

Abstract

Modelling of aerosol particles with chemical transport models is still based mainly on static emission databases while episodic emissions can not be treated sufficiently. To overcome this situation, a coupling of chemical mass concentration modelling with satellite-based measurements relying on physical and optical principles has been developed. This study deals with the observation operator for a component-wise assimilation of satellite measurements. It treats aerosol particles classified into water soluble, water insoluble, soot, sea salt and mineral dust containing aerosol particles in the atmospheric boundary layer as separately assimilated aerosol components. It builds on a mapping of aerosol classes used both in observation and model space taking their optical and chemical properties into account. Refractive indices for primary organic carbon particles, anthropogenic particles, and secondary organic species have been defined based on a literature review. Together with a treatment of different size distributions in observations and model state, this allows transforming the background from mass concentrations into aerosol optical depths. A two-dimensional, variational assimilation is applied for component-wise aerosol optical depths. Error covariance matrices are defined based on a validation against AERONET sun photometer measurements. Analysis fields are assessed threefold: (1) through validation against AERONET especially in Saharan dust outbreak situations, (2) through comparison with the British Black Smoke and Sulphur Dioxide Network for soot-containing particles, and (3) through comparison with measurements of the water soluble components SO_4 , NH_4 , and NO_3 conducted by the EMEP (European Monitoring and Evaluation Programme) network. Separately, for the water soluble, the soot and the mineral dust aerosol components a bias reduction and subsequent a root mean square error reduction is observed in the analysis for a test period from July to November 2003. Additionally, examples of an improved analysis during wildfire and dust outbreak situations are shown.

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1 Introduction

Both, global and regional chemical transport modelling (CTM) have successfully included aerosol modules in recent years (Binkowski, 1999; Ebel, 1997a; Rasch et al., 2000; Chin et al., 2002). Apart from scientific research questions this is motivated mainly by the implications of aerosols on human health and the radiative budget in the atmosphere. While the modelling of physical and chemical processes has improved widely, a specific problem is the timely information of special events with elevated aerosol emissions. Typically, databases show deficiencies in spatial and temporal resolution, quantitative values and their evolution over time (Memmesheimer et al., 2004). Irregular episodic events like fires, dust outbreaks or variable sources in transportation or vegetation can not be modelled on this basis. Therefore, up to date measurements have to be taken into account using data assimilation methodologies. Ground measurements often show a confined geographical coverage which is insufficient compared to the typical horizontal aerosol source variability. Satellite-based sensors provide aerosol measurements on a regional or global scale, covering larger areas within a single overpass. Currently, all satellite-based algorithms provide aerosol optical depth (AOD) observations. Therefore, a coupling between the chemical descriptions in terms of mass concentrations used in air quality models and the physical properties observed by satellites has to be introduced (e.g. Hutchinson et al., 2004). In terms of data assimilation parlance, such a coupling is provided by the observation operator. Besides the spatio-temporal interpolation between a gridded CTM and the irregular distributed satellite observations, the observation operator couples between the different aerosol descriptions in the background model and the observations.

Data assimilation has been widely used in numerical weather prediction (e.g. Daley, 1991; Lorenc, 1986) and oceanography (Ghil, 1989), and chemical transport modelling (Elbern et al., 1997, 1999, 2000, 2001, 2007, 2010; Jeuken et al., 1999; Khattatov et al., 1999; Lahoz et al., 2007a, b; Lamarque et al., 1999; Levelt et al., 1998; van Loon et al., 2000). Most of these studies deal with the assimilation of gas phase species as

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e.g. ozone, nitrogen dioxide or carbon monoxide.

First assimilation studies of satellite based AOD observations have been presented in recent years. During the Indian Ocean Experiment (INDOEX) Collins et al. (2001) used an optimum interpolation method to assimilate AOD observations made by the AVHRR (Advanced Very High Resolution Radiometer) instrument into the Model of Atmospheric Transport and Chemistry (MATCH, Rasch et al., 1997). Due to missing aerosol type information in the observations and the focus of INDOEX on the distribution of maritime aerosols (Rasch, 2001), this study assumes a fixed aerosol type based on optical properties of maritime aerosols. An equivalent approach has been used by Yu et al. (2003) to assimilate AOD observations of the Moderate Resolution Imaging Spectroradiometer (MODIS) into the Goddard Global Ozone Chemistry Aerosol and Transport (GOCART) model. Verver et al. (2002) and van Velthoven et al. (2004) present an assimilation of Along Track Scanning Radiometer (ATSR-2) observations into the chemical tracer model TM3, also based on a similar approach.

While all these studies assume a constant aerosol type for all observations independent of the regional aerosol composition, there are also studies focusing on the assimilation of observations representing a single aerosol component. Wang et al. (2004a, b) assimilate dust aerosol observations taken selectively by the Geostationary Operational Environmental Satellite (GOES-8, Wang et al., 2003) over the Atlantic Ocean. For the Puerto Rico Dust Experiment (PRIDE) in July 2000 a nudging approach was used to assimilate these GOES-8 observations into the Regional Atmospheric Modelling System (RAMS). Niu et al. (2008) use a variational approach for the assimilation of a dust index provided by the geostationary FY-2C satellite into the Chinese Unified Atmospheric Chemistry Environment – Dust (CUACE/Dust) forecast system. Also, in a case study covering East Asia in May 2007, Hara et al. (2009) use observations of the Cloud Aerosol Lidar with Orthogonal Polarization (CALIOP) to simulate dust outbreak events over China and Japan within the RAMS/CFORS dust transport model.

A first assimilation of aerosol size distribution resolving observations was published by Generoso et al. (2007) using “fine” and “coarse mode” observations made over the

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Arctic by the POLarization and Directionality of the Earth's Reflectances (POLDER) instrument and the LMDz-INCA chemical transport model.

Meanwhile the Synergetic Aerosol Retrieval (SYNAER, Holzer-Popp et al., 2002a, b, 2008) provides aerosol optical depth observations from ENVISAT Advanced Along Track Scanning Radiometer (AATSR) and Scanning Imaging Absorption Spectrometer for Atmospheric CHartographyY (SCIAMACHY) separately for the soot, mineral, water soluble, water insoluble and sea salt components. This opens the possibility to assimilate aerosols in a component-wise approach over both land and ocean.

This study focuses on the development of an observation operator for SYNAER observations with their separation of major aerosol components representative for the atmospheric boundary layer and the troposphere. This is a basis for an operational variational assimilation system of SYNAER observations provided by the European Environmental Satellite (ENVISAT) and in the future the Meteorological Operational Polar Satellites (METOP).

Section 2 provides a description of the study setup, the SYNAER observations with their underlying aerosol component definition, the European Air Pollution Dispersion-Inverse Model (EURAD-IM) with its aerosol class definition and the validation data bases. Section 3 describes the concept and development of the observation operator together with a validation of both background model and observations as a basis of error covariance definition. Section 4 presents a validation of analysis fields both in terms of aerosol optical depth and mass concentrations against ground-based measurements. Finally, Sect. 5 concludes the paper with a discussion of its achievements and limitations.

2 Study setup

The study is based on a test period from 1 July to 30 November 2003 with detailed internal EURAD model output available for Europe as a result of the ASSET ("Assimilation of ENVISAT data", Lahoz et al., 2007b) project campaign.

