

**Spatial distribution of  
sulfur SRR in Asia**

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**Spatial distribution of the source-receptor  
relationship of sulfur in Northeast Asia**

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Received: 22 September 2010 – Accepted: 2 December 2010

– Published: 10 December 2010

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Published by Copernicus Publications on behalf of the European Geosciences Union.

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## Abstract

The spatial distribution of the source–receptor relationship (SRR) of sulfur over North-east Asia was examined using an off-line coupled meteorological/chemical transport model (MM5/RAQM). The simulation was conducted for the entire year of 2002. The results were evaluated using monitoring data for six remote stations of the Acid Deposition Monitoring Network in East Asia (EANET). The modeled SO<sub>2</sub> and O<sub>3</sub> concentrations agreed well with the observations quantitatively. The modeled aerosol and wet deposition fluxes of SO<sub>4</sub><sup>2-</sup> were underestimated by 30% and 50%, respectively, whereas the modeled precipitation was overestimated by 1.6 to 1.9 times. The domain was divided into 5 source–receptor regions: I, North China; II, Central China; III, South China; IV, South Korea; and V, Japan. The sulfur deposition in each receptor region amounted to about 50–75% of the emissions from the same region. The largest contribution to the deposition in each region was the domestic origin, accounting for 53–84%. The second largest contribution after the domestic origin was due to region II, supplying 14–43%, outside region II itself. The spatial distributions of the SRRs revealed that subregional values varied by about two times more than regional averages due to nonuniformity across the deposition fields. Examining the spatial distributions of the deposition fields was important for identifying subregional areas where the deposition was highest within a receptor region. The horizontal distribution changed substantially according to season.

## 1 Introduction

Sulfur oxides are one of the most important trace substances in the atmosphere because they cause environmental acidification and aerosol-induced climate changes. The East Asian region has been one of the largest emission source regions in the world (Streets et al., 2003; Ohara et al., 2007; Zhang et al., 2009). Since SO<sub>2</sub> emitted from anthropogenic sources and volcanic eruptions is oxidized in the atmosphere

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during long-range transport and forms  $\text{SO}_4^{2-}$ , which causes acidification of soil, land, water, and vegetation, the issue of transboundary air pollution evokes significant scientific interest and political concern (Carmichael et al., 2002). To better understand the trans-boundary problem and plan environmental policies, inter-political or inter-regional source–receptor relationship (SRR) analyses have been widely conducted (e.g., Carmichael et al., 2002; Lin et al., 2008; Nagashima et al., 2010). State-of-the-art techniques have been established to study SRRs of substances with highly non-linear photochemical chain reactions, such as those of  $\text{NO}_y$  (Lin et al., 2008) and  $\text{O}_3$  (Nagashima et al., 2010), and knowledge of SRRs in Asian regions has been accumulating (Lin et al., 2008).

For the accurate and reliable estimation of an SRR between regions and to plan the most effective strategy for environmental policies, the development of fair and accurate emissions inventories, the formulation of sophisticated elementary processes in three-dimensional chemical transport models, and the extensive and long-term monitoring of air pollutants using accurate instruments are required. In order to reduce uncertainties and to better estimate the impact of trans-boundary air pollution problems, the importance of the ensemble technique, which incorporates several chemical transport models with various implementations and formulations, has been pointed out in various model inter-comparison studies, such as MICS-Asia phase I (Carmichael et al., 2002), MICS-Asia phase II (Carmichael et al., 2008), the joint research on Long-range Transboundary Air Pollutants in Northeast Asia (LTP project) of the Tripartite Environment Ministers Meeting among Japan, China and Korea (TEMM) (Park et al., 2005; Kim et al., 2010).

Discussions of SRR analyses have usually been based on source–receptor tables showing the contributions of the sum of emissions in source regions relative to the aerial summation of deposition in receptor regions. However, due to the high nonuniformity of spatial distributions of deposition, spatial distributions of SRRs should be discussed together with SRR tables. Without discussions of spatial heterogeneity, we cannot identify the true contributions to deposition in some specific subregional areas.

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The SRR over the ocean was rarely focused in the previous studies. Though sulfur oxides may not be seriously hazardous compounds in the ocean environment, other species, emitted together with sulfur from coal combustion and diesel exhaust, such as polycyclic aromatic hydrocarbons (PAHs), are persistent and bioaccumulate, and they can have acute and chronic toxicity to marine life (Livingstone, 1998; Hylland, 2006; Suzuki et al., 2009).

Therefore, we present the horizontal distribution of the SRR of sulfur both over the land and the ocean in Northeast Asia as well as SRR tables using a three-dimensional chemical transport model, the Regional Air Quality Model (RAQM; An et al., 2002; Han, 2007). The RAQM model was developed in the Asian Center for Air Pollution research (ACAP), formerly the Acid Deposition and Oxidant Research Center (ADORC), the network center of the Acid Deposition Monitoring Network Center in East Asia (EANET). The model has been used for various air pollution studies and substantial modifications have been made based on comparison and evaluation with extensive and long-term monitoring data in the Asian region (An et al., 2002, 2003; Han et al., 2004, 2005, 2006, 2008; Kajino et al., 2004, 2005; Han, 2007). The model has been used in the MICS-Asia Phase II study (Carmichael et al., 2008) and the LTP project (Park et al., 2005). Verification of the simulation used in this study was made by Kim et al. (2010) by comparing the model results with observation data from the LTP stations and the results of the other LTP model participants. In this paper, the model results are also compared with EANET observation data. The model formulation and the SRR analysis method are described in Sect. 2. The model results are evaluated using the EANET monitoring data for the concentrations and deposition amounts of air pollutants in Sect. 3, and the spatial and seasonal variations of concentrations, deposition amounts, and SRRs of sulfur compounds are discussed in Sect. 4. Major findings are summarized in Sect. 5.

## 2 Methods

### 2.1 Off-line coupled meteorology-chemical transport model

A three-dimensional Eulerian model, the Regional Air Quality Model (RAQM; An et al., 2002; Han, 2007), which is built on a spherical and terrain-following coordinate system, was used. The model includes a series of major processes for chemical species that are found in the troposphere, such as advection, diffusion, dry deposition, multi-phase chemistry, cloud mixing, and scavenging. A simple but accurate mass conservative advection algorithm is applied to solve the mass conservation equation with a time-splitting technique (Walcek and Aleksic, 1998; Carmichael et al., 1991). The vertical eddy diffusivity is parameterized according to the approach of Byun and Dennis (1995), which has been used in the Regional Acid Deposition Model (RADM 2.6). The dry deposition module for gases is calculated by a modified Wesely's parameterization scheme (Walmsley and Wesely, 1996). For sulfate, the dry deposition velocity is parameterized according to the work of Walcek et al. (1986). Gas phase chemistry can be represented by either a condensed mechanism developed (He and Huang, 1996; An et al., 2002; Han et al., 2006) on the basis of Atkinson et al. (1986) or a complete version of the CB-IV mechanism (Gery et al., 1989), with an updated isoprene mechanism. The model ISSOROPIA1.7 (<http://nenes.eas.gatech.edu/ISORROPIA/>; Nenes et al., 1998) is incorporated into RAQM and applied to account for the partition or transformation between gas and aerosol phases. Subgrid-scale vertical transport, aqueous-phase chemical conversions, and scavenging processes are parameterized by using a one-dimensional dynamical and microphysical cloud model and an aqueous chemistry and scavenging box submodel, which are developed based on RADM (Chang et al., 1987). The photolysis rate is currently calculated according to Madronich (1987), with the effects of clouds and aerosols taken into account. Cloud information such as cloud top, cloud base, liquid water content, and cloud fraction are diagnosed using the same method as that used in RADM.

