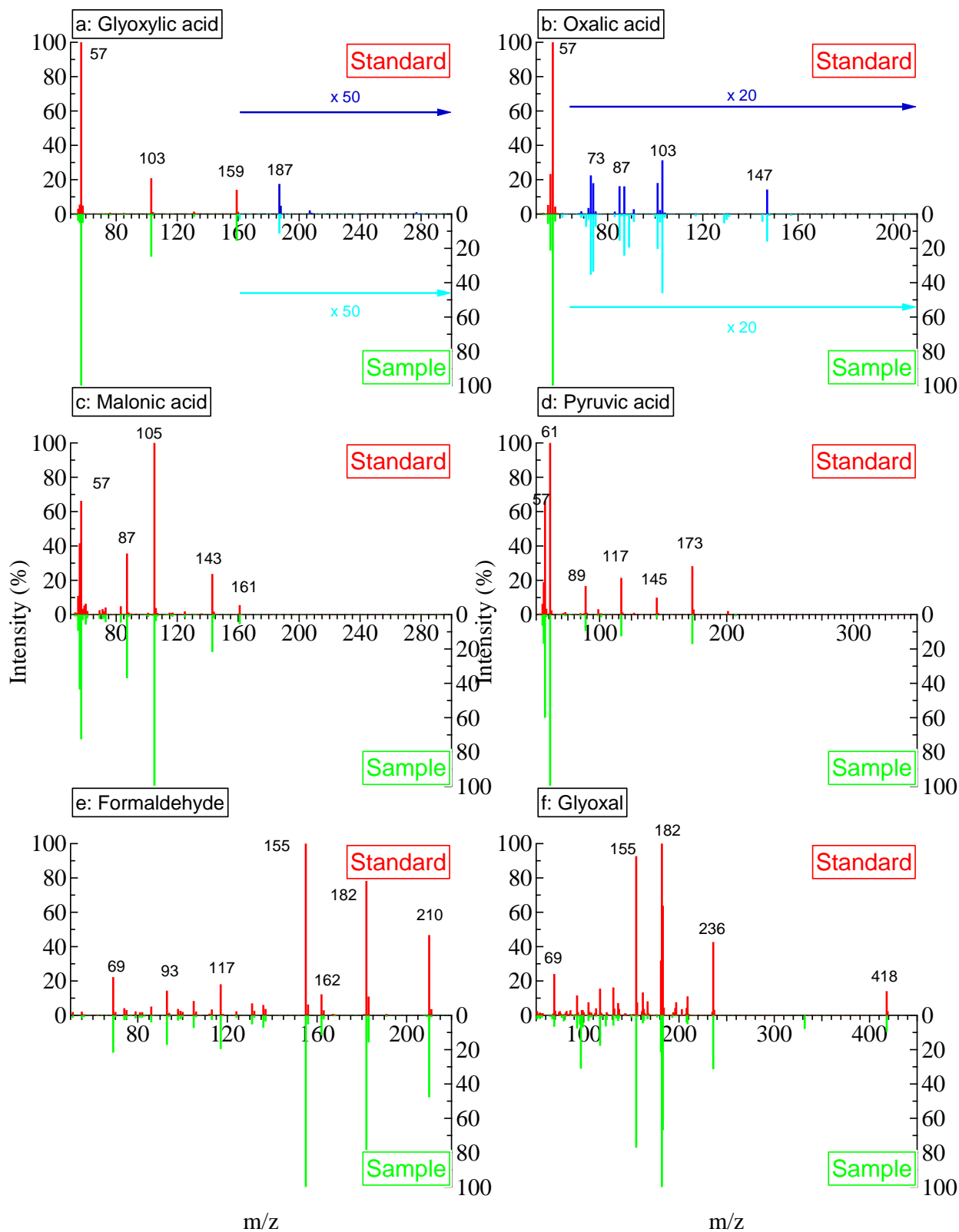
Figure S13. Mass spectra of standard compounds (red) and of products from the samples (green) identified by GC-MS. BF2/butanol method: a – d; PFPH method: e and f.

1 **6. Estimated organic growth factor at 90% RH**

2 The organic growth factors (GF_{org}) were calculated from the measured GF_{90} by assuming a

3 constant GF of ammonium sulfate (AS) of 1.68. They were calculated based on the ZSR rule

4 (Seinfeld and Pandis, 2006) as in Eq. 1 below, with GF_{org} being the GF of organics, GF_{AS} the GF of

5 AS, ε_{org} the volume fraction of organics, and ε_{AS} the volume fraction of AS.

6
$$GF_{org} = \sqrt[3]{\frac{GF_{90}^3 - GF_{AS}^3 \times \varepsilon_{AS}}{\varepsilon_{org}}} \qquad \text{Eq. S1}$$

7 The volume fraction of organics (ε_{org}) was calculated based on the mass fraction of organics

8 measured by the AMS, and the density estimated from H:C and O:C ratios (Kuwata et al., 2012) as

9 shown below.

10
$$\rho_{org} = 1000 \times \frac{12 + H:C + 16 \times O:C}{7.0 + 5.0 \times H:C + 4.15 \times O:C} \qquad \text{Eq. S2}$$

11
$$\varepsilon_{org} = \frac{\frac{f_{org}}{\rho_{org}}}{\left(\frac{f_{org}}{\rho_{org}} + \frac{f_{AS}}{\rho_{AS}}\right)} \qquad \text{Eq. S3}$$

12
$$\varepsilon_{AS} = 1 - \varepsilon_{org} \qquad \text{Eq. S4}$$

13 where ρ_{org} (kg/m^3) is the density of organics, H:C and O:C are the hydrogen-to-carbon ratio and the

14 oxygen-to-carbon ratio, respectively, of organics, f_{org} and f_{AS} are the mass fractions of

15 AMS-measured organics and sulfate, respectively, and $\rho_{AS} = 1769$ kg/m^3 is the density of ammonium

16 sulfate.

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