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2.1 SYNAER measurements

The synergetic aerosol retrieval method SYNAER (Holzer-Popp et al., 2002a, b, 2008) delivers aerosol optical depth (AOD) at 550 nm over both land and ocean and an estimation of the aerosol type in the lower troposphere including a possible elevated dust layer. It exploits a spatially high resolving (typically 1 km²) radiometer allowing accurate cloud detection (Kriebel et al., 1989, 2003) and dark field aerosol assessment together with a spectrally high resolving spectrometer allowing aerosol type retrieval. The type of aerosol is estimated as percentage contribution to AOD of representative components from an extension of the OPAC (Optical Parameters of Aerosols and Clouds, Hess et al., 1998) dataset as described in Holzer-Popp et al. (2008).

The SYNAER aerosol retrieval algorithm comprises of two major parts: (1) A dark field method exploiting single wavelength radiometer reflectances (670 nm over land, 870 nm over ocean) and (2) a least square fit of visible top-of-atmosphere reflectance spectra at 10 wavelengths (415, 428, 460, 485, 500, 516, 523, 554, 615, and 675 nm) with the spectrometer. In the first step, AOD over automatically selected and characterized dark pixels is derived and subsequently a surface albedo correction at 550, 670, and 870 nm is done with the radiometer. The knowledge of atmospherically corrected spectral surface albedo is further used for a characterization of the surface type. This first step is repeated for 40 different pre-defined aerosol mixtures. After spatial integration of the results (40 AOD, 40×3 surface reflectance values, 40 surface types) to the larger pixels of the spectrometer (60×30 km²), these parameters are used in the second step to simulate 40 spectra for the same set of 40 different aerosol mixtures using the same radiative transfer code. A least square fit of these simulated spectra to the measured spectrum delivers the correct AOD value (i.e. the AOD for the selected aerosol mixture) and – if a uniqueness test is passed – the plausible aerosol mixture. All radiative transfer calculations within SYNAER assume spherical particles and Mie theory.

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5 The entire method uses the same basic aerosol components (Holzer-Popp et al., 2008; Table 1). These components, their optical features and log-normal size distribution are taken from the OPAC database (Hess et al., 1998) for the water soluble (WASO), the water insoluble (INSO), the sea salt accumulation and coarse mode
10 (SSAM and SSCM), and the mineral transported (MITR) component. On the basis of more recent campaigns and AERONET data some specific items have been updated. The soot component is split in two components for strongly absorbing diesel soot (DISO) more representative for industrial areas and weakly absorbing biomass burning soot (BISO). For mineral dust an additional component for mineral dust with
15 low absorption (MILO) represents dust sources with lower hematite content. As the insoluble component in OPAC (INSO) is modelled with the identical refractive index as the mineral transported component MITR, also an insoluble component with low absorption (INSL) is included.

These basic components are externally mixed into 40 different aerosol types meant
20 to cover a realistic range of atmospheric aerosol masses (see details in Holzer-Popp et al., 2008). For humidity dependent components two models with 50% and 80% relative humidity are included. In the case of desert dust outbreaks, an elevated dust layer is added in the free troposphere.

25 SYNAER is applied to the radiometer AATSR and the spectrometer SCIAMACHY onboard ENVISAT resulting in a pixel size of $60 \times 30 \text{ km}^2$. The measurement is taken daily at approximately 10:00 local time. Global coverage is achieved every 12 days, but cloud cover above 50% also reduces the number of available retrievals. The number of observations in SYNAER version 1.8 allows the evaluation of an assimilation scheme as presented in this study. Nevertheless, it is expected that an operational and sequential assimilation of SYNAER observations will only be meaningful for future METOP based SYNAER retrievals with a global coverage every 1–2 days.

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2.2 The EURAD model

The European Air Pollution Dispersion Model (EURAD, Ebel, 1989, 1997a, b; Memmesheimer et al., 2004; Builtjes et al., 2003, Elbern et al., 2007) consists of several coupled modules. Meteorological parameters are provided by the mesoscale model MM5 (Grell et al., 1994). The Modal Aerosol Dynamics model (MADE, Ackermann et al., 1998) provides aerosol dynamics and chemistry. Secondary aerosols and their precursor substances are described in the SORGAM module (Schell et al., 2001) and emissions are provided in the EEM module (Memmesheimer et al., 1995) based on existing databases. The National Center for Environmental Protection (NCEP-GFS) analyses are used as meteorological driver for initial and boundary conditions. EURAD provides forecasts of the temporal and spatial distribution of atmospheric species both in the gas and in the particle phase. Explicitly modelled processes are horizontal and vertical advection, turbulent diffusion, dry and wet deposition, sedimentation, coagulation in and between nucleation and accumulation modes, condensation and nucleation of gas phase precursor substances, and particle emission.

For this study EURAD is used on the European scale with a horizontal grid box size of 56 km. EURAD can also be operated on horizontal resolutions ranging from the hemispheric scale down to regional and urban scales with grid box sizes of a few km. EURAD is operated with 23 layers between the surface and 15 km height and provides PM₁₀ and PM_{2.5} as standard output. The MADE module internally provides mass concentrations of SO₄, NH₄, NO₃, primary organic carbon, elemental carbon and not further described anthropogenic particles both in the nucleation and accumulation mode. An additional anthropogenic aerosol class is described in the coarse particle mode. The SORGAM module adds mass concentrations in the nucleation and accumulation mode for secondary organic species resulting from anthropogenic and biogenic aromates, alkanes, alkenes, α -pinene and d-limonene as precursor substances.

Primary aerosol particle emissions are based on a European emission database developed by the Nederlandse Organisatie voor toegepaast-natuurwetenschappelijk

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onderzoek (TNO, Berdowski et al., 1996) and updates within the Co-ordinated Euro-
pean Programme on Particulate Matter Emission Inventories (available at [http://www.
air.sk/tno/cepmeip](http://www.air.sk/tno/cepmeip)). This database provides emission data for carbon monoxide, PM₁₀
and PM_{2.5} particles based on the reference year 1995 in a horizontal resolution of
1×0.5 degree. Anthropogenic sources as for example combustion in different power
plant types, several industrial production and combustion processes, use of petrol and
diesel in motor vehicles, tire abrasion, agricultural production, heating in private house-
holds, and waste combustion are taken into account. The database relies on use and
production statistics, population databases as well as on agricultural food production
statistics from a variety of international organisations. Additionally, the EEM model
uses population databases and a weekly and daily variation distinguishing between
working and weekend day emission levels.

EURAD as a modal model describes aerosol particles as separate log-normal size
distributions in the nucleation, accumulation and coarse mode. Following Whitby et
al. (1978) and Ackermann (1998) the partial differential equations are solved for the
integral moments as prognostic variables. The size distributions' moments represent
the total particle number and the aerosol particle volume. The size distribution width
is kept constant during the model run, while the particle mass and number concentra-
tion are changed by physical and chemical processes. Within each mode the aerosol
particles are modelled as an internal mixture.

Maritime and mineral aerosol components are not included in the EURAD version
used for this study. A comparison against SYNAER observations in the study period
shows an underestimation of the total aerosol optical depth between −0.01 and −0.1
in 25% of 24 747 coincidences over Europe, while an underestimation between −0.1
and −0.2 is found in 7% of all cases and a larger underestimation is found in 3% of all
cases. It has to be noted that in the meanwhile the EURAD model has been extended
by an explicit dust modelling scheme following Nickovich et al. (2001) and a sea salt
scheme following Monahan et al. (1986).