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The MM5 (Grell et al., 1994) version 3.7 (Fifth-Generation NCAR/Penn State Mesoscale Model) was used to provide meteorological fields for RAQM (wind, temperature, water mixing ratio, precipitation, and surface variables). The NCEP FNL reanalysis data with a 1.0-degree resolution was used to provide the initial and boundary conditions for meteorological fields. Four-dimensional data assimilation (nudging by three-dimensional reanalysis data) was utilized to improve model results, especially for the wind and temperature fields. Specific physical options, such as the MRF scheme (Hong and Pan, 1996) for PBL, Betts-Miller (Betts, 1986; Betts and Miller, 1986, 1993) for cumulus, RRTM for radiation, mixed-phase for explicit moisture, and five-layer soil model for land-air processes, were chosen for simulations. The domain for MM5 was larger than that of RAQM, with the center located at 35° N, 125° E. There were 125×95 grid numbers in a plane with a 45-km grid resolution, and 23 sigma layers from the surface to 100 hPa. The MM5 output was adequately interpolated from a Lambert projection to the spherical coordinates of RAQM.

## 2.2 Parameters used in the model simulation

A simulation was performed for the year 2002, from 1 January to 31 December, with a spin-up period of 3 days. March, July, and December were selected for the assessment of seasonality because they represent spring, summer, and winter, respectively. The model domain covered Northeast Asia (100–145° E, 20–50° N) with a horizontal resolution of 0.5° (thus, 90×60 grids, Fig. 1). In RAQM, 12 layers stretched vertically from the surface to 10 km (at about 50, 150, 300, 500, 750, 1500, 2500, 3500, 4500, 6000, 7500, and 8950 m). A 3-day initialization period prior to the formal simulation was carried out to bring each chemical species close to its actual state in the atmosphere. The initial and boundary conditions were taken as the lower end of observations available from recent studies for East Asia (Carmichael et al., 1998; Luo et al., 2000). Side boundary conditions were held fixed during the simulation, whereas the top boundary condition for O<sub>3</sub> was implemented as a constant level with seasonal variation. Topography data were obtained from the US Geological Survey Earth Resources Observation

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System (EROS) Data Center. The land use data came from DeFries and Townshend (1994), in which 12 categories of land use are classified.

Emission inventories of SO<sub>2</sub>, NO<sub>x</sub>, CO, NH<sub>3</sub>, and volatile organic compounds (VOCs) were derived from the common dataset used in the LTP project (Park et al., 2005; Kim et al., 2010). The emission data were available in a grid of 1°×1° and equally allocated to a 0.5° model grid for RAQM. Biogenic VOCs over China, including isoprene, monoterpenes, and other VOCs, were taken into account. Monthly values of biogenic VOC emissions were obtained for model simulations by multiplying annual amounts by temporal scaling factors. All area sources were assigned to the lowest model layer, about 50 m above the ground. In addition to the LTP standard emissions data, volcanic emissions were included from MICS-Asia emissions data (Carmichael et al., 2008), with a release height at an altitudinal range of 750–1500 m. In particular, Mt. Oyama on Miyakejima Island (139.531° E, 34.081° N, 813 m a.s.l.) has a high level of SO<sub>2</sub> emissions that have resulted in substantial environmental acidification in far East Asia (Satsumabayashi et al., 2004; Kajino et al., 2004, 2005). Mt. Oyama is located in the Northwest Pacific Ocean, 180 km south of the Tokyo metropolitan area, and it began to erupt on 8 July 2000. The Seismological and Volcanological Department of the Japan Meteorological Agency (SVD-JMA) has collected continuous measurements of the SO<sub>2</sub> emissions and smoke height above the crater since September 2000 (Kazahaya, 2001). It was reported that the daily amount of SO<sub>2</sub> emissions from Mt. Oyama in 2002 was approximately 9500 t day<sup>-1</sup> on average.

### 2.3 Source–receptor relationship method

Source–receptor relationships were calculated using the following formula:

$$R_{ij}(\%) = \frac{H_{ij}}{\sum_{i=1}^n H_{ij}} \times 100, \quad (1)$$

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where  $R_{ij}$  is the contribution of  $i$ -th emission source to the  $j$ -th receptor and  $H_{ij}$  is the deposition amount at the  $j$ -th receptor due to the  $i$ -th source. Contributions from the respective emission sources were obtained from the difference between the result when considering all of the emission sources and the result when the  $i$ -th source was excluded. This method is simple but useful for compounds involving less non-linearity in their photochemical chain reactions, such as sulfur compounds. In this study, the domain was divided into five source–receptor regions (Fig. 1).

## 2.4 EANET observation

We used the Acid Deposition Monitoring Network in East Asia (EANET) monitoring data (EANET, Guidelines for acid deposition monitoring in East Asia, available at <http://www.eanet.cc/product.html>) for model evaluation. The EANET stations monitor 1-day accumulated once per week (Korea) or 14-day (Japan) accumulated concentrations of gaseous species ( $\text{HNO}_3$ ,  $\text{HCl}$ ,  $\text{NH}_3$ , and  $\text{SO}_2$ ) and aerosol components ( $\text{SO}_4^{2-}$ ,  $\text{NO}_3^-$ ,  $\text{Cl}^-$ ,  $\text{NH}_4^+$ ,  $\text{Na}^+$ ,  $\text{Mg}^{2+}$ ,  $\text{K}^+$ , and  $\text{Ca}^{2+}$ ) using the filter pack (FP) method (EANET, Technical documents for filter pack method in East Asia, available at <http://www.eanet.cc/product.html>). They monitor daily accumulated concentrations in precipitation ( $\text{SO}_4^{2-}$ ,  $\text{NO}_3^-$ ,  $\text{Cl}^-$ ,  $\text{NH}_4^+$ ,  $\text{Na}^+$ ,  $\text{Mg}^{2+}$ ,  $\text{K}^+$ , and  $\text{Ca}^{2+}$ ) using precipitation collection and ion chromatography analysis (EANET, Technical manual for wet deposition monitoring in East Asia, available at <http://www.eanet.cc/product.html>) and hourly  $\text{SO}_2$ ,  $\text{NO}$ ,  $\text{NO}_2$ ,  $\text{NO}_x$ ,  $\text{O}_3$ ,  $\text{PM}_{2.5}$ , and  $\text{PM}_{10}$  concentrations and meteorological data such as wind speed, wind direction, temperature, relative humidity and solar radiation. Quality assurance and quality control (QA/QC) activities are conducted to ensure high quality of monitoring data according to the QA/QC program guidelines (available at <http://www.eanet.cc/product.html>). The FP method is associated with several distinct artifact problems during the long-term sampling period. However, the effects can be neglected in this study because we focus only on sulfur compounds that are not affected significantly compared to other semi-volatile components, such as ammonium, nitrate and chloride.

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Six remote stations located on small islands or isolated capes, depicted in Fig. 1 and listed in Table 1, were selected for comparison with simulation results for model evaluation. These stations were selected because without nearby huge emission sources and complexity of local orographically induced winds, air pollutant transport events mostly coincide with synoptic-scale disturbances and are generally well reproduced by regional-scale models.

### 3 Model evaluation using the EANET monitoring data

The RAQM model performance has been extensively evaluated by comparing to EANET monitoring data sets in previous studies. The model could reproduce monthly  $\text{NO}_3^-$  concentrations in rainwater (An et al., 2002), hourly  $\text{SO}_2$  and monthly  $\text{SO}_4^{2-}$  concentrations in rainwater affected by the Miyakejima Volcano (An et al., 2003; Kajino et al., 2004),  $\text{PM}_{2.5}$ ,  $\text{PM}_{10}$ , and the aerosol size distribution during dust events (Han et al., 2004), 3-hourly concentrations of aerosol  $\text{SO}_4^{2-}$  and  $\text{NH}_4^+$  (Kajino et al., 2004), hourly concentrations of  $\text{SO}_2$ ,  $\text{NO}_x$ , and  $\text{O}_3$  (Kajino et al., 2005; Han et al., 2004, 2006), wet deposition of  $\text{SO}_4^{2-}$ ,  $\text{NO}_3^-$ ,  $\text{NH}_4^+$ ,  $\text{Ca}^{2+}$ ,  $\text{H}^+$ , and the pH of rainwater (Han et al., 2006),  $\text{O}_3$  relevant trace species such as  $\text{NO}_x$  and VOCs measured for the TRACE-P project (Han, 2007), and all of those species along with a model inter-comparison study for MICS-Asia Phase II (Carmichael et al., 2008; Han et al., 2008 and references therein).

Table 2 summarizes statistical analyses for comparisons between daily and monthly observations and simulation data at the six EANET remote monitoring stations depicted in Fig. 1 and described in Table 1. The  $\text{SO}_2$  and  $\text{O}_3$  concentrations, precipitation, and wet deposition of  $\text{SO}_4^{2-}$  were measured at all of the stations, whereas aerosol  $\text{SO}_4^{2-}$  was measured at Jeju (1 day per week for the entire year), Oki (14 days in July 2002) and Rishiri (14 days in March 2002). The filter pack measurements were not yet started at the other stations during the period examined. The table shows the medians and averages of the observations and simulation data, the mean bias (MB)

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of simulations compared to observations, the root mean square errors (RMSEs), the correlation coefficients (Rs), and the FAC2s (number fraction of data which satisfy that the simulation and observation agree within a factor of 2) between simulations and observations.

5 Table 3 summarizes the emission amounts accumulated over the region for the emission inventory used in the study. Because the Asian region is one of the most serious emission source regions of atmospheric pollutants in the world, there have been various studies for the estimation of its emission inventory. For recent years, Streets et al. (2003), for the base year 2000, Ohara et al. (2007), for up to 2003, and Zhang et al. (2009), for the base year 2006, have frequently been used and widely credited. The  
10 estimated Chinese emissions from our emission inventory was 20.7 Tg SO<sub>2</sub>, which is comparable to Streets' inventory (20.4 Tg SO<sub>2</sub>) and relatively smaller than the newer estimates from Ohara's inventory (27.6 Tg SO<sub>2</sub>) and Zhang's inventory (22.9 Tg SO<sub>2</sub> for 2001).

15 In Table 2, the simulated SO<sub>2</sub> concentrations agree with observations quantitatively (median and average), whereas the correlation among them is not large. It is mainly due to the higher detection limit of SO<sub>2</sub> (0.1 ppbv) compared with concentrations measured at the stations (~0.3 ppb). The coarse horizontal resolution of our emission inventory (1°) might cause the relatively small correlation coefficients. It is coarser than  
20 other inventories (0.5°s) and the model grid resolution (60 km). However, it may not be critical because we have selected only remote stations, where we solely capture long-range transport phenomena. The distances between the monitoring stations and the huge emission source regions (the Asian continent) are ~1000 km. Big transport events usually last on the scale of several hours, shorter than 1 day. Though the daily based correlation coefficient is not very large, on an hourly basis, the measured and modeled high concentration peaks are observed to coincide in general. Taken together  
25 with the fact that the daily and monthly mean values are reasonable, the simulation performed successfully.

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Trends of simulated  $O_3$  concentrations were generally well reproduced ( $R=0.4$  to  $0.5$  and  $FAC2>0.9$ ), whereas the mean concentration was lower by 10 ppb. Due to the complexity of  $O_3$  photochemistry and its relatively long lifetime, the regional scale simulation of  $O_3$  is difficult compared with other primary species. It depends highly on the concentrations and speciation of VOCs and on lateral and upper boundary concentrations of the model domain regions, both of which are very much uncertain.

Trends in the simulated monthly mean concentration of aerosol  $SO_4^{2-}$  agreed well with the observations ( $R=0.80$ ). However, the median and mean values were substantially underestimated by about 30%. Atmospheric  $SO_2$  is dissolved efficiently into cloud and rain droplets with fast oxidation occurring in the aqueous phase. Some of the  $SO_4^{2-}$  produced in hydrometeors may remain in the atmosphere after the evaporation of clouds and some may be deposited to the ground surface by precipitation. The high uncertainty in the heterogeneous oxidation and wet scavenging processes contribute to either an increase or a decrease in particulate  $SO_4^{2-}$  concentrations. Heterogeneous  $SO_4^{2-}$  production on sea-salt (Chameides and Stelson, 1992) and dust (Tang et al., 2004) particles should promote the oxidation of sulfur, but these processes are missing in the model.

The simulated monthly amount of accumulated precipitation and wet deposition flux of  $SO_4^{2-}$  were compared with observations. The trend of modeled precipitation generally agreed well with observations ( $R=0.75$ ,  $FAC2=0.59$ ), whereas the model overestimated precipitation by 1.6 to 1.9 times. The modeled wet deposition of  $SO_4^{2-}$  generally underestimated the flux by about 50%, with some exceptions that were overestimations. As a result, the correlation coefficient ( $R$ ) between the modeled and observed data was low (0.29). The underestimations of modeled aerosol  $SO_4^{2-}$  concentrations in the atmosphere ( $\mu\text{g}/\text{m}^{-3}$  air) and in precipitation ( $\mu\text{g}/\text{m}^{-3}$  water; wet deposition flux divided by precipitation rate) are similar to each other. We found that the underestimation of the modeled production rate of  $SO_4^{2-}$  is a main cause for the quantitative discrepancy, whereas the modeled wet scavenging rate is reasonable. At the current stage, however, it is very difficult to identify the reasons for all of the discrepancies because all

of the uncertainties in emissions, gas-phase and aqueous-phase oxidation rates, the dry deposition velocity, and the wet scavenging ratio can together contribute to substantial increases and decreases in the model results. However, it is still essential to evaluate and show the discrepancies in these variables because a discussion of the discrepancies is directly transferable to the uncertainty and reliability of the main topic of this paper, the source–receptor relationship of sulfur and its horizontal and seasonal distribution.

## 4 Results and discussion

### 4.1 Annual mean and seasonal trends in surface meteorology over Northeast Asia

Figure 2a illustrates the mean annual surface wind field and accumulated precipitation for 2002. China, Korea, and Japan are located over the baroclinic zone in the mid-latitudes, where the temperature gradient is steep and synoptic disturbances are active as a driving force of transboundary air pollution. Throughout the year, northerly to northwesterly wind was prevailing over the Yellow Sea and the Sea of Japan. The contribution of those winds to transboundary air pollution in Northeast Asia was substantial. Large amounts of precipitation were observed in the southern areas of the domain, and those areas are mostly distributed in the Pacific Ocean. Inland areas of heavy local precipitation are located in the southern part of China (region III). The precipitation was due to local cumulus convection and mostly happened in summer (June and July). Orographically forced local heavy precipitation near the coastal regions can be seen along the northern and northeastern coast of the Japanese main island and the eastern coast of Taiwan.

Figure 2b–d illustrates the spatial distributions of the monthly mean surface wind fields and accumulated precipitation for (b) March, (c) July, and (d) December, which were selected to assess seasonality over the Northeastern Asian regions in the spring,

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summer, and winter, respectively. The features in autumn are not discussed throughout the paper because they are similar to those in spring. In spring (March), shown in Fig. 2b, westerly and northwesterly winds prevailed in the northern part of the domain over both the continent and the ocean. In contrast, easterly trade winds prevailed in the subtropical regions. Wind is weak in the central and southern part of China. In summer (July), shown in Fig. 2c, southerly and southeasterly winds prevailed over the Pacific Ocean due to the Pacific high pressure system (Pacific High). Transport due to the mid-latitude westerlies in the boundary layers was not very influential in the summer season. Heavy precipitation due to local cumulus convection may substantially affect the wet deposition of air pollutants in the southern part of the region. Precipitation was large only over the northern part of the Asian continent, the Yellow Sea, and the Sea of Japan in summer. In winter (December), shown in Fig. 2d, northerly monsoons are predominant over the Yellow Sea, the East China Sea, and the Sea of Japan. Northerly winds also prevailed over the continent.