2.3 AERONET ground measurements

AERONET (AERosol RObotic NETwork, <http://aeronet.gsfc.nasa.gov>) ground-based sun photometer measurements (Holben et al., 1998) are used for AOD validations at 550 nm. AERONET is a global network with approx. 200 permanent stations, while this study is based on 47 stations providing version 2 datasets for Europe. Level 1.5 datasets include an automatic cloud detection, while level 2 datasets are additionally checked for cloud occurrence manually. AERONET stations at Erdemli (Turkey, close to mountains), Palaiseau and Fontainebleau (France, greater Paris area), Rome Tor Vergata (Italy, next to a busy traffic crossing) and Hamburg (Germany, close to the coastline) are explicitly excluded as these stations are not representative for their surroundings on a satellite pixel or grid box scale. The overall uncertainty of AERONET AOD values in cloud-free conditions is ± 0.01 for wavelengths down to 440 nm and ± 0.02 for shorter wavelengths.

2.4 British Black Smoke and Sulphur Dioxide Network

The UK Black Smoke and Sulphur Dioxide Network (Loader et al., 2003; <http://www.airquality.co.uk>) measures both the black smoke concentrations and the atmospheric SO₂ equivalent of all atmospheric acids as a daily mean at 137 stations in the United Kingdom. Most stations are representing urban background or suburban conditions, but there are also background stations available. A black smoke index is derived by a reflectometer measurement of a darkened filter after 24 h exposition time. This is post processed to black smoke mass concentrations following the “British Standard Smoke Calibration Curve”. Mass concentrations of black smoke can be set equal to mass concentrations of elementary carbon as organic carbon on the filter does not contribute to the blackness of the filter. For typical mass concentrations in the UK, an uncertainty of 30% is reported.

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2.5 EMEP network

The Co-operative Programme for Monitoring and Evaluation of the Long-range Transmission of Air pollutants in Europe with its European Monitoring and Evaluation Programme (EMEP, <http://www.emep.int>) focuses on the chemical composition of gas, particle and liquid phases in the atmosphere. 48 countries are contributing with PM₁₀ and PM_{2.5} measurements. Overall, 16 in-situ stations in Hungary, Italy, Lithuania, the Netherlands, Norway, Poland, Russia and Turkey provide separate daily mean measurements of SO₄, NH₄ and NO₃ based on ion chromatography. These can be compared with mass concentrations as modelled in EURAD. The uncertainty requirement is set at 10% or better (EMEP, 2001).

3 The observation operator

3.1 Two-dimensional variational assimilation

Most data assimilation procedures are based on an optimisation strategy, employing a least square minimisation (e.g. Daley, 1991; Bouttier and Courtier, 1999) and solve the analysis equations

$$\begin{aligned} \mathbf{x}_a &= \mathbf{x}_b + \mathbf{K}(\mathbf{y}_o - \mathbf{H}\mathbf{x}_b) \\ \mathbf{K} &= \mathbf{B}\mathbf{H}^T (\mathbf{H}\mathbf{B}\mathbf{H}^T + \mathbf{R})^{-1} \end{aligned} \quad (1)$$

in order to provide a best linear and unbiased estimator (BLUE). The model state for a specific time is given by the vector \mathbf{x} . The model background state, forecast or first guess field at the beginning of the assimilation process is given by \mathbf{x}_b , while \mathbf{x}_a represents the analysis field at the end of the assimilation procedure. Observations are represented in the vector \mathbf{y}_o , while \mathbf{R} describes the observation error covariance matrix. Matrix \mathbf{B} is the background error covariance matrix. The observation operator

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H maps the model state \mathbf{x} into observation space. Matrix **K**, the “Kalman gain matrix”, describes the weights attributed to the observation increment, given by $(\mathbf{y}_o - \mathbf{H}\mathbf{x}_b)$.

In order to achieve the BLUE characteristics an efficient observation quality control is needed to exclude outliers (see Sect. 3.3.1), both background and observations should be free from systematic errors (discussed in Sects. 3.3 and 5 for this case) and errors in background and observations should be uncorrelated. The latter assumption is justified as the satellite observations are retrieved independently from the chemical transport modelling.

Generally in assimilation theory, the observation operator **H** is non-linear. In our case the optimization is made with respect of particle number and not with respect to the particle diameter distribution. Therefore, **H** becomes a linear operator as the Mie theory is applied.

Both two- and three-dimensional variational assimilation methods define a cost function J as

$$J(\mathbf{x}) = \frac{1}{2}(\mathbf{x} - \mathbf{x}_b)^T \mathbf{B}^{-1}(\mathbf{x} - \mathbf{x}_b) + \frac{1}{2}(\mathbf{y}_o - \mathbf{H}\mathbf{x}_b)^T \mathbf{R}^{-1}(\mathbf{y}_o - \mathbf{H}\mathbf{x}_b) \quad (2)$$

The analysis \mathbf{x}_a is a solution of the minimization problem for J , where the analysis increment $\Delta\mathbf{x}_a$ is defined as $(\mathbf{x}_a - \mathbf{x}_b)$.

Chemical transport models like EURAD provide mass concentrations of a large number of aerosol classes or the particle diameter integrated PM_{10} mass parameter, describing the sum of all particles with a diameter below $10\mu\text{m}$ as a standard output. Having aerosol component resolving SYNAER observations, the assimilation in PM_{10} space would not use this component-wise information. Rather, EURAD mass concentrations are provided on 23 vertical layers and for 29 internal aerosol classes, which is much more detailed information than provided by the SYNAER total column AOD observations provided for 9 different components. Therefore, assimilation of AOD space is chosen to reduce complexity and computational time. As AOD observations are vertically integrated geophysical parameters, the three dimensions used in 3D-Var methodologies are reduced to two dimensions (2D-Var) only.

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As a preprocessing step the separation of $AOD_{total,550nm}$ in the component-wise AOD of SYNAER basic components according to their percentage contribution f_j in each aerosol mixture is needed (Table 1)

$$y_j = f_j AOD_{total,550nm} \quad j=1, \text{ number of components} \quad (3)$$

- 5 For the acceptance as spatio-temporal coincidences between observations and local model states, ENVISAT observations within a ± 2 h window around analysis time (in this study taken at 10:00 UTC) are taken into account.

The following steps are needed in the assimilation procedure:

(A) Observation operator H

- 10 Mapping of vertically resolved mass concentrations $x_{b,mass}$ of all EURAD aerosol classes in all modes (model state) into a vertically integrated AOD_{550nm} for each SYNAER component ($x_{b,AOD}$) following Table 2.

$$x_{b,AOD,j} = \int_{k=1}^{k_{max}} \sum_i \int_0^{\infty} C_{ext,k,i}(D_{k,i}, m_{k,i}) \frac{dN_{k,i}}{d \ln D} d \ln D dz \quad \forall \text{components} \quad (4)$$

- 15 The index i represents a size mode contributing to one of the j components, k represents the index of a vertical layer, C_{ext} describes the extinction cross section of a particle as a function of particle diameter D and complex refractive index m , $N_{k,i}$ denotes the total particle number in mode i and layer k , while z represents height. Section 3.2 describes this step in more detail.