## 4.2 Annual mean and seasonal trends of surface concentrations and deposition of sulfur oxides

Figure 3 illustrates the (a) annual and (b–d) monthly accumulated dry deposition of total S ( $\text{mg S/m}^2$ ). Among dry deposition of S ( $\text{SO}_2$  and aerosol  $\text{SO}_4^{2-}$ ),  $\text{SO}_2$  was predominant and contributed more than 80% over most of the regions such as the North China Plain, the Yangtze Plain, urbanized places in Korea and Japan, the Yellow Sea, the East China Sea, and the Sea of Japan. The contribution becomes smaller over further downwind regions or regions with heavy precipitation, such as the Northwest Pacific Ocean (40–80%), the northern coastal region of the Japanese main island (30–50%), and the southern part of China (30–50%), but the deposition amount of total S is much smaller ( $<200 \text{ mg S/m}^2$  annually) in those regions compared to that in the urbanized regions. Thus, the spatial distribution of total dry S deposition reflects those for the dry deposition of  $\text{SO}_2$  and surface concentrations of  $\text{SO}_2$ .

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In Sect. 3, it is discussed that modeled  $\text{SO}_2$  concentrations agreed quantitatively with the observations, whereas aerosol  $\text{SO}_4^{2-}$  was underestimated by 30%. Considering this discrepancy, however, the main finding that dry deposition of  $\text{SO}_2$  is predominant, especially near source regions, would be unchanged.

In Fig. 3, high deposition, and thus concentrations, of  $\text{SO}_2$  were observed very close to the large emission sources. Due to the activity of the Miyakejima Volcano, peaks of sulfur concentrations were observed near this area ( $139.53^\circ\text{E}$ ,  $34.081^\circ\text{N}$ ). Among the four seasons,  $\text{SO}_2$  concentrations were lower in the summer due to fast photochemical conversion of  $\text{SO}_2$  to sulfate and precipitation scavenging due to the high temperatures, large solar radiations, and high cloud activities. Cloud activities influence the heterogeneous production of sulfate because most of the conversion from  $\text{SO}_2$  to sulfate takes place in cloud and rain droplets. As a result,  $\text{SO}_2$  concentrations in the southern part of China were very low in July (not significant in deposition fields). Turbulent mixing affects surface concentrations in large emission source regions substantially. During the daytime, when the mixing layer is developed, air pollutants are transported (diffused) upward and the concentration within the layer becomes lower. The mixing layer height becomes larger in the summer and smaller in the winter over the inland regions due to surface heating and subsequent buoyancy effects. The effect is seen in the dry deposition field (Fig. 3b–d).

The  $\text{SO}_2$  concentration is largest in winter over the northern part of China and its downstream regions because  $\text{SO}_2$  emissions are highest in the winter due to the high coal consumption for domestic heating (Streets et al., 2003). However, this effect is not included in the simulations as our emission inventory only estimated the annual mean. The emission seasonality will certainly alter the deposition amounts and the source–receptor relationship. The effect could be estimated from this simulation’s results because it is almost linear due to the simple atmospheric reactions for sulfur compounds (i.e., when the emissions of one region are uniformly two times greater, the effects from the region on deposition in other regions can be estimated as about two times greater as well). The volcanic eruptions from Miyakejima have substantially

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decreased in recent years and are about 1000 t per day, 10 times smaller than those in 2002 (August, 2009; recent monitoring data provided by K. Kazahaya using the same method as in Kazahaya et al., 2001) and almost of the same order as other volcanoes in East Asia. The effect of this rapid change in volcanic emissions can also be roughly estimated in the SRR analysis in the same manner as the seasonal variations of emissions were assessed.

Figure 4 illustrates the (a) annual and (b–d) monthly accumulated wet deposition of total S in  $\text{mg S/m}^2$ . The wet deposition of S includes both  $\text{SO}_2$  dissolved into cloud and rainwater droplets and aerosol  $\text{SO}_4^{2-}$  activated as cloud condensation nuclei (CCN) and scavenged by rainwater. Both are observed as the  $\text{SO}_4^{2-}$  ion in the EANET rainwater monitoring samples. The wet deposition amounts appear to be smaller than the dry deposition amounts over the Northeastern Asian region. However, as the modeled  $\text{SO}_4^{2-}$  wet deposition was found to be underestimated, as discussed in Sect. 3, the wet deposition amounts could be comparable to the dry deposition amounts.

Wet deposition depends not only on surface concentrations but also on the concentrations in the entire column up to the cloud top and the water droplet mixing ratios. Wet deposition was larger in the polluted and heavy precipitation regions. The large amount of wet deposition of  $\text{SO}_4^{2-}$  over the southern part of China in summer was related to cumulus precipitation. As precipitation was larger in summer over northern parts of the Asian continent, the amount of wet deposition was larger in summer. Substantial amounts of deposition are found over the ocean for the wet deposition field (Fig. 4a), whereas the amount of dry depositions is relatively smaller over the ocean (Fig. 3a).

### 4.3 Total sulfur deposition and the source–receptor relationship

The annually accumulated SRRs of sulfur compounds are listed in Table 4. The upper part of Table 4 indicates the amounts of annually accumulated dry and wet deposition of sulfur at each receptor region (left panel of Fig. 1) and the oceanic receptor regions (right panel of Fig. 1) in 2002. In the lower part of Table 4, the annual contributions

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of sources (rows) of total S ( $\text{SO}_2$  plus  $\text{SO}_4^{2-}$ ) deposition to land and ocean receptor regions (columns) are given. The impacts of  $\text{SO}_x$  emissions from one source region to total S deposition in the receptor regions are described. As described before, regions I, II, III, IV, and V indicate North China (north of  $40^\circ$  N), Central China (between  $40$  and  $30^\circ$  N), South China (south of  $30^\circ$  N), South Korea, and Japan, respectively.

The annual deposition of sulfur was large in regions II and III (3160 and 2560 kt S/yr, respectively), where  $\text{SO}_x$  emissions were large as well (6020 and 3300 kt S/yr, respectively). The deposition of S in regions II and III accounted for about 50% and 75% of the emissions of S, respectively. Dry deposition was larger in region II, whereas wet deposition was predominant in region III. The domestic contribution for SRR (contribution of one source region to the same receptor region) in regions II and III were the largest among the 5 regions (83% and 84%, respectively), and the second largest contributions to each were from each other's region (14%). Thus, the two regions exchanged sulfur with each other (together accounting for 97% for region II and 98% for region III), mainly via dry deposition to region II and wet deposition to region III. The contributions to regions II and III from the other regions (I, IV, and V) were smaller than 2%.

The deposition of S in region I (623 kt S) was about 60% of the emission amount in the region (1010 kt S), but more than 40% of the S deposition was due to emissions from region II. The contributions of dry and wet deposition were comparable in the region. The contributions from the other regions (III, IV, and V) were less than 2%.

The deposition of S in region IV (273 kt S) was about 50% of the emission amount in the region (558 kt S), with the domestic contribution being the largest (66%). The second largest contribution was due to emissions from region II (18%). The contribution of dry deposition was somewhat larger than that of wet deposition. The contributions from the other regions (I, III and V) were not negligible (4, 6, and 6%, respectively).

The deposition of S in region V (1215 kt S) was about three times larger than the "anthropogenic emissions" from the region (455 kt S) and was equal to 56% of the total emissions (i.e., anthropogenic and the Miyake volcanic emissions, 2188.8 kt S). The domestic contribution was 65%, and the second largest contribution was due to region

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II (19%). Although a substantial amount of the deposition was from emissions from the Miyakejima Volcano in region V, the contribution of region II to region V was as large as that to region IV. Because the Miyakejima Volcano is located in the east of Japan, most of its volcanic sulfur was transported eastward due to westerlies and northwest-  
erlies, except in summer when the Pacific High is predominant and the volcanic plume was transported west to southwestward (Kajino et al., 2004). As a result, the largest contributor to the Northwest Pacific Region (NWP) was region V (54%).