- 20 The extinction efficiency Q_{ext} is calculated using a fast parameterisation of the Mie theory as suggested by Evans and Fournier (1990). This parameterisation combines both the Rayleigh approximation for small particles and the Van de Hulst approximation for large particles with an empirical formula based on explicit Mie calculations. Following Evans and Fournier, the deviation from explicit Mie calculation is below 1% by

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variation of the real part of the refractive index between 1.01 and 2.0 and the imaginary part between 0 and 10. The approximation deficiency can be neglected compared to other uncertainties like in refractive indices introduced by assumptions on the aerosol composition. The parameterization reduces the computational expenditure by approximately a factor of 30 and allows the use of Mie theory for an operational application.

(B) Application of a 2D-Var minimization and derivation of analysis increments for each SYNAER component

$$\Delta x_{a,AOD} = \mathbf{K} (y_o - \mathbf{H}x_{b, \text{mass}}) \quad \forall \text{components} \quad (5)$$

In this study, the 2D-Var assimilation is done separately for the water soluble component (WASO), the insoluble component (sum of SYNAER INSO and INSL) and the soot component (sum of SYNAER BISO and DISO). All components are represented by two-dimensional aerosol optical depth fields. Later in operations this step will be conducted also for the mineral component (sum of SYNAER MITR and MILO) and the sea salt component (sum of SYNAER SSAM and SSCM) as the EURAD model has meanwhile been extended by a dust and sea salt module.

The cost function is minimized using a L-BGFS (Limited memory Broyden-Fletcher-Goldfarb-Shanno) quasi Newton method, adopted from Liu and Nocedal (1989).

Background and observation error covariance matrices are initially set as diagonal with constant values of 0.12 for observations and 0.3 for the model background based on validation results as discussed in Sect. 3.3. A component-wise definition of covariance matrices is not possible due to the lack of component-wise AOD ground measurements. Observations are assumed to be uncorrelated to each other resulting in diagonal covariance matrices.

A diffusion approach described by Weaver and Courtier (2001) is applied to model the background error covariance matrix. Isotropic horizontal diffusion coefficients are used. Correlation lengths of the horizontal diffusion between 0.1 and 3 grid boxes (corresponding to 5 and 150 km with a grid box size of 56 km) were tested following

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Anderson et al. (2003) who found a correlation length for AOD below 200 km in ground, air craft and space-based measurements. Finally, the correlation length is set to 0.5 grid boxes which is also approximately the SYNAER pixel size.

(C) Use of an adjoint observation operator to transfer the new model state $x_{a,AOD}$ from the SYNAER component AOD space into the vertically resolved mass concentrations of EURAD aerosol classes ($x_{a,mass}$)

$$x_{a,mass} = \mathbf{H}^T x_{a,aod} \quad \forall \text{ mass concentrations} \quad (6)$$

It has to be noted that both vertical profile and size distribution of the EURAD aerosol field are not changed by the assimilation as the observations do not provide such information independent from any further assumptions. Therefore, a factor g based on the analysis increment in each aerosol component j can be linearly applied to mass concentrations of all i aerosol size modes mapped to an aerosol component j (Table 2) in each vertical layer. This changes the overall amount of mass concentrations, but keeps the vertical structure of the aerosol profile and the distribution of aerosol classes within a component constant.

$$mk_{i,analysis} = g_j mk_{i,background}$$

$$g_j = \frac{AOD_{j,analysis}}{AOD_{j,background}} \quad (7)$$

3.2 Mapping of aerosol models between observation and model spaces

Mie theory is already used in the SYNAER retrieval to describe optical properties of aerosols. Therefore, the \mathbf{H} operator also applies Mie theory to transform mass concentrations of the EURAD model aerosol classes into aerosol optical depth. Both size distribution and complex refractive index m have to be defined for each EURAD aerosol class as a prerequisite for the derivation of extinction coefficients β_{ext} in each model

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grid box. AOD is finally derived as the vertical integral of β_{ext} , assuming a maximum elevation of the aerosol layer up to 6 km.

EURAD aerosol classes as described in Sect. 2 have to be mapped on SYNAER components. For the inorganic aerosol classes it is known how sulphuric acid, ammonium sulphate, ammonium nitrate and nitric acid with a known refractive index are created in an aqueous solution (Binkowski, 1999). Further, there are also measurements of the refractive index of elemental carbon available (Schnaiter et al., 2003).

Far more complicated is the attribution of refractive indices to those classes which consist of many and even partly unknown chemical species. This applies to the primary organic carbon particles, the not further identified anthropogenic particles class modelled both in the accumulation and coarse modes, and secondary organic species resulting from aromatics, alkanes, alkenes, α -pinene and d-limonene as precursor gas phase substances, where a mass production rate per precursor mass is used in EURAD. Therefore, a literature study has been performed to identify dominating chemical substances observed in in-situ measurement campaigns and to identify appropriate refractive index values from chemical parameter databases (Schroedter-Homscheidt, 2009).

A sensitivity study based on explicit Mie calculations in Schroedter-Homscheidt (2009) shows that the real part of the refractive index needs to be known by an accuracy of within 0.1 or better to achieve an AOD uncertainty below 0.1, which is a reasonable threshold accuracy of satellite-based AOD measurements. Based on the same study, the imaginary part should be known with an accuracy of 0.03.

Primary organic matter consist mainly of alkanes, alkanals, alkanols, fatty acids, fatty alcohols and dicarboxylic acids (Alves et al., 2001; Hahn et al., 1980; Kavaouras et al., 1999; Kendall et al., 2001; Kubatova et al., 2000; Limbeck et al., 1999; Pio et al., 2001; Saxena et al., 1995; Yassaa et al., 2001) which are typically water insoluble due to their long carbon chains. Reaction products of aromatics are typically dicarboxylic acids or substances built by several benzene rings (e.g. Hahn et al., 1980, Kawamura et al., 1999; Koch et al. 2000; Seinfeld and Pandis, 1998; Warscheid and Hoffmann,

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2001). These are found to be water insoluble as well. Reaction products of alkanes are dominated by water soluble dicarboxylic acids with short carbon chains (e.g. Kawamura et al., 1999; Kerminen et al., 2000; Krivacsy et al., 2001; Pio et al., 2001; Röhl et al., 2001; Saxena et al., 1995; Tsapakis et al., 2002; Wang et al., 2002; Yu, 2000).

5 Reaction products of alkenes are typically aldehydes and water soluble organic acids (e.g. Forstner et al., 1997; Lang et al., 2002; Tsapakis et al., 2002), which have also been found as reaction products from α -pinene (e.g. Kavouras et al., 1998; Pio et al., 2001; Warscheid and Hoffmann, 2001) and limonene (Koch et al., 2000). Overall, it is found that aromatic substances typically have a real part of the refractive index between
10 1.5 and 1.6 with an average value of 1.55, while organic substances with linear carbon chains show values between 1.37 and 1.48, with an average value of 1.43 (Schroedter-Homscheidt, 2009). Organic species are generally identified as non-absorbing. The EURAD classes “primary fine aerosol” and “anthropogenic coarse aerosol” are not further described and mapped to the SYNAER insoluble component based on the OPAC description.
15

Overall, Table 2 gives an overview on the mapping of EURAD aerosol classes on SYNAER components together with the refractive index values chosen. This mapping is based on the description of aerosol species as given in the OPAC database (Hess et al., 1998), the description as given in MADE and SORGAM (Ackermann, 1995; Schell, 2001), measurements of water solubility, chemical composition, and the description
20 of size distributions as given both in EURAD and SYNAER, and finally, the refractive index identified for all classes based on a literature overview.