#### 4.4 Total sulfur deposition to the ocean and the contributions of the source regions

Total S deposition to the ocean was assessed in this section. The Yellow Sea and the East China Sea (YEC) are located in upstream regions of the Sea of Japan (SJ), and atmospheric deposition over these oceanic areas is the cause of oceanic pollution over the Northeast Asian region. An examination of the spatial distribution of SRRs of sulfur over these oceanic regions may give some indication of the deposition of the other hazardous compounds such as PAHs.

The amounts of total sulfur deposition to the YEC, SJ, and NWP were 230, 340, and 510 kt S, respectively, which were comparable to or somewhat smaller than deposition amounts over the land receptor regions. It should be noted that the receptor regions and the oceanic regions are not completely separated, and they sometimes overlap. Wet deposition was more predominant than dry deposition over all of the ocean regions. The largest contributor to YEC was region II (53%), and the second and third largest were region III (23%) and region IV (10%). The largest contributor to SJ was also region II (40%), and the second largest was region V (23%), with regions I and IV (16%) next in contribution size. As discussed before, the largest contributor to NWP was region V, which accounted for 54% of the total deposition over the region, mostly of Miyake volcanic origin. The second and third largest were region II (24%) and region III (14%).

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## 4.5 Spatial distributions of total sulfur deposition and the source–receptor relationship and its seasonal variation

Figures 5 to 9 illustrate the horizontal distributions of (a) annual and (b–d) monthly mean contributions of the respective source regions (I to V) to total S deposition (%).

5 The horizontal features are very different than those in the SRR table. They give additional information on SRRs, such as true contributions for specific small subregional areas. Area averages of each receptor region and each oceanic region correspond to the values in Table 4. The solid lines in Figs. 5a, 6a, 7a, 8a, and 9a indicate 40° N and 30° N, the border between regions I and II and that between regions II and III, respectively.

10 region I contributed only a few percent as a source to other receptor regions (Fig. 5). Though the regional average of the contributions from region I to regions IV and V are less than 5%, the region I contributions exceeded 10% in the northern part of regions IV and V (Fig. 5a). As the wind pattern and precipitation vary substantially by season in Northeast Asia, the SRRs vary accordingly. Figure 5 shows that the contributions from sources from region I were similar in March and December, and the values were about 5 to 20% in the northern part of South Korea, the coastal regions of Japan's main island facing the Sea of Japan, and Hokkaido Island. The contribution in July over the area south of 40° N was smaller than 5% due to the Pacific High, and the amount of wet deposition was substantial.

20 Region II was the most influential source to all regions due to its large emissions and its location under the strong influences of mid-latitude disturbances and mid-latitude westerlies (Fig. 6). The contribution of regions II to every other region (outside of region II) exceeded 14%. It was smaller than 20% in regions III, IV, and V. The regionally averaged contribution from region II to region I was 43% (Table 4), with a maximum larger than 60% near the border (Fig. 6a). Region V included large areas of ocean surface. The overall regionally averaged contribution from region II to region V was 19%. The contribution over the coastal regions of the southwestern islands of the Japan

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archipelago exceeded 30%, and that over coastal regions of the main island facing the Sea of Japan was about 20 to 30%. Seasonal variations of the contribution from this major contributor region, region II, to the receptor regions were not very clear in the spatial distribution. A clear seasonal change was found over the ocean, especially for SJ. The contribution to the northern part of SJ was large in terms of its annual mean, spring, and summer (>60%). In winter, however, the northerly monsoon due to the Siberian High prevented the transport of air pollutants from region II northward.

Region III was the second major contributing source region for trans-regional transport, but the contributions were not very large, generally smaller than 10% to the other regions, such as I, IV, and V (Fig. 7). The southern part of region III is under the influence of the trade wind zone, and the contribution to eastern regions may not be large. Averaged over the year, the contributions of region III to North China, South Korea, and the Japanese main islands were small, but the contribution was as large as 20–25% over the southwestern islands of the Japan archipelago (Fig. 7a). The contributions to regions I, IV, and V were small throughout the year, but in spring, as the mid-latitude disturbances moved eastwards with air pollutants, the contribution to the western part of Japan was high (10–30%, Fig. 7b).

The contribution of region IV to other regions was smaller than 2%, except for region V, in which it was 5% on average (Table 4). Figure 8a shows that the deposition in region V occurred mainly over the Sea of Japan and not over the inland area of region V. The contribution of region IV was largest in July. It amounted to 5–10% in the eastern parts of region I and 5–15% in the northern part of Japan. The contribution to SJ was large in July.

The contribution of region V to other regions was small (0–1% for regions I, II, and III, and 6% for region IV) on average (Table 4) as well as in its horizontal distribution (<5% across other regions, Fig. 9a). In summer, however, under the influence of the Pacific High, the contribution was 5–15% over some parts of region I, 5% of region IV, and 10–25% over the northern part of Taiwan and some parts of region III.

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Atmospheric deposition to the ocean varied substantially by subregion and by season. The largest depositions were found over the Yellow Sea, the north of the East China Sea, and the south of the Sea of Japan (Figs. 3a and 4a). The area-averaged contribution of region I to YEC was only 6% (Table 4) but the contributions to the Bo Hai Sea (north of the Yellow Sea) and the Yellow Sea was largest in winter, accounting for 20–30% due to the prevailing northerly seasonal wind (Fig. 5a and d). The contribution of region I to SJ was 16% on average (Table 4), whereas it accounted for up to 30% in the northern part of SJ (Fig. 5a). It varied by season and was 10–25% in March and December, whereas it was smaller than 5% over the area south of 40° N of SJ in July due to the influence of the Pacific High. The mean contribution of region II to YEC was 53% (Table 4); it was larger than 40% in most areas of YEC and 60–80% over the Bo Hai Sea (Fig. 6). The contribution of region II to SJ was 40% on average (Table 4) and was mostly deposited over the northern part (>40%), except in winter due to the above-mentioned influence of the Siberian High (Fig. 6). Region III contributed, on average, 23% and 14% to YEC and NWP, respectively (Table 4). The contribution was larger in the southern part of YEC, especially near the coastal region, and accounted for 10–30% of deposition over the southern part of the Yellow Sea and 20–40% over the East China Sea (Fig. 7). The contribution to NWP was smallest in summer because the Pacific High prevented the transport of air pollutants from region III eastward (Fig. 7c). Region IV contributed 16% and 10% of the deposition over SJ and YEC, respectively (Table 4). The contribution was largest in the eastern part of the Yellow Sea (5–20%) and the western part of SJ (5–25%). The contribution over SJ was highest in the summer (15–40%, Fig. 8c) because the Pacific High prevented airflow from the Asian continent. The contribution of region V to NWP and SJ was 54% and 23%, respectively (Table 4). Figure 9a shows the contribution of region V, especially that of Miyake volcanic origin, was larger over NWP (up to 80%) and SJ near the coastal region of the Japan Archipelago (35–60%). Figure 9b–d shows the contribution to SJ was 20–30% in the eastern part in March, 10–60% over the whole region in July, and 5–30% near the coast in December. The contribution to YEC was largest in July,

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accounting for 5–35% in the eastern and southern parts of the ocean area (Fig. 9c).

## 5 Conclusions

Spatial distributions of the source–receptor relationships (SRRs) of sulfur over North-east Asia were examined using a chemical transport model, Regional Air Quality Model (RAQM), off-line coupled with the Mesoscale Meteorological Model (MM5). The RAQM model was developed at the Asia Center for Air Pollution Research (ACAP), formerly the Acid Deposition and Oxidant Research Center (ADORC). The emission inventory used in this study was developed by the Long-range Trans-boundary Air Pollutants in Northeast Asia (LTP project) of Tripartite Environment Ministers Meeting among Japan, China and Korea (TEMM).

The simulation was conducted for the entire year of 2002. Simulation results were evaluated using monitoring data from six remote stations of the Acid Deposition Monitoring Network in East Asia (EANET). The modeled  $\text{SO}_2$  and  $\text{O}_3$  concentrations agreed well with observations quantitatively. The modeled aerosol  $\text{SO}_4^{2-}$  was underestimated by 30%, but the correlation coefficient was large ( $R=0.80$ ). The modeled and observed amounts of precipitation correlated well ( $R=0.75$ ), whereas the model overestimated precipitation by 1.6 to 1.9 times. The modeled wet deposition flux of  $\text{SO}_4^{2-}$  was generally underestimated by about 50%, with some exceptions of overestimation.

The domain was divided into five source–receptor regions: I, North China; II, Central China; III, South China; IV, South Korea; and V, Japan. The deposition over the oceanic regions was also assessed to provide some indication of emission of other hazardous materials such as PAHs to ocean environment. The oceanic areas were divided into three regions: YEC, the Yellow Sea and East China Sea; SJ, the Sea of Japan; and NWP, the Northwest Pacific Ocean.

The sulfur deposition in each receptor region amounted to about 50–75% of the emissions from the same region. Dry deposition dominated over wet deposition in regions II and IV, whereas wet deposition dominated in regions I, III, and V, and all of

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the ocean regions. The largest contribution to the deposition in each region was the domestic origin, accounting for 53–84% of deposition. The second largest contribution after the domestic origin was due to region II, accounting for 14–43%, outside region II itself. The second largest contribution to region II was region III. Thus, the two regions exchanged sulfur with each other, mainly via dry deposition to region II and wet deposition to region III. The largest contributions to YEC and SJ were from region II (53% and 40%, respectively), whereas that to NWP was from region V (54%), which was mostly of Miyakejima volcanic origin. The second largest contribution to YEC was region III (23%), that to SJ was region V (23%), and that to NWP was region II (24%).

The horizontal distribution of SRRs revealed that subregional values varied about a factor of two from the regionally averaged values. This variation was due to nonuniformity of the deposition field. It was found to be important to examine the spatial distribution fields to identify subregional areas where deposition was highest within a region. The horizontal distribution changed substantially by season. Deposition and SRR features in summer were different than those in spring and winter, due to larger precipitation and different wind patterns in summer. Westerly to northwesterly winds prevailed over the Northeast Asian regions throughout the year, except in summer when southerly to southeasterly winds prevailed under the influence of the Pacific High.

*Acknowledgements.* This research was promoted by the joint research on Long-range Transboundary Air Pollutants in Northeast Asia (LTP project) and Tripartite Environment Ministers Meeting among Japan, China and Korea (TEMM). This research was supported by the Environment Research and Technology Development Fund (Project No. B-0905) of the Ministry of the Environment, Japan. M. Kajino was supported by a research fellowship of the Japan Society for the Promotion of Science (JPSP) for Young Scientists.

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**Table 1.** Description of the EANET remote observation sites used in this study. Each station is depicted in Fig. 1.

	Longitude (E)	Latitude (N)	Altitude (m a.s.l.)	Country
1. Rishiri	141°12′	45°07′	40	Japan
2. Tappi	120°21′	41°15′	105	Japan
3. Sado	138°24′	38°14′	136	Japan
4. Oki	133°11′	36°17′	90	Japan
5. Jeju	126°10′	33°18′	72	Korea
6. Hedo	127°15′	26°52′	60	Japan

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**Table 2.** Statistical analysis for comparisons between daily and monthly observations and simulation data at the six EANET remote stations.

Units Averaging	SO <sub>2</sub>		O <sub>3</sub>		Aerosol	Precipitation	Wet deposition
	Daily	Monthly	Daily	Monthly	nss-SO <sub>4</sub> <sup>2-</sup> μg/m <sup>3</sup> Monthly	Mm Monthly	of SO <sub>4</sub> <sup>2-</sup> μmol/m <sup>2</sup> Monthly
Number of data	1812	50	1788	50	22	59	59
Median (Obs.)	0.17	0.31	42.8	42.6	3.11	101.6	1608
Median (Sim.)	0.20	0.35	32.4	31.1	0.90	168.6	1091
Average (Obs.)	0.35	0.36	42.3	41.1	4.68	108.9	3035
Average (Sim.)	0.40	0.38	31.9	32.0	1.60	190.4	1478
MB	0.045	0.023	-10.4	-9.13	-3.08	81.44	-1557
RMSE	0.66	0.34	5.4	11.8	3.72	139.3	4019
R	0.32	0.11	0.41	0.55	0.80	0.75	0.28
FAC2 <sup>a</sup>	0.38	0.50	0.92	1.0	0.27	0.59	0.48

<sup>a</sup> Number fraction of data within a factor of two.

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**Table 4.** Annual deposition (kt S/yr) of sulfur and contributions of source regions to receptor regions (%) in 2002.

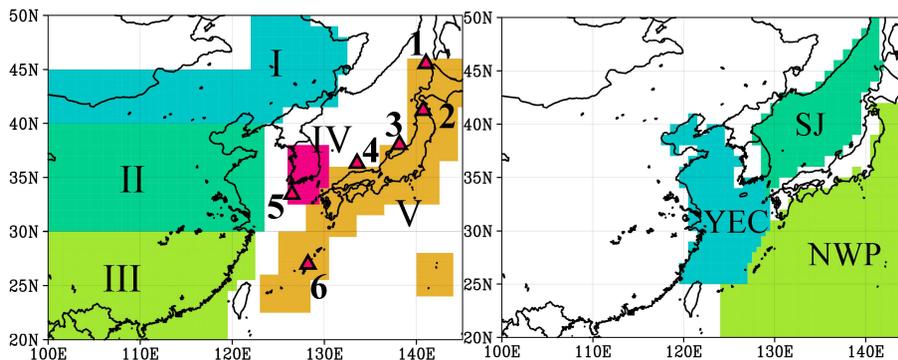
Annual deposition of sulfur (units: kt S/yr)								
Region	I	II	III	IV	V	YEC	SJ	NWP
Dry	294.1	1863.9	989.6	161.1	392.1	75.9	57.8	135.1
Wet	329.3	1295.8	1567.3	112.1	822.8	157.7	279.7	375.5
Total	623.5	3159.7	2557.0	273.2	1214.9	233.5	337.5	510.5

Annual mean contribution of sources to receptors for sulfur deposition (%)								
S \ R	I	II	III	IV	V	YEC	SJ	NWP
I	53	2	1	4	3	6	16	3
II	43	83	14	18	19	53	40	24
III	2	14	84	6	8	23	5	14
IV	2	1	0	66	5	10	16	5
V	1	0	1	6	65	7	23	54

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**Fig. 1.** The model domain and the EANET remote monitoring stations (1–6, pink triangle): 1. Rishiri, 2. Tappi, 3. Sado, 4. Oki, 5. Jeju, and 6. Hedo. Descriptions of the locations are given in Table 1. The color shaded areas (left) I to V represent regions for source–receptor relationship analyses, I: North China ( $>40^{\circ}$  N), II: Central China ( $>30^{\circ}$  N,  $<40^{\circ}$  N), III: South China ( $<30^{\circ}$  N), IV: South Korea, and V: Japan. The areas on the right represent oceanic receptor areas, YEC: Yellow Sea and East China Sea, SJ: Sea of Japan, and NWP: Northwest Pacific Ocean.