The EURAD aerosol model assumes an internal mixture within each mode. Therefore, the refractive index for a EURAD particle is calculated as a volume weighted mean
25 of refractive indices of all classes contributing in a size mode (Ouimette and Flagan, 1982).

Both EURAD and SYNAER describe the aerosol size distribution in a modal log-normal structure. EURAD size distributions in the nucleation, accumulation and coarse mode are defined with initial mean diameters of 0.01, 0.07 and 1 μm , respectively.

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The mean diameter D_m is changing during the model run, e.g. due to coagulation or sedimentation, while the distribution width σ_m is kept constant at 1.7, 2.0 and 2.2 μm , respectively. On the other hand, SYNAER follows the OPAC aerosol models with static size distributions for each component (see Table 1). The size distribution is only implicitly retrieved through the aerosol mixture chosen in the SYNAER procedure. Therefore, changes in the model state's size distributions by the SYNAER assimilation are not an objective of the methodology proposed as it would go beyond the SYNAER limitations.

Model state and observations generally do have different size distributions which affect the aerosol optical depth calculated from the model state. Sensitivity analysis shows that the extinction coefficient at 550 nm can vary between 0.25 and 1.5 km^{-1} for e.g. water insoluble aerosols, if the size distribution diameter is varied between 0.1 and 1 μm while the number of particles and aerosol mass are kept constant.

On the one hand, this calls for the use of a SYNAER equivalent size distribution in the Mie parameterization, which is calculated from the EURAD size distribution assuming a constant aerosol mass. Therefore, the 3rd moments of both log-normal size distributions

$$M_l = N D_m^l \exp\left(\frac{l^2}{2} \ln^2 \sigma_m\right) \quad \text{for } l = 3 \quad (8)$$

are set equal and the equation is solved for the particle number N_{SYNAER} of “equivalent SYNAER particles”.

$$N_{\text{SYNAER}} = \frac{N_{\text{EURAD}} D_{m,\text{EURAD}}^3 \exp\left[\frac{9}{2} (\ln \sigma_{m,\text{EURAD}})^2\right]}{D_{m,\text{SYNAER}}^3 \exp\left[\frac{9}{2} (\ln \sigma_{m,\text{SYNAER}})^2\right]} \quad (9)$$

On the other hand, this sensitivity shows that observations with a large disagreement in their size distribution versus the current model state should be excluded in a pre-processing step. This could be relevant in future once the mineral and/or sea salt components as large particles will be assimilated into a EURAD system.

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ratio.

2. SYNAER is only applied to SCIAMACHY pixels with cloud coverage below 50%. Below this threshold a dependence of deviations between SYNAER and AERONET on cloud cover can not be found.
3. High values of ground albedo also reduce the signal-to-noise ratio.
4. The SYNAER retrieval includes 40 different aerosol mixtures, which can not fully be separated from each other according to an information content analysis (Holzer-Popp et al., 2008). SYNAER defines the ambiguity error E_m as

$$E_m = \sqrt{\sum_{j=1}^{\rho} \frac{(\text{AOD}_j - \text{AOD}_k)^2}{\rho - 1}} \quad (10)$$

for all ρ undistinguishable aerosol mixtures. Undistinguishable mixtures are defined as those, where the least square fit of the j -th aerosol mixture are closer to the least square fit of the retrieved aerosol mixture k than these are different from the satellite measured reflectances R at all l wavelengths ($l=10$)

$$\sum_{l=1}^{10} (R_{\lambda_l, j} - R_{\lambda_l, k})^2 < \sum_{l=1}^{10} (R_{\lambda_l, \text{meas}} - R_{\lambda_l, k})^2 \quad (11)$$

5. The fit error E_f on the basis of all reflectances R at the 10 wavelengths λ_l serves to control retrieval fit quality and is defined in SYNAER as

$$E_f = \sqrt{\sum_{l=1}^{10} (R_{\lambda_l, \text{meas}} - R_{\lambda_l, k})^2} \quad (12)$$

Comparisons against AERONET observations show that large SYNAER overestimations occur if the ambiguity error $E_m > 0.1$ is combined with a land surface albedo above 15%. Additionally, large underestimations are found for $E_m > 0.05$ for water pixels with a surface albedo above 5%. Finally, pixels with $E_f > 0.025$ are also excluded to eliminate outliers.

Using these quality control criteria, the bias for all regions (Europe, Africa and Southern America) remains the same at -0.08 , but the standard deviation $\sigma_{\text{SYNAER-AERONET}}$ is reduced from 0.26 to 0.1. The RMSE decreases from 0.27 to 0.12 and the correlation coefficient increases significantly from 0.3 to 0.73. The number of coincidences in the validation study is reduced by 24% to 105 pixels.

Using this quality control procedure as pre-processing and assuming a diagonal structure, the observation error covariance matrix \mathbf{R} is set to a constant diagonal element value of 0.12 neglecting the rather small bias. Due to the limited number of available validation pixels, a regional variability in \mathbf{R} is not assumed in this study.

3.3.2 Background error quantification

A validation of the model state at 10:00 UTC versus AERONET ground measurements (version 2, level 2) is conducted for the study period, revealing a total of 2268 coincidences with an observation mean value of 0.22. It has to be noted, that AERONET provides ground measurements only in cloud-free situations. Overall an underestimation of -0.15 and a standard deviation $\sigma_{\text{EURAD-AERONET}}=0.17$ is found for the EURAD model. Especially AOD above 0.1 are underestimated. This agrees with other studies, finding PM_{10} and $\text{PM}_{2.5}$ underestimated for Europe in air quality models for example Sartelet et al. (2007) for the POLYPHEMUS model or Vautard et al. (2007) for the models CHIMERE, EMEP, LOTOS, REM-CALGRID, OFIS and CAMx. The EURAD underestimation occurs at all European AERONET stations and reaches its largest values with a bias of up to -0.25 in Italy and Turkey.

A separate monthly analysis reveals strong underestimation in July and August by -0.17 and -0.21 , respectively. In contrast, the bias is reduced to -0.1 in September

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and October and to -0.08 in November 2003. The standard deviation $\sigma_{\text{EURAD-AERONET}}$ shows no clear seasonality with 0.17, 0.17, 0.13, 0.21 and 0.15 for the months July to November, respectively. The large underestimation in July and August is probably caused by dust outbreak events from Sahara, which have not been modelled explicitly in the EURAD version used. Also, the extraordinary hot summer in 2003 may have caused less vegetation cover and increased erosion on agricultural land which is not taken into account in the static aerosol emission databases.

The background validation indicates a remarkable bias which is might be judged as not suitable for data assimilation purposes. However, in the meanwhile a dust model has been included in EURAD and it is expected that the overall bias will be reduced in a foreseen operational assimilation scheme. Additionally, it has to be taken into account that the EURAD model provides a physically and chemically consistent distribution between particle types which is of value in itself. A sustainable bias correction scheme could not be developed in this study as the bias is highly variable in space and time and not known yet due to the restricted data availability of only 5 months. Therefore, it was decided to apply the data assimilation as a correction mechanism without claiming a BLUE analysis. On the other hand, this study deals mainly with the observation operator and a first assessment of the value of the assimilation of aerosol component resolving observations into a CTM by assessing analysis fields versus ground measurements. Therefore, an assimilation experiment is needed as part of the study and for this purpose, an increase of the diagonal element in the initial background error covariance matrix above the standard deviation found in the validation was chosen as a first step and its value is set to 0.3. This reduces the weight of the model state in the assimilation, reflecting implicitly the bias problem found for this study data set.

4 Validation of resulting analysis fields

In the following sections, reports on validation studies either of AOD or particle mass concentrations in the analysis are given.

4.1 Total aerosol optical depth

The comparison of analysed AOD fields versus AERONET ground measurements reveals 189 coincidences for the whole study period in Europe. The validation shows a slight reduction of bias between the background and analysis fields from -0.14 to -0.12 while the standard deviation $\sigma_{x-AERONET}$ remains at 0.16. Due to the low repetition frequency of ENVISAT (12 days with cloud-free conditions) and the restricted number of available observations an overall impact of the assimilation can hardly be seen. It has to be noted that a future METOP assimilation will rely on a significantly larger number of observations than available for this ENVISAT-based study.

Therefore, further validation is performed only in the vicinity of existing SYNAER observations which might cause any positive or negative impact. Only grid boxes with non-zero analysis increment are taken into account. This criterion excludes situations where background and observations agree fully, but this occurs only in 0.04% of all cases. This is also justified as these cases would not contribute to any impact on the assimilation result.

Requiring an analysis increment larger than 0.05 reduces the number of coincidences to 53 with a mean observation AOD of 0.24, but shows also a significant improvement of the overall bias from -0.15 to 0.07 and a reduction in RMSE from 0.25 to 0.2 (Fig. 1). The overall standard deviation reduces only slightly from 0.2 to 0.19. The RMSE is significantly reduced at IMC-Oristano (0.5 to 0.22), El Arenosillo (0.23 to 0.15), Oostende (0.3 to 0.23), Dunkerque (0.19 to 0.13) and Forth Crete (0.11 to 0.05). Also, Lampedusa, Evora and Venice are slightly improved. Only at Toulouse and IS-DGM_CNR the RMSE is slightly increased by 0.03. IMC Oristano, El Arenosillo, Forth Crete, Lampedusa and Venice have been shown to be affected by dust outbreaks in this period (Breitkreuz, 2009). This motivates the validation of specific dust outbreak situations as given in Sect. 4.5.

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4.2 Soot mass concentrations

There is no Europe-wide ground measurement network for mass concentrations of elementary and organic carbon available. National standards can not be easily compared to each other (EMEP, 2007) or the density of ground stations is small. However, in the United Kingdom the Black Smoke and Sulphur Dioxide Network provides a rather dense network with 137 stations using a single measurement standard.

The sum of elementary and organic carbon as modelled in EURAD for 10:00 UTC is compared with the daily mean black carbon (BC) ground measurements. The difference of daily mean measurements and instantaneous model results is neglected. This is justified as the chosen ground stations are located in background conditions without a systematic intra-day variation as typically found in the vicinity of emission sources. Nevertheless, a possible additional gradient due to large scale meteorological conditions is neglected and contributes to the scatter observed.

Overall, 231 coincidences with an analysis increment above 0.001 in soot AOD can be found in the study period. The assimilation causes a bias reduction from -3.66 to $-1.38 \mu\text{g}/\text{m}^3$, but increases the standard deviation from 3.22 to $5.37 \mu\text{g}/\text{m}^3$ (Fig. 2). Single coincidences show overestimations up to $22 \mu\text{g}/\text{m}^3$ after the assimilation.

All outliers with differences $\text{BC}_{x,a} - \text{BC}_{\text{ground}} > 5 \mu\text{g}/\text{m}^3$ are evaluated separately using 300 m spatially resolving MODIS colour composites. All these differences occur at the edge of cloud fields or in areas with small scale clouds. Therefore, it is assumed that errors in cloud detection and clearing in the satellite observations cause an overestimation in AOD and therefore in the analysis field. Unfortunately, these cases show no remarkable patterns either in cloud cover, ambiguity error (Eq. 10) or fit error (Eq. 12), which would allow an automatic exclusion.

After manual exclusion of these outliers, the overall bias in the BC mass concentration is reduced from -3.63 to $-2.73 \mu\text{g}/\text{m}^3$ together with a reduction in standard deviation from 3.25 to $3.08 \mu\text{g}/\text{m}^3$ and a RMSE reduction from 4.87 to $4.12 \mu\text{g}/\text{m}^3$. Ground measurements show a mean value of $4.98 \mu\text{g}/\text{m}^3$ for these cases. It has to be noted

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that this impact is in the range of the uncertainty of the ground measurement network which is given as 30%. On the other hand, a significance test rejects the zero hypothesis, that analysis and background field are originating from the same basic population with a significance level below 1%. Therefore, a positive impact of the assimilation of SYNAER observations on the SOOT component can be stated, together with a strong emphasis on the need of strict cloud clearing in SYNAER observations.

The scatter remains rather large compared to the ground measurement uncertainty. On the other hand, a case study on 15 September 2003 for 3 carbon measuring ground stations in the bay of Gibraltar (Algeciras, La Linea, and Los Barrios) within the COST 633 campaign data base of the European Commission shows the typical variability within a EURAD grid box. The SYNAER assimilation raises the value for the grid box from $0.4 \mu\text{g}/\text{m}^3$ in the background to $4.13 \mu\text{g}/\text{m}^3$ in the analysis which is a positive impact if compared to the $3.1 \mu\text{g}/\text{m}^3$ mean of the 3 stations. On the other hand, the comparison with each single measurement results in a mismatch from -20% underestimation up to 120% overestimation in the analysis field illustrating the intra-grid box variability and explaining the scatter in validation studies using grid boxes or satellite pixels versus in situ measurements.

4.3 Water soluble mass concentrations

A further study focus is placed on the validation of mass concentrations of the water soluble aerosol component versus EMEP mass concentration measurements of SO_4 , NH_4 und NO_3 . EURAD mass concentrations are converted to $\mu\text{g S}/\text{m}^3$ and $\mu\text{g N}/\text{m}^3$ as provided by the EMEP network based on the molar masses of EURAD aerosol species. Finally, the sum of sulphur and nitrogen is compared as the assimilation is performed for the integrated WASO aerosol component and does not distinguish further between water soluble aerosol species. EMEP stations provide also daily mean values typical for background conditions, which are compared with 10:00 UTC EURAD background and analysis fields. The implications of this assumption have been already discussed in Sect. 4.2.

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Overall, 81 coincidences with an analysis increment above 0.01 in AOD at 16 stations in Hungary, Italy, Lithuania, the Netherlands, Norway, Poland, Russia and Turkey can be found for the period July to November 2003. This bias is reduced from 2.11 to 1.38 $\mu\text{g (S+N)}/\text{m}^3$, the standard deviation is decreased from 3.7 to 3.43 $(\text{S+N})/\text{m}^3$ and the RMSE from 4.27 to 3.7 $(\text{S+N})/\text{m}^3$ (Fig. 3) showing also a positive impact of the assimilation for the water soluble component. It has to be noted that this result was derived for ground stations representative for background conditions with a mean value of 2.14 $\mu\text{g (S+N)}/\text{m}^3$ and a maximum value of 8.67 $\mu\text{g (S+N)}/\text{m}^3$. A validation for non-background conditions can not be provided on that basis.