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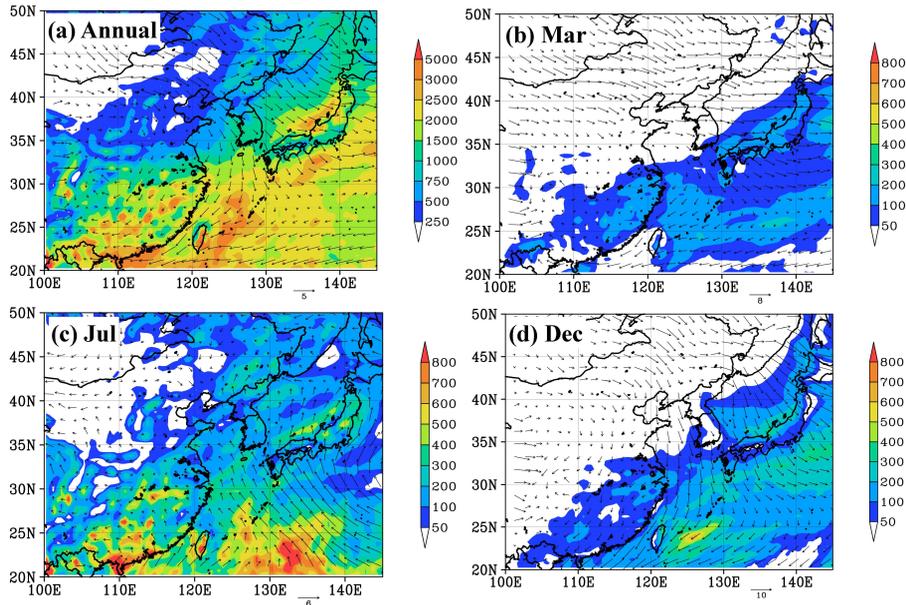
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**Fig. 2.** Spatial distributions of **(a)** yearly and **(b–d)** monthly accumulated precipitation (mm) and mean surface wind field (m/s). The selected months, **(b)** March, **(c)** July, and **(d)** December, represent the spring, summer, and winter seasons, respectively.

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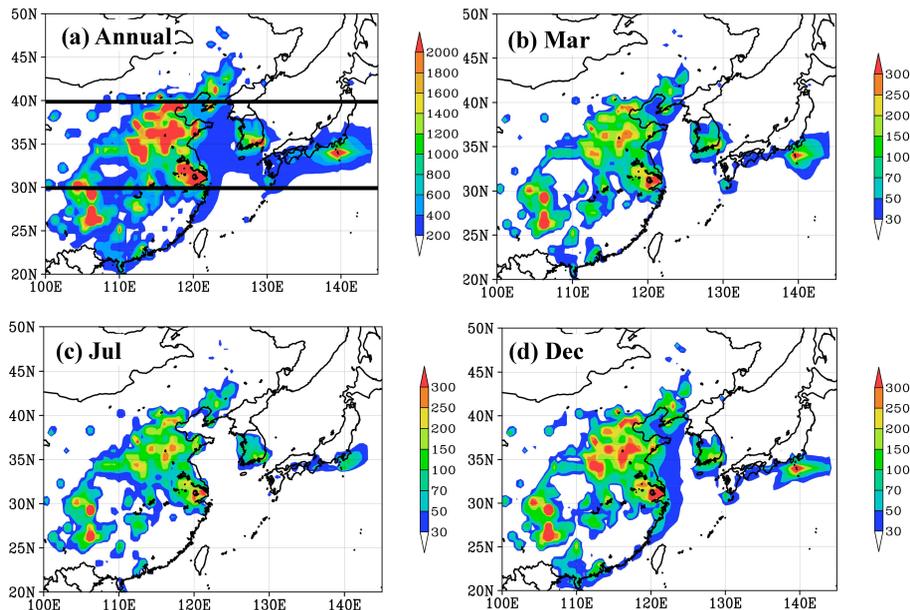
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**Fig. 3.** Spatial distributions of (a) annual and (b–d) monthly accumulated dry deposition of total S ( $\text{mg S/m}^2$ ). The solid lines in (a) indicate  $40^\circ\text{N}$  and  $30^\circ\text{N}$ , the borders between regions I and II and between regions II and III, respectively.

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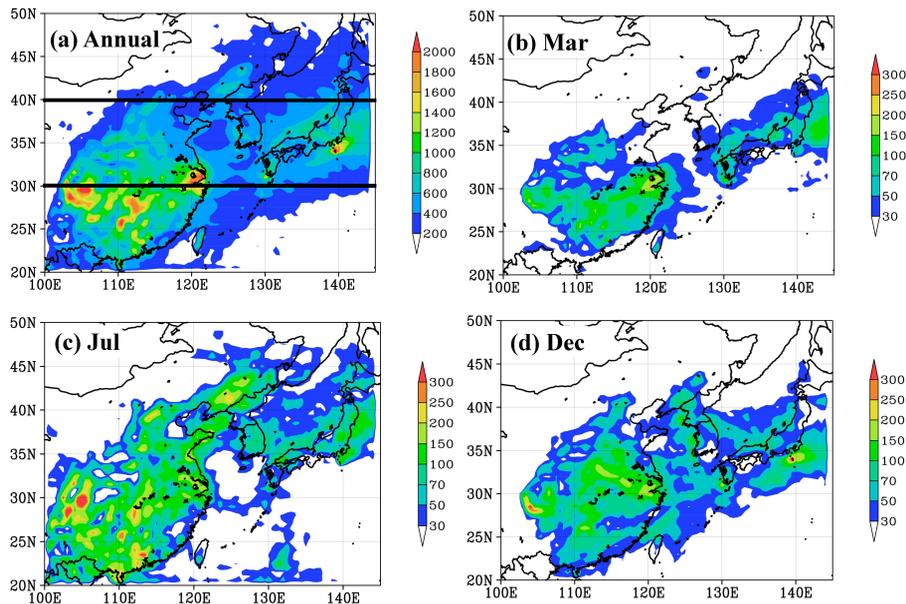
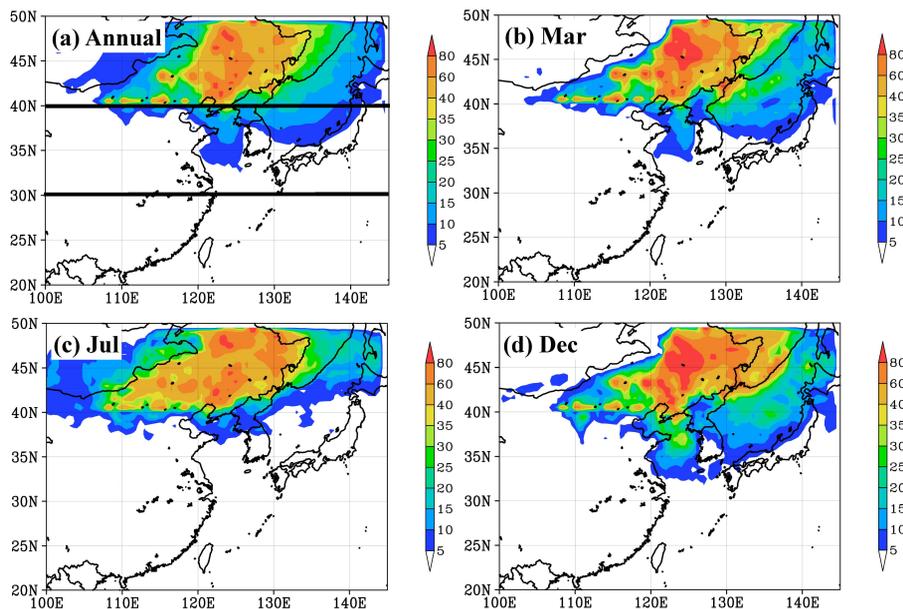


Fig. 4. Same as Fig. 3 but for the wet deposition of total S ( $\text{mg S/m}^2$ ).