4.4 Soot aerosol optical depth

In August 2003 an intensive wild fire and biomass burning period affected Portugal and Southern Spain. Observations show high soot-induced AOD levels, which are not modelled in the EURAD background field. Figure 4 shows differences of soot observations versus the model background (“observations minus forecast”, OmF) of up to 0.15 (left panel) on the Iberian Peninsula due to these wild fires and additional positive OmF values in the Mediterranean area. Also, the water soluble component shows OmF up to 0.4 over the Iberian Peninsula due to the incomplete burning of wild fires. The assimilation includes these observations successfully, showing differences between the SOOT observations and the analysis field (OmA) close to zero.

The validation of analysed total AOD by AERONET ground measurements for all European stations in August 2003 shows an improvement by assimilation. If all coincidences with an analysis increment larger than 0.05 in AOD are taken into account, the bias because of underestimation is reduced from -0.29 to -0.13 , while RMSE is reduced from 0.35 to 0.26, and the standard deviation $\sigma_{\text{ANALYSIS-AERONET}}$ remains nearly constant (0.21 to 0.22). The AERONET mean AOD value for these cases is 0.36. Unfortunately, there are no soot mass concentration ground measurements in the vicinity of ENVISAT overpasses available in August 2003 for any independent quantitative evaluation.

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4.5 Mineral dust aerosol optical depth

Case studies of dust outbreaks are selected based on MODIS colour composites (<http://modis-atmos.gsfc.nasa.gov/IMAGES/index.html>). A number of 29 days with an ENVISAT overpass in a region with a dust outbreak event is selected visually. Overall, 10 coincidences between AERONET ground measurements and corresponding EURAD simulation with an analysis increment above 0.05 in AOD can be found. These cases include the stations El Arenosillo (2 days), Etna, Crete, Lampedusa (3 days), Lecce, and Oristano.

Observed and simulated total AOD values are compared in these cases at locations, where total AOD is dominated by dust events. Therefore, we assume that the results characterize the assimilation impact for dust cases. AOD bias is reduced from -0.26 to -0.07 and RMSE is decreased from 0.38 to 0.21 (Fig. 5) while the AERONET AOD shows a mean value of 0.38 in these cases.

It has to be noted that this positive impact is only attributed to the increased value of the WASO component observation in dust cases as there is no explicit assimilation of the SYNAER mineral component observation performed in this study. This is due to the missing dust module in the EURAD version used. But as SYNAER includes a 25% WASO background in all aerosol mixtures, a large signal in the mineral component is always accompanied by a larger signal in WASO which shows its positive impact in terms of total AOD.

5 Conclusions

An observation operator for a variational 2D-Var assimilation approach using satellite-based SYNAER aerosol component observations and the EURAD chemical transport model has been developed. The study is motivated by an expected positive impact of data assimilation of satellite observations for dynamic aerosol sources like dust outbreaks, fires, vegetation and changing human activities.

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by significant biases which are a risk in terms of the validity of the BLUE approach. It is known that currently most chemical transport models show a similar bias which has not been further defined yet and is therefore not removable by any sophisticated bias correction scheme. On the other hand recent validation of SYNAER version 2.2 observations has shown a reduced bias of -0.01 (instead of -0.08 for the SYNAER version used in this study) and an improved standard deviation of 0.08 which is promising for the future. Therefore, it is decided to use SYNAER observations in an assimilation approach to correct the background state, but without claiming a BLUE analysis. The improvement gained through this approach is shown by an assessment of the analysis. Additionally, the EURAD model is currently being enhanced by including additional processes like secondary organics and wind-blown dust into the aerosol mechanism. For a future operational application of the study results, a background bias correction scheme needs to be developed to ensure the BLUE. It has to be noted that the settings of the background error covariance matrix in this study are preliminary allowing an assimilation experiment within the study and therefore, the assessment of the value of aerosol component distinguishing satellite observations in an assimilation scheme.

Analysis fields are validated for total AOD by AERONET ground measurements both in all situations and in dust outbreak cases separately. To show the positive impact of assimilated SYNAER observations, mass concentration analysis fields are compared with the water soluble aerosol component as measured by the EMEP network and black carbon as measured by the UK Black Smoke and Sulphur Dioxide network.

A positive impact of SYNAER observations on modelled AOD can be observed for nearly all European stations, with special effect by Mediterranean stations. Additionally, a positive impact of the component-wise available SYNAER observations on modelled mass concentrations of water soluble and carbonaceous aerosol components is shown, together with a positive impact on AOD in dust outbreak cases. For a wild fire episode in Portugal a positive OmF can be found indicating a positive impact also for fire events which is confirmed through AOD validation by AERONET.

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It has been shown that the systematic error in all aerosol components can be significantly reduced by the assimilation, while the scatter remains nearly unchanged. The latter finding is not surprising as it represents the error in time and space caused by the comparison of 56 km sized grid boxes versus point ground measurements. A case study of 3 stations within a grid box in Spain shows that the differences between analysis field in the grid box and the individual station can range from -20% underestimation to $+120\%$ overestimation.

Overall, the foundation for the operational use of METOP SYNAER observations with a largely extended temporal and spatial coverage has been laid within this study. Based on longer time periods and more observations as available from METOP, an extended error covariance matrix definition is ongoing (Holzer-Popp et al., 2008; Martynenko et al., 2009) as a further step towards an operational sequential assimilation of SYNAER component AOD observations into the EURAD model.

The estimation of SYNAER equivalent size distributions (Eq. 9) for the EURAD background becomes unreliable if the mode diameter between the two distributions differs largely. In case of a future direct assimilation of mineral dust observations this has to be taken into account and such cases should be excluded in a pre-processing step.

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Table 1. Optical characteristics of basic components used for external mixing of in SYNAER (Holzer-Popp et al., 2008).

Component	Species	Complex refract. index at 550 nm	Mode radius [μm]	Stand. dev. of size distribution	Literature source
WASO, rH=70%	sulfate/nitrate, water soluble	1.53–0.0055 i	0.028	2.24	Hess et al. (1998)
INSO	insoluble particles, absorption as MITR	1.53–0.008 i	0.471	2.51	Hess et al. (1998)
INSL	insoluble particles, low absorption	1.53–0.0019 i	0.471	2.51	Dubovik et al. (2002)
SSAM, rH=70%	sea salt, accumulation mode	1.49–0 i	0.378	2.03	Hess et al. (1998)
SSCM, rH=70%	sea salt, coarse mode	1.49–0 i	3.17	2.03	Hess et al. (1998)
BISO	biomass burning soot	1.63–0.036 i	0.0118	2.0	Dubovik et al. (2002)
DISO	diesel soot	1.49–0.67 i	0.0118	2.0	Schnaiter et al. (2003)
MITR	transported minerals, high hematite content	1.53–0.0055 i	0.5	2.2	Hess et al. (1998)
MILO	transported minerals, low hematite	1.53–0.0019 i	0.5	2.2	Dubovik et al. (2002)

Table 2. Mapping of EURAD aerosol classes to SYNAER aerosol components together with refractive index chosen for EURAD aerosol classes.

EURAD aerosol class	refractive index set for EURAD class	SYNAER aerosol component
sulphuric acid	1.443+i 0.0	water soluble (WASO)
ammonium sulphate	1.53+i 10 ⁻⁷	water soluble (WASO)
ammonium nitrate	1.53+i 0.0	water soluble (WASO)
nitric acid	1.393+i 0.0	water soluble (WASO)
water	1.333+i 1.96×10 ⁻⁹	
primary organic	1.43 + i 0.0	insoluble (INSO and INSL)
reaction products of aromatics	1.55+i 0.0	water soluble (WASO)
reaction products of alkanes	1.43+i 0.0	water soluble (WASO)
reaction products of alkenes	1.45+i 0.0	water soluble (WASO)
reaction products of α -pinen	1.55+i 0.0	water soluble (WASO)
reaction products of limonen	1.55+i 0.0	water soluble (WASO)
elementary carbon	1.75+i 0.44	industrial soot (SOOT/DISO) or biomass burning soot (SOOT/BISO) depending on location
primary fine aerosol	1.53+i 0.08	insoluble (INSO and INSL)
anthropogenic coarse mode	1.53+i 0.08	insoluble (INSO and INSL)
maritime aerosol (in preparation)	1.5+i×10 ⁻⁸	sea salt (SSAM and SSCM)
mineral dust (in preparation)	1.53+i 0.0055	transported mineral aerosol (MITR and MILO)

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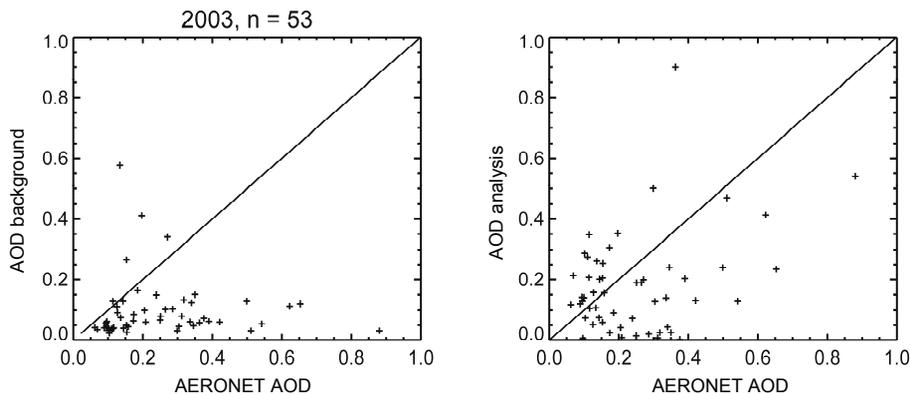
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Fig. 1. Comparison of background field AOD versus AERONET ground measurements (left) and of analysed field AOD versus AERONET (right) for 53 coincidences with a minimum analysis increment of 0.05 during the study period July–November 2003 in Europe.

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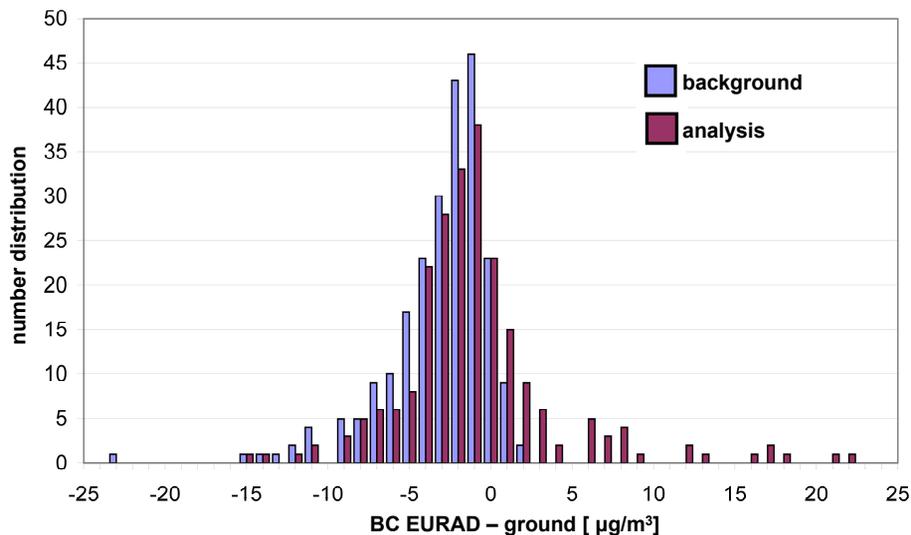
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Fig. 2. Number distribution of differences $BC_{x_b-ground}$ and $BC_{x_a-ground}$ for 231 coincidences in the United Kingdom within July to November 2003 for both the background (blue) and the analysis (purple) field.

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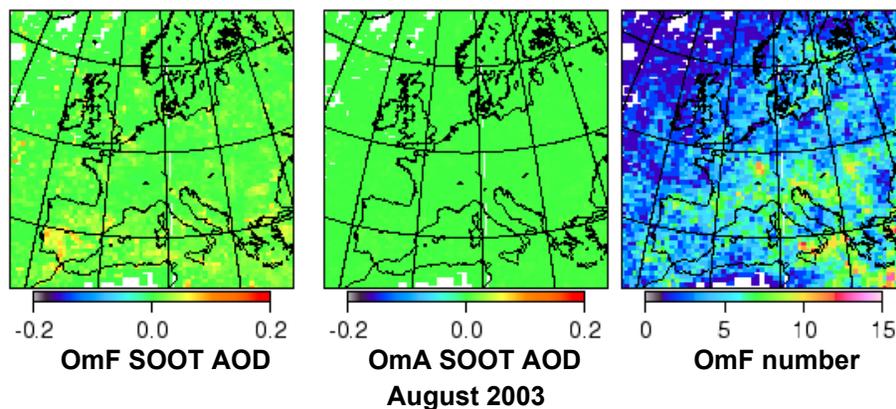
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Fig. 4. Map of SOOT component differences for observations minus forecast (left), observations minus analysis (middle) and number of observations used for each EURAD grid box in August 2003 (right). The wild fire induced signal in the observations is visible over the Iberian Peninsula and North of Algeria (yellow and orange color) and disappears in the analysis field (green colors).

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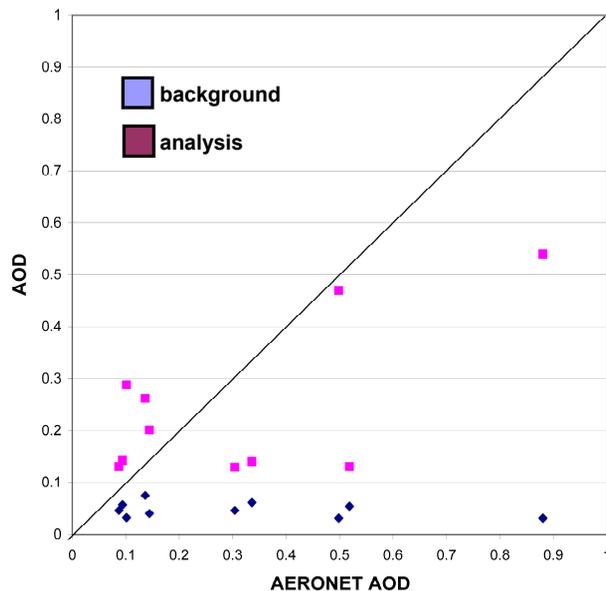


Fig. 5. Background (blue) and analysis (purple) field AOD versus AERONET AOD measurements for 10 dust outbreak cases in the July to November 2003 period.

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