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**Fig. 5.** Spatial distributions of **(a)** annual mean contributions of source region I to total sulfur deposition and monthly mean values for **(b)** March, **(c)** July, and **(d)** December (%). The solid lines in **(a)** indicate 40° N and 30° N, the borders between regions I and II and between regions II and III, respectively.

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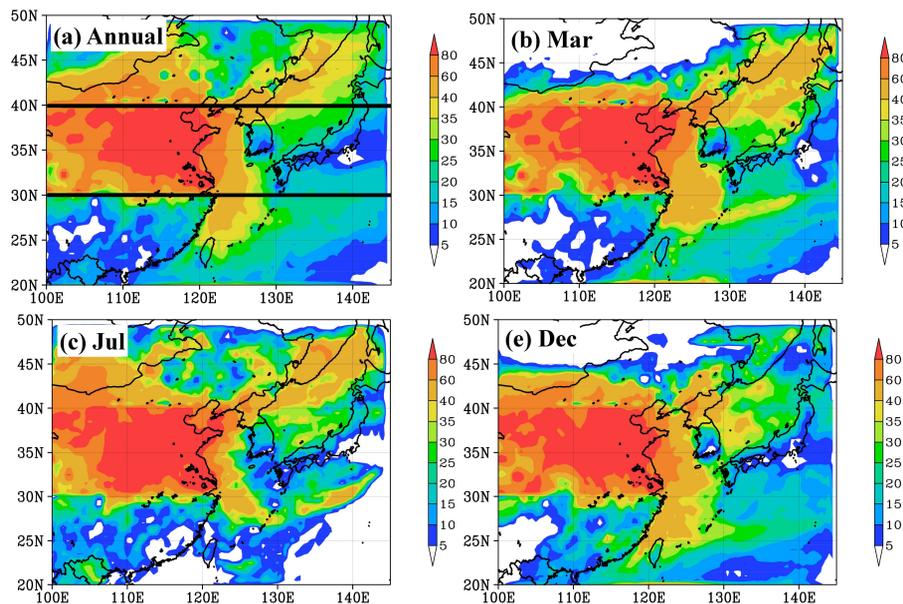


Fig. 6. Same as Fig. 5 but for source region II.

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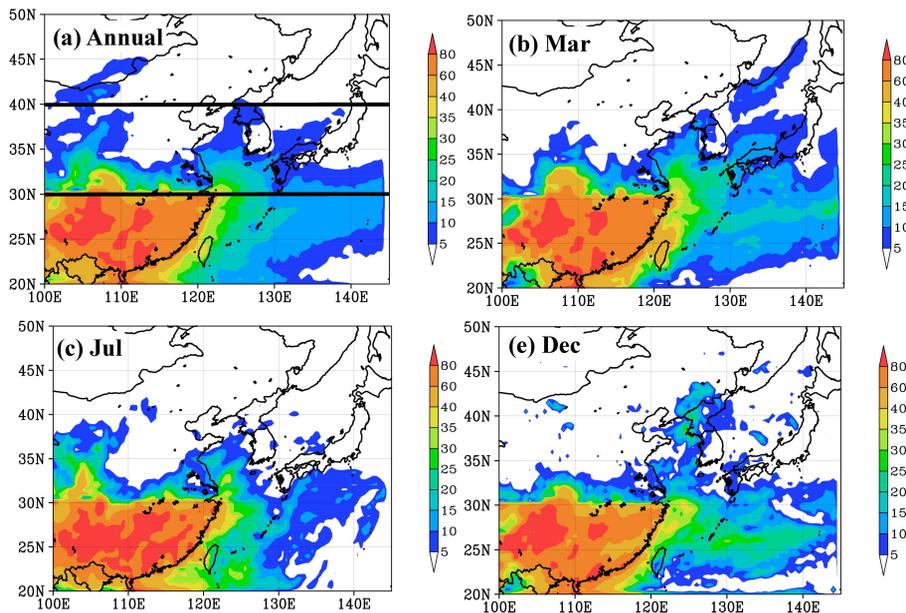
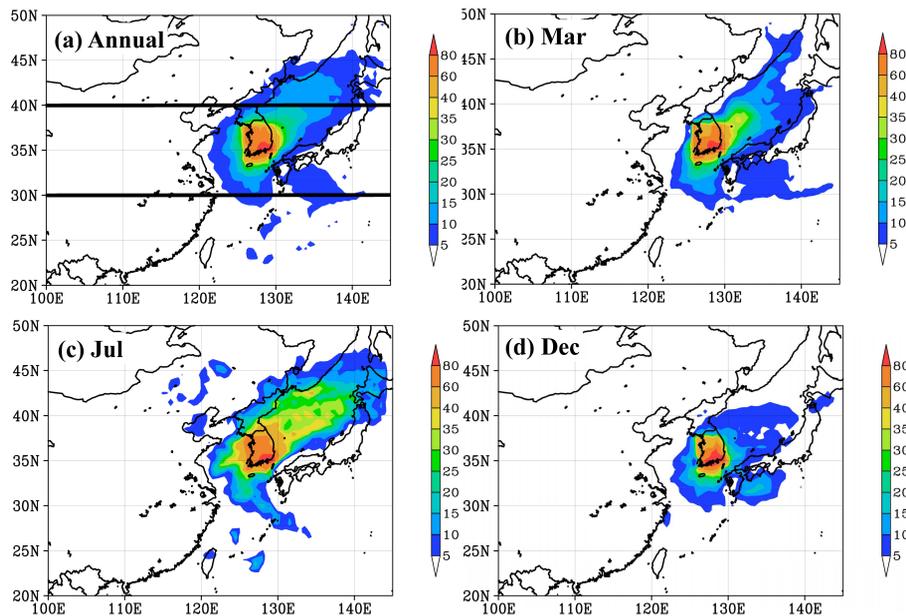


Fig. 7. Same as Figs. 5 and 6 but for source region III.

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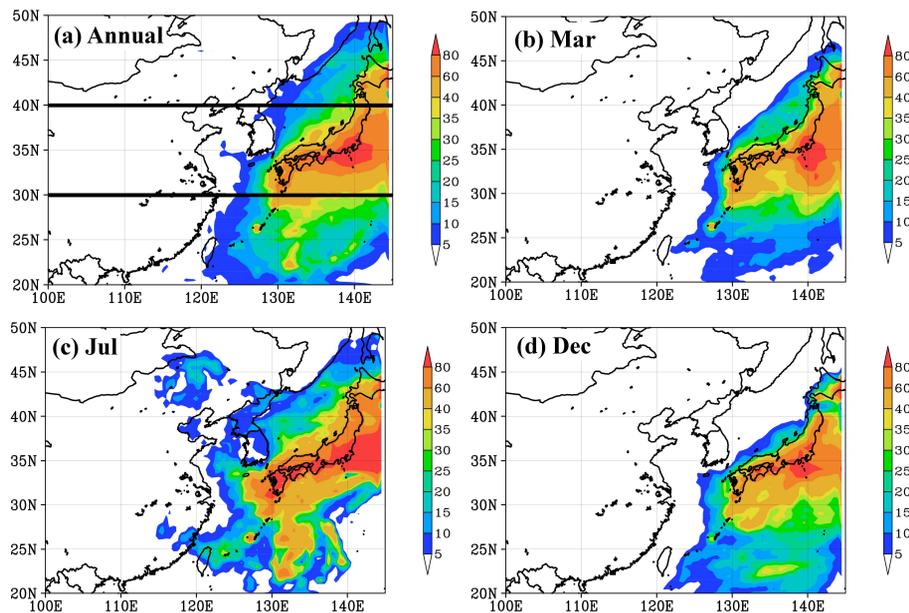


**Fig. 8.** Same as Figs. 5–7 but for source region IV.

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**Fig. 9.** Same as Figs. 5–8 but for source region V.

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